

# **ENVIRONMENTAL SCIENCE**

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## 目 次

厦门大学环境学科创立 40 周年专栏
度门大学环境字科团立40周年专栏 新污染物共排放对生态环境监测和管理的挑战
河口-近海环境新污染物的环境过程、效应与风险 王新红,于晓璇,王思权,殷笑晗,钱韦旭,林晓萍,吴越,刘畅(4810)
海水痕量营养盐和金属的分子光谱分析方法研究进展 袁东星,黄勇明,王婷(4822)
外境水体中硫化物的分析方法:从实验至分析到原位监测 ····································
海洋浪量元素采样技术和分析方法的发展及展望. 厦门大学痕量元素平台建设讲展
聚乙烯微塑料的微生物降解研究进展
聚乙烯微塑料的微生物降解研究进展
水档土中氮素对微生物固砷的扰动及效应机制
中国两件生态母理学研究中的母性侧风生物
2015~2020 年夏漳泉地区大气氨排放清单及分布特征
九龙江口微塑料与抗生素抗性基因污染的分布特征 程宏,陈荣(4924)
厦门湾沙滩沉积物微塑料污染特征
几.无江口-厦门,冯霄项中浴,附念很重金,周的时空分布特仙与影响机制 ·················
厦门西溪河口沉积物活性磷的分布特征及迁移转化机制 · · · · · · · · · · · · · · · · · · ·
改性生物炭固定床对模拟湖库水体中 Mn <sup>2+</sup> 的吸附
基于表面增强拉曼光谱技术的饮用水中痕量恩诺沙星和环丙沙星快速检测 徐婧,郑红,卢江龙,刘国坤(4982)
水稻土中氮素对微生物固砷的扰动及效应机制
研究报告 2019 年秋季海南省 4 次 息菊污染过程特征及港在源区分析····································
2019 年秋季海南省 4 次臭氧污染过程特征及潜在源区分析 ····································
城区与郊区 PM <sub>2.5</sub> 污染及传输特征差异性
南京北郊 BTESX 特征及健康风险评估
东江源流域不同空间尺度景观格局对水质影响分析 ····································
长江与黄河源丰水期地表水中汞的分布特征、赋存形态及来源解析 刘楠涛,吴飞,袁巍,王训,王定勇(5064)
青藏高原湖泊水环境特征及水质评价
伊军河谷夏季 PM <sub>2.5</sub> 和 PM <sub>10</sub> 中水溶性无机离子浓度特征和形成机制 陈巧, 谷超, 徐涛, 周春华, 张国涛, 赵雪艳, 吴丽萍, 李新琪, 杨文(5009) 城区与郊区 PM <sub>2.5</sub> 污染及传输特征差异性 齐鹏, 周颖, 程水源, 白伟超(5018) 南京北郊 BTESX 特征及健康风险评估 冯悦政, 安俊琳, 张玉欣, 王俊秀(5030) 我国陆域水体系统表层水中微塑料生态风险评估 孙晓楠, 陈浩, 贾其隆, 朱弈, 马长文, 叶建锋(5040) 东江源流域不同空间尺度景观格局对水质影响分析 陈优良, 邹文敏, 刘星根, 曾金凤, 李丹, 郑汉奕(5053) 长江与黄河源丰水期地表水中汞的分布特征、赋存形态及来源解析
至了水化子与肌内区系的下怀至重切区石俗水文地环化子行肌及至耐凶系 
铜沸石对磷和重金属的吸附与底泥钝化性能 ····································
- 基十大基内组与太转录组分析有化废水生物处理系统脱级切能闲群
寒冷地区 IFAS+磁混凝污水厂菌群结构和抗生素抗性基因分析
甲国旱作农田一氧化氮排放及减排: Meta 分析 ···································
加州市州州外及四城来生),重、氮化利用华州氧化亚氮城州及西印罗州:Meta 万旬————————————————————————————————————
不同能即堪能对执带地区稻芯轮作休系土壤 CH 和N O排前的影响
不同水分条件下土地利用方式对我国热带地区土壤硝化过程及 NO 和N <sub>2</sub> O排放的影响 ····································
世子文献计量公坛的长江经文带农田土塘黄全屋运洗柱征。————————————————————————————————————
不同水分条件下土地利用方式对我国热带地区土壤硝化过程及 NO 和N <sub>2</sub> O排放的影响————————————————————————————————————
基于 GIS 对宁夏某铜银矿区周边土壤重金属来源解析
基于 GIS 对宁夏某铜银矿区周边土壤重金属来源解析 ····································
老化作用对生物炭钝化白云鄂博矿区碱性土壤中 Cd <sup>2+</sup> 的影响 ····································
磁性氧化铁/桑树杆生物炭的制备及其对砷污染土壤溶解性有机碳和砷形态的影响
世期元粉和石灰改良酸性水稻土对磷有效性、形态和酶活性的影响
磷、锌和镉交互作用对小白菜生长和锌镉累积的影响 帅祖苹,刘汉燚,崔浩,魏世强(5234)
重庆开州区菜地土壤抗生素污染特征及潜在生态环境风险评估 … 方林发,叶苹苹,方标,范晓霞,高坤鹏,李士洋,陈新平,肖然(5244)
基于 InVEST 和 GeoSoS-FLUS 模型的黄河源区碳储量时空变化特征及其对未来不同情景模式的响应
苗十斤防区不同恢复植被类型的固碳特征··········· 许小明 张晓萍 何亭 郭晋伟 薛峒 邻亚东 易海木 智洁 干洪直(5263)
土壤多功能性对微生物多样性降低的响应
氦添加对不同坡度退化高寒草甸土壤真菌多样性的影响····································
碳减排背景下我国与世界主要能源消费国能源消费结构与模式对比 李辉,庞博,朱法华,孙雪丽,徐静馨,王圣(5294)
中国形까用
基于 InVEST 和 GeoSoS-FLUS 模型的黄河源区碳储量时空变化特征及其对未来不同情景模式的响应
我国塑料污染防治政策分析与建议 ————————————————————————————————————
《环境科学》征订启事(4821) 《环境科学》征稿简则(5213) 信息(5052,5191,5273)



# 基于表面增强拉曼光谱技术的饮用水中痕量恩诺沙星 和环丙沙星快速检测

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摘要:近年来,抗生素的滥用引发抗生素抗性基因在环境中的传播和扩散,对生态系统与人类健康构成潜在威胁,特别是饮用水中抗生素污染事件的相关报道引发社会极大关注.因此,如何实现应急公共卫生事件中的痕量抗生素快检成为研究热点.基于表面增强拉曼光谱(SERS)技术,并结合磁性固相萃取(MSPE)样品前处理方法,构建了饮用水水样中ng·L<sup>-1</sup>水平喹诺酮类抗生素的快速检测方法.借助于磁性氧化石墨烯复合纳米材料(Fe<sub>3</sub>O<sub>4</sub>@SiO<sub>2</sub>-GO)的高吸附容量所提供的高富集能力,成功实现了饮用水中1.0 ng·L<sup>-1</sup>恩诺沙星(ENR)和5.0 ng·L<sup>-1</sup>环丙沙星(CIP)的加标检出,回收率在77.5%~91.5%之间,满足当前饮用水水质检测的要求.对于有机基质复杂的湖水等环境水样,萃取材料的选择性尚有待于进一步提升.

关键词:磁性固相萃取(MSPE);复合纳米材料;表面增强拉曼光谱(SERS);喹诺酮类抗生素(QNs);快速检测中图分类号:X830.2 文献标识码:A 文章编号:0250-3301(2022)11-4982-10 **DOI**: 10.13227/j. hjkx. 202205083

### Rapid Detection of Trace Enrofloxacin and Ciprofloxacin in Drinking Water by SERS

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Abstract: In recent years, the abuse of antibiotics has led to the spread and diffusion of antibiotic resistance genes in the environment, which poses a potential threat to the ecosystem and human health. In particular, the related reports of antibiotic contamination in drinking water have aroused great social concerns. Therefore, realizing the rapid detection of trace antibiotics in emergency events has become a research hotspot. Here, in combination with magnetic solid phase extraction (MSPE), we established a rapid detection strategy for  $ng^{\perp}L^{-1}$  level quinolones in drinking water using surface-enhanced Raman spectroscopy (SERS). With the help of the high enrichment capacity provided by the high adsorption capacity of the magnetic graphene oxide composite nanomaterial ( $Fe_3O_4@SiO_2-GO$ ), the spiked detection of 1.0  $ng^{\perp}L^{-1}$  enrofloxacin (ENR) and 5.0  $ng^{\perp}L^{-1}$  ciprofloxacin (CIP) in drinking water was successfully achieved, with recoveries ranging from 77.5% to 91.5%, which met the current requirements of drinking water testing. For environmental water samples such as lake water, the selectivity of extraction materials needs to be further improved due to the strong interference of the complex organic matrix.

Key words: magnetic solid-phase extraction (MSPE); composite nanomaterial; surface-enhanced Raman spectroscopy (SERS); quinolone antibiotics (QNs); rapid detection

抗生素作为一种价格低廉且性质稳定的合成抗菌药物,在临床医疗和畜牧生产中具有很高的应用价值<sup>[1-3]</sup>.在众多种类的抗生素中,喹诺酮类抗生素(quinolone antibiotics, QNs)因其治疗细菌感染的低成本和高效率而成为使用最为广泛的抗菌药物之一<sup>[4]</sup>.鉴于医疗废水与养殖废水等的处理不当,饮用水中抗生素污染事件时有发生,对公众健康和生态系统安全产生威胁<sup>[5-8]</sup>.由此,痕量抗生素的快速定性定量分析成为环境监测领域的热点<sup>[9-11]</sup>.

现有实验室标准检测方法主要是基于色谱和质谱 联 用 技 术,包 括 气 相 色 谱-质 谱 法 (gas chromatography-mass spectrometry,GC-MS)、高效液相色谱法(high performance liquid chromatography,HPLC)和高效液相色谱-串联质谱法(liquid chromatography tandem mass spectrometry,HPLC-MS/MS)等.这些方法兼具灵敏度高、分离效率佳和重现性好等优势,但仪器分析时间长,且需专业技术人员进行操作,在突发的公共卫生应急事件中难以实现快速现场检测[12~15].

近年来,因具有近单分子水平的高灵敏度和指纹图谱的高识别性等优势,表面增强拉曼光谱(surface-enhanced Raman spectroscopy,SERS)已在临床医疗、生命科学和环境监测等多领域中开展了广泛的应用研究 $^{[16~18]}$ . SERS 技术对于 QNs 的检出浓度通常在 $\mu$ g·L $^{-1}$ 水平 $^{[19~21]}$ ,无法满足水体环境中 $^{-1}$ 水平的检测需求 $^{[22]}$ . 为进一步提高检测灵敏度,研究者们围绕对 SERS 基底选择性修饰和通过样品前处理提高待测溶液中目标物浓度等方面开展了相关研究.

从基底修饰角度,研究者们借助于范德华力等加强物理吸附或通过化学成键等化学作用,调控 SERS基底的表面性质,例如采用TiO<sub>2</sub>、介孔分子筛

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SBA-15 和 PEI 纳米纤维等材料修饰 SERS 基底等<sup>[23~26]</sup>,由此获得 QNs 检出浓度降至数百ng·L<sup>-1</sup>水平. 然而相关方法过于复杂,难以在不同实验室获得重现结果.

从样品前处理角度,研究者们尝试使用磁性氧化石墨烯复合金属纳米粒子构筑兼具富集和 SERS 检测能力的纳米材料 $^{[27~31]}$ . 然而,过大的金属纳米粒子间距使得局域表面等离子体共振(localized surface plasmon resonance, LSPR)性能偏弱,从而导致该类材料的 SERS 增强性能偏弱. 因此,这些方法往往仅能实现 $\mu$ g·L $^{-1}$ 水平的检出.

由上可知,目前尚未形成简单有效的基于 SERS 技术的检测策略,以实现ng·L<sup>-1</sup>水平的 QNs 快速高灵敏定性定量分析. 考虑到氧化石墨烯(graphene oxide,GO)本身超高表面积、高亲水性和强极性作用力的特点和 Fe<sub>3</sub>O<sub>4</sub> 纳米 颗粒的磁性富集特质<sup>[32~34]</sup>,本工作制备了一种基于磁性氧化石墨烯纳米材料的磁性固相吸附材料,据此开展磁性固相萃取(magnetic solid-phase extraction,MSPE)前处理,在有效富集 QNs 抗生素的基础上,利用 SERS 技术实现了30 min 内饮用水水样中ng·L<sup>-1</sup>水平的 QNs 定性定量检测.

#### 1 材料与方法

#### 1.1 仪器与试剂

抗生素标准品恩诺沙星(enrofloxacin, ENR)、环 丙 沙星(ciprofloxacin, CIP) 和 磺 胺 嘧 啶(sulfadiazine,SD)均购自上海麦克林生化科技有限公司,纯度均大于98.0%;单层氧化石墨烯(GO)购自苏州碳丰石墨烯科技有限公司,纯度大于98.0%.实验所用超纯水(电阻率 $\geq$ 18.2 M $\Omega$ ·cm)均由 Direct-Q 3UV 超 纯 水 系 统 (美 国 Merck Millipore 公司)制备得到.

1.00 g·L<sup>-1</sup> 抗生素(ENR、CIP和SD)的标准储备液配置于 0.01 mol·L<sup>-1</sup> NaOH 溶液中,避光冷藏保存;其他浓度的抗生素使用液由储备液稀释得到. 王水由浓盐酸和浓硝酸按 3:1(体积比)混合配置而成,实验所用玻璃仪器均经王水浸泡 30 min 后用超纯水清洗干净.

水样来自于福建省厦门市翔安区的茂林人工湖表层水样和厦门大学环境与生态学院实验室的自来水水样,湖水经过 0.45 μm 的滤膜过滤后萃取,自来水样直接进行萃取,然后进行 SERS 检测. 同时,分别向水样中进行不同质量浓度 ENR 和 CIP 的加标实验,考察其加标回收率. 每种水样做 3 组平行实验,水样现取现测.

拉曼光谱通过 B&W TeK 拉曼光谱仪 [B&W TeK 光电科技(上海)有限公司]获得,激发波长为 785 nm,积分时间 2 s,积分次数为 10 次; 粒子的表征均通过 S-4800 扫描电子显微镜(日本 Hitachi 集团)完成.

#### 1.2 实验方法

### 1.2.1 SERS 基底的制备

使用经典柠檬酸钠还原法制备 AgNPs<sup>[35]</sup>,在 100 mL 的双口圆底烧瓶中加入 45 mL 的超纯水和 4.5 mL 的硝酸银 (53 mmol·L<sup>-1</sup>),并以 1500 r·min<sup>-1</sup>的转速持续搅拌溶液并加热煮沸 3~5 min 后,快速加入 5 mL 的 1.00% (质量分数)柠檬酸三钠溶液,溶液由无色逐渐变为金黄色最后转为灰绿色,保持沸腾 1 h 后停止加热,搅拌冷却至室温,由此制备得到银溶胶,4℃下储存,备用.并将制备的AgNPs 以5 000 r·min<sup>-1</sup>的转速离心 5 min 后浓缩 5 倍待用.

### 1.2.2 磁性石墨烯复合材料的制备

Fe<sub>3</sub>O<sub>4</sub>@SiO<sub>2</sub>-GO MNPs(FSGO MPs) 磁性复合材料通过四步制备获得(图 1)<sup>[36,37]</sup>:先采用共沉淀法合成 Fe<sub>3</sub>O<sub>4</sub> MNPs(F MPs), 再包裹 SiO<sub>2</sub> 层形成Fe<sub>3</sub>O<sub>4</sub>@SiO<sub>2</sub> MNPs(FS MPs),然后氨基化制备得到Fe<sub>3</sub>O<sub>4</sub>@SiO<sub>2</sub>-NH<sub>2</sub> MNPs(FSN MPs),最后结合 GO 形成Fe<sub>3</sub>O<sub>4</sub>@SiO<sub>2</sub>-GO MNPs,具体过程如下.

F MPs 的制备: 称取 9.63 g NH<sub>4</sub>Fe (SO<sub>4</sub>)<sub>2</sub>·12H<sub>2</sub>O和 4.93 g (NH<sub>4</sub>)<sub>2</sub>Fe (SO<sub>4</sub>)<sub>2</sub>·6H<sub>2</sub>O,加入至装有 200 mL 超纯水的圆底烧瓶中,保持剧烈搅拌使溶液混合均匀,将混合溶液水浴加热至 50°C,然后以 8 mL·min <sup>-1</sup>的速度加入 24.0 mL 28.0% (质量分数) 氨水,使 Fe<sup>3+</sup>和 Fe<sup>2+</sup>在氨水作用下与 OH <sup>-</sup> 反应生成Fe (OH)<sub>2</sub>和Fe (OH)<sub>3</sub>,这两种沉淀在加热作用下生成Fe<sub>3</sub>O<sub>4</sub>.实验过程中溶液由黄绿色逐渐变为橙黄色,最终稳定为深黑色,以1 500 r·min <sup>-1</sup>转速在 50°C 下持续搅拌 2 h. 反应结束后,待其自然冷却至室温,通过外加磁场收集磁性Fe<sub>3</sub>O<sub>4</sub>,去除剩余混合液,用超纯水反复清洗磁性粒子至溶液呈中性,然后再用无水乙醇清洗 3 次,60°C下干燥 8 h,得到磁性Fe<sub>3</sub>O<sub>4</sub> 颗粒.

FS MPs 的制备:将 2.5 g 已经制得的磁性 $Fe_3O_4$  颗粒加入至装有 20 mL 超纯水和 100 mL 无水乙醇 的圆底烧瓶中,超声(200 W,40 kHz)1 min 使其分散在溶液中,然后加入 10 mL 28.0% (质量分数)氨水和 6 mL TEOS,剧烈搅拌 12 h,溶液由无色逐渐转为奶白色,得到 $Fe_3O_4$ @SiO<sub>2</sub>颗粒.

FSN MPs 的制备: 向上述装有Fe<sub>3</sub>O<sub>4</sub>@SiO<sub>2</sub> NPs 的圆底烧瓶中加入 2 mL APTES,40℃下剧烈 搅拌 24 h,反应结束后自然冷却至室温. 外加磁场筛选磁性粒子,并用超纯水和无水乙醇清洗干净,60°下干燥 4 h,得到磁性  $Fe_3O_4$ @ $SiO_2$ - $NH_2$ 颗粒.

FSGO MPs 的制备:向装有 200 mL 超纯水的圆底烧瓶中加入 0.2 g GO,超声(200 W,40 kHz)30 min 使 GO 均匀分散至水中,然后加入 125 mg EDC

和80 mg NHS,通过 EDC 诱导形成酰胺键,NHS 进一步稳固酰胺键,混合物在室温下剧烈搅拌 1 h,然后加入制得的 $\text{Fe}_3\text{O}_4\text{@SiO}_2\text{-NH}_2$  颗粒,超声(200 W,40 kHz)2 min 使其分散在溶液中,然后在80℃下剧烈搅拌 1 h,外加磁场筛选磁性粒子,并用超纯水和无水乙醇清洗干净,60℃下干燥 4 h,得到磁性  $\text{Fe}_3\text{O}_4\text{@SiO}_2\text{-GO颗粒}$ .

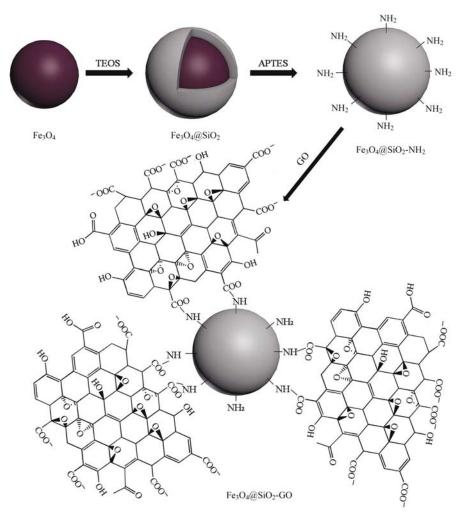


图 1 FSGO MPs 的制备流程示意

Fig. 1 Preparation process of FSGO MPs

#### 1.2.3 FSGO MPs 磁性和形貌表征

对 FSGO MPs 的磁性进行表征,如图 2 (a)所示,可观察到粒子呈深黑褐色,在水中分散性良好,有利于对目标物的吸附.外置磁场后,粒子可在 12 s内迅速聚集,说明粒子磁响应性能良好,有利于萃取过程中固液分离.

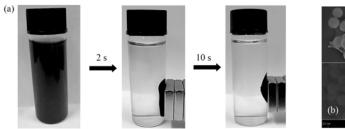
图 2 (b) 为制备的 FSGO MPs 磁性材料的 SEM 图,可明显观察到所制备的材料具有良好的均匀性和分散性.

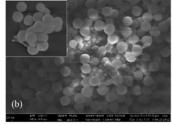
#### 1.2.4 目标物的萃取和检测

目标物的萃取过程如图 3 所示,将一定量的

FSGO MPs 置于梨形瓶中,用少量甲醇水溶液活化后,加入 500 mL 水样,剧烈振荡 10 s,待目标物吸附完全后,借助外置磁铁将粒子与水样分离.加入一定体积的解吸溶液,轻微振荡一定时间,使目标物萃取至解吸溶液中,同样通过外加磁场使粒子与解吸溶液分离,获取的解吸溶液用氮气吹干,再用超纯水定容至 500 μL 获得待测样,FSGO MPs 则先后用解吸溶液和超纯水清洗后供下次使用.

将待测样、Ag NPs 和团聚剂 1  $mol \cdot L^{-1}$  KI 以 200  $\mu$ L: 50  $\mu$ L 50  $\mu$ L 的比例在 96 孔板中进行混合





(a) 外置磁场聚集过程,(b) 磁性材料 SEM 图

#### 图 2 FSGO MPs 磁性材料的外置磁场聚集过程和 SEM 表征

Fig. 2 External magnetic field aggregation process and SEM images of FSGO MPs magnetic materials

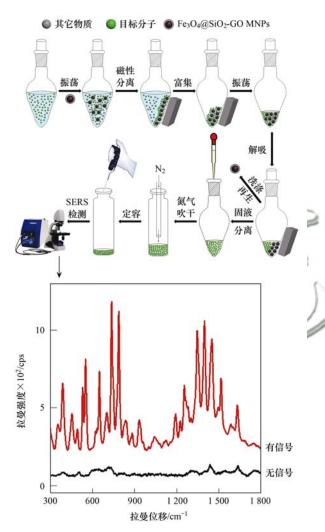


图 3 FSGO MPs 的萃取流程

Fig. 3 Extraction process of FSGO MPs

后进行 SERS 检测.

#### 2 结果与讨论

#### 2.1 萃取性能的考察

萃取效率与待测分子的极性、电负性等性质和所处介质环境密切相关<sup>[38]</sup>. 因此,以 ENR 为探针分子,根据萃取前后溶液中 ENR 的 SERS 谱图变化,系统考察了萃取材料用量、吸附动力学、解吸动力学和解吸溶液中萃取剂的组成等因素对萃取性能的影响.

#### 2.1.1 吸附材料用量

图 4 (a) 为含有 10.0 ng·L<sup>-1</sup> ENR 的水样经FSGO MPs 磁性吸附材料萃取前后的 SERS 谱图.从中可知,在使用该材料萃取后,ENR 的 SERS 信号从无到有,出现了位于 291、388、454、531、737、788、1394 和1 444 cm<sup>-1</sup> 的特征 SERS 谱峰.已有结果表明,在相同测试条件下,ENR 的最低可检出浓度(以质量浓度计)为 100 ng·L<sup>-1[39]</sup>,因此,图 4 (a) 所示结果表明萃取后待测溶液中 ENR 的浓度至少增加了一个数量级.以位于 737 cm<sup>-1</sup> 的 ENR 特征 SERS 谱峰强度为基准,考察 25.0~300.0 mg 范围内FSGO MPs 用量对萃取效果的影响.如图 4 (b) 所示,随着吸附材料用量由 25.0 mg 增加到 50.0 mg,目标物的 SERS 信号迅速增强,并在用量达到 50.0 mg 时获得最强 SERS 信号; 进一步提高用量时,ENR 的 SERS 信号强度逐渐减弱.

萃取效率对吸附材料用量的依赖性可能原因是:随着固相吸附材料用量的增加,一方面提高了吸附过程中 ENR 在固相中的总量,即提高了富集效率;另一方面降低了解吸过程中洗脱液的萃取效率.因此,表现出最佳吸附材料用量.需指出,当吸附材料用量大于 100.0 mg 时,解吸溶液(酸化甲醇)从清澈无色转变为棕黄色.推测可能是由于 FSGO MPs 表面 GO 在酸化甲醇中脱落所致,具体原因尚有待进一步探究.

### 2.1.2 吸附与解吸动力学研究

图 4 (c)以位于 737 cm<sup>-1</sup>的 ENR 的特征 SERS 谱峰强度为基准,考察了 60 min 内,FSGO MPs 材料的吸附动力学曲线. 由图 4 可知,随吸附时间增加到 20 min,目标物的萃取效率线性增加,并在 20 min 左右基本达到吸附饱和. 随着吸附时间继续增加至 60 min,SERS 信号略有波动,表明基本达到了吸附动态平衡.

在该条件下,以位于  $737~cm^{-1}$ 的 ENR 的特征 SERS 谱峰强度为基准,考察  $1\sim15~min$  范围内解吸时间对萃取效果的影响,结果如图 4~(d)所示.可以

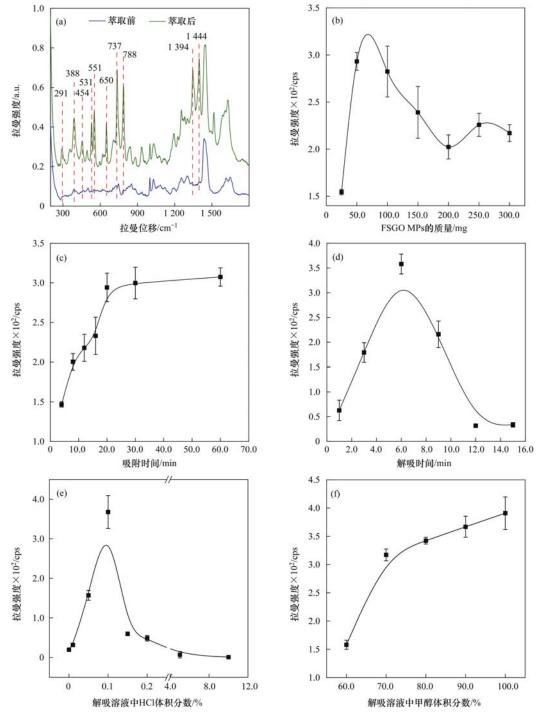
观察到,目标物的 SERS 信号强度随着解吸时间的延长迅速增加,在6 min 时达到峰值;当解吸时间继续延长时,SERS 强度急速减小.此时可明显观察到解吸溶液呈现棕黄色,说明 FSGO MPs 结构可能被解吸溶液所破坏,可能干扰后续 SERS 测试.因此,最佳解吸时间为6 min.

#### 2.1.3 解吸溶液酸度探究

酸化甲醇解吸溶液中酸的用量可影响 QNs 中

极性官能团的存在形态,从而影响萃取效果<sup>[40]</sup>.图 4 (e)为盐酸甲醇溶液中盐酸的体积分数在 0 ~ 10.0% 范围内萃取效果的变化的趋势,以位于 737 cm<sup>-1</sup>的 ENR 的特征 SERS 谱峰强度为基准,考察了解吸溶液中酸度对萃取性能的影响. 结果表明,随着酸体积分数的增加,萃取效率迅速增加,并在 0.1%时达到极值;随后迅速降低.

目标物的萃取效果决定于分子与吸附材料之间



(a)萃取前后的信号强度对比,(b)吸附材料用量,(c)吸附动力学曲线,(d)解吸动力学曲线,(e)解吸溶液酸度探究,(f)解吸溶液甲醇体积分数探究

图 4 萃取体系性能考察(n=3)

Fig. 4 Performance of extraction system

的 π—π 键、氢键、疏水和偶极-偶极作用等多种相互作用的协同效应<sup>[41~43]</sup>. 当解吸溶液中氯化氢体积分数为 0 时,溶液接近中性,ENR 分子主要通过π—π键和疏水作用强吸附于 FSGO MPs 表面. 此时,ENR 难以被甲醇溶液洗脱,因而萃取效率很低. 随着酸体积分数的缓慢增加,pH 逐渐降低,ENR 分子发生质子化,削弱 FSGO MPs 材料之间作用力的同时在酸化甲醇中的溶解度也相应增大,双重作用下显著提高了萃取效率. 随着酸体积分数的进一步增加,pH 进一步降低,FSGO MPs 材料结构被破坏,不利于固液分离,导致萃取效率降低.

#### 2.1.4 解吸溶液甲醇体积分数探究

解吸溶液中的甲醇体积分数影响极性化合物的解吸能力<sup>[44,45]</sup>.如图 4 (f)所示,当甲醇体积分数从60.0%增加至 70.0%时,SERS 信号强度迅速增加,

表明解吸效率显著提高;当继续增加甲醇体积分数时,SERS 信号强度增速趋于平缓,当甲醇体积分数为100.0%时获得最强的SERS 信号,说明高浓度的甲醇有利于解吸效率的提高.

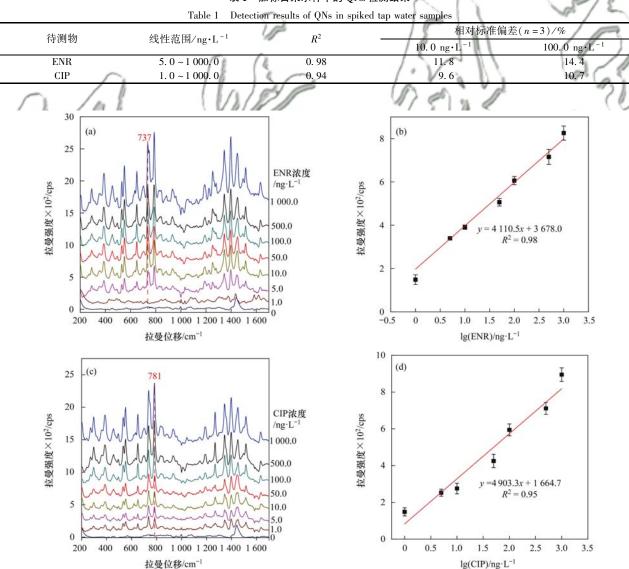
综合考虑,在对于待测水样开展的前处理工作中,采用50.0 mg FSGO MPs 进行吸附,其中吸附时间为20 min;在解吸过程中,使用1.0 mL0.1% 盐酸甲醇溶液作为解吸溶液,解吸时间为6 min.

#### 2.2 标准工作曲线的建立

在建立最优萃取条件萃取方法的基础上,分别考察了 $0.5 \sim 1000.0 \text{ ng} \cdot \text{L}^{-1}$ 范围内 ENR 和 CIP 的萃取效果. 并对方法的线性和平行性分别进行考察,结果见表 1.

由图 5 (a)和图 5 (c)可知,随着浓度的升高, ENR和CIP的SERS谱图的特征峰变化趋势基本保

表 1 加标自来水样中的 QNs 检测结果



(a) ENR 的 SERS 图谱,(b) ENR 的标准工作曲线,(c) CIP 的 SERS 图谱,(d) CIP 的标准工作曲线

图 5 ENR 与 CIP 的 SERS 图谱及其标准工作曲线 (n=3)

Fig. 5 SERS spectra and calibration curves of ENR and CIP (n=3)

持一致,体现为拉曼信号强度的正向增大依赖性. 在一定范围内,二者的 SERS 特征峰的信号强度随其浓度升高呈现出先迅速增大后趋于平缓的趋势,采用 Langmuir 吸附等温式可较好地拟合该变化趋势 $^{[16]}$ . 如图 5 (b)和图 5 (d)所示,以目标物浓度值的对数为横坐标,定量峰 737 cm $^{-1}$ 处的拉曼强度为纵坐标,绘制标准工作曲线,可得到二者的线性范围分别在  $1.0 \sim 1\,000.0\,\mathrm{ng}\cdot\mathrm{L}^{-1}$ 之间.与无萃取结果(最低可检出浓度为  $100\,\mathrm{ng}\cdot\mathrm{L}^{-1}$ )相比,二者的检测灵敏度皆约提高了 2 个数量级 $^{[39]}$ . 对浓度分别为  $10.0\,\mathrm{ng}\cdot\mathrm{L}^{-1}$ 的两种 QNs 标准溶液连续测定 3 次,相对标准偏差均在 15.0% 以内,说明本方法的结果平行性较好.

存在较大误差的主要原因在于: ①FSGO MPs 粒径分布不均匀,每个 MPs 表面 GO 的修饰量存在 差异,导致其对目标分子的吸附能力不同,因而萃取效果也会存在一定的差异;②AgNPs 的粒径差异性所带来的 SERS 增强性能的差异. 尚需要进一步提高 FSGO MPs 和溶胶态 SERS 基底的均匀性以提高 SERS 检测的定量可靠性.

#### 2.3 环境水样中的 QNs 加标检测研究

在建立 QNs 标准工作曲线的基础上,考察了多种环境水样(大洋海水、近岸海水、湖水、鱼塘水、养殖场废水和自来水)中痕量 ENR 和 CIP 的检测. 如表 2 所示,当 QNs 加标量在 100 ng·L<sup>-1</sup>水平及以上时,样品无需富集即可实现检测. 加标回收率在63.9%~115.0%之间,相对标准偏差低于15.0%(n=3). 如表 3 所示,当 QNs 加标浓度低于100 ng·L<sup>-1</sup>时,FSGO MPs 吸附材料在湖水等环境水样中与在简单的自来水基质中体现出较大的萃取能力差异.

表 2 高浓度实际水样中 QNs 的检测结果及加标回收率1)

		_	100			,	45	1 10 1
Table 2	Dotootion ro	culta and	POOCHOWI	of ONa	in high	concentration	of actual	water samples

样品         加标度度 /µg·L·1         ENR         CIP           大洋海水 $hgiqqq$ 加标回收率 /½         相对标准偏差 (n=3)/%         位謝结果 /µg·L·1         加标回收率 /%         相对标准偏差 /µg·L·1         相对标准偏差 /µg·L·1         相对标准偏差 /µg·L·1         相对标准偏差 /µg·L·1         相对标准偏差 /µg·L·1         1         一         —		Table 2	Detection results	and recovery or Q14	3 III IIIgii concen	manon of actual water	a sampics	/ // ) II
特別		加長沈度	(1)	ENR		150	CIP	18/1
大洋海水         0.0         ND         —         ND         —         ND         —         —           大洋海水         0.1         0.09         86.8         2.4         0.09         86.9         2.8           0.5         0.44         89.0         10.8         0.41         82.1         11.4           1.0         1.15         115.0         0.6         0.96         96.0         1.7           0.0         ND         —         —         ND         —         —           0.1         0.07         70.7         3.9         0.08         78.7         3.6           0.5         0.34         68.9         13.6         0.49         98.3         6.4           1.0         0.64         63.9         11.6         0.84         84.2         12.3           0.0         ND         —         —         ND         —         —           湖水         0.1         0.09         91.1         5.5         0.10         99.4         6.2           1.0         0.69         69.2         11.7         0.64         64.5         1.6           0.5         0.36         72.6         2.9         0.37	样品			加标回收率	相对标准偏差		加标回收率	相对标准偏差
大洋海水       0.1       0.09       86.8       2.4       0.09       86.9       2.8         0.5       0.44       89.0       10.8       0.41       82.1       11.4         1.0       1.15       115.0       0.6       0.96       96.0       1.7         厦门近岸海水       0.0       ND       —       —       ND       —       —         0.1       0.07       70.7       3.9       0.08       78.7       3.6         0.5       0.34       68.9       13.6       0.49       98.3       6.4         1.0       0.64       63.9       11.6       0.84       84.2       12.3         0.0       ND       —       —       ND       —       —         3/4       0.1       0.09       91.1       5.5       0.10       99.4       6.2         1.0       0.5       0.36       72.6       2.9       0.37       73.3       11.2         1.0       0.69       69.2       11.7       0.64       64.5       1.6         0.1       0.09       91.6       7.6       0.08       78.8       6.8         0.5       0.40       79.3       14.1		μg·E	/μg•L <sup>-1</sup>	1%	(n=3)/%	/μg·L <sup>-1</sup>	/%	(n=3)/%
No	/ 1	0.0	ND	11/1/	_	ND ND	9 -	
10.8	大洋海水	<b>2</b> / <b>0</b> .1	0.09	86.8	2.4	0.09	86.9	
Descripton			0.44	89.0	10.8	0.41	82.1	41.4
厦门近岸海水     0.1     0.07     70.7     3.9     0.08     78.7     3.6       0.5     0.34     68.9     13.6     0.49     98.3     6.4       1.0     0.64     63.9     11.6     0.84     84.2     12.3       0.0     ND     -     -     ND     -     -       0.1     0.09     91.1     5.5     0.10     99.4     6.2       0.5     0.36     72.6     2.9     0.37     73.3     11.2       1.0     0.69     69.2     11.7     0.64     64.5     1.6       0.1     0.09     91.6     7.6     0.08     78.8     6.8       0.1     0.09     91.6     7.6     0.08     78.8     6.8       0.5     0.40     79.3     14.1     0.37     74.9     7.5       1.0     0.71     71.3     5.5     0.73     73.1     9.7       Africance     0.0     ND     -     -     ND     -     -       Africance     0.0     ND     -     -     ND     -     -       Africance     0.0     ND     -     -     ND     -     -       Africance     0.1     0.09 <td< td=""><td>151</td><td>1.0</td><td>1.15</td><td>115.0</td><td>0.6</td><td>0.96</td><td>96.0</td><td>1.7</td></td<>	151	1.0	1.15	115.0	0.6	0.96	96.0	1.7
B   D近岸海水   0.5	( a 1/1 )	0.0	ND	17# /	\ -	ND	V =	_
13.6	厦门近岸海水	0.1	0.07	70.7	3.9	0.08	78.7	3.6
湖水     0.0     ND     —     —     ND     —     —       0.1     0.09     91.1     5.5     0.10     99.4     6.2       0.5     0.36     72.6     2.9     0.37     73.3     11.2       1.0     0.69     69.2     11.7     0.64     64.5     1.6       0.0     ND     —     —     ND     —     —       6.8     0.1     0.09     91.6     7.6     0.08     78.8     6.8       0.5     0.40     79.3     14.1     0.37     74.9     7.5       1.0     0.71     71.3     5.5     0.73     73.1     9.7       0.0     ND     —     —     ND     —     —       养殖场废水     0.1     0.09     91.4     5.0     0.10     99.0     10.2       养殖场废水     0.5     0.35     70.9     2.0     0.43     86.3     14.0		0.5	0.34	68.9	13.6	0.49	98.3	6.4
湖水     0.1     0.09     91.1     5.5     0.10     99.4     6.2       0.5     0.36     72.6     2.9     0.37     73.3     11.2       1.0     0.69     69.2     11.7     0.64     64.5     1.6       0.0     ND     -     -     ND     -     -       0.1     0.09     91.6     7.6     0.08     78.8     6.8       0.5     0.40     79.3     14.1     0.37     74.9     7.5       1.0     0.71     71.3     5.5     0.73     73.1     9.7       0.0     ND     -     -     ND     -     -       养殖场废水     0.1     0.09     91.4     5.0     0.10     99.0     10.2       养殖场废水     0.5     0.35     70.9     2.0     0.43     86.3     14.0		1.0	0.64	63.9	11.6	0.84	84.2	12.3
樹水     0.5     0.36     72.6     2.9     0.37     73.3     11.2       1.0     0.69     69.2     11.7     0.64     64.5     1.6       鱼塘水     0.0     ND     —     —     ND     —     —       0.1     0.09     91.6     7.6     0.08     78.8     6.8       0.5     0.40     79.3     14.1     0.37     74.9     7.5       1.0     0.71     71.3     5.5     0.73     73.1     9.7       0.0     ND     —     —     ND     —     —       养殖场废水     0.1     0.09     91.4     5.0     0.10     99.0     10.2       养殖场废水     0.5     0.35     70.9     2.0     0.43     86.3     14.0	M	0.0	ND	_	_	ND	_	_
6.5     0.36     72.6     2.9     0.37     73.3     11.2       1.0     0.69     69.2     11.7     0.64     64.5     1.6       0.0     ND     -     -     ND     -     -       0.1     0.09     91.6     7.6     0.08     78.8     6.8       0.5     0.40     79.3     14.1     0.37     74.9     7.5       1.0     0.71     71.3     5.5     0.73     73.1     9.7       0.0     ND     -     -     ND     -     -       **     0.1     0.09     91.4     5.0     0.10     99.0     10.2       **     0.5     0.35     70.9     2.0     0.43     86.3     14.0	湖水	0.1	0.09	91.1	5.5	0.10	99.4	6.2
鱼塘水     0.0     ND     —     —     ND     —     —       0.1     0.09     91.6     7.6     0.08     78.8     6.8       0.5     0.40     79.3     14.1     0.37     74.9     7.5       1.0     0.71     71.3     5.5     0.73     73.1     9.7       0.0     ND     —     —     ND     —     —       养殖场废水     0.1     0.09     91.4     5.0     0.10     99.0     10.2       6.3     0.5     0.35     70.9     2.0     0.43     86.3     14.0	.,,	0.5	0.36	72.6	2.9	0.37	73.3	11.2
鱼塘水     0.1     0.09     91.6     7.6     0.08     78.8     6.8       0.5     0.40     79.3     14.1     0.37     74.9     7.5       1.0     0.71     71.3     5.5     0.73     73.1     9.7       0.0     ND     -     -     ND     -     -       6.1     0.09     91.4     5.0     0.10     99.0     10.2       7.5     0.5     0.35     70.9     2.0     0.43     86.3     14.0				69.2	11.7		64.5	1.6
担端水     0.5     0.40     79.3     14.1     0.37     74.9     7.5       1.0     0.71     71.3     5.5     0.73     73.1     9.7       0.0     ND     -     -     ND     -     -       6.1     0.09     91.4     5.0     0.10     99.0     10.2       7.5     0.5     0.35     70.9     2.0     0.43     86.3     14.0		0.0	ND	_	_	ND	_	_
0.5     0.40     79.3     14.1     0.37     74.9     7.5       1.0     0.71     71.3     5.5     0.73     73.1     9.7       0.0     ND     -     -     ND     -     -       6.1     0.09     91.4     5.0     0.10     99.0     10.2       76     0.5     0.35     70.9     2.0     0.43     86.3     14.0	鱼塘水			91.6	7.6	0.08		6.8
成分     ND     一     一     ND     一     一       参殖场废水     0.1     0.09     91.4     5.0     0.10     99.0     10.2       0.5     0.35     70.9     2.0     0.43     86.3     14.0	,	0.5	0.40	79.3	14.1	0.37	74.9	7.5
养殖场废水     0.1     0.09     91.4     5.0     0.10     99.0     10.2       0.5     0.35     70.9     2.0     0.43     86.3     14.0		1.0	0.71	71.3	5.5	0.73	73.1	9.7
乔组功废水 0.5 0.35 70.9 2.0 0.43 86.3 14.0		0.0		_	_	ND	_	_
0.5 0.35 70.9 2.0 0.43 86.3 14.0	养殖场废水							
1.0 0.95 95.4 4.5 0.81 81.0 7.8					2.0			
		1.0	0.95	95.4	4.5	0.81	81.0	7.8

<sup>1)&</sup>quot;ND"表示未检出,"一"表示未检出物质对应的无加标回收率和无相对标准偏差,下同

表 3 低浓度实际水样中 QNs 的检测结果及加标回收率

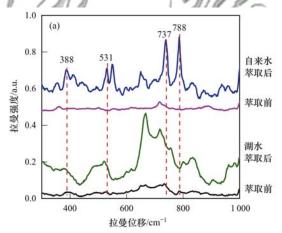
Table 3 Detection results and recovery of QNs in low concentration of actual water sample

Jan de Nota 1950		ENR		CIP			
加怀浓度 ∕ng•L <sup>-1</sup>	检测结果 /ng·L <sup>-1</sup>	加标回收率 /%	相对标准偏差 (n=3)/%	检测结果 /ng·L <sup>-1</sup>	加标回收率 /%	相对标准偏差 (n=3)/%	
0.0	ND	_	_	ND	_	_	
10.0	9.2	91.5	12.1	7.7	77.5	13.8	
50.0	44.6	89.2	14.8	39.3	78.5	14.6	
0.0	ND	_	_	ND	_	_	
10.0	ND	_	_	ND	_	_	
50.0	ND	_	_	ND	_	_	
	0. 0 10. 0 50. 0 0. 0 10. 0	/ng·L -1 極测结果 /ng·L -1	加标浓度 /ng·L <sup>-1</sup> 检测结果 /ng·L <sup>-1</sup> /%  0.0 ND —  10.0 9.2 91.5  50.0 44.6 89.2  0.0 ND —  10.0 ND —	加标浓度 /ng·L <sup>-1</sup>	加标浓度 $holdsymbol{ho$	加标浓度 $/ \text{ng} \cdot \text{L}^{-1}$ 检测结果 $/ \text{ng} \cdot \text{L}^{-1}$ 加标回收率 $/ \text{ng} \cdot \text{L}^{-1}$ 相对标准偏差 $/ \text{ng} \cdot \text{L}^{-1}$ 检测结果 $/ \text{ng} \cdot \text{L}^{-1}$ 加标回收率 $/ \text{ng} \cdot \text{L}^{-1}$ 0.0     ND     —     —     ND     —       10.0     9.2     91.5     12.1     7.7     77.5       50.0     44.6     89.2     14.8     39.3     78.5       0.0     ND     —     —     ND     —       10.0     ND     —     —     ND     —	

在湖水等环境水样中,即便是通过 FSGO MPs 的富集作用也无法获得 QNs 的检出. 如图 6 (a) 所示,以加标浓度为 50.0 ng·L<sup>-1</sup> ENR 为例,可明显观察到在 350~1000 cm<sup>-1</sup>范围内出现多个"峰包",淹没了位于 388、531、737 和 788 cm<sup>-1</sup>等位置的特征峰而无法有效判定 ENR 的检出与否,这些"峰包"可能来源于湖水中藻类和细菌等浮游生物产生的有机质.一方面,这些有机质可有效吸附在 FSGO MPs 而被萃取到待测溶液中,另一方面杂质浓度经富集后大大提升,其非特异性吸附严重干扰了目标物在 SERS 基底表面的吸附,最终导致目标物的 SERS 信号被淹没在杂质信号中.

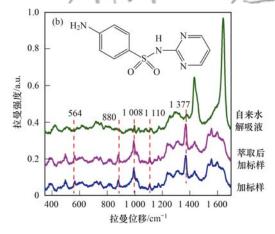
如图 6 (a) 所示, 对于有机质含量低的自来水样, 经 FSGO MPs 吸附材料萃取后可观察到明显的 ENR 特征 SERS 信号, 其加标回收率在 77.5% ~ 91.5%之间, 相对标准偏差低于 15.0% (n = 3), 证明本方法可实现饮用水这一基质简单水样中痕量喹诺酮类抗生素的快速检测.

如图 6 (b) 所示, 当将目标分子改为 1.0  $\mu g \cdot L^{-1}$ 磺胺嘧啶(SD)分子时, 比较 FSGO MPs 吸附 材料萃取前后加标水样和酸化甲醇解吸后水溶液的



SERS 谱图可知:萃取前后加标水样的 SERS 谱图 (蓝线和紫线)特征基本一致,位于 564、880、1008、110和1377 cm<sup>-1</sup>等来自 SD 的特征 SERS 谱峰的拉曼强度几乎不变,意味着 FSGO MPs 吸附材料未能有效萃取水样中的 SD 分子,这也导致洗脱后的溶液 SERS 谱图中未能观察到 SD 的特征谱峰(绿线).需指出,当 SD 浓度提升到 50.0 μg·L<sup>-1</sup>时,洗脱后的溶液 SERS 谱图中可明显观察到 SD 的 SERS 信号.

根据上述结果,推测磺胺类和喹诺酮类抗生素分子间化学结构差异显著,导致了二者与 FSGO MPs 间相互作用力的显著差异. 在中性或弱酸性环境下,SD分子( $pK_a=6.36$ )苯环上的氨基为吸电子状态,降低了苯环的电子云密度,削弱了 SD分子与FSGO MPs 间的  $\pi-\pi$  相互作用力,因此 FSGO MPs 无法有效从水中萃取 SD分子. 与之相反,在中性或弱酸性环境下,ENR分子( $pK_a=2.61$ )苯环上的羧基电离失去质子表现为给电子的状态,提高了苯环的电子云密度,增强了 ENR 分子与 FSGO MPs 间的 $\pi-\pi$  相互作用力,因此 FSGO MPs 可以高效地从水中萃取 ENR 分子.



(a) FSGO MPs 在自来水和湖水中对 50.0 ng·L<sup>-1</sup> QNs 萃取前后的 SERS 谱图,

(b) FSGO MPs 在自来水中对 1.0 µg·L<sup>-1</sup> SAs 萃取前后加标水样和解吸液的 SERS 谱图

#### 图 6 实际水样中 FSGO MPs 对 QNs 和 SAs 萃取前后的 SERS 谱图

Fig. 6 SERS spectra of FSGO MPs before and after extraction of QNs and SAs in actual water samples

#### 3 结论

制备了一种吸附容量高、分散性好且萃取速率快的磁性氧化石墨烯复合材料(FSGO MPs),并将其作为磁固相萃取介质对水样中 QNs 的萃取性能进行研究,结合磁性固相萃取(MSPE)样品前处理方法和表面增强拉曼光谱(SERS)技术,构建了饮用水水样中痕量 QNs 抗生素的快速检测方法.以自来水为例,该方法实现了ng·L<sup>-1</sup>水平 ENR 和CIP 的选择性有效检出,相比于无富集的直接检测

方法,其最低可检出浓度降低了 2 个数量级. 当面临有机质复杂的湖水等天然水体时,尚需要进一步提高 FSGO MPs 对 QNs 的选择性萃取能力. 总之,本方法具备操作简单、萃取性能佳、环境友好且可实现吸附材料重复利用的优势,为基质简单的饮用水中痕量 QNs 的定性和定量分析提供了一种快速、灵敏且高效的检测方法,有望在应急监测中发挥作用.

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### **CONTENTS**

E ' . ID EC' . ID'I CE ' C . ' . ' . I E . C . IE ' . WIANG V' I WILV WIANG C' . I	(4801)
Environmental Process, Effects and Risks of Emerging Contaminants in the Estuary-Coastal Environment  WANG Xin-hong, YU Xiao-xuan, WANG Si-quan, et al.  No. 10 To	
Research Progress of Analytical Methods with Molecular Spectroscopy for Determination of Trace Nutrients and Metals in Seawaters YUAN Dong-xing, HUANG Yong-ming, WANG Ting	
Research Progress on the Determination of Sulfide in Natural Waters: From Laboratory Analysis to In-Situ Monitoring	
Advances in On-site Analytical Methods for Inorganic Arsenic in Environmental Water BO Guang-yong, CHEN Zhao-ying, GONG Zhen-bin, et al.	(4845)
Advances and Prospect of Sampling Techniques and Analytical Methods for Trace Elements in the Ocean; Progress of Trace Element Platform Construction in Xiamen University	( 40 = 0 )
HUANG Yong-ming, ZHOU Kuan-bo, CHEN Yao-jin, et al.	(4858)
Biodegradation of Polyethylene Microplastic: A Review LUO Yuan-rong, QIAN Yi-qian, QI Ya-nan	
Mechanism and Environmental Effect on Nitrogen Addition to Microbial Process of Arsenic Immobilization in Flooding Paddy Soils WANG Feng, ZHANG Jing, ZHOU Shao-yu, et al.	
Toxicity Testing Organisms for Marine Ecotoxicological Research in China	
Estimating Methane Fugitive Emissions from Oil and Natural Gas Systems in China	
$Atmospheric~NH_{3}~Emission~Inventory~and~Its~Tempo-spatial~Changes~in~Xiamen-Zhangzhou-Quanzhou~Region~from~2015~to~2020~\cdots\cdots~UI~Xiang,~WU~Shui-ping,~JIANG~Bing-qi,~\itet~al.~Changes~in~Xiamen-Zhangzhou-Quanzhou~Region~from~2015~to~2020~\cdots\cdots~UI~Xiang,~WU~Shui-ping,~JIANG~Bing-qi,~\itet~al.~Changes~in~Xiamen-Zhangzhou-Quanzhou~Region~from~2015~to~2020~\cdots~\cdots~UI~Xiang,~WU~Shui-ping,~JIANG~Bing-qi,~\itet~al.~Changes~in~Xiamen-Zhangzhou-Quanzhou~Region~from~2015~to~2020~\cdots~\cdots~UI~Xiang,~WU~Shui-ping,~JIANG~Bing-qi,~\itet~al.~Changes~in~Xiamen-Zhangzhou-Quanzhou~Region~from~2015~to~2020~\cdots~\cdots~UI~Xiang,~WU~Shui-ping,~JIANG~Bing-qi,~\itet~al.~Changes~in~Xiamen-Zhangzhou-Quanzhou~Region~from~2015~to~2020~\cdots~\cdots~UI~Xiang,~WU~Shui-ping,~JIANG~Bing-qi,~\itet~al.~Changes~in~Xiamen-Zhangzhou-Quanzhou~Region~from~2015~to~2020~\cdots~\cdots~\cdots~UI~Xiang,~WU~Shui-ping,~JIANG~Bing-qi,~\itet~al.~Changes~in~Xiamen-Zhangzhou-Quanzhou~Region~from~2015~to~2020~\cdots~\cdots~\cdots~UI~Xiang,~MU~Shui-ping,~MU~Shui-p$	(4914)
Distribution of Microplastic and Antibiotic Resistance Gene Pollution in Jiulong River Estuary	(4924)
Pollution Characteristics of Microplastics in Sediments of Xiamen Bay Beach YAO Rui, LIU Hua-tai, LI Yong-yu, et al.	
Spatial and Temporal Distribution and Influencing Factors of Dissolved Trace Metals in Jiulong River Estuary and Xiamen Bay	
Spatiotemporal Characteristics of Dissolved Oxygen and Control Mechanism of Hypoxia (Low Oxygen) in the Watershed-Coastal System in Fujian Province	
YANG Ai-lin, YANG Fang, LI Shao-bin, et al.	
Distribution, Migration, and Transformation Mechanism of Labile Phosphorus in Sediments of Xixi River Estuary, Xiamen	(4961)
Adsorption of Mn <sup>2+</sup> by Modified Biochar Fixed Bed in Simulated Lakes and Reservoir Waters	
Rapid Detection of Trace Enrofloxacin and Ciprofloxacin in Drinking Water by SERS	(4982)
Degradation of Triphenyl Phosphate in Water by UV-driven Advanced Oxidation Processes	(4992)
Characteristics and Potential Sources of Four Ozone Pollution Processes in Hainan Province in Autumn of 2019	(5000)
Characterization and Formation Mechanism of Water-soluble Inorganic Ions in PM <sub>2,5</sub> and PM <sub>3,6</sub> in Summer in the Urban Agglomeration of the Ili River Valley	
CHEN Qiao, GU Chao, XU Tao, et al.	(5009)
Difference in PM <sub>2,5</sub> Pollution and Transport Characteristics Between Urban and Suburban Areas	
Characteristics and Health Risk Assessment of BTESX in the Northern Suburbs of Nanjing	(5030)
Ecological Risk Assessment of Microplastics Occurring in Surface Water of Terrestrial Water Systems across China	
Scale Effects of Landscape Pattern on Water Quality in Dongjiang River Source Watershed	
Mercury Speciation Distribution and Potential Sources in Surface Waters of the Yangtze and Yellow River Source Basins of Tibetan Plateau During Wet Season	
LIU Nan-tao, WU Fei, YUAN Wei, et al.	(5064)
Water Environmental Characteristics and Water Quality Assessment of Lakes in Tibetan Plateau	(5073)
Karst Hydrogeochemical Characteristics and Controlling Factors of Carlin-type Gold Mining Area Based on Hydrochemistry and Sulfur Isotope	(3013)
ZHA Xue-fang, WU Pan, LI Xue-xian, et al.	(5094)
Characteristics of Eukaryotic Phytoplankton Community Structure and Its Relationship with Environmental Factors in Danjiangkou Reservoir HE Yu-xiao, MAI Si-jie, REN Yu-fen, et al.	
Adsorption of Phosphate and Heavy Metals by Lanthanum Modified Zeolite and Its Performance in Sediment Inactivation	
Metagenomic and Metatranscriptomic Analysis of Nitrogen Removal Functional Microbial Community of Petrochemical Wastewater Biological Treatment Systems  WANG Zile, Zilo Juli, Li Weil, et al.  Metagenomic and Metatranscriptomic Analysis of Nitrogen Removal Functional Microbial Community of Petrochemical Wastewater Biological Treatment Systems	(3100)
ZHANG Xu, ZHOU Jia-jia, ZHOU Min, et al.	(5115)
Bacterial Community Structure and Antibiotic Resistance Gene Changes in IFAS + Magnetic Coagulation Process Wastewater Treatment Plant in Cold Regions	(3113)
DU Wen-yan, YAO Jun-qin, MA Hui-ying, et al.	(5122)
Nitric Oxide Emissions from Chinese Upland Cropping Systems and Mitigation Strategies: A Meta-analysis	(3123)
Nitro Oxide Emissions from Chinese Upiand Cropping Systems and Mitigation Strategies; A Meta-analysis  TIAN Zheng-yun, WU Xiong-wei, WU Yuan-yuan, et al.	
HAN Zheng-yun, WU Xiong-wei, WU Yuan-yuan, et al.	(5101)
	(5131)
Impact of Nitrification Inhibitors on Vegetable Production Yield, Nitrogen Fertilizer Use Efficiency and Nitrous Oxide Emission Reduction in China; Meta Analysis	(5131)
LIU Fa-bo, MA Xiao, ZHANG Fen, et al.	(5131)
Effect of Different Fertilization Treatments on Methane and Nitrous Oxide Emissions from Rice-Vegetable Rotation in a Tropical Region, China	(5131) (5140)
Effect of Different Fertilization Treatments on Methane and Nitrous Oxide Emissions from Rice-Vegetable Rotation in a Tropical Region, China  SHAO Xiao-hui, TANG Shui-rong, MENG Lei, et al.	(5131) (5140) (5149)
Effect of Different Fertilization Treatments on Methane and Nitrous Oxide Emissions from Rice-Vegetable Rotation in a Tropical Region, China	(5131) (5140) (5149) (5159)
Effect of Different Fertilization Treatments on Methane and Nitrous Oxide Emissions from Rice-Vegetable Rotation in a Tropical Region, China SHAO Xiao-hui, TANG Shui-rong, MENG Lei, et al.  Effects of Land-use Conversion on Soil Nitrification and NO & N2O Emissions in Tropical China Under Different Moisture Conditions TANG Rui-jie, HU Yu-jie, ZHAO Cai-yue, et al.  Characteristics of Heavy Metal Pollution in Farmland Soil of the Yangtze River Economic Belt Based on Bibliometric Analysis LIU Xiao-yan, FAN Ya-nan, LIU Peng, et al.	(5131) (5140) (5149) (5159)
Effect of Different Fertilization Treatments on Methane and Nitrous Oxide Emissions from Rice-Vegetable Rotation in a Tropical Region, China	(5131) (5140) (5149) (5159) (5169)
Effect of Different Fertilization Treatments on Methane and Nitrous Oxide Emissions from Rice-Vegetable Rotation in a Tropical Region, China	(5131) (5140) (5149) (5159) (5169) (5180)
Effect of Different Fertilization Treatments on Methane and Nitrous Oxide Emissions from Rice-Vegetable Rotation in a Tropical Region, China SHAO Xiao-hui, TANG Shui-rong, MENG Lei, et al.  Effects of Land-use Conversion on Soil Nitrification and NO & N2O Emissions in Tropical China Under Different Moisture Conditions TANG Rui-jie, HU Yu-jie, ZHAO Cai-yue, et al.  Characteristics of Heavy Metal Pollution in Farmland Soil of the Yangtze River Economic Belt Based on Bibliometric Analysis LIU Xiao-yan, FAN Ya-nan, LIU Peng, et al.  Spatial and Temporal Distribution and Source Variation of Heavy Metals in Cultivated Land Soil of Xiangzhou District Based on EBK Interpolation Prediction and GDM Model  GAO Hao-ran, ZHOU Yong, LIU Jia-kang, et al.  Identification of Soil Heavy Metal Sources Around a Copper-silver Mining Area in Ningxia Based on GIS  ZHANG Kou-kou, HE Jing, ZHONG Yan-xia, et al.	(5131) (5140) (5149) (5159) (5169) (5180) (5192)
Effect of Different Fertilization Treatments on Methane and Nitrous Oxide Emissions from Rice-Vegetable Rotation in a Tropical Region, China SHAO Xiao-hui, TANG Shui-rong, MENG Lei, et al.  Effects of Land-use Conversion on Soil Nitrification and NO & N2O Emissions in Tropical China Under Different Moisture Conditions TANG Rui-jie, HU Yu-jie, ZHAO Cai-yue, et al.  Characteristics of Heavy Metal Pollution in Farmland Soil of the Yangtze River Economic Belt Based on Bibliometric Analysis IUU Xiao-yan, FAN Ya-nan, LIU Peng, et al.  Spatial and Temporal Distribution and Source Variation of Heavy Metals in Cultivated Land Soil of Xiangzhou District Based on EBK Interpolation Prediction and GDM Model GAO Hao-ran, ZHOU Yong, LIU Jia-kang, et al.  Identification of Soil Heavy Metal Sources Around a Copper-silver Mining Area in Ningxia Based on GIS ZHANG Kou-kou, HE Jing, ZHONG Yan-xia, et al.  Effect of Aging on Stabilization of Cd <sup>2+</sup> Through Biochar Use in Alkaline Soil of Bayan Obo Mining Area WANG Zhe, CHENG Jun-li, BIAN Yuan, et al.	(5131) (5140) (5149) (5159) (5169) (5180) (5192)
Effect of Different Fertilization Treatments on Methane and Nitrous Oxide Emissions from Rice-Vegetable Rotation in a Tropical Region, China  SHAO Xiao-hui, TANG Shui-rong, MENG Lei, et al.  Effects of Land-use Conversion on Soil Nitrification and NO & N2O Emissions in Tropical China Under Different Moisture Conditions  TANG Rui-jie, HU Yu-jie, ZHAO Cai-yue, et al.  Characteristics of Heavy Metal Pollution in Farmland Soil of the Yangtze River Economic Belt Based on Bibliometric Analysis  LIU Xiao-yan, FAN Ya-nan, LIU Peng, et al.  Spatial and Temporal Distribution and Source Variation of Heavy Metals in Cultivated Land Soil of Xiangzhou District Based on EBK Interpolation Prediction and GDM Model  GAO Hao-ran, ZHOU Yong, LIU Jia-kang, et al.  Identification of Soil Heavy Metal Sources Around a Copper-silver Mining Area in Ningxia Based on GIS  ZHANG Kou-kou, HE Jing, ZHONG Yan-xia, et al.  Effect of Aging on Stabilization of Cd <sup>2+</sup> Through Biochar Use in Alkaline Soil of Bayan Obo Mining Area  Preparation of Marnetic Iron Oxide/Mulberry Stem Biochar and Its Effects on Dissolved Organic Carbon and Arsenic Speciation in Arsenic-Contaminated Soils	(5131) (5140) (5149) (5159) (5169) (5180) (5192) (5205)
Effect of Different Fertilization Treatments on Methane and Nitrous Oxide Emissions from Rice-Vegetable Rotation in a Tropical Region, China SHAO Xiao-hui, TANG Shui-rong, MENG Lei, et al.  Effects of Land-use Conversion on Soil Nitrification and NO & N2O Emissions in Tropical China Under Different Moisture Conditions TANG Rui-jie, HU Yu-jie, ZHAO Cai-yue, et al.  Characteristics of Heavy Metal Pollution in Farmland Soil of the Yangtze River Economic Belt Based on Bibliometric Analysis IUU Xiao-yan, FAN Ya-nan, LIU Peng, et al.  Spatial and Temporal Distribution and Source Variation of Heavy Metals in Cultivated Land Soil of Xiangzhou District Based on EBK Interpolation Prediction and GDM Model GAO Hao-ran, ZHOU Yong, LIU Jia-kang, et al.  Identification of Soil Heavy Metal Sources Around a Copper-silver Mining Area in Ningxia Based on GIS ZHANG Kou-kou, HE Jing, ZHONG Yan-xia, et al.  Effect of Aging on Stabilization of Cd <sup>2+</sup> Through Biochar Use in Alkaline Soil of Bayan Obo Mining Area WANG Zhe, CHENG Jun-li, BIAN Yuan, et al.	(5131) (5140) (5149) (5159) (5169) (5180) (5192) (5205)
Effect of Different Fertilization Treatments on Methane and Nitrous Oxide Emissions from Rice-Vegetable Rotation in a Tropical Region, China  SHAO Xiao-hui, TANG Shui-rong, MENG Lei, et al.  Effects of Land-use Conversion on Soil Nitrification and NO & N2O Emissions in Tropical China Under Different Moisture Conditions  TANG Rui-jie, HU Yu-jie, ZHAO Cai-yue, et al.  Characteristics of Heavy Metal Pollution in Farmland Soil of the Yangtze River Economic Belt Based on Bibliometric Analysis  LIU Xiao-yan, FAN Ya-nan, LIU Peng, et al.  Spatial and Temporal Distribution and Source Variation of Heavy Metals in Cultivated Land Soil of Xiangzhou District Based on EBK Interpolation Prediction and GDM Model  GAO Hao-ran, ZHOU Yong, LIU Jia-kang, et al.  Identification of Soil Heavy Metal Sources Around a Copper-silver Mining Area in Ningxia Based on GIS  ZHANG Kou-kou, HE Jing, ZHONG Yan-xia, et al.  Effect of Aging on Stabilization of Cd <sup>2+</sup> Through Biochar Use in Alkaline Soil of Bayan Obo Mining Area  Preparation of Marnetic Iron Oxide/Mulberry Stem Biochar and Its Effects on Dissolved Organic Carbon and Arsenic Speciation in Arsenic-Contaminated Soils	(5131) (5140) (5149) (5159) (5169) (5180) (5192) (5205)
Effect of Different Fertilization Treatments on Methane and Nitrous Oxide Emissions from Rice-Vegetable Rotation in a Tropical Region, China  SHAO Xiao-hui, TANG Shui-rong, MENG Lei, et al.  Effects of Land-use Conversion on Soil Nitrification and NO & N2O Emissions in Tropical China Under Different Moisture Conditions  TANG Rui-jie, HU Yu-jie, ZHAO Cai-yue, et al.  Characteristics of Heavy Metal Pollution in Farmland Soil of the Yangtze River Economic Belt Based on Bibliometric Analysis  LIU Xiao-yan, FAN Ya-nan, LIU Peng, et al.  Spatial and Temporal Distribution and Source Variation of Heavy Metals in Cultivated Land Soil of Xiangzhou District Based on EBK Interpolation Prediction and GDM Model  GAO Hao-ran, ZHOU Yong, LIU Jia-kang, et al.  Identification of Soil Heavy Metal Sources Around a Copper-silver Mining Area in Ningxia Based on GIS  ZHANG Kou-kou, HE Jing, ZHONG Yan-xia, et al.  Effect of Aging on Stabilization of Cd² + Through Biochar Use in Alkaline Soil of Bayan Obo Mining Area  WANG Zhe, CHENG Jun-li, BIAN Yuan, et al.  Preparation of Magnetic Iron Oxide/Mulberry Stem Biochar and Its Effects on Dissolved Organic Carbon and Arsenic Speciation in Arsenic-Contaminated Soils  LU Lin, YAN Li-ling, LIANG Mei-na, et al.	(5131) (5140) (5149) (5159) (5169) (5180) (5192) (5205) (5214) (5224)
Effect of Different Fertilization Treatments on Methane and Nitrous Oxide Emissions from Rice-Vegetable Rotation in a Tropical Region, China  SHAO Xiao-hui, TANG Shui-rong, MENG Lei, et al.  Effects of Land-use Conversion on Soil Nitrification and NO & N <sub>2</sub> O Emissions in Tropical China Under Different Moisture Conditions  TANG Rui-jie, HU Yu-jie, ZHAO Cai-yue, et al.  Characteristics of Heavy Metal Pollution in Farmland Soil of the Yangtze River Economic Belt Based on Bibliometric Analysis  LIU Xiao-yan, FAN Ya-nan, LIU Peng, et al.  Spatial and Temporal Distribution and Source Variation of Heavy Metals in Cultivated Land Soil of Xiangzhou District Based on EBK Interpolation Prediction and GDM Model  GAO Hao-ran, ZHOU Yong, LIU Jia-kang, et al.  Identification of Soil Heavy Metal Sources Around a Copper-silver Mining Area in Ningxia Based on GIS  ZHANG Kou-kou, HE Jing, ZHONG Yan-xia, et al.  Effect of Aging on Stabilization of Cd <sup>2+</sup> Through Biochar Use in Alkaline Soil of Bayan Obo Mining Area  WANG Zhe, CHENG Jun-li, BIAN Yuan, et al.  Preparation of Magnetic Iron Oxide/Mulberry Stem Biochar and Its Effects on Dissolved Organic Carbon and Arsenic Speciation in Arsenic-Contaminated Soils  LU Lin, YAN Li-ling, LIANG Mei-na, et al.  Effects of Oyster Shell Powder and Lime on Availability and Forms of Phosphorus and Enzyme Activity in Acidic Paddy Soil  ZHAO Li-fang, HUANG Peng-wu, YANG Cai-di, et al.	(5131) (5140) (5149) (5159) (5169) (5180) (5192) (5205) (5214) (5224) (5224)
Effect of Different Fertilization Treatments on Methane and Nitrous Oxide Emissions from Rice-Vegetable Rotation in a Tropical Region, China  SHAO Xiao-hui, TANG Shui-rong, MENG Lei, et al.  Effects of Land-use Conversion on Soil Nitrification and NO & N2O Emissions in Tropical China Under Different Moisture Conditions  TANG Rui-jie, HU Yu-jie, ZHAO Cai-yue, et al.  Characteristics of Heavy Metal Pollution in Farmland Soil of the Yangtze River Economic Belt Based on Bibliometric Analysis  LIU Xiao-yan, FAN Ya-nan, LIU Peng, et al.  Spatial and Temporal Distribution and Source Variation of Heavy Metals in Cultivated Land Soil of Xiangzhou District Based on EBK Interpolation Prediction and GDM Model  GAO Hao-ran, ZHOU Yong, LIU Jia-kang, et al.  Identification of Soil Heavy Metal Sources Around a Copper-silver Mining Area in Ningxia Based on GIS  ZHANG Kou-kou, HE Jing, ZHONG Yan-xia, et al.  Effect of Aging on Stabilization of Cd² + Through Biochar Use in Alkaline Soil of Bayan Obo Mining Area  Preparation of Magnetic Iron Oxide/Mulberry Stem Biochar and Its Effects on Dissolved Organic Carbon and Arsenic Speciation in Arsenic-Contaminated Soils  LIU Lin, YAN Li-ling, LIANG Mei-na, et al.  Effects of Oyster Shell Powder and Lime on Availability and Forms of Phosphorus and Enzyme Activity in Acidic Paddy Soil  ZHAO Li-fang, HUANG Peng-wu, YANG Cai-di, et al.  Effects of Interaction of Zinc and Cadmium on Growth and Cadmium Accumulation of Brassica campestris L.  SHUAI Zu-ping, LIU Han-yi, CUI Hao, et al.  Pollution Characteristics and Ecological Risk Assessment of Antibiotics in Vegetable Field in Kaizhou, Chongqing  FANG Lin-fa, YE Ping-ping, FANG Biao, et al.  Temporal and Spatial Variation Characteristics of Carbon Storage in the Source Region of the Yellow River Based on InVEST and GeoSoS-FLUS Models and Its Response to Different	(5131) (5140) (5149) (5159) (5169) (5180) (5192) (5205) (5214) (5224) (5224) (5234) (5244)
Effect of Different Fertilization Treatments on Methane and Nitrous Oxide Emissions from Rice-Vegetable Rotation in a Tropical Region, China  SHAO Xiao-hui, TANG Shui-rong, MENG Lei, et al.  Effects of Land-use Conversion on Soil Nitrification and NO & N <sub>2</sub> O Emissions in Tropical China Under Different Moisture Conditions  TANG Rui-jie, HU Yu-jie, ZHAO Cai-yue, et al.  Characteristics of Heavy Metal Pollution in Farmland Soil of the Yangtze River Economic Belt Based on Bibliometric Analysis  LIU Xiao-yan, FAN Ya-nan, LIU Peng, et al.  Spatial and Temporal Distribution and Source Variation of Heavy Metals in Cultivated Land Soil of Xiangzhou District Based on EBK Interpolation Prediction and GDM Model  GAO Hao-ran, ZHOU Yong, LIU Jia-kang, et al.  Identification of Soil Heavy Metal Sources Around a Copper-silver Mining Area in Ningxia Based on GIS  ZHANG Kou-kou, HE Jing, ZHONG Yan-xia, et al.  Effect of Aging on Stabilization of Cd <sup>2+</sup> Through Biochar Use in Alkaline Soil of Bayan Obo Mining Area  WANG Zhe, CHENG Jun-li, BIAN Yuan, et al.  Preparation of Magnetic Iron Oxide/Mulberry Stem Biochar and Its Effects on Dissolved Organic Carbon and Arsenic Speciation in Arsenic-Contaminated Soils  LU Lin, YAN Li-ling, LIANG Mei-na, et al.  Effects of Oyster Shell Powder and Lime on Availability and Forms of Phosphorus and Enzyme Activity in Acidic Paddy Soil  ZHAO Li-fang, HUANG Peng-wu, YANG Cai-di, et al.  Effects of Interaction of Zinc and Cadmium on Growth and Cadmium Accumulation of Brassica campestris L.  SHUAI Zu-ping, LIU Han-yi, CUI Hao, et al.  Pollution Characteristics and Ecological Risk Assessment of Antibiotics in Vegetable Field in Kaizhou, Chongqing  FANG Lin-fa, YE Ping-ping, FANG Biao, et al.	(5131) (5140) (5149) (5159) (5169) (5180) (5192) (5205) (5214) (5224) (5224) (5234) (5244)
Effect of Different Fertilization Treatments on Methane and Nitrous Oxide Emissions from Rice-Vegetable Rotation in a Tropical Region, China  SHAO Xiao-hui, TANG Shui-rong, MENG Lei, et al.  Effects of Land-use Conversion on Soil Nitrification and NO & N2O Emissions in Tropical China Under Different Moisture Conditions  TANG Rui-jie, HU Yu-jie, ZHAO Cai-yue, et al.  Characteristics of Heavy Metal Pollution in Farmland Soil of the Yangtze River Economic Belt Based on Bibliometric Analysis  LIU Xiao-yan, FAN Ya-nan, LIU Peng, et al.  Spatial and Temporal Distribution and Source Variation of Heavy Metals in Cultivated Land Soil of Xiangzhou District Based on EBK Interpolation Prediction and GDM Model  GAO Hao-ran, ZHOU Yong, LIU Jia-kang, et al.  Identification of Soil Heavy Metal Sources Around a Copper-silver Mining Area in Ningxia Based on GIS  ZHANG Kou-kou, HE Jing, ZHONG Yan-xia, et al.  Effect of Aging on Stabilization of Cd² + Through Biochar Use in Alkaline Soil of Bayan Obo Mining Area  Preparation of Magnetic Iron Oxide/Mulberry Stem Biochar and Its Effects on Dissolved Organic Carbon and Arsenic Speciation in Arsenic-Contaminated Soils  LIU Lin, YAN Li-ling, LIANG Mei-na, et al.  Effects of Oyster Shell Powder and Lime on Availability and Forms of Phosphorus and Enzyme Activity in Acidic Paddy Soil  ZHAO Li-fang, HUANG Peng-wu, YANG Cai-di, et al.  Effects of Interaction of Zinc and Cadmium on Growth and Cadmium Accumulation of Brassica campestris L.  SHUAI Zu-ping, LIU Han-yi, CUI Hao, et al.  Pollution Characteristics and Ecological Risk Assessment of Antibiotics in Vegetable Field in Kaizhou, Chongqing  FANG Lin-fa, YE Ping-ping, FANG Biao, et al.  Temporal and Spatial Variation Characteristics of Carbon Storage in the Source Region of the Yellow River Based on InVEST and GeoSoS-FLUS Models and Its Response to Different	(5131) (5140) (5149) (5159) (5169) (5180) (5192) (5205) (5214) (5224) (5224) (5234) (5244) (5244)
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