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聚乙烯微塑料的微生物降解研究进展

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摘要:现代工业的发展使得塑料制品的使用量急剧增加,由此产生的大量废旧塑料垃圾在环境中裂解形成粒径更小的微塑料(<5 mm).由于微塑料结构稳定,分布广泛且生物可利用性低,在环境中长期存在,已经逐渐成为对海洋生态和环境造成巨大影响的重要污染物.近年的研究表明,自然环境中存在一些能降解这些难降解微塑料的微生物,微生物降解无二次污染且对环境扰动少,在微塑料的去除中具有很好地应用潜力,但亦有一些局限性.综述了环境中数量最多的聚乙烯微塑料的微生物降解研究现状,着重探讨了降解效果和量化方法.基于微塑料生物分解效率普遍较低的现状,开展进一步的研究还非常有必要.

关键词:塑料垃圾;微塑料;聚乙烯(PE);生物降解;表征方法

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Biodegradation of Polyethylene Microplastic: A Review

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Abstract: Over the recent decades, global plastic production has grown dramatically due to the huge demands of consumption. As a consequence, large amounts of plastic waste have accumulated in the environment and will be cleaved into microplastics. Due to the low bioavailability, the microplastics will exist in the environment persistently and cause massive environmental stress. Plastic pollution is currently one of the biggest environmental concerns. Recent studies have shown the possibility to obtain degrading microorganisms of microplastics from the natural environment. Some microorganisms can break down microplastics into water and carbon dioxide. This paper reviewed the current research on biodegradation of polyethylene (PE), which is the most abundant microplastic type in the environment, and discussed the quantification methods of the degradation effect. Given that current biodegradation efficiency is relatively limited, further research is required.

Key words: plastic waste; microplastics; polyethylene (PE); biodegradation; quantification method

现代工业的发展使得塑料制品的使用急剧增加,全球塑料的产量从1950年的150t剧增到2020年的3.67亿t^[1],由此也造成了大量废旧塑料垃圾的产生.这些合成多聚物由于其化学结构稳定而难以被降解,在环境中经过风吹日晒而裂解,形成稳定的小颗粒,一般把直径5mm以下的塑料碎片、颗粒和纺织纤维统称为微塑料(microplastics).微塑料污染已经成为一个全球性的环境问题.

聚乙烯(polyethylene, PE)是全球生产量最高的塑料^[2],主要用途包括制作薄膜、包装材料和电线电缆等,它是一种以乙烯为原材料聚合加工得到的 热 塑 性 树 脂 材 料, 其 结 构 式 为:一[一CH2—CH2—]"—. 根据聚合方法、相对分子质量和碳链结构的不同,聚乙烯可被分为低密度聚乙烯(low density polyethylene, LDPE)、高密度聚乙烯(high density polyethylene, HDPE)、线性低密度聚乙烯(linear low density polyethylene,LLDPE)和低分子量聚乙烯(low molecular weight polyethylene,LMWPE)等.

合成塑料的产生,仅有不到80年的历史,原来普遍认为尚不足以进化出能降解塑料的微生物或

酶,近年来有报道则显示自然环境中存在着对塑料具有降解能力的微生物.尽管分解速率缓慢,但已有研究预示从自然环境中获得原本认为难以降解的塑料的降解微生物是可能的.由于反应条件温和且产物无污染,微生物降解逐渐被认为是可行的处理微塑料的方法^[3].本文综述了当前生物降解 PE 的研究现状,总结和分析 PE 降解菌株的种类、来源和降解机制并讨论了降解效果的量化方法,以期为科学评估微塑料降解效果及微塑料污染去除提供参考.

1 PE 生物降解效果的表征

由于 PE 的结构特性,生物可利用率较低,现有降解菌普遍分解缓慢,降解现象不够明显.为了确定微生物对 PE 的生物降解程度,一方面可通过监测微生物在以 PE 作为唯一碳源生长时的生长动力学变化,来判断对 PE 的降解能力;另一方面通过分析 PE 材料的性质变化,如表面官能团变化、机械性

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1.1 定性法

1.1.1 显微镜观察法

观察法通常使用原子力显微镜(atomic force microscope, AFM)和扫描电子显微镜(scanning electronic microscope, SEM)两种仪器. AFM 可观察微生物降解前后 PE 本身材质的变化,包括空洞、凹坑的出现和粗糙度的变化等; SEM 在此基础上还可观察微生物在 PE 表面的定殖情况,如生物膜的形成和细胞形态大小等. 因此通过与对照组比较,根据 PE 的变化和微生物的定殖情况,能够直观地判断接种微生物后 PE 表面是否发生降解现象.

Zahra 等^[5] 通过 SEM 发现培养 100 d 后 Aspergillus fumigatus 和 A. terreus 这 2 株真菌在紫外 线照射后的 LDPE 膜表面形成了明显的生物膜. Esmaeili 等^[6]的研究也显示, A. niger F1 与预先照 射紫外线的 LDPE 膜共培养后, SEM 显示其菌丝黏 附并穿透紫外线照射的 LDPE 膜表面. 同一研究中 细菌 Lysinibacillus xylanilyticus S7-10F 也在 LDPE 膜 表面形成了生物膜. Awasthi 等[7] 通过 SEM 图象观 察到,菌株 Klebsiella pneumoniae CH001 培养60 d 后 HDPE 膜表面粗糙,有许多裂缝和凹槽,而对照组则 保持光滑,AFM 图像也观察到表面拓扑结构的类似 变化,证明该菌株能够利用 PE 并形成凹槽. Auta 等^[8]接种 Bacillus 属细菌与 PE 粉末共培养 40 d 后, PE 表面变得粗糙并出现许多孔洞、侵蚀、裂缝和凹 槽. Yang 等^[9]观察到在经高压灭菌的 LDPE 膜表面 生长的 Enterobacter asburiae YT1 和 Bacillus sp. YP1 分别为 0.2 μm × 0.8 μm 的棒状细胞和 0.5 μm × 3μm 的不规则短棒状细胞.

1.1.2 FTIR 法

FTIR 法是利用傅里叶变换红外光谱仪检测微生物处理前后 PE 样品差异,通过官能团对应的吸收峰的变化判断 PE 结构是否因聚合物链断裂或氧化产物形成发生变化.

Skariyachan 等^[10]在3 375.46 cm⁻¹处发现了由于 C—H 键断裂形成的额外的峰,认为可能是由于生物降解导致的长链烃的分解. 经 Aspergillus flavus

PEDX3 处理后的 HDPE 微塑料颗粒与对照样品相 比,在3500~3100 cm⁻¹范围内出现对应羟基 (-OH)的宽吸收峰,还在1113 cm⁻¹和1647~ 1716 cm⁻¹处分别出现对应醚基(C-O-C)和羰基 (一C == O)的吸收峰,3种官能团的出现为HDPE微 塑料颗粒的生物氧化提供了证据[11]. Skariyachan 等[12]报道了处理后 LDPE 条由于羰基的形成在光 谱的1700 cm⁻¹和1600 cm⁻¹处出现新的吸收峰. Chen 等[13] 利用衰减全反射红外光谱(ATR-FTIR) 和光谱图像对在自然环境中暴露的 PE 降解程度进 行建模分析,结果表明可利用1720 cm⁻¹处的羰基 指数表征降解程度,为预测原位暴露微塑料降解程 度提供有效帮助. 而随着生物降解的进行,氧化聚合 物被微生物利用,羰基指数也将降低[14]. 因此使用 FTIR 检测羰基峰的形成对阐明 PE 降解机制有着重 要的作用[15].

1.1.3 GPC 法

GPC 法通过测定数均相对分子质量 (M_n) 、重均相对分子质量 (M_w) 和相对分子质量分布 (MWD),反映 PE 相对分子质量的变化,根据相对分子质量的损失判断 PE 长链结构是否解聚.

Yamada-Onodera 等[16] 通过 3 个月对 PE 相对分 子质量的监测验证了 Penicillium simplicissimum YK 对 PE 的降解作用, 当用热硝酸处理过的 PE 作为唯 -碳源时,相对分子质量高于100 000的 PE 会被真 菌降解为较低的相对分子质量. Zhang 等[11]利用高 温凝胶等效色谱(HT-GPC)对经 Aspergillus flavus PEDX3 处理过的 HDPE 微塑料颗粒样品进行检测 发现,与对照样品的相对分子质量(Mw = 222 003和 $M_n = 55135$)相比,处理后样品的 M_w (89801)和 M_n(26 064)分别下降了132 202和29 069,差异明显. 同样的,用菌株 Enterobacter asburiae YT1 和 Bacillus sp. YP1 培养的 PE 样品使用 HT-GPC 测定培养 60 布(MWD),与对照相比,显示出明显的负趋势.二者 培养的 PE 样品 M_w/M_n 分别为82 500/24 700 和 78 200/23 900, 降解组与对照相比减少了 6%~ 13%[9]. 然而值得注意的是,降解现象的发生也可 能导致相对分子质量增加,这是由于降解菌株优先 降解 PE 样品中低相对分子质量的部分,致使剩下 的PE样品的相对分子质量因生物降解而增 加^[17,18].

1.1.4 GC-MS 法

GC-MS 法是利用气相色谱-质谱联用仪对 PE 降解过程中的代谢产物进行分析.

Albertsson 等^[18] 利用 Arthrobacter paraffineus 降

解高温处理后的 LDPE 膜,3.5 a 后对照组(非生物 降解样品)中检测出一元和二元羧酸和酮酸的同系 物,而这些产物在生物降解样品中则完全被吸收,只 发现疑似醇类的物质. Abraham 等[19] 用曲霉菌处理 LDPE 膜,90 d 后 LDPE 膜上鉴定出烷烃、芳香烃、 饱和脂肪酸和不饱和脂肪酸等无毒化合物. Ambika 等[20]利用气相色谱法洗脱大量挥发性和半挥发性 化合物, GC-MS 分析鉴定了 Achromobacter Denitrificans s1 对 LDPE 的降解产物包括醛、酮、酯 和羧酸. 分别对经 Alternaria alternata FB1 处理 60 d 和120 d的PE 膜进行检测,结果表明处理60 d的 样品产物碳数范围在12~30之间,主要产物碳数为 27,占比为 51.24%,而时间延长到 120 d 后产物碳 数在 3~27 之间, 具有 4个碳的二甘醇胺是 Alternaria alternata FB1 降解 PE 膜的主要产物,占所 有产物的93.28%[3].

1.2 定量法

1.2.1 失重法

通过测定 PE 材料在微生物降解前后的质量 损失,并与对照进行比较,可以直观地以百分比判 断微生物对 PE 的降解效果,有利于横向比较不同 菌株的降解能力,失重法是目前使用较多的一种 方法.

将 Bacillus subtilis 接种在经紫外线照射且加入 表面活性剂的 PE 膜上,30 d 内减重 9.26% [21]. 菌 株 Pseudomonas sp. AKS2 可直接降解未经预处理的 LDPE 膜,45 d 内可降解(5±1)%的原始材料,若添 加矿物油改变疏水性,效果增强至(14±1)%^[22].在 15 d 的培养过程中,接种 Streptomyces albogriseolus LBX-2 菌株的 PE 粉末的重量损失显著增加. 此外, LBX-2 对低相对分子质量的 PE 粉末降解效果更 好[23]. 若是在同一实验体系内,由于降解对象的初 始状态且培养条件相同,失重法是比较不同降解菌 株降解效果的最简单直接的方法. 如对于 HDPE 的 降解, Pseudomonas sp. (15%) 比 Arthrobacter sp. (12%)表现出更好的降解效果^[24]. Devi 等^[25]从沿 海塑料垃圾倾倒地区土壤中分离出 248 株细菌并运 用失重法进行比较,最终得到了10株有效的 HDPE 降解菌.

1.2.2 CO, 生成量法

结合已有研究可知,微生物不论是好氧还是厌氧降解 PE, CO₂ 的产生都是 PE 被完全降解的标志. 有研究通过测量 CO₂ 产生量来表征降解效果. Shah 等 $^{[26]}$ 测定 Fusarium sp. AF4 降解 PE 膜,4 周后 CO₂ 产量 为 1.85 g·L $^{-1}$,显著高于空白对照的 0.45 g·L $^{-1}$. Abraham 等 $^{[19]}$ 的研究显示, Aspergillus

nomius JAPE1 和 Streptomyces sp. AJ1 降解 LDPE, 4 周后的 CO_2 产生量则分别为 2.85 $g \cdot L^{-1}$ 和 4.27 $g \cdot L^{-1}$.

定性法可以直观地表现微生物的生长或 PE 的表形变化,但是判断某一菌株是否为 PE 降解菌,通常需要多种方法确认.且由于定性法难以量化,无法对不同降解菌的降解能力进行横向比较.定量法虽然可以在同一体系中针对不同降解菌的降解能力进行比较,但由于降解对象本身质量很轻,操作过程中易造成较大误差,且不同体系使用的 PE 材质不同(是否含有添加剂)、反应条件不同(温度、时间),无法简单地凭借重量的失去判断降解能力地强弱.因此 PE 降解菌降解能力的量化仍然比较困难.降解效果的表征虽然可以直观体现降解菌对 PE 的降解能力,表征 PE 的生物降解尚需要几种方法结合,如 SEM 结合失重和 FTIR 法,从微生物的生长状态结合 PE 材料的变化,较全面地评价降解菌对 PE 的降解能力.

2 PE 的微生物降解

PE 的微生物降解最早可追溯到 20 世纪 70 年 代,通过检测 CO, 的生成,发现由 C14标记合成的 PE 能被微生物成功降解,并最终矿化为CO₂和H₂O,但 是过程极为缓慢[27]. 近年来,相继有不同的研究从 土壤、海洋和昆虫肠道等生境中筛选出对 PE 具有 降解能力的微生物[25,28],表1和表2分别归纳了目 前为止报道的 PE 降解细菌和降解真菌类群. 如表 中所示,已筛选出的 PE 降解细菌分属 4 个门,其中 变形菌门 γ-变形菌纲和放线菌门放线菌纲所包含 的降解菌种类最多. 相较于细菌, 真菌对于 PE 的降 解速率似乎更高[29]. 真菌能以菌丝体结构在不同基 质上生长,延伸到其他微生物难以生长的地方[2,30], 成为真菌降解 PE 的优势之一[4]. 已报道的 PE 降解 真菌分属子囊菌门和担子菌门下共4个纲,包括链 格孢菌、曲霉菌、镰孢菌和青霉菌等多个属,其中 与曲霉菌相关的研究较多.

Harshvardhan 等^[37]从阿拉伯海沿岸海水中分离出 60 株海洋细菌并从中筛选出 3 株 (Kocuria palustris M16、Bacillus pumilus M27 和 Bacillus subtilis H1584)能够在以 PE 为唯一碳源的培养基中生长的菌株,30 d 共培养后,PE 失重分别为 1%、1.5% 和 1.75%,FTIR 光谱也显示了酯、酮、乙烯基和内部双键的形成. Sangale 等^[46]在印度西海岸红树林根际土壤中得到的 109 株真菌中,从重量变化和拉伸强度变化两个角度,在不同 pH 条件下筛选出了 Aspergillus terreus MANGF1/WL (pH = 9.5) 和

表1 PE 降解细菌类群

Table 1 Bacterial taxa capable of PE degradation

Tane 1 Daceriai taza capanie oi 11 degradation				
类群		降解菌	文献	
变形菌门 (Proteobacteria)	β-变形菌纲 (β-Proteobacteria)	无色杆菌(Achromobacter)、丛毛单胞菌(Comamonas)和代尔夫特菌(Delftia)	[31,32]	
	γ-变形菌纲 (γ-Proteobacteria)	不动杆菌(Acinetobacter)、食碱菌(Alcanivorax)、肠杆菌(Enterobacter)、 微泡菌(Microbulbifer)、寡养单胞菌(Stenotrophomonas)和假单胞菌 (Pseudomonas)	[9,22,24,25,32~35]	
厚壁菌门 (Firmicutes)	芽孢杆菌纲 (Bacilli)	芽孢杆菌(Bacillus)和赖氨酸芽孢杆菌(Lysinibacillus)	[9,21,25,36~39]	
放线菌门 (Actinobacteria)	放线菌纲 (Actinomycetia)	节杆菌(Arthrobacter)、考克氏菌(Kocuria)和诺卡氏菌(Nocardia)、红球菌(Rhodococcus)、链霉菌(Streptomyces)和Leucobacter	[25,25,37,40]	
拟杆菌门 (Bacteroidetes)	鞘脂杆菌纲 (Sphingobacteriia)	鞘氨醇杆菌(Sphingobacterium)	[35]	

表 2 PE 降解真菌类群

Table 2 Fungal taxa capable of PE degradation

类	群	降解菌属	文献
	座囊菌纲 (Dothideomycetes)	链格孢菌(Alternaria)和毛色二孢菌(Lasiodiplodia)	[3,41]
子囊菌门 (Ascomycota)	散囊菌纲 (Eurotiomycetes)	曲霉菌(Aspergillus)、青霉菌(Penicillium)和拟青霉菌 (Paecilomyces)	[5,6,11,14,19,29,41~47]
	粪壳菌纲 (Sordariomycetes)	镰孢菌(Fusarium)、Purpureocillium、木霉菌(Trichoderma)、 炭角菌(Xylaria)和 Zalerion	[5,35,37,41,43,48,49]
担子菌门 (Basidiomycota)	伞菌纲 (Agaricomycetes)	Phanerodontia、Phlebiopsis 和裂褶菌(Schizophyllum)	[50]

Aspergillus sydowii PNPF15/TS (pH = 3.5) 两株 PE 降解真菌. 相较于海洋环境, 更多的 PE 降解菌筛选 自土壤环境. Watanabe 等[38]从田地土壤中筛选到3 株能够降解 LDPE 的菌株,分别属于 Bacillus circulans、B. brevies 和 B. sphaericus. Muhonja 等[29] 从富含塑料垃圾的土壤中分离具有 LDPE 降解能力 的细菌和真菌,培养16周后,其中曲霉菌 (Aspergillus oryzae A5,1)对 LDPE 膜的降解效率最 高,减重 (36.4 ± 5.53)%. Soleimani 等^[40]从含有 塑料垃圾的城市垃圾填埋场土壤中筛选可降解 LDPE 膜的放线菌,获得链霉菌属(Streptomyces)、诺 卡氏菌属(Nocardia)和红球菌属(Rhodococcus)这3 个属共17株降解菌,其中以链霉菌降解效果更好, 经 60 d 培养, Streptomyces sp. IR-SGS-T10 对 LDPE 膜重量降低速率为 1.58 mg·(g·d)⁻¹. Hou 等^[51]从 污水处理厂长期受塑料污染的活性污泥和废水中分 离到 54 株能以 PE 农用地膜为唯一碳源进行生长 的细菌,其中两株假单胞菌(Pseudomonas knackmussii N1-2 和 P. aeruginosa RD1-3)能有效降 解 PE 地膜,在 8 周的培养周期中, PE 地膜分别失重 (5.95±0.03)%和(3.62±0.32)%.此外,昆虫肠 道也是一个特殊的筛选来源. Yang 等[9]以 PE 作为 唯一碳源在食用 PE 膜的印度谷螟幼虫肠道中分离

出 Bacillus sp. YP1 和 Enterobacter asburiae YT1 两株细菌,共培养 28 d 后 LDPE 膜的物理性能、化学结构和相对分子质量等均发生改变. Zhang 等[11]将从蜡蛾 肠 道 分 离 出 的 Aspergillus flavus PEDX3 与HDPE 微塑料共培养 28 d 后,FTIR 检测到新的官能团出现,相对分子质量降低,且两个与降解相关的基因表达上调.

有研究表明,将 PE 膜进行预处理,如紫外线照 射或添加表面活性剂,可以增强降解菌对 PE 的降 解效果. 这是由于紫外线可以使聚乙烯链发生解聚, 生成烷烃、烯烃、酮、醛等低相对分子质量产物[9], 而表面活性剂可以提高 PE 的水溶性,从而提高它 的生物可利用性. Montazer 等[35] 以紫外线处理过的 LDPE 膜为降解对象,利用 Sphingobacterium moltivourum IRN11 和 Acinetobacter pitti IRN19 对其 进行降解,结果表明二者均可在 LDPE 膜表面产生 表面活性剂并定殖,培养 4 周后 Acinetobacter pitti IRN19 降解了 (26.8 ± 3.04)% 的 PE 薄膜. Dwicania 等[52] 发现将铜绿假单胞菌 (Pseudomonas aeruginosa)和短杆菌(Brevibacterium sp.)混合培养, 在 25 °C, pH = 7 的条件下, 30 d 后失重测定, 去除了 2%~7%的经紫外线预处理的 LLDPE. Sowmya 等[49]比较了 PE 膜经高压灭菌、膜表面消毒和紫外

线照射 3 种处理后,木霉菌(Trichoderma harzianum) 对其的降解效果,3 个月后该菌株对这 3 种处理的PE 的降解效率分别是 23%、13% 和 40%,其中漆酶和锰过氧化物酶 2 种降解酶的活性在第 10 周达到最高.将 LDPE 片和 HDPE 片预先用表面活性剂Tween-80 处理,再经 Penicillium oxalicum NS4 和Penicillium chrysogenum NS10 降解,取得较好的降解效果,90 d 后菌株 NS4 对 LDPE 和 HDPE 的降解率分别为 36.60% 和 55.34%,菌株 NS10 的降解率为 34.35% 和 58.60% [14].

3 PE 生物降解影响因子和降解机制

PE 的生物降解受化合物、微生物和环境因子这 3 方面的影响. 从化合物角度, PE 的分子量、表面官能团(羰基、酯、乙烯基和碳碳双键)组成和疏水/亲水性等影响其生物可利用性, 一般来说带氧化基团和亲水性较好的 PE 更易被微生物降解⁵³³. 从微生物角度,微生物能否顺利在 PE 表面定殖, 形成生物膜,并分泌相应的酶对降解至关重要. 而温度、pH 和培养时间等因素则影响 PE 的微生物降解效率.

微塑料 PE 的微生物降解通常包括 4 个阶 段^[3,4,29,50,54]: 劣化 (deterioration)、解聚 (depolymerization)、同化(assimilation)和矿化 (mineralization). 劣化阶段是指 PE 在外因作用下 (如紫外线照射或微生物分泌的胞外酶)物理化学 性质发生改变[55,56],在此过程中引入羰基,随后被 氧化形成羧酸[54,57];解聚阶段 PE 在微生物分泌的 特定酶的作用下进一步被分解成低聚物、二聚体和 单体,即易于吸收的低相对分子质量的片段[2,58]; 这些解聚产物中的一些水溶性短链中间体被受体识 别, 跨膜转运到微生物体内参与多种代谢活动, 并将 这些物质代谢为碳或能量来源[3,56],该过程被称为 同化; 最终微生物通过有氧或无氧代谢将这些短链 低相对分子质量化合物转化为 CO,、H,O 和 CH, 从 细胞中排出,由此产生的能量供给细胞生长[55,56,59]. 由此可见,劣化和解聚是 PE 降解的重要步骤,只有 特定的微生物才能分泌相应的降解相关酶类,以达 到降解的目的. 目前已报道的参与劣化与解聚阶段 的 PE 降解酶类群包括[11,60~64]:单加氧酶(如烷烃 单加氧酶)、过氧化物酶(锰过氧化物酶、大豆过氧 化物酶和谷胱甘肽过氧化物酶等)和漆酶等. Jeon 等[65]发现烷烃单加氧酶基因直接参与热解制备的 低分子量 PE 的生物降解,而红氧还蛋白酶和红氧 还蛋白还原酶通过相关电子的转移间接参与. Santo 等[61] 发现从 Rhodococcus ruber C208 中分离出的漆 酶具有高热稳定性和铜诱导性,与 PE 共培养两周后,羰基指数增加且 PE 分子量减少,证明漆酶参与了劣化与解聚过程. Zhao 等[66]使用大豆过氧化物酶对 HDPE 进行改性,在处理过的 HDPE 表面检测到 羟基和羰基基团,表明酶的处理显著增加了材料亲水性.

尽管在 PE 的微生物降解机制方面已经有研究,但总体而言,当前的研究还不够深入,对于降解中间产物的产生和微生物利用途径仍然未知.

4 结论与展望

相对分子质量高、疏水性强和化学性质稳定等特点使环境中的微塑料难以自行降解,尽管越来越多的 PE 降解菌被筛选出,但是由于降解现象不明显和分解速率普遍比较缓慢等原因,降解菌的筛选难度还比较大,本文认为,在降解菌筛选方法上可以进行适当的改进,如原位环境的定向驯化、稳定同位素探针技术(stable isotope probe, SIP)结合实验室培养和基于合成微生物组的高效降解菌群构建等方法,可以提高从环境中分离降解菌的效率.此外,对PE 降解菌的特性研究、降解条件的优化和降解机制的研究,对于提高降解菌的降解效率,并为其成为潜在的修复菌源也具有很重要的意义.综上,进一步开展对 PE 具有高效降解能力的微生物筛选和特异性研究仍然很有必要.

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