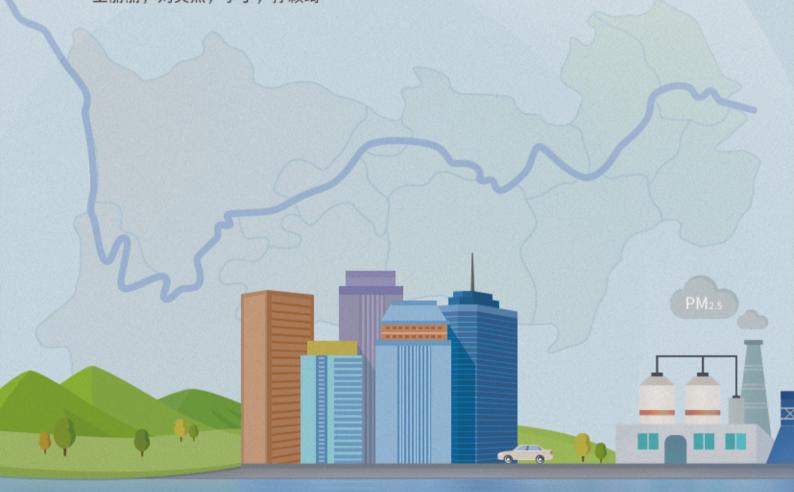




ENVIRONMENTAL SCIENCE

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长江经济带PM2.5空间异质性和驱动因素的地理探测 王丽丽, 刘笑杰, 李丁, 孙颖琦



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基于贝叶斯网络的给水管网消毒副产物生成因素分析

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摘要:给水管网消毒副产物(disinfection byproducts, DBPs)的生成受管网环境因素、微生物群落特征和水厂未完全去除的有机物等多指标共同影响. 各指标间相互关联形成复杂的网络结构,导致影响 DBPs 在管网内生成的主控因子较难确定. 以广州某高校给水管网系统为研究对象,于 2021 年 1~2 月开展终端水质调查,利用吹扫捕集-气相色谱质谱法测定三卤甲烷赋存水平,利用超高效液相色谱-串联三重四级杆质谱法测定抗生素及亚硝胺类 DBPs 质量浓度,利用高通量测序确定微生物群落组成. 基于实测数据,构建贝叶斯网络模型,定量分析各水质指标间的相互关联. 结果表明,管网中三卤甲烷和亚硝胺类 DBPs 质量浓度分别为 18. 33~32. 09 μg·L⁻¹和 13. 08~53. 50 ng·L⁻¹. 共检出 23 种抗生素,质量浓度范围为 47. 92~210. 33 ng·L⁻¹. 高通量测序发现 236 种细菌物种,优势菌群为 Rhizobiales 和 Caulobacterales. 贝叶斯网络推理发现,四环素类、磺胺类和大环内酯类抗生素为三卤甲烷前体物,同时四环素也是亚硝胺类 DBP 前体物. Caulobacterales 和 Corynebacteriales 丰度易受抗生素影响,其胞外聚合物是 DBPs 生成重要前体物. 结果可为 DBPs 的前体物研究和前端控制提供理论支持.

关键词:抗生素: 贝叶斯网络: 消毒副产物: 给水管网: 微生物群落

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Factor Analysis of Disinfection Byproduct Formation in Drinking Water Distribution Systems Through the Bayesian Network

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Abstract: Disinfection byproducts (DBPs) in drinking water distribution systems are affected by multi-factors, such as basic water quality parameters, microbial community structures, and residual organic pollutants that cannot be removed by the water treatment process. The relationship between the above-mentioned factors that forms a complicated network structure, which causes the dominating factor that affects DBPs formation unclear. This study investigated the water quality in regional tap water in January-February 2021. Trihalomethanes were determined using P&T-GC-MS, and antibiotics and nitrosamines were determined using UPLC-MS/MS. Microbial communities were determined using Illumina 16S rRNA gene sequencing. A Bayesian network was constructed to evaluate the intercorrelation between the factors. Three species of trihalomethanes, six species of nitrosamines, 23 types of antibiotics, and 236 OTUs were detected in the tap water. The mass concentrations of trihalomethanes, nitrosamines, and antibiotics were 18. 33-32.09 µg·L⁻¹, 13. 08-53.50 ng·L⁻¹, and 47. 92-210.33 ng·L⁻¹, respectively. The dominant microbial orders were Rhizobiales and Caulobacterales. Based on the Bayesian-network inference, tetracycline, sulfonamides, and macrocyclic antibiotics were precursors of trihalomethanes, whereas tetracyclines were the nitrosamine precursor. The abundances of Caulobacterales and Corynebacteriales were both affected by antibiotics and associated with DBPs formation. The extracellular polymeric substances of these bacteria were highly suspected to be important DBPs precursors. The results of the proposed project revealed the internal relationship between multi-water-quality parameters and DBPs formation, which could provide a theoretical support to guarantee the safety of drinking water.

Key words: antibiotics; Bayesian network; disinfection byproducts; drinking water distribution system; microbial community

饮用水消毒和传输过程中,余氯会和水中多种前体物反应生成具有遗传毒性、细胞毒性、潜在致突变性和致畸性的消毒副产物(disinfection byproducts, DBPs).目前已有700种DBPs被发现.其中,三卤甲烷(trihalomethanes, THMs)为首次发现的含碳氯化消毒副产物,已被列入受控清单;亚硝胺是一类非卤代含氮DBPs,因其致癌风险高而备受关注[1].DBPs前体物包括天然有机质、管壁生物膜分泌的胞外聚合物和水厂未完全去除的微污染物如抗生素等[2~4].因此,DBPs生成潜能受控于前体物特征,包括天然有机质性质、微生物群落结构和抗生素赋存水平[5,6].与此同时,DBPs生成还受到

管网环境因素,如 pH 和温度等影响^[7].

近年来,研究热点逐步从 DBPs 与单一水质指标或环境因素的相关性分析深入至与多水质指标的因果关系分析. 以管网抗生素为例,磺胺嘧啶和环丙沙星的加入使管网出水中总细菌和 Mycobacteria avium 增加,造成具有抗药性的微生物分泌更多的

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胞外聚合物^[8],最终促进 DBPs 生成;即通过微生物,形成了抗生素与 DBPs 的因果链条. 抗生素浓度变化还可能导致生物膜群落结构发生变化,使得具有较强抗生素耐药性的微生物成为主导菌种^[9],进而改变胞外聚合物的组分并最终影响 DBPs 的生成种类. 有研究发现, Pseudomonas putida 和Pseudomonas aeruginosa 的胞外聚合物主要成分分别为蛋白质和聚多糖^[10,11]. 由于蛋白质较多糖更易生成卤乙酸,因此 P. putida 胞外聚合物的卤乙酸生成势高于 P. aeruginosa^[12]. 由此可见,抗生素-微生物-DBPs 间不仅相互关联,还可形成具有一定逻辑关系的网络结构. 由于实验研究难以厘清上述结构,因此亟需更加灵活和精确的方法来描述 DBPs 与各水质指标的因果关系.

贝叶斯因果概率推理模型,可以在复杂的关系 网络中,基于算法评估各输入因素的不确定性[13], 推理出局部因素的相互关联和整体网络的层次关 系[14,15];可以快速有效地建立含有大量输入参数的 复杂网络系统,并将关系网络的逻辑性可视化[16]. 目前已有研究将贝叶斯网络应用于环境科学研 究^[17~19]. 尽管如此, Aguilera 等^[20]的研究认为贝叶 斯网络在解决环境类问题上的应用还不够充分,在 给水管网水质分析中的应用潜力还有待开发. Li 等[7]通过大量文献调研进一步指出,给水管网中的 环境因素、微生物指标和 DBPs 赋存水平间存在复 杂关系网,贝叶斯网络是适合梳理该网络中因果关 系的模型工具. 综上,本研究选择三卤甲烷和亚硝胺 为目标 DBPs,构建管网基本水质参数、抗生素浓度 和微生物群落结构的 4 层贝叶斯网络,依靠模型解 析影响 DBPs 生成的主控因子,以期为给水管网的 优化运行提供科学依据和理论支持,对保障用户健 康具有十分重要的现实意义.

1 材料与方法

1.1 样品采集

本研究选取广州市大学城某高校的给水管网为研究对象,依据均匀布点原则进行水样采集和检测.研究区域内共设置7个采样点,其中4个采样点位于教学区内,3个采样点位于生活区内.教学区采样点为泵房、教学楼一栋(教一)、教学楼六栋(教六)和理学楼三栋(理三);生活区采样点为北区超市(北超)、南区食堂(南食)和南宿舍八栋(南八).校区总面积约1 km²(约1500亩),教学区管网总长7.5 km,生活区管网总长6.3 km.

采样时间为2021年1~2月,所有采样点均未配置终端净水装置.采样前,打开水阀5 min 以上至



图 1 给水管网采样点分布示意

Fig. 1 Distribution map of sampling sites in the study area

水温恒定. 抗生素水样采集储存于 1 L 棕色玻璃瓶中,并加入 50 mL 甲醇和 0.4 mL 4 mol·L⁻¹ 硫酸/水溶液. 亚硝胺水样采集储存于 2 L 棕色玻璃瓶中,加入适量碳酸氢钠调节 pH = 8. 三卤甲烷水样采集储存于 40 mL 玻璃瓶中,装流溢满,加入 15 mg 无水硫代硫酸钠. 上述指标均设置 3 个平行和质控样. 采集15 L 水样于 70% 乙醇消毒的聚丙烯塑料采样桶中,用于微生物水样处理. 抗生素和 DBPs 水样于 4℃保存,并在 48 h 内进行前处理. 微生物样品采集后送至实验室及时处理. 常规水质参数如溶解氧、电导率、水温和 pH 等通过便携式水质检测仪(HACHHQ30D)现场测定. 余氯和总氯使用便携式分光光度计(HACH DR900)现场检测.

1.2 化学试剂

抗生素内标磺胺甲嘌唑-D₄、红霉素-¹³C-D₃、噻 苯咪唑-D₄、环丙沙星-D₈、甲氧苄啶-D₃和林可霉 素-D, 购自 Toronto Research Chemicals 公司(加拿 大),磺胺甲基嘧啶-D4 购自 Dr. Ehrenstorfer 公司 (德国),磺胺二甲嘧啶-13C6 购自 Cambridge Isotope Laboratories 公司(美国),甲氯环素购自 Sigma-Aldrich 公司(美国);亚硝胺替代物 N-亚硝基二甲 胺-D₆(NDMA-D₆, 1000 μg·mL⁻¹)和内标 N-亚硝基 正丙胺-D₁₄ (NDPA-D₁₄, 1000 μg·mL⁻¹)购自 Accustandard 公司(美国); 乙二胺四乙酸四钠(AR, 纯度 > 99%) 和甲醇(AR,纯度 > 99.5%) 购自天津 大茂试剂公司;硫酸(AR,纯度>95%)购自广州化 学试剂公司;无水硫代硫酸钠(纯度>99%)和碳酸 氢钠(ACS)购自上海阿拉丁生化科技股份有限公 司;乙酸乙酯(HPLC)、二氯甲烷(HPLC)、甲酸 (HPLC)和甲醇(HPLC)购自 Merck 公司(德国);

正己烷(HPLC)购自 CNW 公司(德国);超纯水使用 MilliQ 超纯水系统生产(德国 Merck Millipore 公司). 甲醇(AR)仅用于样品采集,样品前处理和上机检测使用甲醇(HPLC).

1.3 测试方法

抗生素和亚硝胺检测均采用固相萃取法对样品进行前处理^[21]. 抗生素水样中加入 0.5 g 乙二胺四乙酸四钠和 100 μL 1 mg·L⁻¹混合内标;依次通过10 mL 甲醇和 10 mL 超纯水活化 Oasis HLB 小柱(6 mL, 200 mg);以 5~10 mL·min⁻¹流速加载水样;上样结束后,额外加入 2次 25 mL 5% 甲醇/水溶液于样品瓶中润洗,并过柱;然后加入 10 mL 超纯水于小柱中,洗去残留的乙二胺四乙酸四钠;真空抽干小柱中残留水分;完成后向小柱内依次加入 3 mL 二氯甲烷,4 mL 乙酸乙酯和 5 mL 甲醇洗脱;氮气吹干洗脱液,并用 0.22 μm 聚醚砜滤膜以 1 mL 甲醇定容于棕色进样小瓶用于上机检测.

亚硝胺前处理方法参考 US EPA521 方法 $^{(22)}$ 进行改进. 水样中加入 50 μ L 200 μ g·L $^{-1}$ NDMA-D₆替代物; 依次通过 10 μ L 正己烷、10 μ L 二氯甲烷、12 μ L 甲醇和 12 μ L 超纯水活化冲洗椰子壳活性炭小柱 (CNWBOND Coconut Charcoal, 6 μ L, 2 g, 80~120 目); 以 15 μ L· μ min $^{-1}$ 流速加载水样; 上样结束后加入 10 μ L 三氯甲烷用于洗脱柱上富集的亚硝胺物质; 洗脱液中加入 400 μ L 超纯水混合; 氮气洗脱液至 400 μ L 后加入 100 μ L 超纯水定容, 过 0. 22 μ m 聚四氟乙烯滤膜后加入 25 μ L 1 μ mg·L $^{-1}$ 内标NDPA-D₁₄, 最后保存于棕色进样小瓶.

抗生素和亚硝胺检测采用 ACQUITY 超高效液相色谱-串联三重四级杆质谱联用仪 (ACQUITY UPLC I-Class-Xevo TQ-S micro, 美国 Waters 公司). 抗生素检测色谱柱选用 Waters ACQUITY UPLC BEH C18 柱(50 mm × 2.1 mm, 1.7 μ m);流动相为 0.1%甲酸水 A 和甲醇 B,流速为 0.3 mL·min⁻¹;梯度洗脱,即 0 ~ 0.5 min 为 80% A, 0.5 ~ 6 min 为 0% A, 6 ~ 6.5 min 为 80% A;柱温为 40℃;样品室温度为 10℃;进样量为 5 μ L. 质谱采用电子轰击源负离子模式(ESI –)进行检测,离子源温度为 150℃,脱溶剂温度为 400℃,脱溶剂气流速为 800 L·h⁻¹,毛细管电压为 3.5 kV,碰撞气体为高纯 氩气.

亚硝胺检测色谱柱选用 Waters ACQUITY UPLC HSS T3 柱(100 mm×2.1 mm, 1.8 μm); 流动相为 0.1% 甲酸水 A 和甲醇 B,流速为 0.4 mL·min⁻¹; 梯度洗脱,即 0~1 min 为 90% A, 1~3 min 为 88% A, 3~9 min 为 10% A, 9~11 min 为 95% A. 质谱

采用电子轰击源正离子模式(ESI+)进行检测;其他条件与抗生素质谱条件相同.

三卤甲烷检测采用吹扫捕集-气相色谱质谱联用仪(P&T-GC 7890A-MS 7000,美国 Aglient 公司). 吹扫流量为 $40 \text{ mL} \cdot \text{min}^{-1}$; 吹扫温度为 20°C ,吹扫 11 min; 解析温度为 220°C ,解析 2 min; 烘烤温度为 260°C ,烘烤 2 min. 色谱条件为毛细管色谱柱 HP-5UIms($30 \text{ m} \times 250 \text{ } \mu\text{m} \times 0.25 \text{ } \mu\text{m}$),进样口温度为 220°C ,分流比为 20:1; 程序升温设置为: 35°C 保持 5 min,然后以 $6 \text{ $^{\circ}\text{C}} \cdot \text{min}^{-1}$ 的速度升至 100°C ,再以 $20 \text{ $^{\circ}\text{C}} \cdot \text{min}^{-1}$ 的速度升至 210°C 保持 1 min. 质谱采用 100°C ,溶剂延迟 100°C ,溶剂延迟 100°C ,容别延迟 100°C ,不可以 1000°C ,不可

1.4 微生物高通量测序

微生物水样富集于孔径为 $0.22~\mu m$ 的水相滤膜,滤膜放置于 2~mL 无菌离心管中,避光保存于 -80° C,送至上海美吉生物医药科技有限公司进行微生物 PCR 扩增和高通量测序. 扩增引物为:338F (5'-ACTCCTACGGGAGCCAGCAG-3')和 806R (5'-GGACTACHVGGGTWTCTAAT-3'),扩增区间为 V3 ~ V4 可变区. PCR 扩增产物在 Illumina-MiSeq 平台进行测序.

1.5 贝叶斯网络

利用贝叶斯网络对多水质指标相关性建立有向层次网络结构. 贝叶斯网络是一种用于因果概率推理的有向概率图模型,模型通过有向无环图定性和条件概率分布定量. 有向图由含有随机变量的节点和表示节点间的概率相依性的有向边组成. 有向边的尾部定义为父节点,而头部为子节点. 每个子节点 X_i 使用条件概率分布来列出其相对于父节点 X_i^i , …, $X_i^{n_i}$ 的所有可能的条件概率. 对于离散变量使用条件概率表来表示条件概率分布,而连续变量子节点则使用线性高斯模型函数计算见式(1).

$$(X_i \mid X_i^1, \dots, X_i^{n_i}) \sim N(b_0 + b_1 X_i^1 + \dots + b_{n_i} X_i^{n_i}, s^2)$$
(1)

式中, b_{n_i} 为 X_i 对第 n_i 个父节点的回归系数,s 为标准差. 贝叶斯网络定义了从全体节点的联合概率分布或全局概率分布到每个节点间的局部概率分布的因子化过程. 因子化由马尔科夫性质表示,即每个子节点的概率只依赖于父节点,关系式见式(2)和式(3):

$$P(X_1, X_2, \dots, X_n) = \prod_{i=1}^n P[X_i \mid \pi(X_i)]$$
 (2)

或
$$f(X_1, X_2, \dots, X_n) = \prod_{i=1}^n f[X_i \mid \pi(X_i)]$$
 (3)
式中, P 和 f 分别为离散变量和连续变量概率分布;

 $\pi(X_i)$ 是节点 X_i 的父节点集合,若节点 X_i 没有父节点,则 $\pi(X_i) = P(X_i)$ 或 $f(X_i)$; $P[X_i \mid \pi(X_i)]$ 或 $f[X_i \mid \pi(X_i)]$ 是父子节点间的局部概率.

本研究首先通过相关性分析和文献信息确定贝叶斯网络基本结构,然后利用结构学习算法获得最优网络结构,并建立最终的贝叶斯网络结构.结构学习算法选用的 tabu 搜索算法;该算法是一种评分类亚启发式随机算法^[23],相较于其他算法,如约束类或评分-约束混合类算法,在贝叶斯网络结构建立的准确度和速度方面具有显著优势^[24].

1.6 数据处理

抗生素和亚硝胺数据利用 Masslynx 软件进行处理. 三卤甲烷质谱数据通过 MassHunter 软件进行处理. 采用 Excel 2010 和 Origin 2018 软件进行数据分析. ANOVA TEST ($\alpha=0.05$)通过 IBM SPSS Statistics 25 软件计算. 应用 Chao 指数、Shannon 指数和覆盖率进行微生物 α 多样性分析. 利用上海美吉生物医药科技有限公司信息平台, 采用 Bray-

Curtis 距离算法,以样品平均层级聚类方法对微生物群落相对丰度进行聚类分析. 相关性分析和贝叶斯网络构建及分析使用 R-4.0.6 进行编译计算. 相关性分析利用 Pearson 相关系数衡量基本水质参数、抗生素、三卤甲烷和亚硝胺间相关性; 采用 Mantel 检验计算各微生物群落矩阵与各基本水质参数和微污染物之间的相关性.

2 结果与讨论

2.1 给水管网水质参数、抗生素和 DBPs 分布特征 给水管网水样为弱碱性 (pH 7.88~8.01),各 采样点水温稳定在 14.6~15.9℃之间, ρ (溶解氧) 范围为 9.82~10.01 mg·L⁻¹,电导率范围 327~332 μ S·cm⁻¹. ρ (余氯)和 ρ (总氯)在泵房最高,分别为 0.61 mg·L⁻¹和 0.72 mg·L⁻¹.生活区管网 pH 略高于教学区,其他基本水质参数在教学区和生活区未见明显差异(ANOVA TEST,P > 0.05),如表 1 所示.

表 1 给水管网各采样点基本水质参数

	Tabl	e I Primary water	quality parameters of t	ne arınkı	ng water in each samplin	g site	100
区域	地点	ρ(溶解氧)	电导率	水温	(1) (ff) "	ρ(余氯)	ρ(总氯)
	地点	/mg•L ⁻¹ //	/μS•cm ⁻¹	℃	7/ 11/0	/mg·L ⁻¹	/mg·L ⁻¹
3	泵房	9.87	329	15. 1	7. 94	0.61	0. 72
教学区	教一	9. 82	328	14. 6	7.91	0. 44	0. 56
教士区	教六	9. 99	332	15. 3	7. 88	0.46	0. 59
(a 1/1	理三	10. 01	330	15. 9	7. 99	0.51	0. 61
N3 11	北超	9. 96	331	15. 2	7. 90	0. 58	0. 70
生活区	南食	9. 90	327	15. 4	7. 99	0. 56	0. 68
	南八	9. 83	329	14.8	8. 01	0. 54	0.65

饮用水样品中共检出 23 种抗生素. 由图 2 可见,检出ρ(抗生素)为 47.92~210.33 ng·L⁻¹. 其中以氯四环素和甲烯土霉素为主的四环素类抗生素,占总抗生素质量浓度的 68.6%~85.6%. 氟喹诺酮类和磺胺类抗生素分别占 1.6%~5.6% 和 0.1%~12.6%. 检测出的氟喹诺酮类主要有培氟沙星和达诺沙星等,而磺胺类抗生素主要为磺胺甲基嘧啶、磺胺二甲嘧啶和磺胺间甲氧嘧啶等. 以脱水红霉素和罗红霉素为主的大环内酯类占 1.0%~12.8%. 除四环素类抗生素外,其它各类抗生素质量浓度略低

对饮用水中 4 种三卤甲烷和 9 种亚硝胺进行检测, 其 中 三 卤 甲 烷 类 DBPs 包 括 三 氯 甲 烷 (trichlormethane, TCM)、二 氯 一 溴 甲 烷 (bromodichloromethane, BDCM)、一氯二溴甲烷 (dibromochloromethane, DBCM) 和 三 溴 甲 烷 (tribromomethane, TBM); 亚硝胺包括 N-亚硝基甲基乙基胺(N-nitrosomethylamine, NMEA)、N-亚硝基

于国内饮用水抗生素水平[25]. 抗生素在教学区和生

活区间未发现显著性差异(P>0.05).

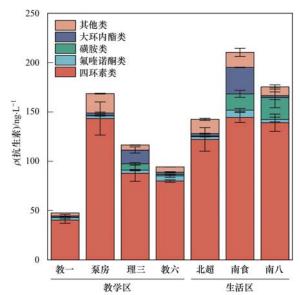


图 2 给水管网末端出水中抗生素的空间质量浓度分布

Fig. 2 Spatial distribution of the antibiotics mass concentrations $in \ the \ drinking \ water \ distribution \ system$

二甲胺(N-nitrosodimethylamine, NDMA)、N-亚硝基吡咯烷(N-nitrosopyrrolidine, NPYR)、N-亚硝基二乙

胺(N-nitrosodiethylamine, NDEA)、N-亚硝基二正丙胺(N-nitrosodi-n-propylamine, NDPA)、N-亚硝基哌啶(N-nitrosopiperidine, NPIP)、N-亚硝基二正丁胺(N-nitrosodi-n-butylamine, NDBA)、N-亚硝基吗啉(N-nitrosomorpholine, NMOR)和N-亚硝基二苯胺(N-nitrosodiphenylamine, NDPhA).除TBM、NDEA、NDPA和NDBA外,其他物质均有检出.

如图 3 所示,三卤甲烷质量浓度高于亚硝胺质量浓度. $\rho(TCM)$ 、 $\rho(BDCM)$ 和 $\rho(DBCM)$ 平均值分别为(15.97 ± 3.40)、(6.71 ± 1.40)和(1.96 ± 0.36) $\mu g \cdot L^{-1}$ 均未超过国家标准限值[26]。此外,亚

別为(15.97 ± 3.40)、(6.71 ± 1.40)和(1.96 ± >0.36)μg·L⁻¹,均未超过国家标准限值^[26].此外,亚 区 10.36 μg·L⁻¹ μg·L⁻¹

硝胺中 $\rho(\text{NDMA})$ 最高,为(22.20±13.24)ng·L⁻¹, 紧随之后的是 $\rho(\text{NMEA})$ [(2.94±1.78)ng·L⁻¹]、 $\rho(\text{NPIP})$ [(0.98±0.60)ng·L⁻¹]、 $\rho(\text{NMOR})$ [(0.78±0.33)ng·L⁻¹]和 $\rho(\text{NPYR})$ [(0.50±0.17)ng·L⁻¹].所有样品中 $\rho(\text{NDPhA})$ 均小于0.10ng·L⁻¹.本研究区域亚硝胺平均质量浓度与我国某中部城市饮用水中实测亚硝胺浓度^[27]相当.在空间分布上,教学区内三卤甲烷平均质量浓度较高,而亚硝胺相反,但空间差异性不显著(ANOVA TEST, P>0.05). DBPs 在两区域内的浓度变化可能由生活区和教学区的用水量和水力停留时间差异造成^[28].

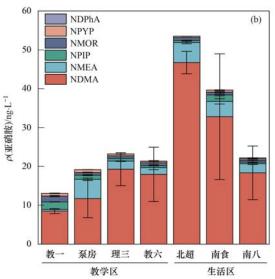


图 3 给水管网末端出水中三卤甲烷和亚硝胺的空间质量浓度分布

南八

南食

生活区

Fig. 3 Spatial distribution of the trihalomethane and nitrosamine mass concentrations in the drinking water distribution system

2.2 微生物群落多样性和组成结构

教一

泵房

理三

教学区

教六

北超

根据给水管网微生物群落的 α 多样性分析(表 2), 泵房的 Chao 和 Shannon 指数普遍很高, 说明管 网集中流动区域具有较高的微生物群落复杂度和群落稳定度. 空间分布上, 教学区的 Shannon 指数基本高于生活区, 反映教学区管网中微生物具有更高的群落多样性.

表 2 给水管网各采样点 α 多样性指数

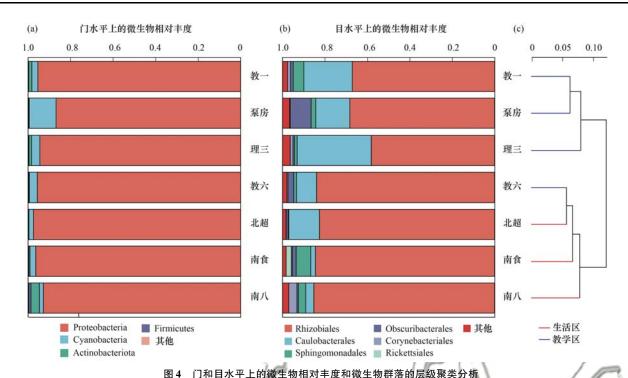
Table 2 The α diversity index for the drinking water

		in each sampling	point	
区域	地点	Chao 指数	Shannon 指数	覆盖率
	泵房	114. 00 ± 26. 57	1. 33 ± 0. 01	0. 999 6
教学区	教一	94. 50 ± 9. 88	1. 17 \pm 0. 01	0. 999 8
扒子匹	教六	105.15 ± 4.42	0.85 ± 0.01	0. 999 9
	理三	62.00 ± 8.57	1.07 ± 0.01	0. 999 9
	北超	81. 06 ± 11. 30	0.75 ± 0.01	0. 999 7
生活区	南食	79. 00 ± 8.57	0.88 ± 0.01	0. 999 9
	南八	123.33 ± 8.32	0.97 ± 0.01	0. 999 8
		_		

调查区内共发现 236 种细菌物种(OTU),其中 泵房、教学区和生活区共有 OTU 97 种,采样点间共 有 OTU 为 30 种. 如图 4(a) 所示, 在微生物的门分类 水平上, Proteobacteria 为所有样品中最主要的菌门, 占86.6%~97.4%,这与已报道的给水管网微生物多 样性研究结论相同[8,29]. 此外, 管网中含有较多的 Cyanobacteria (1.9% ~ 12.6%), Actinobacteriota (0.3% ~ 3.9%) 和 Firmicutes (0.1% ~ 1.0%). 在 Proteobacteria 中 Rhizobiales、Caulobacterales 和 Sphingomonadales 为优势菌目「图 4 (b)],而 Cyanobacteria 的优势菌目为 Obscuribacterales. 所有样 品中, Rhizobiales 相对丰度最高(58.2%~85.3%), 在空间分布上, Rhizobiales 在生活区的相对丰度要 高于教学区. Caulobacterales 发现在教学区的相对丰 度为9.5%~34.9%,高于生活区(2.2%~14.4%). 通过图 4(c)层级聚类分析结果可见,微生物群落相 对丰度在教学区和生活区具有差异性,但整体微生 物群落结构分布的差异性变化较小.

2.3 DBPs 生成因素的贝叶斯网络

构建贝叶斯网络需首先通过 Pearson 相关性分析和文献信息确定基础网络. 如图 5(a) 所示,四环



Relative abundance of the microbial phyla and orders with Bray-Curtis distancing cluster analysis of microbial community in drinking water distribution systems of the study area

素类抗生素与余(总)氯正相关,与三卤甲烷显著负相关(P<0.05).左下角连线图为微生物群落与抗生素、基本水质参数的距离矩阵 Mantel 相关性检验.结果显示 Caulobacterales 与氟喹诺酮类和磺胺类抗生素有显著相关性(P<0.05).基于 Pearson 相关性分析和 Mantel 检验,利用 tabu 网络结构学习算法寻找最优贝叶斯网络结构,完善多层关联性信息.考虑到基本水质参数在样品间的差异性较小,除余(总)氯外,贝叶斯网络将忽略其他基本水质参数如pH,水温等环境因素关联性.图5(b)显示了以DBPs作为目标的贝叶斯网络结构. 网络分成了4层,分别为实测余(总)氯质量浓度(第一层)、抗生素质量浓度(第二层)、微生物群落丰度(第三层)和DBPs质量浓度(第四层).

第一层贝叶斯网络提示,余氯与三卤甲烷和四环素类抗生素关联,与实验结论一致,即氯四环素容易发生氯化反应并生成三氯甲烷^[30,31].此外,余氯仅与三卤甲烷关联,与亚硝胺无关联,符合基于实验得出的亚硝胺与余氯投加量无相关性的结论^[32].第二层网络显示,磺胺类和大环内酯类抗生素均与三卤甲烷直接相关.这一结果已得到多项实验研究证实,即大环内酯类和磺胺类等抗生素会与氯消毒剂发生氧化还原反应,产生大量 DBPs^[33,34].其中磺胺类抗生素含有靠近苯环的氧、氮类基团,具有较强的氯反应活性,因此更容易生成三卤甲烷等DBPs^[35].贝叶斯网络还发现四环素类抗生素与亚硝

胺具有关联性. 前期调查发现,四环素在氯和氯胺化条件下是亚硝胺的潜在前体物^[3,36]. 需要指出的是,给水管网中抗生素的质量浓度较低(处于ng·L⁻¹水平),而且四环素类抗生素的亚硝胺转化率和磺胺类抗生素的三氯甲烷转化率分别约为 1% 和6%^[3,5],这说明给水管网中抗生素氯化产生的 DBPs 质量浓度有限,仅是给水管网中 DBPs 生成的因素之一.

抗生素不仅是 DBPs 前体物,而且会选择性抑制微生物菌群,改变微生物群落在管网饮用水中的结构 分布^[37]. 贝 叶 斯 网 络 发 现 了 抗 生 素 与 Caulobacterales 和 Corynebacteriales 丰 度 间 的 关 联 性. 值得注意的是,采样点余氯变化波动小且并未发 现与微生物丰度有关联,说明余氯在给水管网中的空间差异性对微生物群落组成的影响有限.

除了抗生素以外,微生物有机质也是 DBPs 的重要前体物之一. 微生物胞外聚合物中的色氨酸类物质被发现与多种 DBPs 具有很显著的正相关性 $(R_{\min}^2 \ge 0.76, P < 0.05)^{[38]}$. 微生物有机质会受到微生物群落结构影响,因此微生物群落丰度会间接与DBPs 关联. 第三层贝叶斯网络发现 Caulobacterales 和 Obscuribacterales 与 亚 硝 胺 具 有 相 关 性; Corynebacteriales 与三卤甲烷显著相关. 有研究证实 $[^{29,39]}$,以 Caulobacterales 为代表的 Proteobacteria 和以 Corynebacteriales 为代表的 Actinobacteria 是建立管网生物膜结构并影响胞外聚合物的化学组分的

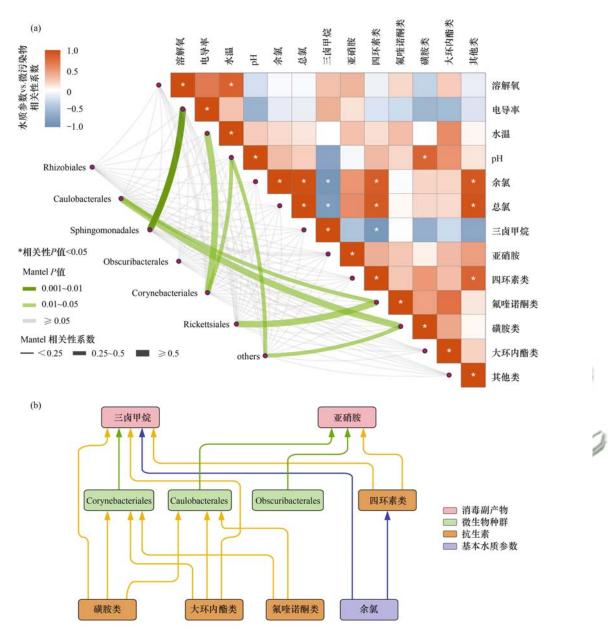


图 5 给水管网目水平上的消毒副产物、微生物目水平丰度、抗生素和管网基本水质参数的相关性分析和贝叶斯网络

Fig. 5 Correlation analysis and Bayesian network structure corresponding to DBPs, abundance of the microbial orders, antibiotics, and basic water quality parameters

重要菌种. 其次, Obscuribacterales 作为蓝藻菌门微生物,会产生胞内和胞外藻类有机物,是水环境中重要的 DBPs 前体物^[40]. 由于微生物结构组成复杂,通过贝叶斯网络挖掘管壁生物膜中重点菌群,可为进一步分析管网生物膜对 DBPs 生成贡献提供参考.

3 结论

- (1)给水管网中三卤甲烷、亚硝胺、抗生素质量浓度空间上未发现明显差异. ρ (三卤甲烷)、 ρ (亚硝胺)和 ρ (抗生素)分别为 18.33 ~ 32.09 μ g·L⁻¹、13.08 ~ 53.50 η g·L⁻¹和 47.92 ~ 210.33 η g·L⁻¹.
 - (2) Rhizobiales 和 Caulobacterales 为调查区域

给水管网的优势菌目. 由于管网的水力条件变化, 微生物群落丰度具有空间分布差异性.

(3)基本水质参数、抗生素质量浓度、微生物群落组成和 DBPs 质量浓度之间具有层次、有向性的网络结构. 四环素类、磺胺类和大环内酯类抗生素与三卤甲烷相关, 氟喹诺酮与亚硝胺相关. Caulobacterales 和 Corynebacteriales 丰度均受抗生素赋存水平影响.同时, Caulobacterales和 Obscuribacterales与亚硝胺生成相关, Corynebacteriales与三卤甲烷生成显著相关.

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