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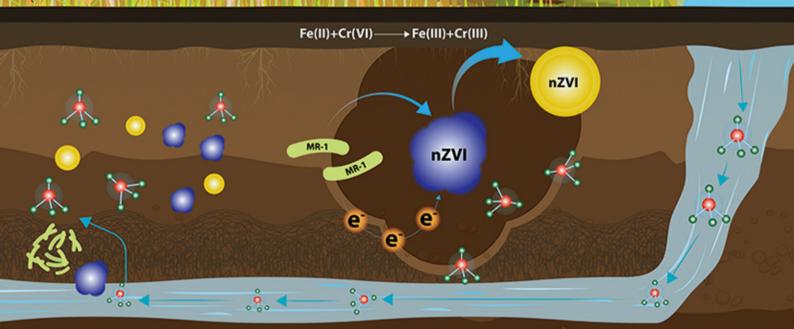
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电活性微生物激活生物质炭/零价铁协同钝化Cr(VI)及机制

廖聪坚,赵晓蕾,刘凯,钟松雄,李芳柏,方利平,叶挺进,石虎砚



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电子垃圾拆解区土壤-农作物系统中镉元素的空间分布特征及其风险评价

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摘要:随着电子垃圾的与日俱增,电子垃圾拆解活动导致的土壤重金属污染受到众多研究者的关注.为了探究电子垃圾拆解区周边土壤-农作物系统中镉的污染状况和空间分布特征,本研究采集了171对土壤和农作物样品进行系统解析.结果表明,根茎类、叶菜类、茄果类和水果类农作物土壤中镉含量分别为(1.292±0.647)、(1.010±0.201)、(0.921±0.125)和(0.861±0.135)mg·kg⁻¹,均值分别是浙江省土壤镉含量背景值的10.0、7.8、7.1和6.3倍,是农用地土壤污染风险筛选值的4.31、3.4、3.07和2.72倍,说明镉在土壤中累积明显.而农作物只有少部分超过食品安全限值,而且不同种类农作物对镉的富集能力的大小顺序为:叶菜类>根茎类>茄果类>水果类.单因子污染指数和潜在生态风险评价结果显示,研究区土壤镉污染严重且有高度的潜在生态风险.健康评价模型发现,手-口摄人重金属是造成当地居民健康风险的主要途径,而且重金属镉暴露对儿童造成的健康风险高于成人,但是该地区单一重金属镉污染暂时不会威胁到当地居民的身体健康.Moran's I 指数以及克里格插值结果揭示了镉具有显著的空间自相关性,存在明显的空间分布格局和局部空间聚集现象,且高值主要集中在电子垃圾拆解区周围,与电子垃圾拆解活动存在显著相关性.

关键词:电子垃圾拆解;土壤-农作物;镉污染;潜在生态风险;人体健康风险;空间分布 中图分类号: X171.5; X820.4 文献标识码: A 文章编号: 0250-3301(2021)09-4432-09 **DOI:** 10.13227/j. hjkx. 202101212

Spatial Distribution Characteristics and Risk Assessment of Cadmium Pollution in Soil-crops system of an E-waste Dismantling Area

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Abstract: With the rapid development of electronic technology, soil heavy metal contamination caused by electronic waste dismantling activities has attracted the attention of many researchers. To investigate the contamination status and spatial distribution of Cd in soil-crop systems around an e-waste dismantling area, 171 pairs of soil and crop samples were collected for analysis. The concentrations of cadmium in root vegetable soil, leaf vegetable soil, solanaceous vegetable soil, and orchard soil were (1.292 ± 0.647) , (1.010 ± 0.201) , (0.921 ± 0.125) , and (0.861 ± 0.135) mg·kg⁻¹, respectively. The average values of cadmium in these four soil types were (1.00, 7.8, 7.1), and (0.301 ± 0.135) mg·kg⁻¹, respectively. The average values of cadmium in these four soil types were (1.001 ± 0.135) , and (0.301 ± 0.135) mg·kg⁻¹, respectively. The average values of cadmium in these four soil types were (1.001 ± 0.135) , and (0.301 ± 0.135) mg·kg⁻¹, respectively. The average values of cadmium in these four soil types were (1.001 ± 0.135) , and (0.301 ± 0.135) , and (0.301 ± 0.135) mg·kg⁻¹, respectively. The average values of cadmium in these four soil types were (1.001 ± 0.135) , and $(0.301 \pm 0.135$

Key words: E-waste dismantling activities; soil-crop systems; cadmium contamination; potential ecological risk; human health risk; spatial distribution

土壤是人类生存和农业生产的重要资源,而农田土壤污染一直受到环境研究者的高度关注.其中,土壤重金属污染因其毒性大和持久性强等特点成为了长久且严峻的热点环境问题[1~3].农田土壤中重金属累积主要来源于农药和化肥的不当施用、污水灌溉和固体废物的堆放等[4~6].随着社会的发展,有研究发现农田周边的冶炼采矿活动[7.8]和废旧电器

拆解活动^[9,10]也是农田重金属污染的重要来源. 尤其是随着电子科技的飞速发展和电子产品的广泛普及,电子垃圾以一种前所未有的速度逐年递增. 根据

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重金属污染, E-mail; mail_zhangly@163.com * 通信作者,E-mail;fuweijun@zafu.edu.cn 文献[11], 2019 年全球产生了 5.36 × 107 t 电子垃 圾,在短短5a内增长了21%,而且2030年全球电 子垃圾将达到 7.40×10^7 t. 电子垃圾的粗放式回收 方式,例如手工拆除、剪断、熔化、燃烧和化学溶解, 致使大量重金属和持久性有机污染物释放到周边环 境[12~14]. 这不仅对环境构成潜在威胁, 也会对农作 物安全造成影响,并且通过多种途径,包括直接摄 人、呼吸吸入和皮肤接触等损害当地居民健 康[15,16]. 虽然近年来政府部门加大对非法拆解电子 垃圾的打击力度,但土壤污染具有一定的滞后性、隐 蔽性和复杂性,因此拆解场地周边的土壤污染在短 时间内难以消除,拆解场地的环境污染"历史遗留 问题"需持久性关注. 国内外的学者对电子垃圾拆 解区重金属污染研究表明,电子垃圾拆解造成了研 究区的土壤重金属污染,并且污染程度与从事拆解 时间长短有关[17]. 任露陆等[18]的研究发现电子废 物拆解区周边农田土壤中的 Cd、Pb、Cu 和 Zn 等重 金属含量显著增加且有较强的潜在生态风险. 但是 对电子垃圾拆解的区域性重金属分布特征及通过食 物链传递的潜在危害研究相对缺乏.

近年来,传统的环境污染评价方法,如单因子污染指数法、地累计指数法、内梅罗综合指数法和潜在生态风险指数法只是对点污染状况进行评价^[19,20].为系统了解区域重金属污染状况,需结合空间分析方法如地统计学和 Moran's I 等对污染物空间分布特征进行定量描述^[21,22].与其它重金属相比,镉有较强的生物迁移性和毒性,更容易被植物吸收,且易通过食物链进入人体,损害人类健康^[23,24].有研究表明,长期接触 Cd 会导致肺癌、肺腺癌、肾功能不全和乳腺癌^[25].因此,本文选取浙江省东南部电子垃圾拆解区周边的农用地为研究对象,对土壤以及不同类型农作物进行采样分析,对镉污染进行风险评价,揭示其空间分布特征,以期为优化受污染农田的安全利用、保障当地居民身体健康和农作物质量提供切实可行的支撑.

1 材料与方法

1.1 研究区概况

研究区位于中国浙江省的东南部,该地区属亚热带季风气候,气候温暖,光照适宜,年平均降水量为1583 mm,年平均气温为17.6℃.研究区以种植水稻为主,兼有蔬菜和经济作物.该地区从20世纪90年代到21世纪前15 a存在着众多电子垃圾回收区,许多以家庭为基础的电子垃圾处理车间和非正式的拆解活动无处不在,造成了严重的土壤污染,并对当地居民的身体健康造成潜在危害.

1.2 样品采集和分析

本研究以100 m×100 m 网格为基础,结合实地 农作物种植情况,在电子垃圾拆解区周围农田区域 共采集了171 对表层土壤(0~20 cm)及其对应农作物样品(图1).农作物样品包括叶菜类蔬菜(小白菜、苋菜、荠菜、空心菜、木耳菜、青菜、生菜和番薯叶)46 个、茄果类蔬菜(菜瓜、冬瓜、红豆、葫芦、黄瓜、豇豆、辣椒、南瓜、茄子、丝瓜、四季豆、西红柿和玉米)74 个、根茎类蔬菜(莲藕和葱)6 个和水果类(橙子、梨、葡萄、桃子和西瓜)45 个.每一个样点都是在半径为10 m 的区域内至少采集5 个子样品,然后均匀混合形成一个单独的复合样品.

土壤样品摊在风干盘中自然风干,然后去除土壤中的石块和动植物残渣等异物并研磨通过 2 mm 尼龙筛子,储存在聚乙烯袋中. 再从 2 mm 的土壤样品中取部分样品研磨至完全通过 0.149 mm 筛子,保存并用于检测. 农作物样品先用自来水冲洗干净,再用超纯水清洗,然后吸干农作物表面的水分,打成匀浆,存储于样品瓶并存放冰箱中待测. 重金属镉含量选用电感耦合等离子体质谱仪(ICP-MS, Thermo Fisher Scientific,美国)测定. 为了保证分析准确度,每个样品都进行重复检测,每 20 个样品后再进行一个空白样分析,相对偏差控制在 10% 以内,并且采用标准物质(GSS-30、GSS-32、GSB-5 和 GSB-27)控制分析精密度,加标回收率在 95%~110%.

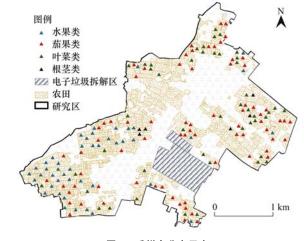


图 1 采样点分布示意

Fig. 1 Location of sampling sites

1.3 污染评价方法

1.3.1 单因子污染指数法

采用单因子污染指数 P_i 对单一重金属的污染程度进行评价 $^{[26,27]}$,其计算公式见式(1).

$$P_i = C_i / S_i \tag{1}$$

式中, C_i 为污染物 i 的含量 $(mg \cdot kg^{-1})$; S_i 为污染物 i 的限量标准 $(mg \cdot kg^{-1})$,本研究参照浙江省土壤重

金属背景值^[28]和土壤环境质量中农用地土壤污染风险筛选值(GB 15618-2018)^[29]. 通常 $P_i \le 1.0$ 表示无污染; $1.0 < P_i \le 2.0$ 表示轻微污染; $2.0 < P_i \le 3.0$,表示轻度污染; $3.0 < P_i \le 5.0$,表示中度污染; $P_i > 5.0$ 表示重度污染.

1.3.2 潜在生态风险评价指数

Hakanson^[30]提出的潜在生态风险指数,综合考虑了重金属的毒性和环境的效应,评价研究区土壤重金属含量对生态的危害程度^[31].单个重金属的潜在生态风险指数 E_i 计算公式见式(2).

$$E_i = T_{\rm r}^i \times C_i / C_{\rm n}^i \tag{2}$$

式中,n 为重金属的数量(本研究中 n = 1), T_r 为重金属生物毒性响应系数(T_{Cd} = 30), C_i 为污染物 i 的实际值($mg \cdot kg^{-1}$), C_n^i 为污染物 i 的背景参比值($mg \cdot kg^{-1}$).潜在生态风险指数可划分为 5 个等级: E_i < 30表示轻微生态风险; $30 \le E_i < 60$ 表示中度生态风险; $60 \le E_i < 120$ 表示较高生态风险; $120 \le E_i < 240$ 表示高度生态风险; $E_i \ge 240$ 表示重度生态风险^[32,33].

1.3.3 健康风险评估

本研究采用致癌和非致癌风险评价模型评估了 土壤重金属对人体健康的潜在危害.成人和儿童经 手口摄人、呼吸吸人和皮肤接触某种重金属的日平 均暴露量(ADD)计算公式和相关参数的取值见参 考文献[34~36].

单一重金属的致癌和非致癌风险的计算见式 (3)和(4).

$$CR = ADD \times SF$$
 (3)

$$HO = ADD/RfD$$
 (4)

式中, CR 表示单个重金属致癌健康风险指数, SF 表示致癌斜率因子, HQ 表示单个重金属非致癌健康风险指数, RfD 为毒性参考剂量. 不同暴露途径的 SF 和 RfD 值见参考文献[34]. 如果 HQ≤1 表示重金属对人类没有非致癌健康风险, HQ>1 表明重金属对人类有非致癌健康风险,并且随着 HQ 值的增加,非致癌风险越大. 致癌健康风险是指一个人在一生中由于暴露于致癌性风险重金属而患上任何一种癌症的概率^[36]. CR < 10⁻⁶,表明重金属的致癌风险可以忽略不计, CR > 10⁻⁴,表明人类患癌症的风险较高. CR 值处于 10⁻⁶~10⁻⁴之间,代表重金属的致癌风险是可以接受的.

1.3.4 富集系数

富集系数(EF)可用来衡量农作物对土壤重金属的迁移累积能力,表达见式(5).

$$EF = C_{x} / C_{\pm}$$
 (5)

式中, C_{x} 和 C_{\pm} 分别为农作物和土壤中重金属的含量,富集系数越大表明农作物吸收累积土壤中重金

属的能力就越大[37].

1.3.5 空间自相关

Moran's I 是一个典型的空间自相关指标,包括全局 Moran's I 和局部 Moran's I. 空间关联局部指标 (local indicator of spatial association, LISA) 主要通过局部 Moran's I 测量每个特定空间位置的相关程度,识别局部空间聚类模式和空间离群值 [38]. 当 Moran's $I \geq 0$,说明样点具有空间集聚性;而当 Moran's I 的 <0 时,说明该样点是空间离散值 [39].

1.3.6 克里格空间插值

普通克里格法是地统计学中最常用的一种无偏最优估计的一种插值法,利用变量原始数据和半方差函数的结构性,来估算其他位置上的区域化变量的数据.本研究通过 ArcGIS 软件,采用普通克里格插值法进行重金属空间分布图的绘制,直观地展现空间变异性.

此外,用析取克里格法估计目标位置重金属含量超过特定阈值的概率(本文以 GB 15618-2018 中的风险筛选值为阈值),计算见式(6).

$$\Omega[z(x_0) \geqslant Z_c] \approx P[z(x_0) \geqslant Z_c|z(x_1), z(x_2), \cdots, z(x_N)]$$
 (6)
式中, $z(x)$ 为随机变量,即重金属含量, x 为二维空

式中,z(x)为随机变量,即重金属含量,x为二维空间坐标, Z_c 为重金属阈值含量. 对任意点 x_0 近似条件概率的估测值 $z(x_0)$ 等于或超过 $Z_c^{[40]}$.

2 结果与讨论

2.1 镉含量描述统计分析

由表1可见,根茎类、叶菜类、茄果类和水果类 农作物土壤中镉含量分别为(1.292 ± 0.647)、 (1.010 ± 0.201) 、 (0.921 ± 0.125) 和 $(0.861 \pm$ 0.135) mg·kg⁻¹,均值分别是浙江省土壤镉含量背 景值的 10.0、7.8、7.1 和 6.3 倍,是农用地土壤污 染风险筛选值的 4.31、3.4、3.07 和 2.72 倍. 与浙 江省土壤 Cd 含量背景值比较,不同作物土壤的超 标率(样品数)分别为: 叶菜作物土壤 98% (45)、茄 果作物土壤 95% (70)、根茎作物土壤 100% (6) 和 果园土壤98%(44).以农用地土壤污染风险筛选值 为限量标准,根茎作物土壤的超标率(样品数)依然 为 100% (6),其次是果园土壤 93% (42)、茄果作物 土壤 85% (63) 和叶菜作物土壤 76% (35). 这说明 重金属镉在电子垃圾拆解区周边的农田土壤中大量 积累,危害到该区域的生态环境和农作物生长.任露 陆等[18]和尹伊梦等[35]对广东省电子垃圾拆解区周 边的农田土壤进行了调查,土壤镉含量超过土壤环 境标准限制,超标严重.分析不同类别农作物中镉含 量结果发现,只有极少部分的叶菜类农作物和茄果

类农作物中镉的含量是超过了食品安全国家标准 (GB 2762-2017)中的限量值,故当地居民应尽量避免在污染严重的农田土壤种植叶菜类和茄果类农作物.此外,根茎作物土壤中的镉含量平均值高于叶菜作物土壤,而根茎类农作物中镉含量却低于叶菜类农作物且都属于安全食用范围,这可能是由于不同农作物对重金属的富集能力不同造成的差异^[26].

变异系数越大,元素在土壤中的分布越不均匀. 根据 Zhao 等^[27]的研究,变异系数 < 10% 表示弱变异,意味着重金属含量主要受自然来源影响; 10%~90%为中等变异;变异系数>90%为强变异,主要受人为因素影响. 由表 1 结果可知,不同种类农作物种植下的土壤和农作物中镉含量的变异系数远高于90%,均属于极高程度的变异. 说明该地区的土壤及农作物中镉含量受人类活动的影响很大.

2.2 土壤-农作物中重金属镉的富集情况

镉的富集系数是指农作物中重金属镉含量和对应土壤中重金属镉含量的比值,反映了不同种类农作物对土壤中镉的吸收和富集能力. 富集系数越大,农作物富集镉能力越强,对人体健康带来的潜在危害也就越大. 如表 1 中富集系数的平均值所示,不同种类农作物镉富集能力的排序为:叶菜类 > 根茎类 > 茄果类 > 水果类,与农作物中镉含量的多少排序结果一致. 多重比较结果发现叶菜类、根茎类和水果类作物的富集系数存在显著性差异,茄

果类作物与叶菜类作物的富集系数存在显著性差异,与根茎类作物和水果类作物的富集系数不存在显著性差异(P<0.05).相关研究指出蔬菜不同品种间的重金属累积差异是由于生物机制不同,对重金属表现出不同的吸收能力^[26].叶菜类的富集系数明显高于非叶菜类,这可能是因为相较于非叶菜农作物来说,叶菜类农作物有着较高的转运速率和生长速度^[37].

2.3 土壤重金属镉的单因子污染风险评价和潜在 生态风险评价

表 2 为重金属镉的单因子污染风险评价结果. 以浙江省土壤元素背景值作为限量标准计算的单因子污染指数表明,研究区土壤遭受了不同程度的镉污染,其中100%的根茎作物土壤样点,77%的茄果作物土壤样点,84%的果园土壤样点和51%的叶菜作物土壤样点遭受到中度及重度镉污染.根茎作物土壤污染指数平均值高达9.94,其次是叶菜作物土壤、茄果作物土壤和果园土壤,单因子污染指数平均值分别为8.46、7.09和6.28,均属于重度污染.说明该研究区土壤镉积累明显,污染严重.张金莲等[41]的单项污染指数结果表明,镉污染的占比高达72.7%.而以风险筛选值为评价标准时,虽然污染指数平均值的排序依然是根茎作物土壤(4.31)>叶菜作物土壤(3.67)>茄果作物土壤(3.07)>果园土壤(2.27),但是超过50%样点主

表1 土壤及农作物镉含量描述性统计分析¹⁾/mg·kg⁻¹

	1	Tabl	e 1 Descrij	ptive statistic	es of cadmium	in soil and	agricultural	products/mg	·kg ⁻¹		
项	i 🗏	范围	平均值	标准误	变异系数 /%	pH 均值	背景值	超标率 ²⁾ /%	筛选值 限量值	超标率 ³⁾ /%	富集系数 均值
叶菜类	土壤 农作物	0. 177 ~ 4. 860 0. 004 ~ 0. 420	1.010 0.074	0.201 0.015	124 136	6.065 —	0. 129 —	98 —	0.300 0.200	76 7	0.096a
茄果类	土壤 农作物	0. 118 ~ 6. 830 0. 001 ~ 0. 140	0.921 0.016	0.125 0.003	116 155	6.288	0.129 —	95 —	0.300 0.050	85 7	0.026bc
根茎类	土壤 农作物	0.491 ~4.510 0.003 ~0.091	1.292 0.033	0.647 0.016	123 120	5.620	0.129 —	100	0.300 0.100	100 0.00	0.050b
水果类	土壤 农作物	0. 109 ~ 7. 660 0. 001 ~ 0. 044	0.816 0.007	0.135 0.001	111 125	5.794	0.129	98 —	0.300 0.050	93 0.00	0.012e

1) 背景值:浙江省土壤重金属背景值(中国土壤元素背景值,1990年);筛选值:农用地土壤污染风险筛选值(GB 15618-2018);限量值:食品中污染物限量(GB 2762-2017);"一"表示文章中没有相关数据;不同小写字母表示存在显著性差异(P<0.05);2)土壤中重金属含量超过浙江省土壤元素背景值的占比;3)土壤中重金属含量超过农用地土壤污染风险筛选值的占比,农作物中重金属含量超过食品中污染物限量的占比

表 2 不同农作物地土壤镉污染的单因子污染指数和潜在生态风险指数/%

Table 2 Single factor pollution index and potential ecological risk index of cadmium pollution in soil of different agricultural products/%

頂目		P _i (背景值)							P_i (筛选值)					
项目	均值	< 1.0	1.0 ~ 2.0	2.0 ~ 3.0	3.0 ~ 5.0	>5.0	均值	< 1.0	1.0 ~ 2.0	2.0 ~ 3.0	3. 0 ~ 5. 0	>5.0	(均值)	
叶菜类	8.46	0	17	32	15	36	3. 67	22	39	10	6	23	253. 83	
茄果类	7.09	2	6	16	38	39	3.07	11	46	15	12	16	212. 64	
根茎类	9. 94	0	0	0	50	50	4. 31	0	50	16	17	17	298. 23	
水果类	6. 28	1	0	15	43	41	2. 72	6	49	18	18	9	188. 25	

要受到轻微和轻度镉污染,30%左右的样点受到中度及重度镉污染.故以土壤背景值作为参比值估量的单因子污染评价结果显示研究区土壤镉污染程度更严重.

以浙江省土壤元素背景值作为评价标准对研究 区土壤镉含量进行潜在生态风险评价,结果见表2 和图 2. 由表 2 可知, 4 种农作物土壤的镉生态风险 平均值均远超过了120,属于高度生态风险,叶菜类 作物土壤和根茎类作物土壤镉污染的生态风险平 均值甚至高达 253.83 和 298.23,存在重度生态风 险. 由图 2 可知,农田土壤镉含量的潜在生态风险 指数大小范围约为25~1122,高风险区主要分布 在研究区的东部,在电子垃圾拆解区周围,说明这 一带主要受电子垃圾拆解影响,存在严重的生态 风险. 北部和西南部的农田距离电子垃圾拆解区 较远,生态风险低,表现为中度或轻微生态风险. 相关研究表明,在电子垃圾拆解区附近,重金属镉 对潜在生态风险有明显的贡献,因此要高度重视 电子垃圾拆解点周围农田土壤中镉污染积累带来 的生态危害[18].



Fig. 2 Potential ecological risk index of Cd in soil

2.4 重金属镉污染的人体健康评价

利用健康风险评价模型,分析电子垃圾拆解区周边农田重金属镉经由手-口摄入、呼吸吸入和皮肤接触这3种途径对成人和儿童造成的致癌暴露风险(表3)和非致癌暴露风险(表4).

表 3 中不同类别农作物土壤镉含量对儿童的致癌暴露风险均值明显大于其对成人的致癌暴露风险^[37],且致癌风险大小排序为根茎类作物土壤 > 叶菜类作物土壤 > 茄果类作物土壤 > 果园土壤. 成人和儿童不同暴露途径的致癌风险值表明, 手-口摄人重金属镉的风险值(CR_{ing})处于 10⁻⁶ ~ 10⁻⁴ 水平,是可接受范围,其次是皮肤接触暴露途径,致癌风险值(CR_{derm})在 10⁻⁸ 水平,最后是呼吸吸入途径,致癌风险值(CR_{inh})基本在 10⁻¹⁰ ~ 10⁻⁹ 水平,重金属镉的致癌暴露风险可忽略不计. 总体而言,研究区土壤镉污染对人体造成的致癌暴露风险处于可接受范围内.

表 4 中非致癌风险评价的均值结果表明,成人 和儿童不同暴露途径非致癌风险中占主导作用的是 手-口摄入造成的暴露风险,且与致癌风险的结果一 致,表现为手口摄入风险 > 皮肤接触风险 > 呼吸吸 人风险. 这与众多学者研究中的暴露风险评估一致, 即非致癌风险与暴露途径有关,并且经手-口摄人是 非致癌风险的主要途径[34]. 重金属镉对儿童的非致 癌暴露风险均明显大于对成人的非致癌风险,并且 儿童的 HQ 值比成人多了一个数量级,意味着相同 污染物同样的暴露途径对儿童造成的非致癌危害可 能更大[31]. 不同种类农作物土壤镉污染的 HQing、 HQinh、HQderm和HQ均值都有一个相同的趋势,即根 茎作物土壤 > 叶菜作物土壤 > 茄果作物土壤 > 果园 土壤. HQ 的平均值和最大值均小于1,说明单一的 镉导致的非致癌风险尚且不会对当地居民(成人和 儿童)的健康造成威胁.

表 3 土壤镉污染的致癌健康风险指数

Table 3 Carcinogenic health risk index of cadmium pollution in soil of different agricultural products

项	i II		成	人			J	L童	
坝	! 🗏	CR_{ing}^{-1}	CR _{inh} ²⁾	CR _{derm} ³⁾	CR	CR_{ing}	CR_{inh}	$\mathrm{CR}_{\mathrm{derm}}$	CR
叶菜类	平均值	3.93E -06	4.33E - 10	1.40E -08	3.95E -06	1.07E -05	6.25E - 10	5.51E - 08	1.07E -05
171020	最大值	1.74E - 05	1.91E -09	6.17E – 08	1.74E - 05	4.72E - 05	2.76E - 09	2.43E - 07	4.74E - 05
茄果类	平均值	3.29E - 06	3.63E - 10	1.17E -08	3.31E -06	8.94E -06	5.24E - 10	4.61E -08	8.99E - 06
加水人	最大值	2.44E - 05	2.69E - 09	8.68E - 08	2.45E - 05	6.63E - 05	3.88E - 09	3.42E - 07	6.66E -05
根茎类	平均值	4.62E -06	5.09E - 10	1.64E -08	4.64E -06	1.25E - 05	7.34E – 10	6.47E - 08	1.26E - 05
似至天	最大值	1.61E -05	1.78E -09	5.73E - 08	1.62E - 05	4.38E - 05	2.56E - 09	2.26E - 07	4.40E -05
水果	平均值	2.92E -06	3.21E – 10	1.04E -08	2.93E - 06	7.92E - 06	4.64E - 10	4.08E -08	7.96E -06
71676	最大值	2.74E - 05	3.02E - 09	9.73E - 08	2.75E - 05	7.44E - 05	4.35E - 09	3.84E - 07	7.47E - 05

¹⁾ 手-口摄入途径的致癌风险值; 2) 呼吸吸入途径的致癌风险值; 3) 皮肤接触途径的致癌风险值

表 4 土壤镉污染的非致癌健康风险指数

Table 4 Non-carcinogenic health risk index of cadmium pollution in soil of different agricult

项	iн		成	人			J	L童	
27.	! 🗎	HQ _{ing} ¹⁾	HQ _{inh} ²⁾	HQ _{derm} ³⁾	HQ	HQ_{ing}	$\mathrm{HQ}_{\mathrm{inh}}$	$\mathrm{HQ}_{\mathrm{derm}}$	HQ
叶菜类	平均值	1.86E - 03	1. 98E - 05	2. 64E - 04	2. 14E - 03	2. 10E - 02	1. 19E – 04	4. 33E – 03	2. 55E - 02
門本天	最大值	8. 20E – 03	8.75E - 05	1. 17E – 03	9.46E -03	9. 28E – 02	5. 26E – 04	1. 91E – 02	1.12E - 01
茄果类	平均值	1. 56E - 03	1.66E -05	2. 21E – 04	1. 79E - 03	1. 76E – 02	9. 98E - 05	3. 63E – 03	2. 13E – 02
加木天	最大值	1.15E - 02	1.23E - 04	1.64E -03	1.33E - 02	1. 30E – 01	7. 39E – 04	2. 69E – 02	1.58E - 01
根茎类	平均值	2. 18E - 03	2. 33E - 05	3. 10E - 04	2. 52E - 03	2. 47E - 02	1. 40E - 04	5. 09E - 03	2. 99E – 02
似至天	最大值	7.61E - 03	8. 12E – 05	1.08E - 03	8.78E - 03	8.61E - 02	4. 88E – 04	1.78E - 02	1.04E - 01
水果	平均值	1. 38E - 03	1.47E - 05	1.96E -04	1. 59E - 03	1.56E - 02	8. 83E - 05	3. 21E - 03	1.89E - 02
/\tau_	最大值	1. 29E – 02	1. 38E – 04	1.84E - 03	1.49E - 02	1.46E -01	8. 29E – 04	3.02E - 02	1.77E - 01

¹⁾ 手-口摄入途径的非致癌风险值; 2) 呼吸吸入途径的非致癌风险值; 3) 皮肤接触途径的非致癌风险值

2.5 重金属镉的空间自相关特征和空间分布

土壤和农作物中镉的 Moran's *I* 值分别为 0.408 和 0.148(图 3),说明土壤和农作物中的重金属镉有显著的空间自相关性(*P* < 0.01).土壤镉含量的高值集聚区主要分布在研究区的东部,接近电子垃圾拆解区,低值集聚区主要分布在研究区的北部,以

及少数研究区西南部,距离电子垃圾拆解区较远.农作物中镉含量的高值集聚区主要分布在研究区的东部区域,与土壤镉的高值分布类似,低值集聚区主要分布在研究区的西南部.镉含量高值和低值的分布与电子垃圾拆解点的位置有关,说明电子垃圾拆解活动可能对镉含量空间分布有影响.

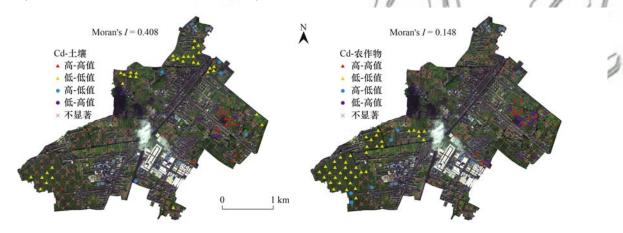


图 3 土壤和农作物中 Cd 含量的 LISA 局部空间自相关类型

Fig. 3 Local indicators of spatial association for Cd in soil and crops

通过普通克里格插值得到土壤和农作物中镉含量的空间分布(图4).土壤镉含量高的地区主要分布在研究区的西南部和东南部,其余的地区含量低,

呈现了局部高度集聚特征. 农作物中镉含量高的地区位于研究区的东部, 其对应的地区也正是土壤镉含量高的区域, 而低值区主要在研究区的西南部, 呈

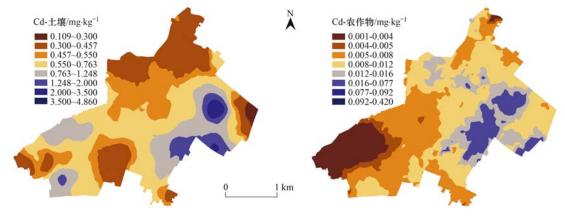


图 4 土壤和农作物中 Cd 含量空间分布

Fig. 4 Spatial distribution map of Cd in soil and crops

现了农作物镉含量从西南到东北递增的趋势. 镉含量的空间分布图与 Moran's *I* 所展现的空间自相关特征较为一致,高值分布区距离电子垃圾拆解区较近,而低值区主要分布在距离电子垃圾拆解点较远的农田. 国外学者 Amphalop 等^[42] 也表明电子垃圾的拆除活动增强了表层土壤中 Cd 的污染水平.

农田土壤镉含量和对应农作物镉含量空间上的差异,可能会与农作物富集重金属镉能力的大小有关(图5). 研究区的北部和东部的富集系数较高,研究区西部的富集系数明显较低. 因此土壤镉含量高的地区(研究区东部),由于农作物的富集系数较高,所以农作物中镉的含量也属于高值集聚区;土壤镉含量较低的区域(研究区西南部),由于农作物的富集系数较低,故农作物中重金属镉含量属于低集聚区;而在研究区北部,虽然土壤镉的含量较低,但由于该地区的农作物有较高的重金属镉富集能力,所以该地区农作物中镉的含量要高于研究区西南部农作物中镉的含量.

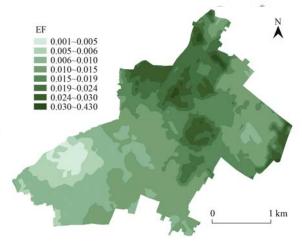


图 5 Cd 富集系数的空间分布

Fig. 5 Spatial distribution map of Cd enrichment factor

图 6 是利用 ArcGIS 软件分析得到的研究区隔含量超过农用地土壤污染风险筛选值的估计概率空间分布图. 根据土壤环境质量标准农用地土壤污染筛选值标准的含义,当土壤中镉的含量超过 0.3 mg·kg^{-1} 的时候(5.5 < $\text{pH} \le 6.5$),意味着该区域对农作物质量安全、农作物生长或土壤生态环境可能存在风险^[29]. 土壤镉的高风险区覆盖了研究区的大部分地区,其估计概率 Ω [Cd \ge 0.3 mg·kg^{-1}]为 0.80 ~ 1.00,说明研究区土壤存在普遍且严重的镉污染,该地区应引起重视,减少农作物的种植.

因此为了农作物质量和居民健康,在该研究区, 应当在距离电子垃圾拆解区近的农田种植茄果类的 蔬菜和水果等农作物,而在距离拆解区较远的农田 种植居民所需的叶菜类和根茎类蔬菜.

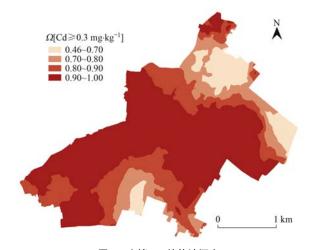


图 6 土壤 Cd 的估计概率

Fig. 6 Estimated probability map of Cd in soil

3 结论

- (1)研究区土壤镉含量普遍超标,污染严重,且 该地区镉含量受外界人为活动的影响较大. 农作物 中镉含量只有部分叶菜类和茄果类蔬菜超标,与农 作物的富集能力有关,且叶菜类农作物富集镉的能 力明显高于非叶菜类农作物.
- (2)研究区土壤镉积累明显且存在严重的潜在生态风险,而且根茎类作物土壤>叶菜类作物土壤> 加果类作物土壤 > 加果类作物土壤 > 加果类作物土壤 > 加里金属是造成当地居民健康风险的主要途径,而且重金属镉的暴露对儿童造成的健康风险高于成人.
- (3)土壤和农作物中镉含量的高值区主要分布 在电子垃圾拆解区周围,其空间分布的差异性主要 受电子垃圾拆解活动的影响.

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