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针铁矿改性生物炭对砷吸附性能

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摘要:为了提高生物炭(BC)对砷的吸附能力,本研究选取小麦秸秆作为原料,采用共沉淀方法制备了针铁矿(Goethite)改性生物炭材料(Goethite@BC). 比较了 BC、Goethite 和 Goethite@BC 对As(\blacksquare)的吸附特性,同时使用 SEM-EDS、BET、FT-IR、XRD 和 XPS 等技术对改性吸附剂的理化性质和吸附机制进行表征. 结果表明,扫描电子显微镜分析显示有纳米级针铁矿附着在生物炭表面,可有效提高生物炭的比表面积和总孔容;3 种吸附剂对As(\blacksquare)的吸附符合伪二级动力学模型和 Langmuir 等温吸附模型,Goethite@BC 对As(\blacksquare)的最大吸附量为65.20 mg·g⁻¹,与BC 相比吸附量提高了62.10 倍. Goethite@BC 吸附机制包括非特异性吸附(静电引力)和特异性吸附(配位、络合、离子交换等),纳米针铁矿颗粒在 Goethite@BC 表面对污染物的吸附起到重要作用. Goethite@BC 在污染物修复领域具有很好地应用前景.

关键词:生物炭(BC);针铁矿;改性;砷(As);吸附

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Application of Goethite Modified Biochar for Arsenic Removal from Aqueous Solution

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Abstract: To improve the adsorption capacity of wheat biochar (BC) for arsenic (As), wheat stalks were selected as biomass to generate nano-sized goethite modified biochar (Goethite@BC) by co-precipitation. The adsorption capacities of BC, Goethite, and Goethite@BC for As(\blacksquare) were compared. The samples were analyzed by scanning electron microscopy (SEM) along with energy dispersive spectrometry (EDS), Brunauer-Emmett-Teller (BET), Fourier transform infrared (FT-IR), X-ray diffraction (XRD), and X-ray photoelectron spectroscopy (XPS) techniques. The results showed that the nano-goethite coating was uniformly attached to the surface of the BC and improved the surface area and total pore volume of the biochar. The adsorption of As(\blacksquare) by the three adsorbents was proved to fit well with the pseudo-second-order kinetic model and the Langmuir model. Compared to BC, the Goethite@BC increased the adsorption rate of As(\blacksquare) by 62. 10 times, and the maximum adsorption capacity of Goethite@BC was 65. 20 mg·g⁻¹. The adsorption mechanism of Goethite@BC included non-specific adsorption (electrostatic attraction) and specific adsorption (coordination, complexation, ion exchange, etc.), and nano-goethite particles on the Goethite@BC surface played an important role in the adsorption of As. Goethite@BC has a good application prospects in the field of environmental remediation.

Key words: biochar (BC); goethite; modification; arsenic (As); adsorption

神是世界上最常见的环境污染物之一,被国际癌症研究机构(IARC)归为第一类致癌物^[1]. 神广泛分布在地壳中,是几百种矿物的组成成分^[2]. 自然环境中可通过风化作用使之活化,导致土壤和水体污染^[3]. 此外,工业废弃物的排放、矿物的冶炼和饲料添加剂的使用也会导致神污染^[4,5]. 神在土壤和水中毒性取决于它的形态,其中无机砷毒性强于有机砷,亚砷酸盐[AsO₃³ 和As(Ⅲ)]的毒性强于砷酸盐[AsO₄³ 和As(V)]^[6]. 目前,土壤和水体除砷的方法主要包括:原位化学钝化法、膜分离法、离子交换法、沉淀法、氧化-凝聚法和吸附法等^[7~9]. 在这些技术中,吸附法具有简易高效、操作简便、绿色环保等特点^[10]. 在过去的几十年中,已开发出多种高效吸附剂,例如:碳质材料、黏土

矿物材料和金属氧化物等[11].

生物炭(biochar, BC)是生物质在限氧条件下热解制备的碳质材料,具有成本低廉、来源丰富等优点^[12].由于其独特性质(官能团含量丰富、阳离子交换能力强、化学性质稳定等),常被用于去除土壤和水体中重金属和有机污染物,可降低污染物在土壤中的迁移率,降低生物有效性^[13,14].但由于大多数生物炭表面带有负电荷,因此对As(Ⅲ)、As(Ⅴ)等阴离子吸附效率较低^[4].

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目前,主要改性方法包括:生物炭活化、表面官能团改性和金属氧化物改性 $^{[15^{-17}]}$. 其中将铁氧化物(针铁矿、磁铁矿、赤铁矿、水铁矿等)和生物炭制备复合材料研究对砷的吸附效果较好 $^{[18]}$. Zhang 等 $^{[16]}$ 的研究表明,用 FeCl₃处理过的生物质经热解可直接制备负载 Fe₂O₃ 颗粒的生物炭,复合材料对水中As(V)的最大吸附量为 3. 15 mg·g $^{-1}$. 然而,这些生物炭复合材料制备方法相对复杂且成本高,其负载铁氧化物颗粒较大,对砷的吸附效果受到限制. 因此,需要开发简单易行、成本较低的方法来制备负载铁氧化物的生物炭.

本实验以小麦生物炭为原料,采用共沉淀方法,在生物炭(BC)表面生成纳米针铁矿(Goethite),制备针铁矿负载生物炭材料(Goethite@BC).开展Goethite@BC对As(Ⅲ)的批量吸附实验,同时利用扫描电子显微镜-能谱仪(SEM-EDS)、全自动比表面积与孔隙度分析仪(BET)、振动样品磁强计(VSM)、傅里叶红外光谱分析仪(FT-IR)、X射线衍射仪(XRD)和 X 射线光电子能谱(XPS)对样品结构进行表征分析,探究其对水溶液中As(Ⅲ)的吸附特征和影响因素,并分析不同 pH 条件下As(Ⅲ)和As(V)形态间转化的关系与 Goethite@BC对As(Ⅲ)的吸附机制.

1 材料与方法

1.1 实验试剂与仪器

实验试剂:氢氧化钾(KOH)、硝酸铁 $[Fe(NO_3)_3]$ 、盐酸(HCl)、硼氢化钾(KBH₄)、硫 $W[(NH_2)_2CS]$ 、抗坏血酸($C_6H_8O_6$)均为分析纯,购自国药集团化学试剂有限公司,亚砷酸钠 $(NaAsO_2)$ 、硝酸(HNO_3)为优级纯,购自国药集团化学试剂有限公司;实验用水为去离子水和超纯水.

主要仪器:电子天平(AL104,梅特勒-托利多集团,瑞士);恒温烘箱(DHG-9030,上海一恒科学仪器有限公司);箱式电阻炉(2.5-10,上海光地仪器设备有限公司);振荡器(HY4A,国华电器有限公司);pH计(FE20,梅特勒-托利多集团,瑞士).

1.2 针铁矿改性生物炭的制备

小麦秸秆产自河南郑州,使用去离子水将小麦秸秆清洗、干燥并粉碎后过 10 目筛,在 600℃限氧条件下热解 1 h,制备小麦秸秆生物炭. 先将 BC 分散在 1 mol·L^{-1} KOH 溶液中使之活化,后在混合液中加入 5 mol·L^{-1} Fe(NO₃)₃ 溶液,产生红色沉淀后移至烘箱,在 70℃条件下合成 60 h. 合成材料转移至透析袋中透析至电导率 < 30 μ S·cm⁻¹,低温干

燥,即得到针铁矿改性生物炭材料(Goethite@BC). 按上述步骤除去 BC 环节以制得针铁矿(Goethite).

1.3 吸附材料结构表征

Geothite@ BC 表面特征测定:样品形貌特征用扫描电子显微镜-能谱仪(SEM-EDS, Nova NanoSEM 430, FEI 公司,美国);比表面积与孔径分布采用全自动比表面积与孔隙度分析仪(ASAP2020,麦克仪器公司,美国);样品的磁化强度使用振动样品磁强计(VSM, LakeShore7404,美国);吸附材料鉴定分析采用傅里叶变换红外光谱仪(FT-IR Nixolet iS50,赛默飞世尔科技有限公司,美国)、X射线衍射仪(D8,布鲁克科技有限公司,德国)、纳米粒度和 Zeta 电位仪(Zetasizer Nano ZS 90,马尔文仪器有限公司,英国)和X射线光电子能谱(XPS ESCALAB 250Xi,赛默飞世尔科技有限公司,美国)进行分析.

1.4 吸附实验

1.4.1 吸附动力学实验

使用 NaAsO₂ 配置质量浓度为 500 mg·L⁻¹ As(Ⅲ)储备液. 分别称取 0.01g 3 种吸附剂(BC、Goethite 和 Goethite@BC)置于 15 mL 离心管中,加入 10 mL 10 mg·L⁻¹的 As(Ⅲ)溶液,使用 0.1 mol·L⁻¹的 HCl 调节溶液 pH 为 4.0 ± 0.1. 在室温下以 120 r·min⁻¹的转速分别振荡 0.5、2、4、6、8、12、24 和 48 h(每组实验重复 3 次),用 0.22 μm 水系滤膜过滤后使用原子荧光光度计(AFS-920,北京吉天仪器有限公司)测定溶液中As(Ⅲ)的浓度,计算公式如式(1):

$$q_{e} = \frac{(c_{0} - c_{e})V}{m_{0}} \tag{1}$$

式中, q_e 为振荡结束时吸附剂对As(\blacksquare)的吸附量, mg·g⁻¹; c_0 和 c_e 分别为吸附前后和吸附平衡时As(\blacksquare)的质量浓度(mg·L⁻¹); V 为溶液的体积(mL); m_0 为吸附剂的投加量(g).

1.4.2 吸附等温实验

分别称取 0.01 g 3 种吸附剂置于 15 mL 离心管中,加入 10 mL 一系列 5、10、20、40、60、80 和 100 mg·L⁻¹ 的 As(III) 溶液,使用 0.1 mol·L⁻¹ 的 HCl 调节溶液 pH 为 4.0 ± 0.1 . 在室温下以 120 r·min⁻¹的转速振荡 24 h(每组实验重复 3 次),用 0.22 μ m 水系滤膜过滤后使用原子荧光光度计测定溶液中As(III)的浓度.

1.4.3 溶液初始 pH 值的影响

分别称取 0.01 g 3 种吸附剂(BC、Goethite 和Goethite@BC)置于 15 mL 离心管中,加入 10 mL 质量浓度 10 mg·L⁻¹ 的 As(Ⅲ)溶液,使用 0.1

mol·L⁻¹的 HCl 调节溶液 pH 值为 2.0 ~ 9.0. 在室温下以 120 r·min⁻¹的转速振荡 24 h(每组实验重复 3 次),用 0.22 μ m 水系滤膜过滤后使用原子荧光光度计测定溶液中剩余总 As 浓度 [剩余As(Ⅲ)与 As(V)浓度之和].

As(Ⅲ)的测定:使用超纯水将样品稀释后加入2 mL 6 mol·L⁻¹HCl 溶液, 摇匀后放置 30 min 后上机测定; 总 As 的测定:称取1 g KOH 溶于超纯水中,将2 g KBH₄ 溶解于 KOH 溶液中,定容至 200 mL 作为还原剂. 将5 g (NH₂)₂CS 和 C₆H₈O₆ 溶解在超纯水中后定容至 100 mL,配置成 5%的 (NH₂)₂CS-C₆H₈O₆ 混合溶液.向稀释的待测样品中加入2 mL 6 mol·L⁻¹HCl 溶液和2 mL 的(NH₂)₂CS-C₆H₈O₆ 混合溶液,摇匀,静置 30 min 后上机测定;样品中As(\mathbf{V})浓度 = 总 As 浓度 – As(\mathbf{III})浓度.

1.4.4 吸附解吸实验

称取 0.10 g Goethite@ BC 置于 150 mL 三角瓶中,加入 100 mL 质量浓度为 10 mg·L⁻¹的 As(III) 溶液,使用 0.1 mol·L⁻¹的 HCl 调节溶液 pH 为 4.0 ± 0.1,在室温下以 120 r·min⁻¹的转速振荡 24 h 进行吸附实验. 使用 0.1 mol·L⁻¹ NaOH 对吸附后的Goethite@ BC 进行再生实验. 加入 0.10 g 吸附剂,在室温下以 120 r·min⁻¹的转速振荡 12 h. 对完成再生的吸附剂水洗、烘干后进入下一个吸附-解吸实验,共循环 4 次. 每次吸附、解吸实验结束后,用 0.22 μ m 水系滤膜过滤,滤液使用原子荧光光度计

测定溶液中As(Ⅲ)的浓度.

1.5 数据分析

使用 SPSS 20.0 进行数据分析, 使用 Sigmaplot 12.0 进行模型拟合.

2 结果与讨论

2.1 生物炭和其改性材料表征

通过 SEM 可以清晰直观地反映 BC 和其改性材料之间的形态变化. 从图 1(a)可观察到 BC 表面为光滑致密,有孔隙均匀分布;如图 1(b)所示,Goethite 为针状,长度在 1~3 μm 间,宽度在 100 nm 左右,并伴随聚集现象;图 1(c)显示 Goethite@BC 表面较为粗糙,有不规则的针形颗粒(400~700 nm)分布在其表面;图 1(d)显示,对 Goethite@BC 使用 EDS 分析后发现,在其表面存在 Fe、C、O、Mg 和 K 等元素.

通过 BET 分析得出,BC 的比表面积(SSA)为 44. 97 $\text{m}^2 \cdot \text{g}^{-1}$,总孔容(TPV)为 0. 04 $\text{cm}^3 \cdot \text{g}^{-1}$,平均孔径(AP)为 3. 85 nm. Goethite@ BC 的 SSA 为 276. 24 $\text{m}^2 \cdot \text{g}^{-1}$,TPV 为 0. 19 $\text{cm}^3 \cdot \text{g}^{-1}$,AP 为 2. 78 nm. 与 BC 相比,Goethite@ BC 的 SSA 增大了 6. 14 倍和 TPV 增大了 4. 75 倍,而 AP 减小. 这是因为改性材料制备前期 BC 经过 KOH 活化有助于微孔的增加,提高 SSA 和 TPV,但这种微孔、大孔的组合会导致平均孔径的减小 [12]. Huang 等 [19] 和 Luo 等 [20] 的研究发现,经过 KOH 活化的 BC,SSA 和

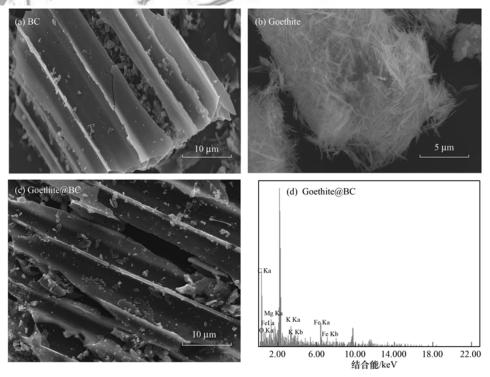


图 1 BC、Goethite 和 Goethite@BC 的 SEM-EDS 图像

Fig. 1 SEM-EDS images of BC, Goethite, and Goethite@ BC

TPV 均有明显的增加,这种改性不仅增加了 BC 表面吸附位点和与污染物接触能力,还有利于新官能团的引入. Mohan 等 $^{[21]}$ 和 Hu 等 $^{[22]}$ 的研究发现,BC 直接浸渍铁盐 $[FeCl_2, Fe(NO_3)_3\cdot 9H_2O$ 等]改性后会导致 SSA、TPV 和 AP 减小,是因为在没有活化的 BC 上生成的铁矿颗粒可能造成 BC 的微孔堵塞.

利用 VSM 对 Goethite@ BC 的磁学性能进行分析测试,其磁滞回线如图 2 所示. 在室温条件下,外加磁场强度在 \pm 10 000 Oe(\pm 25 × 10⁵ · π ⁻¹ A·m⁻¹) 范围内,Goethite@ BC 的饱和磁化强度为 6.96 emu·g⁻¹. 磁滞回线过原点且以原点成中心对称,剩余磁感应强度(B_r)和矫顽力(H_s)为 0,说明改性材料具有超顺磁性[11,23]. 利用外磁场可对 Goethite@ BC 进行分离和回收,以达到反复循环使用的效果,大幅度降低污染物吸附提取过程能耗.

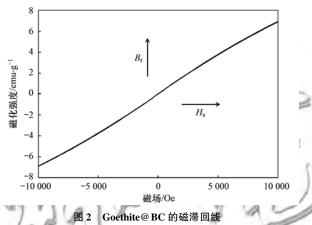


Fig. 2 Magnetization curves of Goethite@ BC

图 3 是 BC 和 Goethite@ BC 的 XRD 图谱, BC 的 2θ 衍射峰 28.32° 、 40.46° 、 50.13° 和 66.39° 分别 对应(200)、(220)、(222)和(420)晶面,与钾盐 (PDF No. 41-1476)相符,这可能与制备 BC 过程中产生的灰分有关^[24].相比于 BC,Goethite@ BC 的XRD 图谱出现新的衍射峰 2θ 为 17.77°、21.13°、33.21°、36.61°和 53.21°,与针铁矿 (PDF No. 29-0713)相符,分别对应晶面(020)、(110)、(130)、(111)和(221),表明通过共沉淀方式在 BC 表面生成了针铁矿.

使用 FT-IR 对 BC 和 Goethite@ BC 表面官能团进行分析,如图 4 所示.在3 308 cm⁻¹处宽峰是由BC 和 Goethite@ BC 中羟基(—OH)形成的,且Goethite@ BC 中峰透过率较低,这表明改性后形成了大量的—OH 官能团^[25].1531 cm⁻¹处的特征峰主要是由于酯内物质中酯基(C =O)的振动引起^[22].BC 中1 375 cm⁻¹处的特征峰主要与碳羟基(C—OH)的伸缩振动有关,Goethite@ BC 中此峰跃

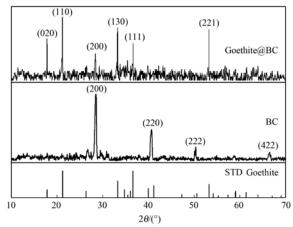


图 3 BC 和 Goethite@BC 的 XRD 图谱

Fig. 3 XRD patterns of BC and Goethite@ BC

迁为1 323 cm $^{-1[4]}$. BC 在1 006 cm $^{-1}$ 出现特有宽峰,是由碳氧键(C—O)伸缩振动引起的 $^{[26]}$. Goethite@BC 在 540 cm $^{-1}$ 处出现与铁氧键(Fe—O)弯曲振动有关的特征峰,这是低结晶水铁矿典型特征峰 $^{[27]}$.

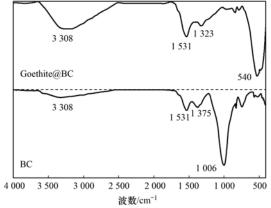


图 4 BC 和 Goethite@BC 的 FT-IR 图谱

Fig. 4 FT-IR spectrum of BC and Goethite@ BC

2.2 吸附动力学实验

BC、Goethite 和 Goethite@ BC 对As(Ⅲ)的吸附是个复杂的固液间反应过程,采用伪一级动力学[式(2)]、伪二级动力学[式(3)]和颗粒内扩散模型[式(4)]对吸附数据进行拟合.

$$\frac{\mathrm{d}q_{t}}{\mathrm{d}t} = k_{1}(q_{e} - q_{t}) \tag{2}$$

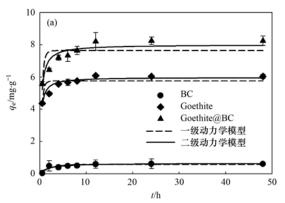
$$\frac{\mathrm{d}q_t}{\mathrm{d}t} = k_2 (q_e - q_t)^2 \tag{3}$$

$$q_t = k_{\rm p} t^{0.5} + C {4}$$

式中, t 为吸附反应时间(h); q_e 为达到平衡时的吸附量(mg·g⁻¹); q_t 为 t 时刻的吸附量(mg·g⁻¹); k_1 、 k_2 和 k_p 分别为伪一级吸附速率常数(h⁻¹)、伪二级吸附速率常数[g·(mg·h⁻¹)] 和颗粒内扩散常数[mg·(g·h^{0.5})⁻¹]; C 是常数,为颗粒内扩散方程的截距,其与膜扩散边界层的厚度有关,C 值越大,边界层厚度越大^[28].

图 5(a)显示与伪一级动力学模型相比,伪二级动力学模型能更好地描述 3 种吸附材料对 As(III)的吸附过程($R^2 > 0.845$). 表 1 显示,吸附剂对As(III)的吸附速率依次为: Goethite > Goethite

@ BC > BC, BC 表面负载针铁矿可有效提高其对As(Ⅲ)的吸附速率. 平衡吸附量依次为: Goethite@ BC > Goethite > BC, 相比于 BC, Goethite@ BC 对As(Ⅲ)的平衡吸附量分别提高了11.05 倍.



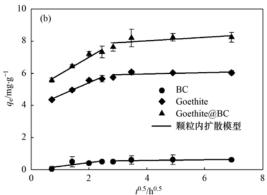


图 5 BC、Goethite 和 Goethite@BC 对As(III)的吸附动力学曲线和 q_e 与 $t^{0.5}$ 曲线

Fig. 5 Adsorption kinetics curves and q_e versus $t^{0.5}$ curves of As(\mathbb{II}) by BC, Goethite, and Goethite@ BC

图 5(b)显示使用颗粒内扩散模型对动力学数据进行拟合. 3 种吸附剂对As(III)的吸附呈现多段性:快速吸附阶段(0~6h),吸附量达到饱和吸附量的85%以上;吸附平衡阶段(6h之后).快速吸附阶段

斜率较大,吸附速率较快,吸附平衡阶段斜率较小,属于颗粒内扩散阶段.拟合直线的反向延长线不通过原点,说明颗粒内扩散不是唯一的速率控制步骤,可能是由表面吸附和颗粒内扩散共同控制.

表 1 伪一级和伪二级动力学模型拟合参数

Table 1 Adsorption kinetic models for As(III) adsorption on three materials

吸附剂 -	61 1	伪一级动力学模型	1	/ 1	伪二级动力学模型	1
	k_1/h^{-1}	$q_{ m e}/{ m mg}\cdot{ m g}^{-1}$	R ²	$k_2/g \cdot (mg \cdot h^{-1})$	$q_{ m e}/{ m mg}\cdot{ m g}^{-1}$	R^2
BC	0. 517	0. 58	0. 891	0. 414	0.72	0. 907
Goethite	2. 781	5. 76	0. 649	0.810	5. 96	0. 886
Goethite@ BC	2. 524	7. 63	0. 582	0. 496	7. 96	0. 845

2.3 吸附平衡实验

BC、Goethite 和 Goethite@ BC 对As(Ⅲ)的吸附量与平衡浓度之间关系如图 6 所示. 对As(Ⅲ)的吸附能力为: Goethite@ BC > Goethite > BC. 吸附平衡浓度在 0 ~ 40 mg·L⁻¹时,Goethite 和 Goethite@ BC 对As(Ⅲ)的吸附量随其平衡浓度增加而迅速增加,之后逐渐达到吸附平衡.

吸附平衡采用 Langmuir [式(5)]和 Freundlich [式(6)]等温吸附模型对图 6 实验数据进行拟合分析(表2):

$$q_{\rm e} = \frac{k_{\rm L}Qc_{\rm e}}{1 + k_{\rm L}c_{\rm e}} \tag{5}$$

$$q_e = k_F c_e^n \tag{6}$$

式中, q_e 为吸附达到平衡时的吸附量($\operatorname{mg} \cdot \operatorname{g}^{-1}$); Q 为吸附材料对 $\operatorname{As}(\coprod)$ 的最大吸附量($\operatorname{mg} \cdot \operatorname{g}^{-1}$); c_e 为平衡时溶液中 $\operatorname{As}(\coprod)$ 的浓度($\operatorname{mg} \cdot \operatorname{L}^{-1}$); k_L 为 Langmuir 吸附平衡常数($\operatorname{L} \cdot \operatorname{mg}^{-1}$); k_F 和 n 为 Freundlich 吸附平衡常数.

两种模型中,3种吸附剂对As(Ⅲ)吸附更符合

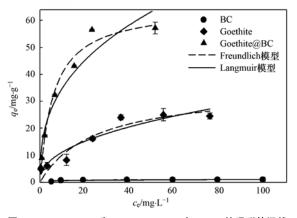


图 6 BC、Goethite 和 Goethite@BC 对As(III)的吸附等温线 Fig. 6 Adsorption isotherms of As(III) on BC, Goethite, and Goethite@BC

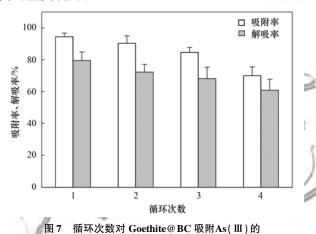
Langmuir 等温吸附模型. Goethite@ BC 对As(\blacksquare) 的最大吸附量为 65. 20 mg·g⁻¹,是 BC 的 62. 10 倍. Freundlich 等温吸附模型中 $k_{\rm F}$ 值越大,吸附能力越大^[29],表 2 中显示 Goethite@ BC 的 $k_{\rm F}$ 大于 BC,n 小于 BC,进一步说明 Goethite@ BC 对As(\blacksquare) 的吸附固定能力强于 BC.

表 2 生物炭和其改性材料对As(III)的吸附等温线模型参数

吸附剂		Langmuir 模型			Freundlich 模型	
火門刑	$k_{\rm L}/{ m L}\cdot{ m mg}^{-1}$	$Q/\text{mg}\cdot\text{g}^{-1}$	R^2	$k_{\rm F}/({\rm mg}^{n+1}\cdot {\rm L}^n)\cdot {\rm g}^{-1}$	n	R^2
BC	0. 113	1.05	0. 834	0.30	3. 828	0. 910
Goethite	0.053	32. 53	0. 929	4. 72	2. 475	0. 929
Goethite@ BC	0. 165	65. 2	0. 983	14. 94	2. 705	0. 937

2.4 吸附解吸实验

以 0.1 mol·L⁻¹的 NaOH 作为解吸剂,使已被吸附的As(Ⅲ)从 Goethite@ BC 表面解吸出来.图 7显示,Goethite@ BC 对As(Ⅲ)的吸附率随循环次数增加而降低,从 94.47%降低到 70.05%,而解吸率也表现出相同的规律,从 79.61%下降到 61.23%.表明 Goethite@ BC 中的活性位点在解吸过程中不能完全逆转^[30,31].虽然 Goethite@ BC 的吸附能力随循环次数增加而减弱,但经过 4 次循环后其对As(Ⅲ)的去除率也较高,因此 Goethite@ BC 具有较高的重复利用性.



影响(pH = 5.0 ± 0.1)

Fig. 7 Regeneration cycles of Goethite@ BC for As(Ⅲ) removal (pH = 5.0 ± 0.1)

2.5 不同 pH 值时砷的形态变化

图 8 表示在不同 pH 条件下 Goethite@ BC 对 As(Ⅲ)去除率的变化. 使用纳米粒度仪在 pH 2.0~9.0 下测定 Goethite@ BC 的等电点 pH_{PZC} = 5.2^[32]. 结果如图 8(a)所示,当 pH < pH_{PZC}时,去除率随 pH 升高而缓慢降低,由 92.15%下降到 86.65%;当 pH > pH_{PZC}时,去除率随 pH 升高而快速降低,由 88.30%下降到 76.27%.图 8(b)显示,在不同 pH 条件下,Goethite@ BC 吸附后溶液中As(Ⅲ)和As(V)发生了相互转化^[33].As 是一种氧化还原能力较强的元素,在氧化条件下As(V)占优势,而在充分还原条件下As(Ⅲ)占优势。氧化还原反应通过影响 As 的形态,从而影响吸附剂对 As 的吸附能力^[34].针铁矿自身具有一定的还原性,可在溶液中将 AsO^{3,-}氧化为 AsO^{3,-[33]}.这种反应不

仅将毒性较大的As(\blacksquare)转化为毒性较小的As(V),并且会产生 FeAsO₄ 沉淀,降低溶液中总 As 的浓度^[35]. As(\blacksquare)转化为As(V)的转化率随 pH 的升高而降低,当 pH < pH_{PZC}时,降低幅度较大,由49.98%降低到 33.44%,当 pH > pH_{PZC}时,转化率变化较为平稳,其中 pH = 7 时转化率达到最低,为30.65%.有研究表明,Fe(\blacksquare)氧化物在 pH 低于 4~5 时,可氧化As(\blacksquare),但当 pH 高于 8 时,氧化As(\blacksquare)能力减弱或不再具有氧化能力^[34].

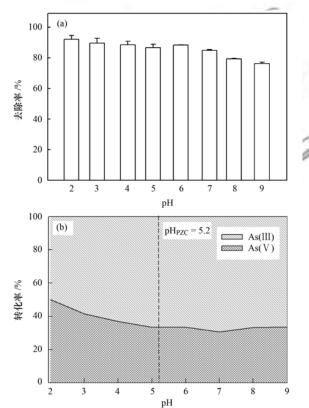


图 8 在不同溶液 pH 值条件下 Goethite@ BC 对As(Ⅲ)的 去除率和As(Ⅲ)和As(V)的转化率的影响

Fig. 8 Effect of Goethite@ BC on the As(III) removal rate and the conversion between As(III) and As(V) under different solution pH conditions

2.6 吸附机制分析

Goethite@ BC 对As(Ⅲ)吸附前后的 FT-IR 图谱如图 9 所示. 吸附后在 981 cm⁻¹处出现特征峰,有研究者表明,吸附砷酸盐后在 650 ~ 950 cm⁻¹间出现特征吸收带,这是由于 As—O 的伸缩振动所致^[22]. 与吸附前相比,吸附后主峰透过率都发生了不同程度的升高和偏移,这表明As(Ⅲ)可以通过

离子交换作用吸附在 Goethite @ BC 表面,其中—OH、C =O、C—OH 和 Fe—O 的变化表明吸附质在生物炭和针铁矿表面发生羟基络合反应并与生物炭表面芳香族化合物发生亲电取代反应^[36].

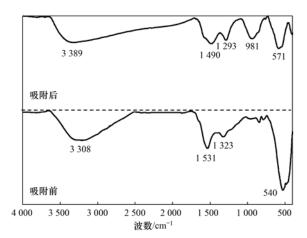


图 9 Goethite@BC 对As(III)吸附前后的 FT-IR 图谱

Fig. 9 FT-IR spectra of Goethite@ BC before and after adsorption of As($\rm III$)

为了进一步探究 Goethite @ BC 对As(III) 的吸附机制,本研究使用 XPS 对 Goethite @ BC 吸附As(III) 前后C1s、O1s、Fe2p 和 As3d进行分析(图10).图 10(a)显示吸附前后 Goethite @ BC 的C1s高

分辨谱进行分峰拟合,将谱线分成 3 组峰,在结合能 284.8 eV 处出现的峰为 C—C/C—H 的特征峰,结合能 286.1 和 288.7 eV 处出现的峰分别为 C—O 和 O—C =O 的特征峰^[37]. 吸附 As(Ⅲ)后这些官能团发生了变化,C—C/C—H 的相对含量由72.41%降低至 67.63%,C =O 的相对含量由14.42%增加至 16.47%,O—C =O 的相对含量由13.17%增加至 15.90%. 这表明吸附后 Goethite@BC 表面形成了 C =O 和 O—C =O,而 C—C/C—H 的数量减少了.

图 10(b) 为吸附前后 Goethite@ BC 的 01s 高分辨谱.由于含氧化合物组成繁多复杂,通常依据样品的组成特点和化学键的成分再进行分峰拟合 $[^{38]}$. 01s 的峰较宽,表明 Goethite@ BC 表面存在不同化学价态的 0,例如:有机氧(羧基、羰基、醇基和醚基中的氧)和无机氧(氧化铁中的氧) $[^{37]}$.结合能 $532.9~{\rm eV}$ 处出现的峰为 C=0 键,在结合能 $531.4~{\rm an}$ $530.4~{\rm eV}$ 处出现的峰为 $Fe=0-{\rm cm}$ $Fe=0-{\rm$

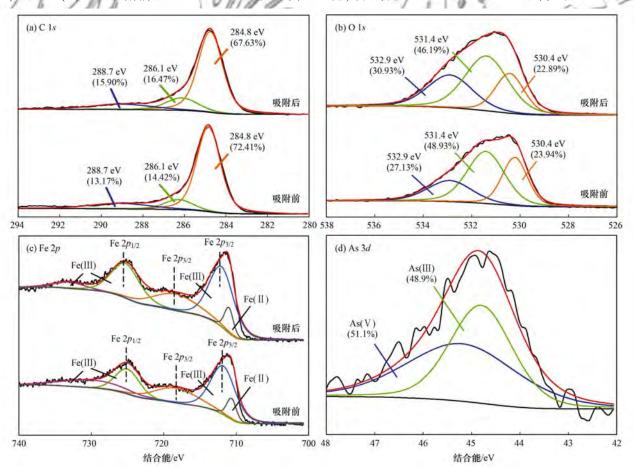


图 10 C1s、O1s、Fe2p 和 As3d的 XPS 图谱分析 Fig. 10 XPS analysis of C1s, O1s, Fe2p, and As3d

22.89%, 这是可能是因为 Goethite@ BC 表面针铁 矿吸附As(Ⅲ)所引起的^[22].

图 10(c) 显示吸附 As(III) 前后 Goethite@ BC 的 Fe2p 分峰拟合结果. Goethite@ BC 表面 Fe 形态主要为 Fe(III) 和 Fe(III),且 Fe(III) 多于 Fe(III),Goethite@ BC 吸附 As(III) 后 $Fe2p_{3/2}$ 光电子峰向低结合能方向移动,这表明在吸附过程中 Fe(III) 发生了夺电子反应. Fe(III) 可从吸附过程夺电子,从 As(III) 夺电子反应,使 As(III) 向 As(V) 进行了转化,这与图 8(b) 所得结论一致.

图 10 (d) 显示,在 Goethite @ BC 存在下,As(Ⅲ)的形态发生了变化. As 3d结合能为 44.8 和 45.2 eV,分别代表As(Ⅲ)和As(V)的特征峰^[39],二者占比分别为 48.9% 和 51.1%,其他研究也得到了相似的结论^[40].

Goethite@ BC 对As(Ⅲ) 吸附是一个复杂的过程,这与 As 形态、浓度、吸附剂表面性质、溶液

pH 值和竞争离子浓度有关^[41]. Goethite@ BC 吸附 As(Ⅲ)和As(Ⅴ)机制包括:非特异性吸附和特异性 吸附. 非特异性吸附:主要是吸附剂的静电引力引起的,吸附剂和吸附质间距离较远^[34]. 特异性吸附:吸附剂会与吸附质形成配位络合物,这部分反应和吸附剂表面含氧官能团减少有关^[41,42],这与FT-IR 和 XPS 图谱观察到的结果一致.

2.7 针铁矿改性生物炭与其他材料对 As 的吸附性 能比较

为了评价 Goethite@ BC 对 As 的吸附性能,将本材料与其他研究者制备材料进行比较. 表 3 显示,使用铁氧化物对碳质材料进行改性,制备原料包括松树^[3,21]、甜根子草^[41]、稻壳^[43]、小麦^[44]和氧化石墨烯^[36]等. 与其他材料相比, Goethite@ BC对砷具有良好的吸附性能,但小于石墨烯类改性材料. 石墨烯类改性材料成本较高,而生物炭具有较低的成本.

表 3 铁氧化物改性碳质材料对 As 吸附性能比较

Table 3	Summary of the	adsorption capacity	of As in iron	oxide modified	carbonaceous materials

吸附剂	吸附质	吸附质浓度 /mg·L ⁻¹	pH 值	吸附量 /mg·g ⁻¹	文献
针铁矿改性生物炭	As(III)	5 ~ 100	4. 0	As(III) :65. 20	本研究
松树生物炭/松树皮生物炭	As(III)	0.01 ~ 0.1	5. 0	2.62/13.1	[21]
磁铁矿改性生物炭	As(V)	1 ~ 50	7. 0	0. 492 9	[3]
磁性铁-甜根子草生物炭	As(Ⅲ)和As(Ⅴ)	0.4 ~ 0.8	7. 0	As(III) :2.0; As(V) :3.1	[41]
钙-铁改性稻壳生物炭	As(V)	0.8	大约7.0	1.0	[43]
Fe ₃ O ₄ 改性小麦秸秆生物炭	As(Ⅲ)和As(Ⅴ)	1 ~ 28	原始 pH	As(III):8.062; As(V):3.898	[44]
$CuFe_2O_4$ -氧化石墨烯泡沫复合材料	As(Ⅲ)和As(Ⅴ)	5 ~ 500	7. 0	As(III):51.64; As(V):124.69	[39]
钙基磁性生物炭复合材料	As(III)	0.5~20	6.00 ± 0.04	6. 34	[36]

3 结论

- (1)本研究以小麦秸秆为原料,采用共沉淀方法在小麦秸秆生物炭表面生成纳米针铁矿,制备针铁矿改性生物炭(Goethite@BC)材料,对As(Ⅲ)具有很好地吸附效果.
- (2)通过 SEM-EDS、BET、FT-IR、XRD 和 XPS 分析表明, 纳米针铁矿成功负载到生物炭上, 并且对As(Ⅲ)的吸附起到重要作用.
- (3)伪二级动力学和 Langmuir 等温吸附方程能够很好地描述 Goethite@ BC 对As(Ⅲ)的吸附过程,最大吸附量可达为 65. 20 mg·g⁻¹,在低 pH 值下,改性材料对As(Ⅲ)吸附效果较好.
- (4) Goethite@ BC 对As(Ⅲ)的吸附机制主要包括:非特异性吸附(静电引力),特异性吸附(配位、络合、离子交换等),改性材料表面的纳米针铁矿颗粒对 As 的吸附起到重要作用,吸附过程以化学吸附为主.本研究开发的 Goethite@ BC 作为低成本吸附材料,对 As 的吸附效果良好,具有较大的环

境修复应用潜力.

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