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# 博斯腾湖流域大气多环芳烃污染特征、干沉降通量及 来源

宋世杰1,黄韬1\*,赵留元1,毛潇萱1,穆熙1,高宏1,马建民1,2

(1. 兰州大学资源环境学院甘肃省环境污染预警与控制重点实验室, 兰州 730000; 2. 北京大学城市与环境学院地表过程分析与模拟教育部重点实验室, 北京 100871)

摘要:本研究使用大气被动采样器 (PAS-PUF) 和干沉降被动采样器 (PAS-DD),分别于 2016 年采暖期和 2017 年非采暖期对新疆博斯腾湖流域及周边地区 15 种 USEPA 优控多环芳烃 (PAHs) 大气浓度和干沉降进行了观测,并对其污染特征和来源进行了研究. 结果表明,采暖期和非采暖期博斯腾湖流域 PAHs 大气浓度范围分别为 6. 38 ~ 245. 43  $\,\mathrm{ng} \cdot \mathrm{m}^{-3}$  和 2. 33 ~ 74. 76  $\,\mathrm{ng} \cdot \mathrm{m}^{-3}$ ;采暖期与非采暖期均呈现出居民区 > 湖泊周边 > 塔中的空间分布. 采暖期和非采暖期 PAHs 大气干沉降通量范围分别为 0. 45 ~ 18. 10  $\,\mathrm{\mug} \cdot (\mathrm{m}^2 \cdot \mathrm{d})^{-1}$  和 0. 25 ~ 8. 15  $\,\mathrm{\mug} \cdot (\mathrm{m}^2 \cdot \mathrm{d})^{-1}$ ;采暖期居民区 PAHs 干沉降通量比湖泊周边和塔中采样点高,但在非采暖期塔中采样点高于其它采样点. 整体而言,博斯腾湖流域大气及干沉降中 PAHs 在采暖期显著高于非采暖期,在采暖期与非采暖期均以非 (Phe)、芴 (Flu)、荧蒽 (Flua) 和芘 (Pyr)等 3 ~ 4 环 PAHs 为主. 比值法源解析结果显示,博斯腾湖流域大气和干沉降中 PAHs 主要来源于煤炭和生物质燃烧;HYSPLIT 前向和后向轨迹模拟结果表明,非采暖期居民区较高 PAHs 排放通过大气传输到达博斯腾湖区,经大气干沉降进入水体,可能会对博斯腾湖水生环境造成影响.

关键词:博斯腾湖流域; 大气; 多环芳烃(PAHs); 干沉降; 来源

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# Pollution Characteristics, Dry Deposition Fluxes, and Sources for Atmospheric Polycyclic Aromatic Hydrocarbons in the Bosten Lake Watershed

SONG Shi-jie<sup>1</sup>, HUANG Tao<sup>1</sup>\*, ZHAO Liu-yuan<sup>1</sup>, MAO Xiao-xuan<sup>1</sup>, MU Xi<sup>1</sup>, GAO Hong<sup>1</sup>, MA Jian-min<sup>1,2</sup> (1. Key Laboratory for Environmental Pollution Prediction and Control, Gansu Province, College of Earth and Environmental Sciences, Lanzhou University, Lanzhou 730000, China; 2. Laboratory for Earth Surface Processes, College of Urban and Environmental Sciences, Peking University, Beijing 100871, China)

Abstract: Passive air samplers (PAS-PUF) and passive dry deposition (PAS-DD) samplers were applied in the Bosten Lake watershed located in Xinjiang to estimate the atmospheric concentrations and dry deposition fluxes for 15 US EPA priority polycyclic aromatic hydrocarbons (PAHs) during a heating period in 2016 and no-heating period in 2017, respectively. The results showed that the atmospheric PAHs concentrations in the Bosten oasis area ranged from 6.38 ng·m<sup>-3</sup> to 245.43 ng·m<sup>-3</sup> during the heating period and 2.33 ng·m<sup>-3</sup> to 74.76 ng·m<sup>-3</sup> during the non-heating period. The highest atmospheric PAHs concentrations were found in the residential area, followed by regions near Bosten Lake and Tazhong. The atmospheric dry deposition fluxes of PAHs in the Bosten Lake watershed ranged from 0.45  $\mu g \cdot (m^2 \cdot d)^{-1}$  to 18.10  $\mu g \cdot (m^2 \cdot d)^{-1}$  during the heating period and 0.25  $\mu g \cdot (m^2 \cdot d)^{-1}$  to 8.15 μg·(m²·d) -1 during the non-heating period. During the heating period, the atmospheric dry deposition fluxes in the residential area were significantly higher than those in the regions near Bosten Lake and Tazhong. However, the atmospheric PAHs dry deposition flux in Tazhong was much higher than that in other sites during the heating and no-heating periods. In general, the atmospheric PAHs dry deposition fluxes during the heating period were significantly higher than those during the non-heating period. PAH composition for the atmosphere and dry deposition were dominated by 3 and 4 ring congeners, especially by phenanthrene, fluorine, fluoranthene, and pyrene during two sampling periods. In addition, the congener diagnostic ratio shows that coal and biomass combustion were possible sources of atmospheric PAHs in the Bosten Lake watershed. The forward and backward trajectory analysis based on the HYSPLIT model demonstrated that the higher atmospheric PAH emissions from the residential area would be transported to Bosten Lake, which can affect the aquatic environment of this lake by dry deposition.

Key words: Bosten Lake watershed; atmosphere; polycyclic aromatic hydrocarbons (PAHs); dry deposition; source

多环芳烃 (polycyclic aromatic hydrocarbons, PAHs)是由两个或两个以上苯环组成的芳烃化合物,是一类持久性有机污染物 (persistent organic pollutants, POPs) [1],因其具有毒性,广泛存在于环境中会对生态系统和人体健康产生影响 [2].环境中PAHs 的来源主要分为天然源和人为源,其中人为源主要由生物质和化石燃料在人类生产或生活中不

完全燃烧产生<sup>[3]</sup>. 许多研究表明, 大气中 PAHs 的 类型和浓度与当地能源结构密切相关, 排放也表现 出季节性变化<sup>[4]</sup>.

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作者简介: 宋世杰(1995~), 男, 硕士研究生, 主要研究方向为有机污染物区域环境过程, E-mail; songshj17@ lzu. edu. cn

\* 通信作者,E-mail:huangt@lzu.edu.cn

博斯腾湖流域地处我国西北干旱区,位于世界 第二大沙漠的边缘[5],自古以来是丝绸之路上的重 要节点,同时博斯腾湖也是国家 5A 级旅游景区, 作为新疆南疆地区最重要的生态系统, 博斯腾湖的 生态和环境安全对整个南疆的生态、社会、经济和 人民生活的改善和发展均有突出的意义. 早期对天 山南侧积雪中 PAHs 的研究发现, 燃煤和生物质燃 烧是积雪中 PAHs 的主要来源, 并且 PAHs 含量随 海拔升高而增加,并存在潜在风险,可能是由于该 地区频繁的人类活动和长途运输排放的 PAHs 经大 气沉积进入天山积雪中[1,6]. 另外, 博斯腾湖沉积 岩芯中 PAHs 近年来也呈上升趋势[7]. 博斯腾湖流 域是一个典型的绿洲-沙漠交错带, 生态环境脆弱. 然而,对于绿洲地区大气中的多环芳烃鲜有报道, 可能源于偏远地区大气 PAHs 污染研究由于缺乏电 力和人力等因素而受限于采样技术. 另外, 不同于 我国中东部地区的气候条件, 西北地区总悬浮颗粒 物(total suspended particulate, TSP)及降尘污染较 为严重,因此干沉积对大气污染物的清除具有重要 作用,而关于博斯腾湖流域及周边地区大气 PAHs 干沉降的研究尚未发现公开报道.

大气被动采样技术(passive air sampler, PAS)由于其简易的设计及可在边远地区和条件恶劣的地区使用,因而广泛应用于大气中有机污染物的采样<sup>[8]</sup>. 考虑到操作使用、价格以及处理等多方面原因,聚氨酯软性泡沫材料的被动采样装置(PAS-PUF)近年被广泛应用<sup>[9]</sup>. 本研究利用大气被动采样技术(PAS-PUF)和干沉降被动采样技术(passive dry deposition, Pas-DD)分别研究博斯腾湖流域大气 PAHs 和干沉降通量的污染特征和季节差异,并探讨其主要来源,以期为博斯腾湖流域及周边地区大气 PAHs 的污染排放控制及治理提供理论依据,也为该地区人群健康和大气环境管理提供基础科学数据.

#### 1 材料与方法

#### 1.1 样品采集

本研究共布设7个被动采样点,主要覆盖博斯腾湖及周边经济活动较频繁的地区,形成一个居民区-湖泊-沙漠剖面<sup>[5]</sup>.其中和静县(HJX)、焉耆县(YQX)和博湖县(BHX),人口较为密集,作为居民区采样点;选择博斯腾湖东部(BHD)、博斯腾湖南部(BHN)和博斯腾湖北部(BHB)布设采样点以监测湖泊周围环境的污染水平;在塔里木油田作业区(TZ)布设采样点以考察该地区污染对博斯腾湖流域 PAHs 污染的潜在影响.采样点周围均无直接的人类活动干扰,具体采样点见图1.

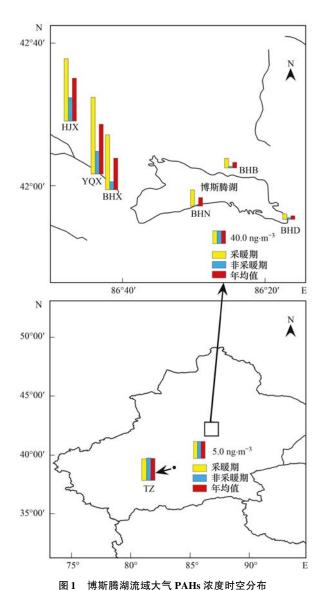


Fig. 1 Spatial distributions and seasonal variations in ambient atmospheric PAHs concentrations in the Bosten Lake watershed

大气和干沉降样品的采集使用加拿大环境部 Tom Harner 团队研发和改进的大气被动采样器 (PAS-PUF)<sup>[10]</sup> 和大气被动干沉降采样器(PAS-DD)[11], 图 2 中展示了两种采样器的结构. 样品采 集时间为 2016-11-01~2017-01-10(采暖期)和 2017-07-22~2017-10-16(非采暖期). 使用聚氨酯 泡沫塑料(PUF 规格为: 直径14 cm, 厚度1.35 cm, 体积为 207 cm<sup>3</sup>,表面积为 365 cm<sup>2</sup>,密度为21 300 g·m<sup>-3</sup>)采集大气和干沉降中化合物. 其中 PAS-PUF 主要以吸附气相有机物为主, 但也会有少许颗粒相 物质被捕集, PAS-PUF 的观测结果主要反映的是大 气气相 PAHs 的污染水平. PAS-DD 采样器的平行 板相较于 PAS-PUF 双圆顶外壳具有更开放的设计, 能够捕获沉积速率较高的粗颗粒(粒径, <200~ 250 μm), 颗粒物通过 PUF 表面的细孔而被捕集于 PUF内,并很少会出现再悬浮以及化学解析的现

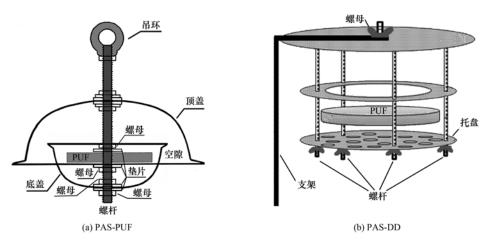


图 2 PAS-PUF 和 PAS-DD 构造示意

Fig. 2 Schematic of the PAS-PUF and Pas-DD

象<sup>[11]</sup>. 采样速率参考 Shoeib 等<sup>[10]</sup>和 Wilford 等<sup>[12]</sup>的研究报道. 采样前, PUF 经过 3 次 48 h 索氏提取净化处理, 先后使用丙酮(Ace)、二氯甲烷(DCM)、正己烷(Hex)溶剂. 采样结束后, PUF 密封, -20℃保存至分析, 在每个采样时段将 2 个干净的 PUF 带至现场, 打开密封包装, 作为现场空白随即带回.

#### 1.2 仪器与试剂

仪器: Trace1300GC-ISQLT 气相色谱-质谱联用仪(配备 EI 离子源), Thermo Fisher 公司; 色谱柱为 Thermo Fisher TG-5MS(30 m×0. 25 mm×0. 25 μm); 旋转蒸发仪, EYELA 公司 OSB-2100 型; 氮吹仪, Organomation 公司 N-EVAP 型; 超声波清洗器, KQ-500E 型; 150 mL 烧瓶; 150 mL 索氏提取器; 2 mL 棕色样品瓶; 20 mL 棕色样品瓶; 层析柱(规格: 长15 cm, 内径8 mm).

试剂: RBS 碱性洗液(购自 Rue Bollinckxstraat); 无水  $Na_2SO_4(AR, 购自国药)$ ;  $Al_2O_3(购自阿拉丁)$ ;  $SiO_2(100 \, \text{B}, 购自 \, \text{Merck})$ ; 二氯甲烷(DCM)、正己烷 (Hex)、丙酮(Ace)均为色谱纯(购自 OCEANPAK 公司, 纯度 99.99%); 16 种优控 PAHs 混标; 6 组分氘代 PAHs(1,4-Dcb-D8、萘-d8、苊-d10、菲-d10、d12 和苝-d12, 购自 Accustandard 公司); 六甲基苯(纯品, 购自 Dr. Ehrenstorger 公司).

#### 1.3 样品前处理与仪器分析

样品 PUF 置于 150 mL 索氏提取器,同时加入 6 组分氘代1 000 ng作为回收率指示物以控制样品处理过程中目标物的损失.使用 DCM 索氏提取 24 h,将抽提液旋转蒸发至 2 mL,用 Hex 置换溶液,浓缩至约1 mL. 浓缩液过中性 Al<sub>2</sub>O<sub>3</sub>-SiO<sub>2</sub> 柱分离净化,浓缩样用 DCM、Hex 混合液(1:1,体积比)淋洗,收集 20 mL 淋洗液于棕色样品瓶中,氮吹浓缩

至 0.5 mL,转样至进样瓶内,氮吹浓缩至 200 μL,加入内标 200 ng 六甲基苯后,使用 GC-MS 测定 15 种 PAHs,分别为苊烯(Acy)、苊(Ace)、芴(Flu)、菲(Phe)、蔥(Ant)、荧蔥(Flua)、芘(Pyr)、苯并 [ghi]蔥(BaA)、茄(Chr)、苯并[b] 荧葱(BbF)、苯并[k] 荧蔥(BkF)、苯并[a] 芘(BaP)、茚并 [1,2,3-cd]芘(IP)、二苯并[a,h]蔥(DiB)、苯并 [ghi] 苝(BghiP). 进样口温度 280℃,恒流无分流进样为 1.0 μL,高纯氦恒流 1.2 mL·min <sup>-1</sup>,离子源温度 250℃. 色谱柱升温程序:60℃保留 5 min,以 3℃·min <sup>-1</sup>的速度升温至 290℃,保留 5 min,以 25选择离子扫描(SIM),特征离子和保留时间进行定性,用内标法定量.

#### 1.4 质量控制与质量保证

样品定量采用内标 6 点校正曲线,各组分标准曲线  $R^2$  除个别外均大于 0.99.为保证观测数据的准确性,在实验室分析阶段增加实验室空白及平行样,实验室空白均未检出,平行样未超过 ± 10%;在样品处理过程中加入 6 组分氘代 PAHs 作为回收率指示物了解目标物损失情况,其范围分别为:54.2% ± 7.3%、67.3% ± 13.2%、83.3% ± 11.7%、100.4% ± 17.1%、109.5% ± 20.0%和97.7% ± 21.4%;并在样品上机测定前加入已知量的六甲基苯作为内标,以消除进样体积误差对样品测定的影响;连续分析 7 个空白样品,取目标物出峰处仪器响应的 3 倍标准偏差作为方法检出限,得到 PAHs 各组分的方法检出限为 0.05 ~ 0.79 ng.

#### 2 结果与讨论

#### 2.1 大气中 PAHs 污染特征

#### 2.1.1 大气 PAHs 空间分布特征

本研究除个别采样点外所有样品中 16 种 PAHs

均全部检出. 由于萘(Nap)在环境中以气相形式存 在, 其辛醇-气分配系数较小, 具有高挥发性, 因此 Nap 的观测结果具有较大不确定性[13], 故对 Nap 没有进行定量检测分析. 从图 1 可见, 研究区 7 个 采样点大气 > 15PAHs 含量年均浓度范围为 6.50 ~158.93 ng·m<sup>-3</sup>, 均值为65.68 ng·m<sup>-3</sup>. 研究区域 大气中 PAHs 空间分布存在较大差异, 其中焉耆县 (YQX)浓度最高,年均值为158.93 ng·m<sup>-3</sup>,同样 和静县(HJX)(137.19 ng·m<sup>-3</sup>)和博湖县(BHX) (100.83 ng·m<sup>-3</sup>)也有较高浓度.以上地区较高的 人口分布可能是造成局部地区出现高污染的原因, 尤其在我国西部干旱地区水资源对人口空间分布和 经济发展有着重要影响,县城居民区是流域人口主 要聚居地[14,15],家用燃煤、汽车尾气及农业生产过 程中秸秆的焚烧对居民区大气中 PAHs 有较高贡 献. 塔中(TZ)大气 PAHs 浓度最低,为 6.50 ng·m<sup>-3</sup>, 与焉耆县(YQX) 相差近 24.4 倍. 塔中 (TZ)位于沙漠腹地,远离人口聚居地,石油开采是 塔中(TZ)仅有的人类活动,但其周围环境空旷有 利于污染物扩散稀释,该地区的 PAHs 排放不会对 周边的大气环境造成影响. 在湖泊周边采样点博斯 腾湖东(BHD)、博斯腾湖南(BHN)和博斯腾湖北 (BHB), 大气 PAHs 也表现出远低于居民区的污染 水平, 博斯腾湖北侧为天山山脉, 南边为霍拉山-库 鲁克塔格低矮山地[5],湖区及周边的人为活动主要 以渔业及旅游业为主. 整体而言, 博斯腾湖流域7 个采样点表现出居民区 > 湖泊周边 > 塔中的空间分 布,较大的PAHs空间分布差异主要受到周边人为 活动的影响. 这与沈贝贝等[16]对博斯腾湖绿洲河 流沉积物中 PAHs 的空间分布研究一致, 居民区附 近的环境介质中有较高含量 PAHs.

### 2.1.2 大气 PAHs 季节性差异

大气 PAHs 的季节性差异如图 1 所示,研究区 8 个采样点除塔中(TZ)外,PAHs 浓度均表现出采暖期高于非采暖期. 博斯腾湖流域 PAHs 大气平均浓度在采暖期为 103. 84 ng·m<sup>-3</sup>,非采暖期为 23. 30 ng·m<sup>-3</sup>,采暖期 PAHs 大气浓度约为非采暖期的 4. 46 倍. 此季节变化与我国北方地区的研究结果一致,主要是由于在采暖期北方地区冬季家庭采暖造成燃煤量的增加<sup>[17]</sup>,造成大气中 PAHs 浓度升高. 其中,居民区和静县(HJX)、焉耆县(YQX)和博湖县(BHX)采暖期大气 PAHs 平均浓度是非采暖期的 4. 26 倍. 值得注意的是,在博斯腾湖周边的博斯腾湖东(BHD)、博斯腾湖南(BHN)和博斯腾湖北(BHB)大气 PAHs 的季节变化更为显著,相较于居民区,湖泊周边没有明显的大气污染排放,在采暖

期较大的增加可能来自周边大气污染的扩散和远距 离传输对湖泊的潜在影响.

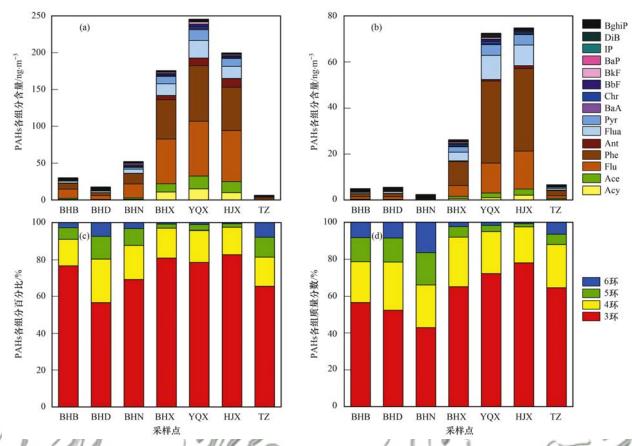
#### 2.1.3 大气 PAHs 组成特征

各采样点大气中 PAHs 组成特征如图 3 所示. 在采暖期,博斯腾湖流域大气 PAHs 主要由 Phe、 Flu、Flua 和 Pyr 组成,占 \( \sum 15PAH 的质量分数为 67.87%~84.06%. 将 15 种 PAHs 按照分子环数可 划分为 3 环(Acy、Ace、Flu、Phe 和 Ant)、4 环 (Flua、Pyr、BaA 和 Chr)、5 环(BbF、BkF、BaP 和 IP)和6环(DiB、BghiP)PAHs. 如图3(c)所示,采 暖期研究区 PAHs 大气中的环数分配与国内其他地 区的研究结果一致[17~19], 主要以3环为主, 3环、4 环、5 环和 6 环 PAHs 质量分数平均值分别为 78.50%、16.52%、3.60%和1.38%. 在非采暖期 观察到与采暖期相同的组成特征, 其中 Phe、Flu、 Flua 和 Pyr 是主要污染单体 PAHs, 占 ∑ 15PAH 质 量分数为 42.50%~88.08%. 非采暖期 3 环、4 环、 5 环和 6 环 PAHs 质量分数平均值分别为 71.38%、 22.4%、3.95%和2.28%[图3(d)].此外,BaP由 于其致癌性,可作为人类暴露于 PAHs 的指示 物<sup>[20]</sup>, 博斯腾湖流域采暖期(0.14~0.89 ng·m<sup>-3</sup>) 和非采暖期(0.11~0.38 ng·m<sup>-3</sup>)浓度均处于较低 水平, 年均浓度为 0.30 ng·m⁻³, 占 ∑15PAHs 的 质量分数为0.56%,低于国家环境空气质量标准中 规定的日均暴露限值 2.5 ng·m<sup>-3[21]</sup>.

#### 2.1.4 国内外大气 PAHs 污染水平对比研究

本研究将博斯腾湖流域大气 PAHs 浓度与国内外(使用 PAS-PUF)已有研究进行了对比,结果见表 1. 可以看出,绿洲地区 PAHs 污染水平低于兰州及哈尔滨,与长江三角洲一些城市(无锡市、苏州市和南通市)的污染水平相当,但却比珠江三角洲地区(广州、深圳、佛山、东莞、肇庆、珠海、惠州、江门和中山)高,是神农架大九湖景区污染浓度的两倍.与国外研究对比,高于尼泊尔(Kathmandu、Pokhara、Hetauda)和哥伦比亚卡里等地,但低于土耳其布尔萨.

而在新疆,对于大气中 PAHs 的研究主要集中在北疆乌鲁木齐以及石河子等城市大气颗粒物的研究.如吾拉尔·哈那哈提等[18] 在乌鲁木齐市一个高车流量采样点检测出大气 PM<sub>2.5</sub>中  $\sum$  13PAHs 浓度为 247.2 ng·m<sup>-3</sup>; 吴甜等[22] 的研究发现石河子市春、夏、秋、冬四季大气总悬浮颗粒物 (TSP) 中 $\sum$  16PAHs 平均浓度分别为 108.14、74.05、31.13、48.96 ng·m<sup>-3</sup>. 此外,也有学者对南疆塔里



(a) 采暖期各组分含量; (b) 非采暖期各组分含量; (c) 采暖期各环数所占质量分数; (d) 非采暖期各环数所占质量分数 图 3 博斯腾湖流域大气 PAHs 组成特征

Fig. 3 Composition of atmospheric PAHs in the Bosten Lake watershed

木盆 地 的 喀 什 ( 390. 5  $\,\mathrm{ng\cdot m^{-3}}$  )、铁 干 里 克 ( 388. 65  $\,\mathrm{ng\cdot m^{-3}}$  )、民 丰 ( 84. 54  $\,\mathrm{ng\cdot m^{-3}}$  ) 和塔中 ( 74. 07  $\,\mathrm{ng\cdot m^{-3}}$  ) 大气 TSP 中  $\sum$  16PAHs 进行了研 究  $\,\mathrm{[}^{14]}$ . 可见博斯腾湖流域大气 PAHs 污染相较于以上地区颗粒物中浓度处于较低水平,但新疆地

区干旱少雨,缺乏污染物的大气清除机制,导致大气颗粒物中更高的 PAHs 浓度.而本研究中被动采样器主要采集的是气态及部分细颗粒物<sup>[23]</sup>,因此,对博斯腾湖流域颗粒物中 PAHs 污染还需进一步研究.

表 1 国内外大气 PAHs 污染对比

Table 1 Comparison of atmospheric PAHs concentration domestically and overseas

研究区域	采样日期(年-月)	PAHs	浓度范围 /ng·m <sup>-3</sup>	均值 /ng·m <sup>-3</sup>	文献
博斯腾湖流域	2016-11 ~2017-10	15	6. 50 ~ 158. 93	65. 68	本研究
珠江三角洲	2015-02 ~ 2015-04	15	3.57 ~37.2	19.8 $\pm$ 9.69	[13]
兰州盆地	2013 夏季	16	_	214	[19]
	2013 冬季	16	_	302	
哈尔滨	2007-01 ~ 2007-04	16	162. 65 ~ 463. 44	_	[24]
长三角城市	2011-07 ~ 2012-06	15	6. 45 ~ 154	$56.8 \pm 14.8$	[ 25 ]
申农架大九湖	2012-04 ~ 2013-03	16	6. 94 ~ 184. 23	30. 36	[26]
尼泊尔	2014-08 ~ 2016-08	15	4. 1 ~ 38. 0	_	[27]
哥伦比亚卡里	2011-05 ~ 2011-09	14	25 ~66	_	[28]
土耳其布尔萨	2013-02 ~ 2014-02	15	6.4 ~1 100	_	[29]

#### 2.2 PAHs 干沉降通量时空分布特征

#### 2.2.1 PAHs 干沉降通量空间分布特征

博斯腾湖流域 7 个采样点大气  $\sum$  15PAHs 干 沉降年均通量为 0.70 ~ 9.22  $\mu$ g·( $m^2$ ·d)  $^{-1}$ ,均值

为 3. 19  $\mu g \cdot (m^2 \cdot d)^{-1}$ . 图 4 为博斯腾湖流域大气 PAHs 干沉降通量空间分布,可以看出,最大值出现在焉耆县(YQX),其浓度为 9. 22  $\mu g \cdot (m^2 \cdot d)^{-1}$ , 而最小值出现在博斯腾湖东(BHD),其年均通量为

0.70 μg·(m²·d)  $^{-1}$ . 类似在大气中 PAHs 空间分布,居民区 PAHs 干沉降通量整体较高,为 5. 26  $^{-1}$  ~ 9. 22 μg·(m²·d)  $^{-1}$  , 均值为 7. 42 μg·(m²·d)  $^{-1}$  . 博斯腾湖周边较低,为 0. 70  $^{-1}$  ~ 2. 44 μg·(m²·d)  $^{-1}$  ,均值为 1. 57 μg·(m²·d)  $^{-1}$  . 此外,在所有采样点中,塔中(TZ) PAHs 年均干沉降通量为 4. 30 μg·(m²·d)  $^{-1}$  , 大于湖泊周边,这主要是由于塔中地处空旷的戈壁沙漠,干沉降速率较大,导致该地区干沉降通量较大.

#### 2.2.2 PAHs 干沉降通量季节性差异

从图 4 看出, 大气 PAHs 干沉降通量表现出明显的季节性变化, 在采暖期为 0.45 ~ 18.10 μg·( $m^2$ ·d)  $^{-1}$ , 均值为 6.77 μg·( $m^2$ ·d)  $^{-1}$ , 非采暖期为 0.25 ~ 8.15 μg·( $m^2$ ·d)  $^{-1}$ , 均值为 2.16 μg·( $m^2$ ·d)  $^{-1}$ . 研究区属于暖温带大陆性干旱气候,全年干旱少雨,夏季干燥炎热,冬季寒冷  $^{[30]}$ .

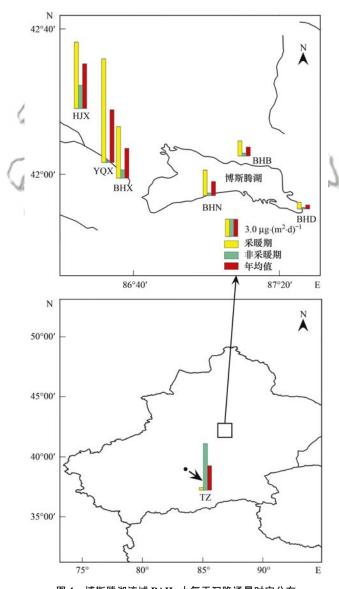


图 4 博斯腾湖流域 PAHs 大气干沉降通量时空分布

Fig. 4 Spatial distributions and seasonal variations in dry deposition fluxes of PAHs in the Bosten Lake watershed

在采暖期,随着周围环境气温的降低,居民供暖燃煤的消耗也在不断增加,从而增加了大气 PAHs 沉降量. 在非采暖期,大气 PAHs 干沉降通量较低,与采暖期相差近 3. 42 倍. 此外,博斯腾湖流域非采暖期风速高于采暖期,有利于污染物的远距离扩散迁移,并在一定程度上降低 PAHs 通量. 特别地,塔中(TZ)采样点非采暖期干沉降通量达到 8. 15  $\mu$ g·(m2·d) $^{-1}$ , 远 大 于 采 暖 期 的 0. 45  $\mu$ g·(m2·d) $^{-1}$ , 这可能是由于夏季沙漠地区沙尘天气较为频繁,导致干沉降通量增加.

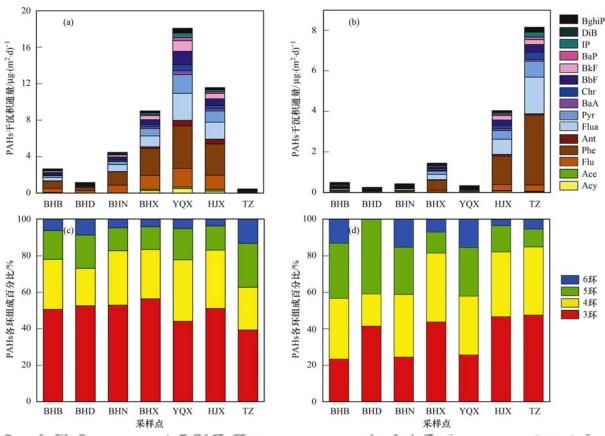
#### 2.2.3 大气 PAHs 干沉降组成特征

博斯腾湖流域大气 PAHs 干沉降组成特征如图 5 所示. 可以看出, 采暖期干沉降中 Phe 为主要单 体 PAHs, 占总通量的比例达到 25.78%~32.95%. 此外 Flu、Flua 和 Pyr 也是干沉积通量占比高的单 体 PAHs. 3 环为干沉降中主要污染组成,占 ∑15PAHs 的质量分数为 39.44% ~56.43% ,4 环 PAHs 次之, 为20.47%~33.79%.此外, 采暖期各 采样点干沉降中 PAHs 环数和各单体具有相似的百 分比分配特征. 在非采暖期, 各采样点干沉降中 PAHs 单体与采暖期表现出相同的组成特征, 主要 以 Phe、Flu、Flua 和 Pyr 为主. 非采暖期 PAHs 干沉 降同样主要以3环和4环为主,其中3环占比为 21.37%~47.66%,4 环为17.73%~41.81%. 但 不同的是,5~6环较采暖期百分比出现上升,这与 国内其他城市 PAHs 干沉降在夏季高环 PAHs 占比 上升的变化特征基本一致[31],大多数高分子量多 环芳烃化合物以固态形式吸附在大气中的细颗粒, 低分子量3环PAHs主要以气态存在于空气中,夏 季高温有利于气态污染物的扩散迁移.

#### **2.3** PAHs 来源解析

#### 2.3.1 比值法来源解析

来自不同污染源的 PAHs 各组分的比例不同,不同污染源释放的 PAHs 也有各自的特征指标. 使用特征化合物比值, IP/(IP + BghiP)和 Flua/(Flua + Pyr)可对包括石油燃料、生物质和煤燃烧等排放的潜在 PAHs 污染来源进行判断<sup>[32]</sup>: IP/(IP + BghiP)小于 0. 2 说明 PAHs 可能来源于石油或与石油有关的污染源,比值为 0. 20 ~ 0. 50 之间表示PAHs 来源于液态化石燃料的燃烧,比值大于 0. 50时 PAHs 来源于煤燃烧和生物质燃烧;Flua/(Flua + Pyr)小于 0. 40,说明 PAHs 主要来源于石油以及与石油有关的污染,若比值为 0. 40 ~ 0. 50 之间则PAHs 主要来自化石燃料的燃烧,当比值大于 0. 50时表示 PAHs 来源于生物质和煤的燃烧。图 6 为采暖期和非采暖期大气中 PAHs 特征化合物比值,可



(a) 采暖期各组分比例; (b) 非采暖期各组分比例; (c) 采暖期各环数比例; (d) 非采暖期各环数比例 图 5 博斯腾湖流域大气 PAHs 干沉降组成特征

Fig. 5 Composition of atmospheric dry deposition of PAHs in Bosten Lake watershed

以看出, 所有采样点在采暖期和非采暖期 IP/(IP + BghiP)比值均大于 0.5, Flua/(Flua + Pyr)比值也大于 0.5, 说明博斯腾湖流域大气中 PAHs 主要来源于生物质和煤的燃烧<sup>[25,32]</sup>.

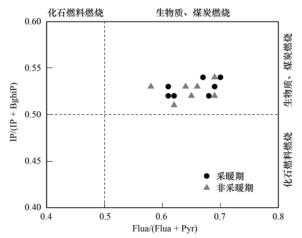


图 6 博斯腾湖流域大气 PAHs 特征化合物比值来源解析 Fig. 6 Source analysis based on the characteristic ratios of atmospheric PAH congeners in the Bosten Lake watershed

#### 2.3.2 后向轨迹分析

本研究采用 HYSPLIT 模型,对采暖期和非采暖期研究区大气传输后向轨迹进行模拟及轨迹聚类分析,通过气象轨迹分析研究区域大气中 PAHs 污

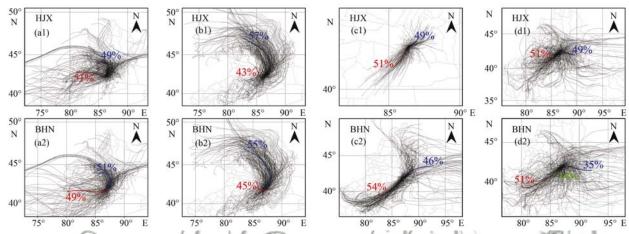
染物的可能来源;对大气传输前向轨迹进行模拟及聚类分析,评估可能受研究区大气 PAHs 污染影响的下风向区域.模拟计算采暖期和非采暖期的后向轨迹和前向轨迹,选取15 m作为模拟初始高度,模型高度10000 m,轨迹计算起始时间为每日00:00、06:00、12:00 和18:00<sup>[33]</sup>,轨迹起始点选择居民区和静县(HJX)和位于湖泊周边的博斯腾湖南(BHN),轨迹聚类结果如图7所示.

后向轨迹聚类结果显示:在采暖期主要受西方向(红色线)和东北方向(蓝色线)气流的影响.其中,来自西方向的气流轨迹,表现出受到来自库尔勒以及塔里木盆地西部区域铁门市、轮胎、库车等地的气流影响[14],这些气流轨迹分别占和静县(HJX)和博斯腾湖南(BHN)总轨迹的51%和49%;而东北方向的气流主要来自北疆地区如北乌鲁木齐等地绕经天山东口到达博斯腾湖流域,其分别占和静县(HJX)和博斯腾湖南(BHN)总轨迹的49%和51%.在非采暖期主要受到来自西北方向(红色线)和东北方向(蓝色线)气流的影响,其中东北方向的轨迹走向与采暖期相同,分别占和静县(HJX)和博斯腾湖南(BHN)总轨迹的57%和55%;而来自西北方向的聚类轨迹显示受到来自焉耆盆地气流的影

响,分别占和静县(HJX)和博斯腾湖南(BHN)总轨迹的43%和45%.

前向轨迹聚类结果显示:在采暖期博斯腾湖流域气流轨迹对东北方向(蓝色线)延伸表现出对新疆东部地区的影响;对西南方向(红色线)延伸,则表现出对塔克拉玛干沙漠地区大气环境的影响.而类似地,在非采暖期,气流传输轨迹则主要表现出向新疆东南部地区和西南方向的塔克拉

玛干沙漠地区的影响.此外,还发现在非采暖期和静县(HJX)的前向轨迹(蓝色线)与博斯腾湖南(BHN)后向轨迹(红色线)基本吻合,表现出在非采暖期居民区较高的 PAHs 排放通过大气传输到达博斯腾湖区,经大气干沉降进入水体,可能会对博斯腾湖水生环境造成影响.而在采暖期由于轨迹方向并不契合,湖区受到来自居民区大气污染传输的影响较小.



(a) 采暖期后向轨迹; (b) 非采暖期后向轨迹; (c) 采暖期前向轨迹; (d) 非采暖期前向轨迹 **图 7** 和静县(HJX) 和博斯腾湖南(BHN) 大气后向和前向轨迹

Fig. 7 Atmospheric backward and forward trajectory analysis for the HJX and BHN sampling sites in the Bosten Lake watershed

## 3 结论

(1)博斯腾湖流域大气中  $\sum$  15PAHs 的质量浓度为 6.50~158.93 ng·m<sup>-3</sup>,均值为 61.29 ng·m<sup>-3</sup>;空间分布上呈现出居民区污染最高(132.32 ng·m<sup>-3</sup>),其次为湖泊周边(18.78 ng·m<sup>-3</sup>)和塔中(6.50 ng·m<sup>-3</sup>);季节差异性表现为采暖期远大于非采暖期,并且两个采样期大气中 PAHs 组成以 3~4 环为主,分别占总 PAHs 的 80.29%~97.42%和 66.10%~97.51%.

- (2) 博斯腾湖流域  $\sum$  15PAHs 大气干沉降通量 为 0.70 ~ 9.22  $\mu g \cdot (m^2 \cdot d)^{-1}$ , 均 值 为 2.95  $\mu g \cdot (m^2 \cdot d)^{-1}$ ; 空间分布上呈现出居民区通量最高 [ 7.42  $\mu g \cdot (m^2 \cdot d)^{-1}$  ],其 次 为 塔 中 [ 4.30  $\mu g \cdot (m^2 \cdot d)^{-1}$  ]和湖泊周边[1.57  $\mu g \cdot (m^2 \cdot d)^{-1}$ ];季节差异性表现为采暖期远大于非采暖期,并且两个采样期干沉降中 PAHs 组成以 3 ~ 4 环为主,分别占总 PAHs 的 62.76% ~ 83.54%和 56.76% ~ 84.91%.
- (3)特征比值法源解析结果表明,博斯腾湖流域采暖期和非采暖期大气中 PAHs 主要来源于生物质和煤的燃烧;后向轨迹分析表明新疆北部和塔里木盆地西部气团对博斯腾湖流域大气 PAHs 具有潜在影响;前向轨迹分析表明博斯腾湖流域大气

PAHs 会对博斯腾湖西南及东部地区大气环境造成影响;另外,非采暖期来自居民区较高 PAHs 排放通过大气传输到达博斯腾湖区,经大气干沉降进入水体,可能会对博斯腾湖水生环境造成影响.

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