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水体模拟颗粒物对四环素的吸附特性及基本规律

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摘要:为了更深入了解水环境中颗粒物对抗生素的吸附规律,选用典型抗生素——四环素(TC)和悬浮态的高岭土模拟天然水体中颗粒物对四环素的吸附过程,探究颗粒物对 TC 的吸附规律及不同颗粒物投加量、pH、温度和阳离子对颗粒物吸附 TC 的影响. 结果表明颗粒物对 TC 在混合 4 h 之前快速吸附,之后单位吸附量随时间较小波动,在 12 h 后基本达到吸附平衡. 在溶液中颗粒物对 TC 的单位吸附量随着颗粒物投加量的增大而减小;颗粒物对 TC 抗生素的吸附更符合 Langmuir 等温吸附规律;颗粒物对 TC 的吸附在 pH = 4.5 附近达到最大值,强酸(pH < 4)或强碱(pH > 9)环境均抑制颗粒物对 TC 的吸附;低价态的阳离子如 Na⁺、Ca²⁺(浓度在0.000 1 ~ 0.1 mol·L⁻¹范围)等对颗粒物吸附 TC 均产生抑制作用,且随着离子浓度的增加,抑制作用增强,但三价阳离子的作用却非常特别,如低浓度的 Al^{3+} (0.000 1 mol·L⁻¹)会促进吸附作用,随着 Al^{3+} 浓度增加,促进作用减弱,直到 Al^{3+} 达到较高浓度(0.01 mol·L⁻¹),又会抑制颗粒物对 TC 的吸附. 综合本实验获得的颗粒物吸附 TC 的基本特征和规律,可以初步推断:在实际水环境中由于颗粒物对 TC 的快速吸附和低浓度 Al^{3+} 的促进作用,TC 在水环境和饮用水处理工艺中更易随颗粒物的运动发生同步迁移,颗粒物的归宿主要决定了 TC 的归宿,如进入天然水体沉积物中或饮用水处理工艺的污泥中,这也进一步揭示了天然水体中沉积物往往易被检出较高含量 TC 的根本原因.

关键词:四环素; 高岭土; 吸附; 阳离子; 等温吸附

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Adsorption of Tetracycline on Simulated Suspended Particles in Water

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Abstract: The mechanism of adsorption of a typical antibiotic (tetracycline, TC) on particles in the aquatic environment and the parameters affecting adsorption were investigated. Experiments were conducted to elucidate the effects of pH and various cation species with different concentrations. The results show that the adsorption of TC on particles is rapid during the first four hours in the mixing stage. The adsorption process becomes slow after the first four hours. The adsorption of TC on particles can be described well by a Langmuir equation. The maximum adsorption of TC on particles occurs at pH 4.5, however it is reduced by strongly acidic (pH <4) or alkaline (pH >9) conditions. Moreover, the adsorption process is also inhibited by various cations (e. g. Na⁺ and Ca²⁺) in the range of 0.000 1-0.1 mol·L⁻¹ ionic concentrations. A special finding concerns Al³⁺ ions; at a low concentration of these ions (0-0.000 1 mol·L⁻¹) the adsorption of TC on particles improves, whereas at increased concentrations the adsorption is weakened. In summary, an effective removal of the particles is critical to control TC pollution in natural waters because of the rapid adsorption of TC on particles.

Key words: tetracycline; kaolin; adsorption; cations; isothermal adsorption

近几十年来抗生素在我国的一些主要河流中陆续被检测出来,如黄河[1]、九龙江[2]、巢湖[3]、珠江[4]和洪泽湖[5]等水体. 2009 年中国抗生素产量已经高达 14.7万 t^[6],抗生素用于人体和动物后,随着食物排出体外进入环境. 其中黄浦江作为上海重要的水源地之一,上游曾有大量规模化养殖场及数量庞大的零散养殖场,甚至曾发生过千头死猪浮现黄浦江的事件(2013年)^[7]. 2015年,黄浦江上游水体共被检出 2 类 6 种抗生素残留,TC 是上游主要抗生素污染物之一^[8]. 抗生素进入环境后很容易在环境中累积并且长期存留^[9,10]. 此前四环素类抗生素在黄浦江流域的养猪场、水产养殖场废水中

检测值就达 27.5 ng·L^{-1[11]},四环素在天然水体中残留和污染现状已经十分严峻.四环素类抗生素广泛用于人类医疗和动物养殖中,它可以提高饲料利用率及促进动物生长,使用后随着粪便排出体外^[12],同样很可能进入食物链,危害人类的健康^[13,14],也通过很多途径进入环境,造成污染.

抗生素与多种颗粒物作用研究已有报道,朱齐

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齐等^[15]利用河边的土壤代替天然水体中的颗粒物研究了颗粒物对于磺胺类抗生素的吸附特征. 武庭瑄等^[16]研究了膨润土对四环素吸附影响,发现阳离子对于膨润土吸附四环素与阳离子的价态有关. Chang 等^[17]研究了四环素在蒙脱石和膨润土上的吸附机制,在研究不同 pH 和温度的影响时发现阳离子交换是吸附的主要作用机制.

但是实际环境中如土壤、水体及沉积物中的颗粒物主要成分之一还有高岭土类黏土矿物.如靖青秀等^[18]发现南方离子型稀土矿区土壤氨氮污染严重,其中高岭土是矿区土壤的主要黏土成分之一;席雅娟等^[19]在调查杭州湾潮滩沉积物成分发现表层沉积物主要是伊利石和高岭石,赵德博等^[20]在提取海洋沉积物中黏土矿物发现含量最多的主要是蒙脱石和高岭石.可见高岭土类成分在土壤、沉积物及水体颗粒物中分布广泛,因此本研究选用悬浮态高岭土模拟水体悬浮颗粒物.

抗生素在水环境中与颗粒物的相互作用可能是 影响其迁移转化的关键环节之一, 因为已有大量研 究发现水体中颗粒物常与污染物相互作用并成为其 载体, 如吸附结合水中的重金属离子、持久性有机 污染物(POPs)等微污染物^[21],并在很大程度上决 定着这些污染物在环境中的迁移转化和循环规 律[22~24]. 但是, 对于典型抗生素(如四环素)排放 至湖泊河流等天然水体后, 其发生怎样的迁移转化 过程并不完全清楚,例如是更多的抗生素立即被颗 粒物吸附还是立即进入沉积物,还是更多的溶于水 体,或者是容易随颗粒物迁移,还是容易随水本身 迁移等等具体的过程目前均有待深入揭示. 因此, 存在着研究典型抗生素四环素在水中与悬浮颗粒物 之间吸附特征及规律的必要. 本文采用悬浮态的高 岭土模拟水体颗粒物,四环素为目标化合物,探究 它在悬浮态高岭土上的吸附作用特性及其环境影响 因素, 研究结果有助于初步揭示实际水体中抗生素 的迁移归趋规律,并对开发针对天然水体中抗生素 污染的控制与处理技术有一定的指导意义.

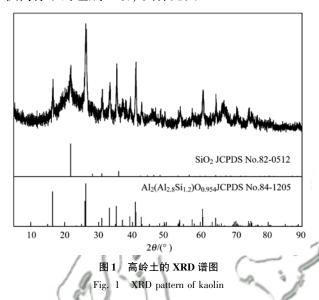
1 材料与方法

1.1 材料与仪器

盐酸四环素购自德国 Dr. Ehrenstorfer 公司. 高岭土为分析纯,购自天津市致远化学试剂有限公司,主要成分有 SiO₂ 45.38%,Al₂O₃ 38.46%等. 盐酸、氢氧化钠和六水氯化铝均为分析纯,购自国药集团化学试剂有限公司,氯化钠和无水氯化钙均

为分析纯,购自上海凌峰化学试剂有限公司. 紫外分光光度计(U-2910 日本日立集团),电热恒温鼓风干燥箱(上海精宏实验设备有限公司),pH 计(上海梅特勒-托利多仪器公司),多点磁力搅拌器(上海科升仪器有限公司),恒温摇床(上海精宏实验设备有限公司),0.45 μm 滤膜购自上海安谱实验科技股份公司.

取高岭土粉末进行 XRD 衍射, 分析结果表明高岭土主要含有 SiO₂, Al₂O₃ 成份, 与上述试剂提供商标示的组成一致, 具体见图 1.



1.2 仪器分析方法和 TC 浓度测定

取 20 mg·L⁻¹ TC 溶液,用紫外分光光度计扫描峰谱图,选择 275 nm 波长作为 TC 测定波长(图 2).测溶液 TC 浓度前均使用 0.45 μm 滤膜过滤溶液,每次实验均通过测定 TC 吸光度与浓度的标准曲线进行未知样中 TC 的定量.每批滤膜使用前均通过预实验来确认滤膜对 TC 无吸附作用.

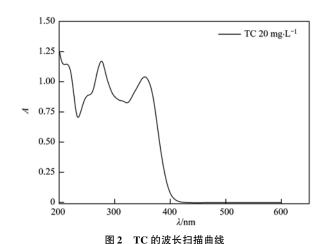


Fig. 2 Wavelength scanning of tetracycline

1.3 实验方法

1.3.1 模拟颗粒物悬浮液的制备和使用

取一定质量的高岭土于1 L 纯净水中,用磁力搅拌器连续搅拌 2 h,静置沉淀 24 h,取溶液上方稳定悬浮液部分作为本研究的模拟悬浮颗粒储备液.悬浮储备液中颗粒物浓度的定量方法:取 100 mL 的悬浮储备液,过 0.45 μm 膜(已烘干称重),截留了颗粒物的滤膜再烘干称重,差重法得到 100 mL 储备液中颗粒物质量为 1.7 g,因此本研究使用的悬浮颗粒物储备液的颗粒物浓度为 17 g·L⁻¹.

模拟悬浮颗粒物使用液的配制:在进行后续吸附实验时,根据所需颗粒物浓度和体积的需要,将悬浮颗粒物储备液进行一定比例的稀释即得悬浮颗粒物使用液.

为了使本研究的结果更利于实际应用和参考,本实验采用同样的差重法测定了实际黄浦江水体悬浮颗粒物浓度(6次平均值为5 mg·L⁻¹),以使本研究一系列实验中的模拟悬浮颗粒物使用液浓度范围尽可能覆盖一般实际水体悬浮颗粒物浓度(由于实际水体颗粒物成分的混杂性,本研究所有实验均未采用黄浦江水体悬浮颗粒物,除特别指明,后续实验及讨论涉及的颗粒物,均指由高岭土配制的模拟颗粒物).

1.3.2 颗粒物吸附 TC 动力学实验

分别配制颗粒物浓度为 0、0.34、0.68 和 2 g·L⁻¹的悬浮颗粒物使用液(0 g·L⁻¹为无颗粒物对照样本),分别准确量取 50 mL 置于锥形瓶中,每种浓度平行准备至少 10 个锥形瓶,再分别投加同样体积的 TC 储备液于锥形瓶中,使溶液中 TC 的最终浓度为 5 mg·L⁻¹,然后同时置于恒温摇床中以120 r·min⁻¹的速度连续振荡,在 0.08、0.25、0.5、1、2、4、8、12、18 及 24 h 采集样品,取上清液用0.45 μm 滤膜过滤.同时还进行无 TC 的空白对照实验,实验均做 3 个重复,测定了初始溶液 pH 值均为 6.1.

颗粒物对 TC 的吸附量由以下公式进行计算:

$$Q_{t} = (c_0 - c_e) \times V/M$$

式中, Q_ι 为颗粒物对 TC 的吸附量 $(mg \cdot g^{-1})$, c_0 为 TC 初始浓度 $(mg \cdot L^{-1})$, c_e 为上清液中 TC 浓度 $(mg \cdot L^{-1})$;V 为溶液体积(L);M 为吸附剂颗粒物质量(g).

1.3.3 温度对吸附影响实验

分别取一定体积的悬浮颗粒物使用液置于 50 mL 锥形瓶中, 其中颗粒物的浓度为 0.34 g·L^{-1} , 并

分为三组(每组6瓶),投加 TC 使其初始浓度分别为 0.5、2.5、10、15、20 和 30 mg·L⁻¹,调节 pH 为 4.5(预实验发现颗粒物吸附 TC 的最适 pH 为 4.5,详见 2.3 节),用锡箔纸将 3 批锥形瓶密封后分别在 10、20 和 35℃ 的温度条件(模拟春、夏、秋环境)下于恒温摇床内恒温振荡至吸附平衡,取上清液过 0.45 μm 滤膜后测定其中 TC 浓度.吸附前后溶液中 TC 的质量差即为溶液中颗粒物对 TC 的吸附量,参照上述实验方法进行实验和分析.

由此绘制等温吸附线并且用 Langmuir 和 Freundlich 方程对等温吸附线进行拟合.

1.3.4 初始 pH 值对高岭土吸附 TC 的影响

分别配制颗粒物浓度为 0.0.34.0.68 和 2 g·L⁻¹的悬浮颗粒物使用液,设置 TC 的起始浓度均为 5 mg·L⁻¹,设置起始 pH 分别为 2.5.3.4.5.6.7.8.9 和 10 在 25 ∞ 室温条件下进行恒温振荡实验,达平衡吸附。同上方法测定颗粒物对 TC 的单位吸附量和吸附率,上述实验均设置对应的无颗粒物和无 TC 的平行对照实验.

1.3.5 阳离子对高岭土吸附 TC 的影响实验

使用无水氯化钠、氯化钙和氯化铝试剂准确配制浓度分别为 1 mol·L^{-1} 和 0.1 mol·L^{-1} 的储备液,备用. 通过稀释法分别配制浓度为 0.1、0.01、0.001 和 0.0001 mol·L⁻¹的 Na⁺、Ca²⁺和 Al³⁺阳离子使用液,分别移取阳离子使用液于 4 组 50 mL 锥形瓶中,同时配置空白对照一组(不加入阳离子). 分别投加同样体积 TC 储备液使其初始浓度均为 5 mg·L⁻¹,均在 25°C 室温条件下进行吸附平衡后取上清液过 0.45 μ m 滤膜,并用分光光度法测定溶液中颗粒物对 TC 的吸附量和去除率.

2 结果与讨论

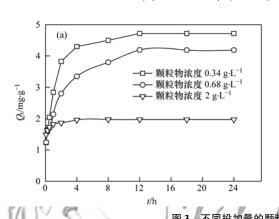
2.1 不同投加量的颗粒物对 TC 吸附平衡时间的确定

颗粒物对 TC 的吸附特征见图 3, 由图 3(a)可知, 颗粒物在前 4 h 快速吸附, 4 h 后吸附量随着时间有微小波动,在 12 h 后基本达到吸附平衡, 颗粒物对 TC 的吸附量达到最大值.

对于固定投加浓度的 TC(5 mg·L⁻¹),随着颗粒物投加量的增加,对 TC 的总吸附量和总去除率在增加[图3(b)],同时颗粒物对 TC 的单位吸附量却降低[图3(a)]. 因为固定 TC 浓度的溶液中有更多高岭土颗粒时,相同时间内 TC 与颗粒物的接触几率增大,这使得颗粒物对 TC 的总吸附量也增

大. 但是, 每增加单位质量颗粒物其吸附的 TC 量 逐渐下降, 因为从浓度平衡和浓度梯度驱动来看, 随着颗粒物的增加和对 TC 的不断吸附, 溶液中 TC 总浓度下降, TC 浓度对颗粒物浓度的比值也 在下降, 因此从溶液到颗粒表面上的 TC 浓度梯度 下降, TC 往颗粒物表面迁移的驱动力在下降, 从 而导致随着投加颗粒物浓度的增加, 颗粒物对 TC 的单位吸附量下降. 这一点, 与后面的等温吸附 结果(图4)完全一致, 即其符合 Langmuir 单分子 层吸附规律.

对于一般河流水体,如黄浦江,5~10月为汛 期, 其水流速度为5 400 m·h⁻¹, 颗粒物浓度大约 5 $mg \cdot L^{-1}(m1.3.1 \, \overline{)}$,那么如果某个时刻有养殖 场排放一定量的 TC 进入水体(排放区域的 TC 浓度 若接近本实验的 TC 投加浓度),根据图 3(a)可知,



起初 TC 浓度相对颗粒物浓度的比值较大, 颗粒物 对 TC 的单位吸附量也较大. 再又根据去除率图 3 (b)可估算, 在前 1 h(将 1 h 作为一个微分时间单 位考虑)约有 20%~30%的 TC 被颗粒物吸附,同 时,这20%~30%被吸附的TC也随水流扩散了大 约5 400 m, 之后 3 h 又增加吸附 20% 左右 TC, 且 吸附范围扩大到 18 km 左右. 实际上, TC 浓度一直 随水流迁移时不断下降,遇到新的颗粒物一定发生 新的吸附和平衡, 且 TC 浓度相对颗粒物浓度的比 值越来越小, 越有利于 TC 被颗粒物吸附去除, 这 由图 3(b)清晰可见. 由此可知, 在天然水体中, TC 进入水体后,大部分(超50%)TC 更可能随颗粒物 进行迁移. 这也给出一个启示, 就是在治理水体 TC 污染或制定相关污染控制政策时, 重点可关注颗粒 物的控制和去除.

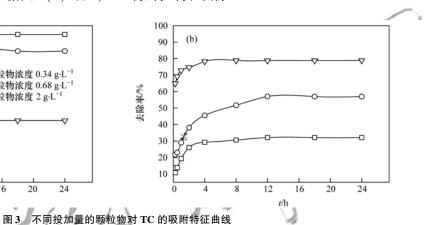


Fig. 3 Adsorption of tetracycline on kaolin in different dosages

温度对吸附效果的影响

将颗粒物浓度固定为 0.34 g·L⁻¹, 对 5 mg·L⁻¹ 的 TC 在 3 个不同温度下的吸附等温曲线如图 4 所 示,分别用 Langmuir 和 Freundlich 等温方程对数据 进行拟合. Langmuir, Freundlich 等温吸附模型的数 学方程如下.

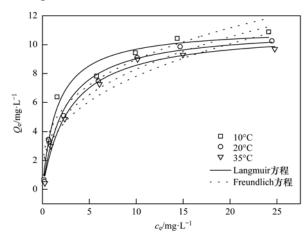
(1) Langmuir 等温式:

$$Q_e = \frac{K_1 Q_{\rm m} c_e}{1 + K_1 c_e}$$

(2) Freundlich 等温式:

$$Q_{\rm e} = K_{\rm f} c_{\rm e}^{1/n}$$

式中, Q_e 表示平衡吸附量, 单位 $mg \cdot g^{-1}$; c_e 表示溶 液达到平衡后,溶液中四环素的浓度,单位为 mg·L⁻¹; Q_m 表示最大吸附量,单位mg·g⁻¹. K₁ (Langmuir)常数表示吸附剂对吸附质结合能力的大 小, $K_{\rm f}$ (Freundlich)常数表示吸附容量, 反映了吸附 能力的强弱; n 表示特征常数. 表示吸附过程的非 线性程度和吸附机制上的差异. 四环素在高岭土上 的 Langmuir 和 Freundlich 吸附拟合参数如表 1.



不同温度下 TC 在颗粒物上的的吸附等温线

Fig. 4 Adsorption isotherms of tetracycline on kaolin

由图 4 可知, 3 种不同的环境温度下, TC 在颗 粒物上的吸附等温线均可以用 Langmuir 和 Freundlich 等温吸附模型拟合, 其中 Langmuir 方程 的拟合效果更好, R^2 均大于 0.97. 在不同的温度环

境下其中随着温度的升高,颗粒物对 TC 的吸附量降低[²⁵].

表 1 不同温度下四环素在高岭土上的 Langmuir 和 Freundlich 吸附拟合参数

Table 1	Parameters of the	Langmuir and	Freundlich e	equations for	tetracycline	adsorption on kao	lin

温度/℃ -	Langmuir 吸附等温方程			Freundlich 吸附等温方程			
血皮/ C	$Q_{ m max}$	K_1	R^2	$K_{ m f}$	1/n	R^2	
10	10. 885	0. 646	0. 972	4. 340	0. 315	0. 910	
20	10. 262	0. 427	0. 973	4. 686	0. 262	0. 927	
35	9. 695	0. 349	0. 989	4. 191	0. 286	0. 933	

图 5 给出了用 Zeta 值表示其特征的 Zeta-pH 关系,说明颗粒物在 pH 值为 3~11 时,表面带负电荷,电荷量的大小依赖于溶液 pH 值的变化,在 H⁺和 OH⁻作用下,表面电荷具有相应的变化.

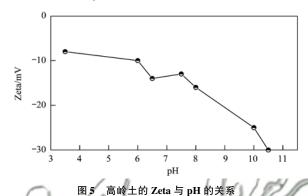


Fig. 5 Relation between Zeta and pH of kaolin

2.3 pH 对颗粒物吸附 TC 的影响

初始 pH 对颗粒物吸附 TC 的影响如图 6,随着溶液中 pH 从 2.5 开始不断加大即溶液中 H⁺的减少,颗粒物对 TC 的吸附量不断加大并且在 pH = 4.5 左右达到最大吸附量,而颗粒物浓度别为 0.34、0.68 和 2 g·L⁻¹时,颗粒物对 TC 的吸附量随 pH 变化的规律均情况相似.

四环素带有多种官能团, 在不同的 pH 条件下

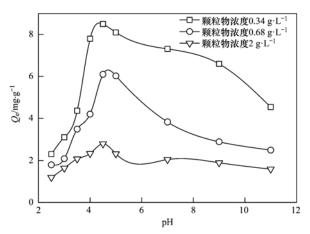


图 6 初始 pH 对颗粒物吸附 TC 的影响

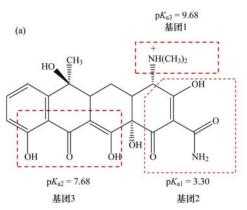
Fig. 6 Adsorption of tetracycline at different initial pH

会在水体里以不同的离子形态存在,高岭土颗粒物表面带负电荷,同时还有一部分可变电荷^[26,27],TC的 pK_a 值分别为 3. 30、7. 69 和 9. $69^{[28,29]}$,分别与溶液中的 H^+ 与 OH^- 结合,使其在不同 pH 值溶液中以阳离子,兼性离子或者阴离子形态存在,从而影响 TC 在颗粒物上的吸附作用.

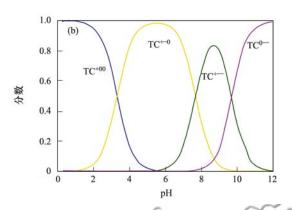
由图 7 可知, pH 小于等于 2 时四环素主要以 TC+00的形式存在(这时基团1带1正电荷,基团2 和3电荷均为0,因此图7(b)中标示为TC+00,因 此过多的 H+与基团 1 的竞争作用反而降低了颗粒 物对 TC 的吸附作用, 所以图 6 中起点的单位吸附 量较低; 当 pH 值为 2~4.5 时, 四环素的主体由 TC +00 逐渐变为 TC + -0 (TC + -0 即由 TC +00 的基团 2 逐渐脱离一个 H⁺而来), TC⁺⁻⁰可是一个带正电荷 的二甲氨基基团和一个带负电荷的酚羟基的兼性离 子(+-0). 这区间 H⁺浓度逐渐下降,对 TC 的竞 争逐渐减弱,同时,TC 自身仍带有正电荷,与带有 负电荷的颗粒物表面仍具有较强的吸附势能(颗粒 物表面约-10 mV, 图 5), 到 pH 为 4.5~5 附近, TC 大部分变为 TC+-0, 从而在颗粒物上具有最高 的单位吸附量. 随着 pH 值不断升高即 pH > 5 以 后,TC+-0比例下降[30],基团三开始逐步脱H+离 子, 其负电荷分子形态(如 TC+--和 TC0--)所占 比例不断增加[31],同时,颗粒物表面负电荷量也在 增加(如图 5),颗粒物与 TC 表现逐步相斥,从而 导致颗粒物对 TC 的单位吸附量减小.

在 pH < 4.5 时,随着 pH 的增大即溶液中 H⁺ 的减少,颗粒物对 TC 的吸附量快速增加;当 pH > 4.5 时,随着 pH 的增大即溶液中 OH ⁻ 的增多,颗粒物对四环素的吸附量减少,这可能是由于在颗粒物上 TC 的吸附主要发生在高岭土矿物的表面层,在高岭土矿物的表面层 TC 与颗粒物的吸附过程主要是由其各自带有的表面电荷影响^[32]. 其中颗粒物表面带负电荷,并且存在着特殊离子可以在不同环境下改变其带电性. 而 TC 中含有二甲胺基

 $N(CH_3)_2^{[33]}$ 、酰氨基 $CONH_2$ 、酚羟基 C-OH 和两个含有酮基和烯醇基的共轭双键系统可分别与溶液中的 H^+ 和 OH^- 结合,使四环素分子在不同 pH 环境下的溶液中以阴离子、兼性离子的形态存在,从而改变其表面电性,影响 TC 与颗粒物的表面吸附特性 [34].



综上可知,最容易被颗粒物吸附的主要形态是TC⁺⁻⁰,且吸附的机制主要是电荷作用,但需求较好的电荷匹配特性,即除了TC主要以TC⁺⁻⁰形态存在外(一般溶液 pH 为 4.5~5 范围),还要求颗粒物表面的负电荷量不能太高,本实验条件显示 – 10 mV 左右为宜.



(a)四环素的结构,(b)四环素在不同 pH 的形态分布

图 7 四环素的结构及在不同 pH 时的形态分布 Fig. 7 Tetracycline structure and speciation as a function of pH

2.4 阳离子对颗粒物吸附 TC 效果的影响/

不同阳离子与阳离子浓度对颗粒物吸附 TC 的影响如图 8, 其中颗粒物浓度为 0.34 g·L⁻¹, 此时 pH 为 6.1. 对于 Na⁺和 Ca²⁺而言, 在低浓度时候 (浓度小于等于0.000 1 mol·L⁻¹), 此两种阳离子对 吸附作用几乎无影响. 随着离子浓度的增高, 此两种阳离子对于颗粒物吸附 TC 出现了同等程度的抑制作用, 这可能是溶液中的 TC 由于阳离子的竞争 吸附作用而导致了颗粒物对 TC 吸附量发生了变化. 例如 Ca²⁺可以形成 CaCl⁺形态而易在颗粒物表面吸附,竞争了吸附位点, 从而降低了颗粒物对 TC 的吸附量.

而三价阳离子 Al³+对高岭土吸附四环素的影响非常特别. 有研究表明四环素与 Al³+容易形成稳定络合物^[35]. 相对于没有投加任何阳离子的时候(如图 8 中 None 所对应单位吸附量),当 Al³+浓度为0.000 1 mol·L⁻¹, Al³+的存在明显促进了颗粒物对 TC 的吸附. 而当浓度为0.001 mol·L⁻¹时及更高时, Al³+的存在却明显抑制颗粒物对 TC 的吸附. 从2.3 节可知,最容易被颗粒物吸附的 TC 形态是TC+⁻⁰,而且最佳 pH 范围为4.5 左右,因此本实验特别测定了投加0.000 1 mol·L⁻¹的 Al³+时溶液 pH值为4.3,可见0.000 1 mol·L⁻¹的 Al³+促进颗粒物吸附 TC 的原因可能是因为 Al³+水解改变溶液 pH值进而改变了 TC 的优势形态(TC+⁻⁰),同时使 TC

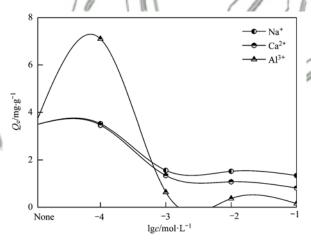


图 8 不同阳离子的不同浓度对颗粒物对 TC 吸附曲线

Fig. 8 Effect of various cations and their concentrations on the adsorption of tetracycline on kaolin

和 pH 都达到最佳条件. 而且, 此时由于 Al³⁺浓度相对较低, 且部分 Al 离子水解为 Al (OH)²⁺或 Al (OH)²⁺等形态^[36], 从而与 TC 的竞争吸附也相对较弱. 甚至部分 Al 离子水解形态与 TC⁺⁻⁰结合,增加 TC⁺⁻⁰的正电荷量, 从而使 TC 更被表面带负电荷的颗粒物吸附. 还有可能, Al³⁺水解形态进入颗粒物表面,部分改变颗粒物表面电荷分布特征. 总之结果表现为促进作用.

而当 Al^{3+} 浓度较高时 ($0.001~mol \cdot L^{-1}$),因 Al^{3+} 水解使 pH 值下降到 4 以下,这时 TC 主体变为 不易被吸附的 TC^{+00} 形态,且颗粒物表面负电荷量

也有减少,这都不利于颗粒物对 TC 的吸附. 同时, Al³⁺水解形态和组成也发生改变,这里面涉及太多复杂变化,未来将有一个专题研究进行揭示,总之, Al³⁺浓度升高到 0.001 mol·L⁻¹时结果表现为抑制颗粒物吸附 TC 的作用.

不过,由于此次实验中发现的关于 Al³⁺影响最特别的现象是较低浓度的 Al³⁺对颗粒物吸附 TC 产生了促进作用,为了更好地探究 Al³⁺对颗粒物吸附 TC 的影响,特又设置了不同低浓度的 Al³⁺梯度进行更详细的实验研究,具体见 2.5 节.

2.5 低浓度铝离子对高岭土吸附四环素的影响

颗粒物浓度分别为 0. 34、0. 68 和 2 g·L⁻¹,溶液中添加的铝离子分别为 0. 1、0. 2、0. 3、0. 4、0. 5 mmol·L⁻¹,由图 9 可知,当 Al³⁺的浓度达到 10⁻⁴ mol·L⁻¹时,Al³⁺的存在促进了颗粒物对 TC 的吸附,但随着 Al³⁺浓度的逐步增加,促进作用减弱,逐渐变为抑制颗粒物对 TC 的吸附。这进一步验证了 2. 4 节的主要现象和结论,当 0. 1 mmol·L⁻¹ Al³⁺ 对促进颗粒物吸附 TC 的原因可能是因为 Al³⁺水解改变溶液 pH 值进而改变了 TC 的优势形态(TC⁺⁻⁰),同时使 TC 和 pH 都达到最佳条件.

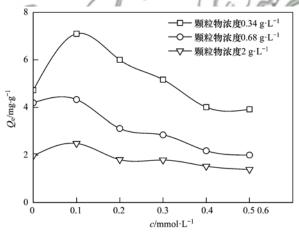


图 9 低浓度铝离子对高岭土吸附四环素的影响

Fig. 9 Effect of Al³⁺ concentrations on the adsorption of tetracycline on kaolin

高岭土吸附 TC 时 pH 值的影响如图 6 所示, pH 对颗粒物吸附抗生素的独立影响为:当 pH 值从 6 下降到 4.5 时,颗粒物对抗生素的吸附容量增加,可见 Al³⁺对减弱它们之间的吸附作用更加强烈,总体表现为单位吸附量随 Al³⁺浓度的增加而下降. 因此在天然水体环境下 pH 值不发生剧烈波动时,Al³⁺对他们相互吸附作用的负影响会更加强烈. 正因为考虑到 Al³⁺浓度的影响,本实验所选用的 Al³⁺浓度已经很低.

3 结论

- (1)颗粒物对 TC 的吸附分为快速吸附和缓慢吸附两个阶段,颗粒物对 TC 抗生素的吸附更符合 Langmuir 等温吸附规律,当颗粒物浓度不变时,随着 TC 浓度的减少, TC 浓度对颗粒物浓度比值下降时,颗粒物对 TC 的单位吸附量也下降.
- (2) 在 pH 为 4.5 左右时, H⁺浓度逐渐下降, 对 TC 的竞争逐渐减弱,同时, TC 自身仍带有正电荷,与带有负电荷的颗粒物表面仍具有较强的吸附势能(颗粒物表面约 10 mV),到 pH 为 4.5 附近, TC 大部分变为 TC⁺⁻⁰,从而在颗粒物上具有最高的单位吸附量.
- (3)不同的阳离子对与颗粒物吸附 TC 的影响与浓度有关,其中 Ca^{2+} 、 Na^+ 随着离子浓度的增加,与颗粒物竞争吸附 TC 的能力增加,但对 Al^{3+} 来说浓度为0.000 1 $mol \cdot L^{-1}$, Al^{3+} 的存在明显促进了颗粒物对 TC 的吸附. 而当浓度为 0.001 $mol \cdot L^{-1}$ 时及更高时, Al^{3+} 的存在却明显抑制颗粒物对 TC 的吸附.
- (4)对实际水环境的启示:由于颗粒物对TC的快速吸附和低浓度 Al³⁺的促进作用,TC 在水环境和饮用水处理工艺中更易随颗粒物一起发生同步迁移,这也进一步说明了天然水体中沉积物往往易被检出较高含量TC的原因.

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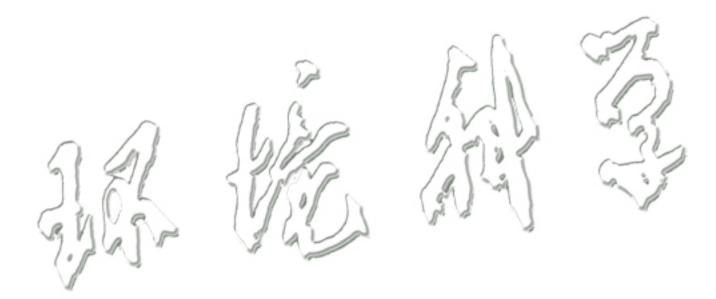
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