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典型土壤双季稻对 Cd 吸收累积差异

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摘要:用盆栽试验方法研究典型土壤双季稻条件下水稻对 Cd 的吸收累积差异. 选取典型水稻土黄泥田(板页岩母质发育)和麻砂泥(花岗岩母质发育),通过添加不同浓度梯度外源 Cd,进行盆栽试验,研究双季稻不同生育期土壤有效态 Cd(DTPA-Cd)、水稻植株各部位以及糙米 Cd 累积情况. 结果表明,双季稻晚稻生育期土壤有效态 Cd 大于早稻,黄泥田大于麻砂泥,其差异性均达极显著水平(P<0.01). 水稻植株各器官(根、茎、叶、壳和糙米)Cd 累积量随外源 Cd 增加和生育期的延长而呈现递增的趋势. 不同生育期、不同土壤水稻糙米与植株各器官 Cd 累积量差异显著,具体表现为:早稻小于晚稻,黄泥田小于麻砂泥. 水稻各器官(根、茎、叶、壳和糙米)中 Cd 含量与土壤有效 Cd 含量呈显著或极显著正相关关系. 应用稻米 Cd 含量预测模型及水稻累积 Cd 的特征方程推算出土壤 Cd 安全阈值为:黄泥田早稻 0.98 mg·kg⁻¹和晚稻:0.83 mg·kg⁻¹; 麻砂泥分别为 0.86 mg·kg⁻¹和 0.56 mg·kg⁻¹. 不同母质土壤的安全阈值与环境容量不同,其环境质量标准与污染修复控制措施应该有所区别.

关键词:Cd; 母质类型; 双季稻; Cd 累积; Cd 安全阈值

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Differences in Cd Accumulation in Typical Soils Under the Double Rice System

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Abstract: Pot experiments were used to study the differences of Cd uptake and accumulation in double-cropping rice in typical soil types. To analyze the soil availability of Cd (DTPA-Cd) in soils and the Cd accumulation in double-cropping rice at different growth stages of the rice, we conducted pot experiments that selected the yellow clayey soil (paddy soil developed from plate shaley parent materials) and the granitic sandy soil (paddy soil developed from granitic parent materials). Exogenous Cd was added with gradients of 0, 0.5, 1.0, 2.0, 5.0, and 10.0 mg·kg⁻¹. Results showed that, during the rice growth period, the available Cd in the yellow clayey soil was higher than that in the granitic sandy soil, and the difference was significant (P < 0.01). This showed that the content of Cd in rice (roots, shoots, leaves, rice shells, and brown rice) increased along with the treatment level and with the extension of the rice growth period. The accumulation characteristics of Cd in rice grains and other tissues of rice indicated differences between two seasons and two soil types, that is, late rice was higher in Cd than was early rice, and reddish yellow clayey soil was higher in Cd than granitic sandy soil. Significant positive linear correlations were found between the effective contents of Cd in soils and those in rice tissues (roots, shoots, leaves, and brown rice). The prediction model of Cd in rice and the characteristic equation for rice accumulation of Cd were applied to calculate the critical values of Cd; 0.98 mg·kg⁻¹ for early rice and 0.83 mg·kg⁻¹ for late rice in reddish yellow clayey soil, and 0.86 mg·kg⁻¹ for early rice and 0.56 mg·kg⁻¹ for late rice in granitic sandy soil. These threshold values are higher than the National Standards given in "farmland environmental quality evaluation standards for edible agricultural products (HJ 332-2006)." The soil security threshold values and the soil environmental capacities of the two different parent materials varied greatly; therefore, different environmental quality standards may be formulated and different measures may be needed to control Cd pollution in different parent materials.

Key words: Cd; different types of paddy soil; double-rice system; accumulation of Cd; Cd security threshold values

工农业及城市化的快速发展加速了土壤和环境中 Cd 的污染程度^[1].水稻是我国第一大粮食作物,而 Cd 是一种极具生物毒性的重金属元素,作为有色金属之乡的湖南已经广泛关注水稻土 Cd 的危害^[2,3].由于土壤 Cd 污染不仅造成稻米品质下降,影响我国农业的可持续发展,还能通过食物链累积,危害人身健康^[4].多年来,国内外进行了大量的土壤-水稻体系中 Cd 的迁移研究,水稻品种、土壤类

型对重金属的累积存在较大差异^[5~10]. 甘国娟^[11]的研究表明不同母质土壤-水稻系统 Cd 的迁移系数、积累系数存在明显差异. 成颜君等^[12]、赵步洪

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等[13]的研究表明不同母质类型土壤水稻糙米对 Cd 的累积大于水稻品种基因型. Ding 等[14]和 Sun 等[15]的研究表明不同土壤 Cd 阈值有所不同. 因此关注典型成土母质稻田土壤的 Cd 污染尤为重要,不同土壤对 Cd 吸收累积差异以及土壤重金属 Cd 安全阈值是一项重要研究内容. 本文通过盆栽试验方法研究湖南两种典型成土母质水稻土双季稻 Cd 吸收累积差异,推算水稻土 Cd 安全阈值,以期为土壤环境质量评价与污染修复控制提供参考依据.

1 材料与方法

1.1 供试材料

供试土壤黄泥田为板页岩母质发育的水稻土, 采自长沙县路口镇燕窝屋(地理坐标:北纬 28°26′46″,东经113°19′13″);供试土壤麻砂泥为花 岗岩母质发育的水稻土,采自长沙县金井镇脱甲村(地理坐标:北纬28°33′31″,东经113°20′5″). 两种土壤均取自耕作层(0~15 cm),土壤采回后,捡出肉眼可见的石粒、根系碎屑等杂物,经风干、研磨后混合均匀备用. 两种土壤的基本理化性质如表1所示.

1.2 试验设计

盆栽试验用具为无盖圆形胶质盆,高 20 cm,桶 底半径 10 cm. 每盆装土 4.0 kg. 加入 $CdCl_2$ 溶液, Cd 水平为:0、0.5、1.0、2.0、5.0、10.0 mg·kg⁻¹ (以下图表中均用 0、0.5、1、2、5、10 表示),平衡 老化 30 d. 按 N 0.15 g·kg⁻¹, P_2O_5 0.1 g·kg⁻¹, K_2O 0.15 g·kg⁻¹,以尿素 [$CO(NH_2)_2$]、磷酸铵 [$(NH_4)_3PO_4$]和碳酸钾(K_2CO_3)的水溶液加入作基肥.

表1 供试土壤基本理化性质

		Table 1	Basic physical an	d chemical propert	ies of soil		
土壤名称	母质类型	Cd 含量	DTPA 浸提 Cd	pH 值	CEC	有机质	黏粒(< 0.002 mm)
工場名你	母灰矢型	/mg•kg ⁻¹	/mg·kg ⁻¹	рп 1	∕cmol·kg ⁻¹	/g•kg ⁻¹	//%
黄泥田	板页岩	0.09	0.06	6. 13	8. 97	14. 03	25.37
麻砂泥	花岗岩	0.05	0.03	5. 79	7. 85	20. 98	14. 28

早晚稻盆栽试验分别进行.供试水稻品种早稻为湘早籼45号,晚稻为湘晚籼13号.每盆种植水稻2穴,每穴2株,按大田处理要求进行施肥,整个生育期保持田面水层2cm,每个处理设5个重复.早稻于2016年5月25日移栽,7月15日收获;晚稻于2016年7月15日移栽,10月9日收获.

1.3 采样与分析

本试验采用破坏性取样法(即采集整盆全部样品). 早稻分别于分蘖盛期(5月25日)、抽穗期(6月10日)、成熟期(7月15日);晚稻分别于分蘖盛期(8月10日)、抽穗期(9月10日)、灌浆期(9月25日)、成熟期(10月9日)采集水稻全株与土样. 去离子水洗净植物后分离根、茎、叶、谷壳与糙米, 并于105℃杀青30 min后,70℃烘干至恒重,然后磨碎、过70目筛、称量,备用. 土样自然风干,磨碎, 过10目筛用于pH值、可提取态重金属浓度测定;过100目筛用于土壤总重金属测定.

样品分析:土壤理化性质采用常规分析方法测定^[16];土壤有效态 Cd 采用 DTPA 浸提法测定^[17];土壤总 Cd 采用王水-高氯酸混酸消解,火焰原子吸收分光光度法测定^[17]. 植物样品 Cd 采用干灰化法消解,石墨炉原子吸收分光光度法测定^[17]. 为保证数据的可靠性和稳定性,植株 Cd 含量测定时每个

样测 3 次,并以国家标准植物样品 GSB-23(湖南大米)进行质量控制,相对标准偏差(RSD)低于 5%.

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1.4 数据分析

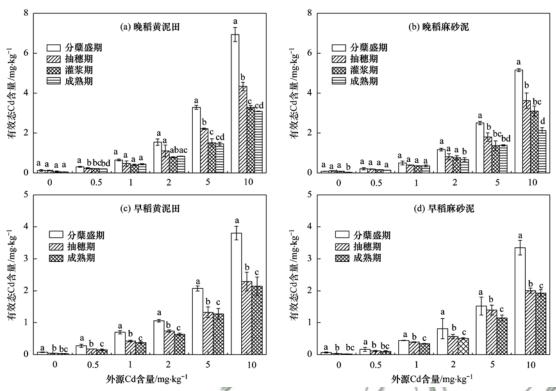
本试验中的数据结果采用显著性 F 测验和 Duncan 多重比较法(*P* < 0.05) 进行统计分析,应用 Excel 2016、SPSS 22.0 和 Origin 9.0 进行处理.

2 结果与分析

2.1 水稻生育期土壤有效态 Cd 变化

水稻各生育期土壤有效态 Cd 含量测定结果表明, 黄泥田 Cd 胁迫处理早稻土壤有效态 Cd 含量在 $0.15 \sim 3.81 \text{ mg} \cdot \text{kg}^{-1}$, 平均含量 $1.16 \text{ mg} \cdot \text{kg}^{-1}$, 晚稻土壤有效态 Cd 含量在 $0.20 \sim 7.29 \text{ mg} \cdot \text{kg}^{-1}$, 平均含量 $1.67 \text{ mg} \cdot \text{kg}^{-1}$, 黄泥田早晚稻平均 $1.45 \text{ mg} \cdot \text{kg}^{-1}$; 麻砂泥 Cd 胁迫处理早稻土壤有效态 Cd含量在 $0.10 \sim 3.36 \text{ mg} \cdot \text{kg}^{-1}$, 平均含量 $0.98 \text{ mg} \cdot \text{kg}^{-1}$, 晚稻土壤有效态 Cd含量在 $0.13 \sim 5.22 \text{ mg} \cdot \text{kg}^{-1}$, 平均含量 $1.34 \text{ mg} \cdot \text{kg}^{-1}$, 麻砂泥早晚稻平均 $1.19 \text{ mg} \cdot \text{kg}^{-1}$. 晚稻土壤有效态 Cd大于早稻; 黄泥田土壤有效态 Cd均大于麻砂泥.

随着水稻生育期的延长,土壤有效态 Cd 含量 呈持续降低趋势,并从分蘖盛期到抽穗期下降最多, 该趋势在高浓度添加 Cd 条件下更明显(图1);同



不同小写字母表示同一母质相同 Cd 水平不同生育期土壤有效态 Cd 含量差异显著(P<0.05)

图 1 水稻各生育期土壤有效态 Cd 含量变化

Fig. 1 Change of available Cd in the soil at different growth stages of rice

时,随着外源 Cd 含量的增加,双季稻各生育期土壤有效态 Cd 含量提高,土壤有效态 Cd 含量均与外源 Cd 含量呈极显著相关关系(P<0.01). 相比麻砂泥,黄泥田土壤有效态 Cd 含量提高更多,表现出显著差异(P<0.05); 当外源 Cd 含量相同时,各时期晚稻土壤有效态 Cd 大于早稻,黄泥田大于麻砂泥,其差异性均达到显著水平(P<0.05).

2.2 水稻植株 Cd 吸收累积的生育期变化

从图 2 可以看出,水稻各生育期中,早、晚稻根部 Cd 含量在分蘖盛期不高,在抽穗、灌浆期迅速上升,达到最高并随后降低;茎部 Cd 含量在分蘖盛期处于较高水平,并在抽穗期有所下降,到灌浆期达到最高,之后在成熟期 Cd 又明显下降(图 3),可见根、茎、叶中 Cd 含量受生育期变化影响明显.此外,不同土壤和不同季度下生长的水稻根、茎部 Cd 的累积特征有所不同,表现为晚稻高于早稻,麻砂泥高于黄泥田,且差异均达到显著水平(P<0.05).水稻叶部对 Cd 的累积特征基本与根、茎一致(图 4),表现为随生育期增延长 Cd 含量呈增加的趋势,并在灌浆期达到最高,成熟期下降.高水平外源 Cd 下该趋势更明显.不同母质与不同季度条件下,叶对 Cd 的累积存在差异,不同母质发育土壤上晚稻叶部 Cd 的累积存在差异,不同母质发育土壤上晚稻叶部 Cd

含量在各生育期均显著高于早稻;麻砂泥发育下的水稻叶部 Cd 含量亦高于黄泥田.

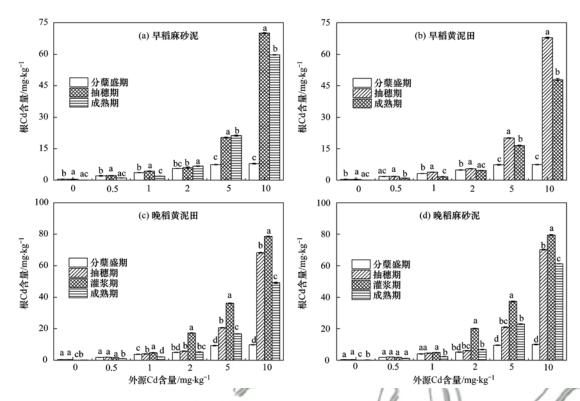
2.3 糙米 Cd 的累积

从图 5 可知,当外源 Cd 在 0~5 mg·kg⁻¹,早晚稻糙米 Cd 含量均表现出随外源 Cd 含量的增加而上升的趋势. 当外源 Cd 含量为 5 mg·kg⁻¹时,水稻糙米中的 Cd 含量达到最大值,而当外源 Cd 含量为 10 mg·kg⁻¹时,糙米中的 Cd 比外源 Cd 含量为 5 mg·kg⁻¹时有一定程度的降低. 说明低浓度的 Cd 胁迫有促进水稻生长,而 Cd 浓度过高时抑制水稻生长而产生毒害作用.

不同季水稻及不同成土母质类型条件下,水稻糙米对 Cd 的累积存在显著差异(P<0.05),表现为晚稻高于早稻,麻砂泥高于黄泥田.相同 Cd 处理水平条件下,外源 Cd 水平越高差异越明显,如外源 Cd 含量为 5 mg·kg⁻¹时,晚稻糙米 Cd 含量在黄泥田和麻砂泥条件下显著高于早稻,分别是早稻的 2.8 倍和 1.7 倍;早稻糙米 Cd 含量在麻砂泥条件下显著高于黄泥田条件下的,是其 1.3 倍.

2.4 水稻全株 Cd 累积量

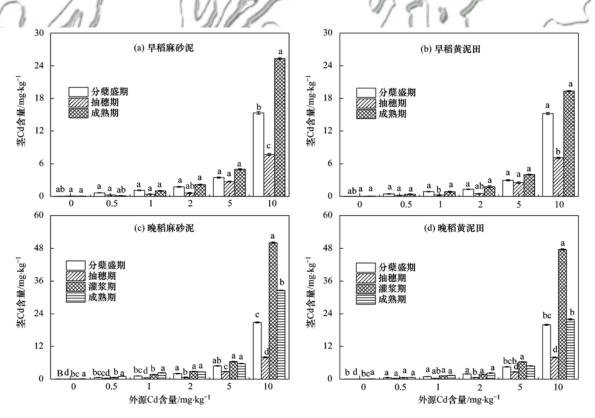
通过测定水稻各部位(根、茎、叶、壳和糙米) Cd含量及干重,汇总计算出水稻全株 Cd 累积量



不同小写字母表示同一母质相同 Cd 水平不同生育期水稻根部 Cd 差异显著(P<0.05)

图 2 水稻各生育期根 Cd 含量的变化

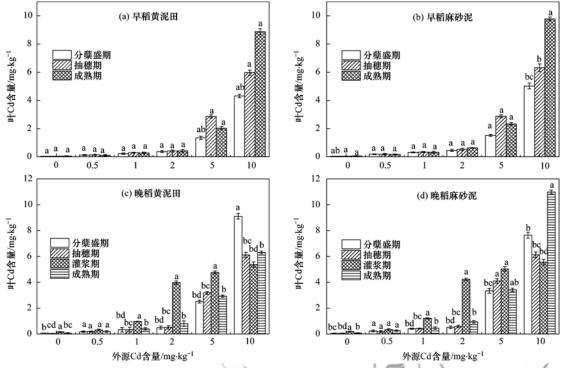
Fig. 2 Change of the concentration of Cd in roots at different growth stages of rice



不同小写字母表示同一母质相同 Cd 水平不同生育期水稻茎部 Cd 差异显著(P<0.05)

图 3 水稻各生育期茎 Cd 含量的变化

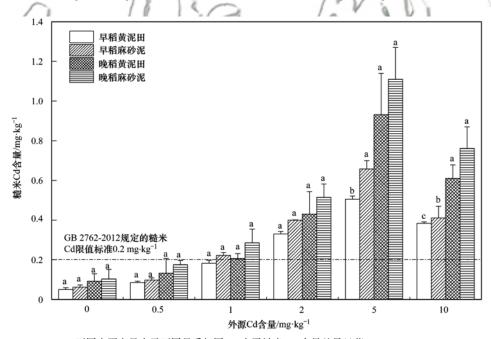
Fig. 3 Change of the concentration of Cd in stems at different growth stages of rice



不同小写字母表示同一母质相同 Cd 水平不同生育期水稻叶部 Cd 差异显著(P<0.05)

图 4 水稻各生育期水稻叶 Cd 含量的变化

Fig. 4 Change of the concentration of Cd in leaves at different growth stages of rice



不同小写字母表示不同母质相同 Cd 水平糙米 Cd 含量差异显著(P<0.05)

图 5 添加不同浓度 Cd 对糙米 Cd 含量的影响

Fig. 5 Effects of different concentrations of Cd on the Cd content in brown rice

(表2). 结果表明:随着水稻生育期的延长,水稻全株 Cd 累积量表现出从分蘖盛期到成熟期的持续增加的趋势,且随外源 Cd 含量的提高而逐渐增加,这种趋势在高浓度外源 Cd 处理下更明显,这主要是由于根、茎、叶生长率加快以及不断的累积 Cd 而

使得 Cd 累积量始终保持增加的趋势. 水稻全株 Cd 累积量在不同季水稻及不同母质条件下有所不同,按不同土壤和早、晚稻统计 Cd 胁迫处理水稻全株 Cd 累积量,早稻平均为 1.54 mg·pot⁻¹,而晚稻为 3.04 mg·pot⁻¹,晚稻大于早稻,约其 2 倍;黄泥田平

表 2 水稻全株 Cd 累积量¹⁾/mg·pot ⁻¹

Table 2	Total	accumulation	of Cd	in	rice/	mg • pot	- 1
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外源 Cd	早	.稻	晚	超
/mg•kg ⁻¹	黄泥田	麻砂泥	黄泥田	麻砂泥
0	$0.03 \pm 0.01 df$	$0.03 \pm 0.01 df$	$0.05 \pm 0.01 \mathrm{cf}$	$0.09 \pm 0.02 df$
0. 5	0. 15 \pm 0. 04de	$0.12 \pm 0.01 ce$	$0.30 \pm 0.01 ce$	$0.42 \pm 0.15 ce$
1	$0.30 \pm 0.01 \mathrm{cd}$	$0.35 \pm 0.10 \mathrm{cd}$	$0.69 \pm 0.01 \mathrm{cd}$	$0.90 \pm 0.23 \mathrm{cd}$
2	$0.57 \pm 0.01c$	$0.77\pm0.10\mathrm{bc}$	$1.12 \pm 0.04c$	$1.95 \pm 0.43c$
5	1. 74 ± 0.06 b	1. $28 \pm 0.42b$	$2.91 \pm 0.75 $ b	3.73 ± 0.63 b
10	$4.78 \pm 0.41a$	$5.37 \pm 0.56a$	$7.88 \pm 0.95 a$	$10.53 \pm 1.32a$

¹⁾ 同列比较,不同小写字母代表差异显著(P<0.05)

均为 2. 04 mg·pot⁻¹, 麻砂泥为 2. 54 mg·pot⁻¹, 麻砂泥大于黄泥田.

2.5 水稻各部位 Cd 与土壤有效态 Cd 及外源 Cd 水平的关系

相关分析结果表明,不同成土母质双季稻根、茎、叶 Cd 含量均与土壤有效态 Cd 及外源 Cd 水平呈极显著的正相关关系(表3). 早稻外源 Cd 水平与糙米 Cd 含量和土壤有效 Cd 含量的相关系数范围在 0.640~0.994 之间,其中与土壤有效 Cd

含量均呈极显著相关水平(P<0.01),而 Cd 水平与糙米中 Cd 含量相关性不显著.晚稻在不同成土母质条件下,Cd 水平与糙米 Cd 含量和土壤有效 Cd 含量的相关系数范围在 0.699~0.997 之间,其中与土壤有效态 Cd 含量均达到了极显著相关水平(P<0.01),麻砂泥条件下土壤重金属 Cd 有效态含量与糙米中 Cd 含量达到了显著相关水平(P<0.05),而外源 Cd 水平与糙米中 Cd 含量相关性不显著.

表 3 水稻各部位 Cd 含量、有效态 Cd 含量和外源 Cd 水平的相关性1

Table 3 Correlation among Cd concentration in rice, the effective concentration of Cd and exogenous Cd addition

项目	1	早	稻		10	晚程	Í	93)
700	黄泥田		麻砂泥		黄泥田		床	形砂泥
60	外源 Cd	有效态 Cd						
糙米	0. 724	0. 794	0. 640	0. 711	0.719	0. 699	0.738	0. 830 *
根	0. 987 **	0. 964 **	0. 989 **	0. 969 **	0. 987 **	0. 985 **	0. 991 **	0. 959 **
茎》,	0. 958 **	0. 925 **	0. 956 **	0. 924 **	0. 960 **	0. 965 **	0. 944 **	0. 885 **
(原性)/	0. 962 **	0. 928 **	0. 965 **	0. 934 **	0. 997 **	0. 991 **	0. 981 **	0. 938 **
有效态 Cd	0. 994 **	1.000	0. 994 **	1.000	0. 997 **	1.000	0. 987 *	1. 000

1) * * 表示 Pearson 相关性在 0.01 水平之上极显著, * 表示 Pearson 相关性在 0.05 水平之上显著, n = 12

3 讨论

3.1 不同母质土壤水稻对 Cd 吸收累积差异的原因

一般认为,土壤-水稻系统 Cd 的迁移以及糙米 Cd 累积受 pH 值、土壤机械组成、黏粒矿物含量、阳离子交换量、有机质、土壤湿度和土壤温度等因素影响^[18,19].土壤淹水,氧化还原电位(Eh)下降并形成还原环境,土壤阳离子浓度有所减少,导致其与 Cd 的吸附竞争降低;土壤有机质的积累使水稳定性聚集体更加稳定,均使得 Cd 在土壤颗粒上的吸附作用得到促进^[20~23].试验结果早稻糙米 Cd 比晚稻低,可能与早稻时期降雨量大且土壤温度相对较低有关.土壤温度会影响水稻对 Cd 的累积,土壤温度升高,在 Cd 的刺激下,水稻根系土壤的微生物含量、C 含量和酶活性增加,促进了 Cd 向水稻体内的转运^[24].土壤温度升高导致土壤孔隙水 pH 值降

低,水溶性 Cd 浓度增加^[25],有利于 Cd 在水稻植株的迁移累积.

土壤有效态 Cd 麻砂泥低于黄泥田,但水稻糙 米中 Cd 高于黄泥田,可能与黄泥田黏粒含量高有 关.麻砂泥土壤颗粒较大,比表面积小,因此其吸附 Cd 的能力较弱^[27].两者差异性形成原因可能还与黏粒矿物类型有关^[28].花岗岩母质发育而来的麻砂泥土壤黏粒矿物主要是1:1型(高岭土),而板页岩发育而来的黄泥田土壤黏粒矿物还含有2:1型(云母、伊利石等).1:1型黏土矿物,无膨胀性,带电荷少,胶体特性差,CEC含量低;2:1型黏土矿物则带电量较大,CEC含量较1:1型高.CEC反映了土壤胶体的负电荷量,CEC 越高,负电荷越高,能够提供更多的吸附位点吸附 Cd²⁺,使土壤对 Cd 的吸附量增加,可能导致进入糙米中的 Cd 含量降低^[27-31].供试土壤黄泥田 pH 值高于麻砂泥,pH 值

对土壤重金属有效态含量影响显著,pH 值越高,带正电荷的重金属离子被带负电荷的土壤胶体吸附越多、与黏土矿物发生共沉淀的量也越多,越容易被固定而降低其活性;同时,pH 值越高,土壤中 Fe、Mn 等离子形成的羟基化合物为 Cd 提供了越多的吸附位点,从而使有效态 Cd 含量降低,因此进入糙米中的 Cd 含量有所降低^[32,33].

3.2 不同母质土壤稻米 Cd 安全阈值

不同成土母质条件下,通过糙米 Cd 含量与外源 Cd 含量的对应关系模拟出 Cd 累积方程(图 6),并将国家大米镉限值标准(GB 2762-2012, 0.2 $mg \cdot kg^{-1}$)代入方程,得到土壤 Cd 安全阈值分别为,早稻:0.98 $mg \cdot kg^{-1}$ (黄泥田)与 0.86 $mg \cdot kg^{-1}$ (麻砂泥)以及晚稻:0.83 $mg \cdot kg^{-1}$ (黄泥田)与 0.56 $mg \cdot kg^{-1}$ (麻砂泥),均高于国家标准(HJ 332-2006). 范中亮等[34]研究了碱性潮土(pH = 7.50)和酸性水稻土(pH = 5.94)土壤 Cd 安全阈值分别为 1.49 $mg \cdot kg^{-1}$ 和 0.79 $mg \cdot kg^{-1}$;陈齐等[27]研究得出

碱性紫色土(pH = 7.70)和酸性红黄泥(pH = 5.10)土壤 Cd 安全阈值分别为 6.36 mg·kg⁻¹和 3.08 mg·kg⁻¹. 对比可知,碱性土壤 Cd 安全阈值远高于酸性土壤,说明 pH 值是影响土壤 Cd 阈值的重要因素,这与李志博等^[35]根据稻米 Cd 的预测模型和稻米摄入风险研究的估算结果一致. 进行土壤修复时,应考虑施加酸碱调节剂调控土壤 pH 值. 因此本研究为湖南稻米 Cd 限量标准和生态系统健康的土壤 Cd 安全阈值和大田施用降 Cd 措施提供了一定的科学依据.

值得注意的是,本试验研究只是盆栽试验一年的结果,虽然本试验水稻土老化时间为 30 d,添加的外源 Cd 在土壤中的行为可能与田间条件有所差异. Ma 等^[36]和刘斌等^[37]的研究发现土壤添加不同水平外源 Cd 后的土壤有效态 Cd 随老化时间的延长而降低,影响其生物有效性. 因此 Cd 安全阈值的研究仍然需要多年多点与田间试验验证.

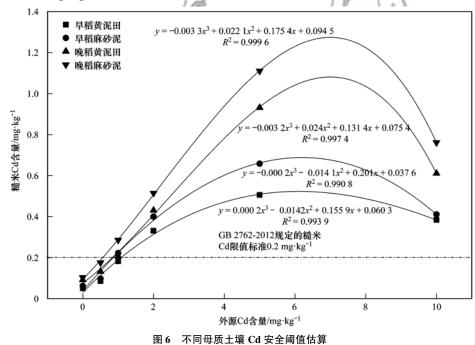


Fig. 6 Calculated safety threshold of Cd in different parental soils

4 结论

(1)晚稻各生育期土壤有效态 Cd 大于早稻,黄泥田大于麻砂泥,其差异性均达到极显著水平(P < 0.01). 随着外源 Cd 浓度的增加,水稻植株各器官(根、茎、叶、壳和糙米)Cd 累积量随生育期的延长而呈现递增的趋势. 不同母质土壤 Cd 累积量差异显著,具体表现为:早稻小于晚稻; 黄泥田小于麻

砂泥.

(2)双季稻糙米 Cd 含量随添加 Cd 浓度的增加表现出增加的趋势,外源 Cd 浓度为 5 mg·kg⁻¹时,水稻糙米 Cd 含量达到最大值,外源 Cd 浓度为 10 mg·kg⁻¹时,水稻糙米 Cd 有一定程度的降低.不同母质土壤水稻糙米 Cd 差异显著,具体表现为:早稻小于晚稻,黄泥田小于麻砂泥. 糙米中 Cd 含量与土壤有效 Cd 含量呈显著正相关;根、茎、叶中 Cd 含

量与土壤有效 Cd 呈极显著正相关.

(3)本试验得到的湖南典型成土母质水稻土安全阈值为,早稻:0.98 mg·kg⁻¹和 0.86 mg·kg⁻¹(黄泥田和麻砂泥,下同);晚稻:0.83 mg·kg⁻¹和 0.56 mg·kg⁻¹.不同母质土壤的安全阈值与环境容量不同,其环境质量标准与污染修复控制措施应该有所区别.

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