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区域点源和非点源磷入河量计算的二元统计模型

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摘要:基于流域或区域点源和非点源磷入河过程的水文学差异,以及影响河流持留作用的主要机制,建立了 描述河流段末磷负荷量与流量和水温之间定量关系的二元统计模型;通过 逐月的河流水文水质监测数据对模型中 4 个系数的有效校正和验证,实现了对点源和非点源磷入河过程的准确定量.与现行的水文估算法相比,该模型既考虑了河流磷的持留能力及其时间变异性,也考虑了上游水体输入的磷负荷量,推进了对磷污染过程的定量认识,满足了我国以行政区为主要水污染控制管理单元的现实需要.应用该模型,计算了浙江长乐江集水区 2004~2009 年的 总磷(TP)人河量.结果表明,TP 年入河总量为(54.6±11.9)t·a⁻¹,其上游水体输入、点源和非点源的入河量贡献率分别为 5%±1%、12%±3%和 83%±3%.夏季5~6月和 8~9月的 非点源 TP 累计入河量占其全年的 50%±9%,增加了引起下游水体藻类暴发的风险.河流 TP 持留量为(4.5±0.1)t·a⁻¹,占年入河总量的 9%±2%;5~9月的 TP 累计持留量占全年的 55%±2%,表明河流持留能力对流域或区域磷素迁移转化过程的调控作用不容忽视.本研究建立的二元统计模型仅需常规的河流水文水质监测数据,无需专业软件知识,且计算结果直接来源于实际的河流水文水质测算值,为实施流域或区域磷污染总量控制策略提供了一种简便、实用、可靠的定量工具.

关键词:点源; 非点源; 河流持留能力; 入河量; 二元统计模型; 磷中图分类号: X52 文献标识码: A 文章编号: 0250-3301(2013)01-0084-07

Bivariate Statistical Model for Calculating Phosphorus Input Loads to the River from Point and Nonpoint Sources

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Abstract: Based on the hydrological difference between the point source (PS) and nonpoint source (NPS) pollution processes and the major influencing mechanism of in-stream retention processes, a bivariate statistical model was developed for relating river phosphorus load to river water flow rate and temperature. Using the calibrated and validated four model coefficients from in-stream monitoring data, monthly phosphorus input loads to the river from PS and NPS can be easily determined by the model. Compared to current hydrological methods, this model takes the in-stream retention process and the upstream inflow term into consideration; thus it improves the knowledge on phosphorus pollution processes and can meet the requirements of both the district-based and watershed-based water quality management patterns. Using this model, total phosphorus (TP) input load to the Changle River in Zhejiang Province was calculated. Results indicated that annual total TP input load was (54.6 ± 11.9) t·a⁻¹ in 2004-2009, with upstream water inflow, PS and NPS contributing to 5% ±1%, 12% ±3% and 83% ±3%, respectively. The cumulative NPS TP input load during the high flow periods (i. e., June, July, August and September) in summer accounted for 50% ±9% of the annual amount, increasing the alga blooming risk in downstream water bodies. Annual in-stream TP retention load was (4.5 ± 0.1) t·a⁻¹ and occupied 9% ±2% of the total input load. The cumulative in-stream TP retention load during the summer periods (i. e. , June-September) accounted for 55% ±2% of the annual amount, indicating that in-stream retention function plays an important role in seasonal TP transport and transformation processes. This bivariate statistical model only requires commonly available in-stream monitoring data (i. e., river phosphorus load, water flow rate and temperature) with no requirement of special software knowledge; thus it offers researchers and managers with a cost-effective tool for quantifying TP pollution processes in both district and watershed scales.

Key words: point source; nonpoint source; in-stream retention capacity; input load to river; bivariate statistical model; phosphorus

由氮、磷等营养物过剩引起的湖库、河口、海岸等水体富营养化是国内外普遍关注的水污染问题

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之一^[1-3]. 大量实践证明,实施总量控制是遏制水污染、改善水质的有效手段^[4,5]. 然而,总量控制方案的制定首先必须明确到底有多少营养物从陆域进入水体、点源和非点源的贡献率各为多少^[6-8]. 河流作为上游流域或区域的营养物向下游水体输送的主要通道^[5],使得如何准确计算流域或区域的点源和非点源营养物入河量成为实施总量控制策略所需解决的关键科学问题之一.

理论上,点源营养物入河量可直接采用监测手 段获得. 然而,许多企业为了降低治污成本,污水超 标排放和偷排现象较为普遍[9,10]. 在许多城镇地 区,由于缺乏生活污水集污管网和污水处理厂,使得 居民区通常分布着为数众多、排污量小的独立排污 口. 因此,由于人力和物力的限制,实际中尚很难对 流域或区域内所有的点源进行有效监测. 对非点源 营养物入河量的计算目前主要依赖数学模型方法. 降雨量差值法[11]、水文分割法[12,13]、水质水量相 关法[14] 等水文估算方法是应用较为广泛的模型方 法之一. 这类水文估算方法主要基于点源与非点源 营养物入河过程的水文学差异而建立的,即点源的 营养物入河量稳定且与径流过程无关,而非点源营 养物入河量则主要依赖于径流过程. 也有学者基于 两类污染源的这一水文学差异,建立了分割河流点 源和非点源营养物负荷量的统计模型,解决了其他 水文估算法在水文期划分或基流分割过程中存在的 主观性问题[15,16]. 总体而言,与输出系数法、机理 性模型等其他方法相比[8],水文估算方法不仅所需 的数据资料较少(仅需河流水文水质监测数据)、操 作简单,而且能同时估算点源和非点源污染引起的 河流营养物负荷量及其相对污染贡献率. 相对于我 国基础数据资料缺乏的现状而言,这无疑提供了一 种较为可行的非点源污染定量方法.

然而,现行的水文估算方法仍存在一定的不足. 一方面,这些水文估算方法往往忽略了河流营养物的持留过程.河流持留能力是指河流生态系统通过自然生态过程及物质循环作用,将水体中氮、磷、有机物等富营养化物质或有害物质予以吸收、转化、再分配,从而使水体净化的能力[1,4].大量研究表明,河流年均氮、磷持留量可占入河总量的1%~80% [6,7,17].显然,如果忽略了河流营养物持留过程,计算得到的应为点源和非点源污染引起的河流表观负荷量,即低估了实际的人河量.由于河流水文生态条件的变化,河流营养物持留能力又存在着显著的季节性差异[4,17,18],因此,现行的水文估算法 计算得到的点源和非点源污染相对贡献率也存在较大的误差.另一方面,由于未考虑上游水体输入的营养物负荷量,使得现行的水文估算方法往往只能适用于流域尺度的研究,却不能适用于区域尺度的研究,很难满足我国以行政区为主要水污染控制管理单元的现实需要.随着政府和公众对保护水环境的需求增加,许多国家和地区对河流的水文水质监测工作在不断加强,使得仅需河流监测数据的水文估算方法将具有更大的应用潜力.因此,针对以上的不足,对现行水文估算方法进行必要改进,对推进营养物总量控制策略实施具有重要现实意义.

本研究针对点源和非点源营养物入河量计算问题,取现行水文估算方法之长并对其不足进行必要改进,构建一个能同时适用于流域和区域的、考虑河流磷持留能力的二元统计模型,以期为实施河流营养物总量控制提供简便、实用的定量工具.

1 区域点源和非点源磷入河量计算的二元统计模型

根据河段营养物输入-输出平衡原理(图 1),某一研究区域河流出口断面中第 i 月的磷负荷量(L_i , $kg \cdot f^{-1}$)是各点源入河量(P_i , $kg \cdot f^{-1}$)、各非点源入河量(N_i , $kg \cdot f^{-1}$)、上游水体输入(U_i , $kg \cdot f^{-1}$)和支流汇入带来的磷负荷量总和减去由取水渠系带走的磷负荷量(C_i , $kg \cdot f^{-1}$),并经过河流持留过程作用后的最终结果。由于支流输入的磷负荷量往往也主要来自该区域相应的点源和非点源,故而可不单独考虑。因此, L_i 可表示为 $^{[6,7]}$:

 $L_i = (P_i + N_i + U_i - C_i)R_i$ (1) 式中, R_i 为第 i 月的河流磷持留因子(无量纲).

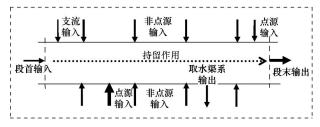


图 1 河段营养物输入-输出平衡示意

Fig. 1 Conceptual diagram of nutrient input-output balance in a river reach

由于非点源污染的发生主要受径流过程的影响,某一区域非点源磷的入河量时间变异性与径流量的变化密切相关^[18~21],因而,非点源磷入河量可以近似地表达为河流流量的幂函数^[15,16].与之相反,来自于工业企业和城镇生活以及大型畜禽养殖场等的点源磷排放量相对稳定,可将一定研究时期

内(即点源磷的排放和污染治理措施无显著变化情况下)逐月的点源磷入河量近似地看成为一个常数^[13,15,16,19,22].因此,点源和非点源磷入河量分别可以表示为:

$$P_i = A; \quad N_i = BQ_i^c \tag{2}$$

式中, $A \setminus B \setminus C$ 为模型系数(无量纲); Q_i 为第 i 月的平均流量($\mathbf{m}^3 \cdot \mathbf{s}^{-1}$).

河流营养物持留能力的大小主要取决于河流水体的滞留时间长短、水体与河床的接触面积大小以及水生生物的活性强弱[17,18,21~24].由于河流流量的增加不仅降低了营养物参与各种物理作用的几率和生物地球化学反应的时间,而且也减少了单位水体中营养物与河床接触面积,使营养物持留效率呈现随流量增加而降低的趋势[6,17].河流水生生物(包括大型水生植物)的生长速率以及各种生物地球化学反应速率均会随着水温的增加而提高,使营养物持留效率呈随水温增加而提高的趋势[4,21~24].因此,式(1)中的河流磷素持留因子 R_i 可表示为一个河流流量和水温的单调递减的自然指数函数[7].

$$R_i = \exp(-Dq_i t_i) \tag{3}$$

式中, q_i 和 t_i分别表示归一化后的河流流量和水温变量(无量纲); D 为模型系数(无量纲). 与水温不同,河流流量与营养物持留能力之间表现为负相关关系,因此,归一化处理前首先对实测的流量值取倒数. 本研究中采用逐月平均流量的倒数和水温分别与研究期内逐月流量的倒数最大值和水温最大值相

比的方法,进行数据归一化处理[7].

因此,式(1)可以进一步表示为:

 $L_i = (A + BQ_i^c + U_i - C_i) \exp(-Dq_it_i)$ (4) 式(4)包含4个模型系数,即A、B、C、D. 系数 A 表示来自研究流域或区域所有点源的磷入河量. 系数 B 表示来自研究区域内所有非点源磷的入河潜力. 系数 C 表示河流流量或径流量对非点源磷入河量的作用大小. 系数 D 表示河流对磷的持留潜力大小. 某一流域或区域在研究时期内对点源和非点源污染调控措施未发生显著变化的情况下,以上4个模型系数可认为是常数. 作为常规监测内容,河流流量、磷浓度以及水温数据获取较为容易. 因此,式(4)实质上是一个含有4个待定系数的、描述河流段末磷负荷量与流量和水温之间定量关系的二元统计模型. 当考虑上游水体输入项 U_i 时,式(4)适用于区域和流域尺度;当不考虑时,则适用于流域只要

本研究采用 Microsoft Excel[©]软件中的规划求解模块对二元统计模型系数 A、B、C、D 进行求解.为了减少求解误差绝对值大小对收敛效率的影响,求解前首先对式(4) 左右两边同时取自然对数^[7].求解的目标函数(O)设为出口断面磷输出负荷量测算值的对数值与模拟值的对数值之间误差绝对值的平均值,根据研究时期内 n 个月的出口断面磷输出负荷量以及流量和水温测算值,并限定各模型系数均大于等于 0,求解得到使目标函数 O 达到最小值下的模型系数值:

$$\operatorname{Min}(O) = \frac{\sum_{i=1}^{n} \operatorname{abs}[\ln(L_i) - \ln(A + BQ_i^C + U_i - C_i) - Dq_i t_i]}{n}$$
(5)

2 实例研究

以浙江省嵊州市的长乐江为研究对象^[4],河流"源短流急",是我国东南地区农业小流域的典型代表.本研究河段从位于上游南山水库坝下的河流起始断面至下游雅致桥的出口断面^[4,18],集水区面积641 km²,河长70.5 km,河宽40~70 m,多年平均流量11.8 m³·s⁻¹,沿河无明显的取水渠系. 2004~2009年^[6,18],集水区内的农地(水田、旱地、茶园、苗圃、林地、竹园)、人居地、林地分别占了集水面积的43%±2%、7%±1%、49%±3%,居住人口为(25.7±1.1)万人,农地施用的磷肥量(纯磷)为(1997±124) $t\cdot a$ ⁻¹.

2.1 数据来源和分析

2004-01~2009-12,对长乐江干流的起始和段末断面进行了以月为单位的水质监测 [4.6.18]. 水样的 TP 含量采用过硫酸钾消解、钼酸铵分光光度法测定 (GB 11893-89). 两个监测断面的逐日流量和月平均水温由浙江省水文局提供. 每个监测日的 TP 负荷量根据测定的 TP 浓度与日平均流量乘积获得,从而建立各断面每日的流量与 TP 负荷量的回归统计方程(起始断面 $R^2=0.86^{**}$,n=72),再根据河流逐日流量推算两个断面中逐月的累积负荷量 [18.21.25]. 由浙江省环境保护科学设计研究院提供的数据资料显示,长乐江集水区内已受监测的 11 家排污企业和 16 家规模化畜禽养殖场年均污水排放总量为 200 × 10^3 10^3 10^3 10^3 10^3 10^3 10^3 10^3 10^3 10^3 10^3 10^3 10^3 10^3

2.2 模型校验结果

本研究首先采用长乐江 2004~2007 年逐月的 TP 负荷量、流量、水温数据,根据式(5),对二元统计模型系数 A、B、C、D 进行了校正,再采用 2008~2009 年逐月数据进行了验证.校正得到的模型系数如表 1 所示,河流段末 TP 负荷量的校验结果如图 2 所示.结果表明,模型对河流 TP 负荷量模拟值能很好地反映测算值的按月变化,两者的决定系数 R^2 >0.97 **, Nash-Sutcliffe 系数 [7] >0.90,相对误差 <10%.参考已有研究对 SWAT、AGNPS、HSPF等机制性模型应用结果的总结评价 [26],当河流逐月氮、磷负荷量模拟结果的 Nash-Sutcliffe 系数 >0.65时,模型的校验结果为相当满意.因此,本研究建立的二元统计模型具有较强的数据同化能力,其校验结果可以满足实际研究的需要.需要指出的是,本

实例研究中根据各监测日流量与当日 TP 负荷量之间的回归统计方程推算河流逐月的 TP 负荷量方法[18,21],由于每月一次的监测结果尚很难反映研究期内所有水文事件的情况(尤其是在河流特别枯水和丰水的情况),使计算得到的逐月 TP 负荷量存在着一定的误差[25].因此,尽管模型的校验结果相当满意,但存在着一定的不确定性.在今后的研究中,有必要通过提高水质监测频率以提高河流 TP 负荷量的测算精度[25],从而进一步降低本模型的不确定性.总体而言,与输出系数法、机制性模型等相比,该模型无需详细的流域性状资料和数据以及专业软件知识,且其模拟结果直接来源于实际的河流水文水质测算值,为我国基础资料缺乏现状条件下实现水污染污染过程定量提供了一种简便、可靠的工具.

表 1 长乐江集水区校正的模型系数、各类源逐年的 TP 入河量以及河流 TP 持留量

Table 1 Calibrated model coefficients, annual TP input load to the river, and in-stream TP retention load in the Changle River catchment

		,		,			U	
模型系数			TP 人河量/t∙a ⁻¹				河流 TP 持留量	
A	В	С	D	年份	点源	非点源	上游断面	/t•a ⁻¹
'				2004		29.1	1.7	4.7
				2005		55.9	3.3	4.4
526.4	306.1	1.01	0.67	2006	6.3	37.9	2.2	4.6
320.4	300.1	1.01	0.07	2007	0.3	53.0	3.1	4.5
				2008		41.2	2.4	4.5
				2009		56.4	3.4	4.4

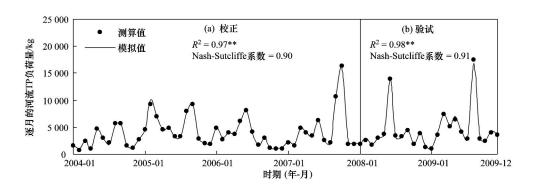


图 2 长乐江集水区应用二元统计模型模拟逐月河流段末 TP 负荷量的校正和验证结果

Fig. 2 Calibration and validation results of the bivariate statistical model for monthly river TP load in the Changle River catchment

2.3 集水区的磷入河量

2004~2009年,长乐江集水区 TP 年均入河总量为(54.6±11.9)t·a⁻¹(表1),上游断面输入的 TP 负荷量占5%±1%.尽管本研究中上游断面输入的 TP 负荷量贡献率较小,然而本研究建立的二元统计模型由于考虑了这一输入项,使其能同时满足流域和区域尺度的研究需要,对于我国以行政区为主要水污染控制单元的现状而言具有重要现实意义.

长乐江集水区点源的 TP 年均入河量为 6.3

t·a⁻¹(表 1),占总人河量的 12% ± 3%. 从逐月的点源 TP 人河量贡献率变化看(图 3),点源的贡献率随着河流流量的增加而显著降低($R^2 = 0.98^{**}$). 在河流月平均流量 < 4.0 m³·s⁻¹的情况下,点源的贡献率 > 30%(图 3),因此,枯水期控制点源污染仍有必要. 然而,对长乐江集水区内 11 家工业排污企业和 16 家规模化畜禽养殖场的监测结果显示,已监测的点源 TP 年均排放总量 < 2.0 t·a⁻¹,远小于本研究的计算结果,这表明实际中存在着企业超排、

偷排以及生活污水直排等现象. 实地调查结果也表明本研究区内尚无集污管网和相应的污水处理厂,使得沿河的居民区存在着一些独立的生活污水直排口. 随着人们生活水平的提高,家用化粪池在村镇逐渐普及,而这些化粪池往往直接与就近的溪流、沟渠相连,使其成为营养物排放稳定、量少、数量众多的点源^[16]. 因此,本研究建立的二元统计模型,能帮助相关管理部门识别可能存在的点源超排、偷排等现象,推进对点源污染的监控.

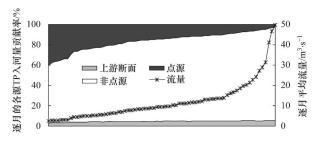


图 3 长乐江集水区河流平均流量增加情况下相应的点源、 非点源、上游水体 TP 入河量贡献率月变化趋势

Fig. 3 Changes of monthly TP input contributions for the point source, nonpoint source and upstream inflow with increasing average river water flow rate in the Changle River catchment

长乐江集水区非点源的 TP 年均入河量为 (45.6 ± 11.2) t·a⁻¹(表1),占了总入河量的83% ±3%,这与实地调查显示长乐江受非点源污染为 主的结果相一致[4,6],使得每月的 TP 总入河量分 布规律与非点源的相似(图4). 本模型计算得到 的 2004~2006 年非点源 TP 年均入河量(40.1 t·a⁻¹)与采用输出系数法的估算结果 38.3 t·a⁻¹ 相接近[4],这也进一步表明了本模型的可靠性. 从 每月的变化看,非点源 TP 入河量在夏季 5~6 月 和8~9月的丰水期达到最大值(图4),这4个月 的累计入河量占了全年的50%±9%. 由于每年5 ~9月的夏季不仅降雨量集中,产生了较大的营养 物入河量,而且水温较高,使得夏季成为我国东南 地区下游湖库、河口以及海岸带藻类暴发的主要 时期^[1~3]. 非点源 TP 入河量的这种时间分布规律 无疑增加了引起下游水体藻类暴发的风险. 逐月 的非点源 TP 入河量贡献率为 57% ~91% (图 3), 且随着流量的增加而显著增加($R^2 = 0.98^{**}$). 因 此,非点源是长乐江 TP 污染的关键源,而夏季是 TP 污染控制的关键时期.

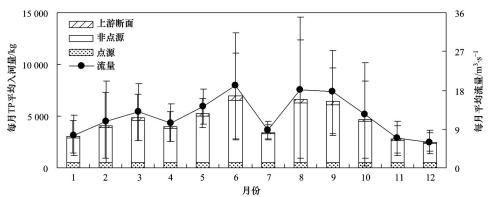


图 4 长乐江集水区点源、非点源、上游水体的 TP 每月平均入河量与河流平均流量的变化关系 Fig. 4 Monthly average TP input loads from the point source, nonpoint source and upstream inflow with changing average river water flow rate in the Changle River catchment

2.4 河流对磷的持留能力

2004~2009年,长乐江年均 TP 持留量为(4.5±0.1)t·a⁻¹,占了 TP 入河量的 9%±2%,分别与上游水体和点源的 TP 入河量相当(表1),因此,河流的 TP 持留能力不容忽视. 从每月的分布看(图 5),TP 持留量随着水温的增加而显著增加(R²=0.99**).由于河流中包括水生植物在内的水生生物活性及其生长速率和其他反应速率随着温度的增高而提高,促进了对磷的吸收利用^[18,21],使得生物吸收利用的部分磷通过食物链传递等方式而最终离开河流生态系统^[24]. 因而,河流在每年5~9月的夏季(大型水生植

物生长旺季^[18,23])具有较高的 TP 持留量(图 5),其 5 个月的累计持留量占了全年的 55% ±2%,这对缓解下游水体由养分过剩而引起的藻类暴发风险具有现实意义^[6,18]. 然而,随着流量的增加,不仅减小了单位水体体积的河床接触面积,而且减少了河流中养分参与各种物理作用几率和生物地球化学反应的时间,从而降低了单位河流养分负荷的持留率^[6,17]. 因此,河流逐月的 TP 持留率(r_i ,%)随着平均流量(Q_i , $m^3 \cdot s^{-1}$)的增加而显著降低(r_i =0.5279 $Q_i^{-0.7678}$, R^2 =0.70),且流量的变化能表达 70%的持留率变异性,这一方面表明由流量增大引起的 TP 人河量增量大于

TP 持留量增量;另一方面,与水温相比,流量是决定TP 持留率变化的主要因素,使得夏季 5~6 月和 8~9 月水温较高的丰水期(图 4)具有较低的持留率,从而仍然产生了较大的段末TP 输出负荷量.在点源贡献

率较大的低流量时期(图 3),河流具有的较高 TP 持留效率对缓解点源污染具有现实意义. 因此,河流持留能力对流域或区域磷素迁移转化过程具有重要调控作用.

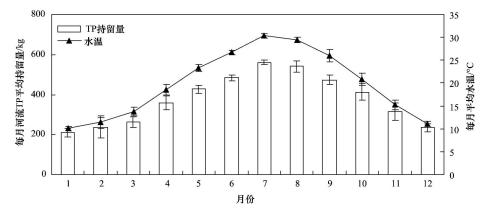


图 5 长乐江集水区每月河流 TP 平均持留量与平均水温的变化关系

Fig. 5 Monthly average in-stream TP retention load with changing average river water temperature in the Changle River catchment

3 结论

- (1)建立的二元统计模型仅需常规的河流水文水质监测数据,无需专业软件知识进行求解,且计算结果直接来源于实际的河流水文水质测算值,为研究流域或区域基础资料数据和点源排污监测不足条件下实现非点源污染过程定量、明确点源和非点源污染贡献大小、识别可能存在的点源超排、偷排等现象,提供了一种简便、实用、可靠的方法.
- (2)与现行的水文估算法相比,本模型既表达了河流磷持留能力及其时间变异性,也考虑了上游水体输入的影响,不仅深化了对流域或区域磷污染过程的定量认识,而且可以满足我国以行政区为主要水污染控制管理单元的现实需要.
- (3)长乐江集水区 TP 年入河总量为(54.6±11.9)t·a⁻¹,上游水体、点源和非点源的入河量贡献率分别为5%±1%、12%±3%和83%±3%.逐月的非点源入河量贡献率均大于57%,夏季5~6月和8~9月的非点源 TP 累计入河量占其全年的50%±9%,增加了引起下游水体藻类暴发的风险.
- (4)长乐江 TP 持留量为(4.5±0.1)t·a⁻¹,占年入河总量的9%±2%.每年夏季5~9月的 TP 累计持留量占全年的55%±2%,而点源污染贡献率较大的低流量时期具有较高的 TP 持留率.河流持留能力在流域或区域磷素迁移转化过程中具有重要调控作用.

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