

三元缔合物抑制分光光度法测定痕量氟^{*}

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摘要 利用氟离子对锆(IV)-水杨基荧光酮(SAF)-溴化十六烷基三甲基铵(CTMAB)三元缔合物形成的竞争抑制作用, 建立了测定痕量氟的抑制分光光度法, 确定了测定的最佳条件为: $[\text{SAF}] = 6.0 \times 10^{-5} \text{ mol} \cdot \text{L}^{-1}$, $[\text{CTMAB}] = 8.0 \times 10^{-4} \text{ mol} \cdot \text{L}^{-1}$, $[\text{Zr(IV)}] = 1.0 \mu\text{g} \cdot \text{ml}^{-1}$, $[\text{HCl}] = 0.08 \text{ mol} \cdot \text{L}^{-1}$, 25°C . 线性测定范围为 $0.08 - 0.48 \mu\text{g} \cdot \text{ml}^{-1}$, 方法检测限为 $2 \text{ ng} \cdot \text{ml}^{-1}$.

关键词 三元缔合物抑制, 光度法, 氟, 测定.

常见用于测定微量氟的方法主要有离子选择性电极法^[1, 2], 分光光度法^[3, 4]及离子色谱法^[5, 6]等. 在光度法中, 基于氟离子对双元配合物形成的竞争、抑制或催化反应而建立的分析法占有重要地位^[7].

本文研究了氟离子对 Zr(IV)-水杨基荧光酮(SAF)-溴化十六烷基三甲基铵(CTMAB)三元缔合物形成的竞争抑制作用^[8], 基于三元缔合物的颜色消褪建立了测定痕量氟的抑制分光光度法. 本法反应快, 重现性好, 可望在不经分离的情况下直接测定水中的氟离子含量.

1 实验部分

1.1 主要仪器及试剂

722 型分光光度计(上海第三分析仪器厂).

氟标准溶液($500 \mu\text{g} \cdot \text{ml}^{-1}$): 快速准确称取 0.5525 g 分析纯 NaF , 用适量蒸馏水溶解并定容至 500 ml , 将此溶液储于聚乙烯塑料瓶中, 使用时适当稀释.

锆标准溶液($10.0 \mu\text{g} \cdot \text{ml}^{-1}$): 将 0.0707 g 分析纯 $\text{ZrOCl}_2 \cdot 8\text{H}_2\text{O}$ 用 $2 \text{ mol} \cdot \text{L}^{-1} \text{HCl}$ 加热溶解, 并煮沸数分钟, 冷却后用 $2 \text{ mol} \cdot \text{L}^{-1} \text{HCl}$ 稀释至 100 ml , 得 $200 \mu\text{g} \cdot \text{ml}^{-1} \text{Zr(IV)}$ 溶液; 使用时用 $2 \text{ mol} \cdot \text{L}^{-1} \text{HCl}$ 稀释成 $10.0 \mu\text{g} \cdot \text{ml}^{-1}$ 的工作液.

水杨基荧光酮(SAF)溶液($1.0 \times 10^{-3} \text{ mol}$

$\cdot \text{L}^{-1}$): 将 0.336 g SAF 溶于 50 ml 乙醇及 4.0 ml HCl ($6.0 \text{ mol} \cdot \text{L}^{-1}$) 的混合液中, 用乙醇稀释至 100 ml 并避光保存.

溴化十六烷基三甲基铵(CTMAB)溶液($1.0 \times 10^{-2} \text{ mol} \cdot \text{L}^{-1}$): 称取 1.817 g CTMAB, 加水溶解并稀释至 500 ml .

HCl 溶液: $1.0 \text{ mol} \cdot \text{L}^{-1}$.

1.2 实验方法

向 25 ml 刻度比色管中, 依次加入 $10.0 \mu\text{g} \cdot \text{ml}^{-1}$ 锆标准溶液 2.5 ml , 2.0 ml HCl , 1.5 ml SAF 溶液及 2.0 ml CTMAB 溶液, 并用蒸馏水稀释至刻度, 以不加锆的体系(试剂空白)为参比测定非抑制体系吸光度; 另向一系列比色管中依次加入相同量的上述各试剂后, 再加入一定量的氟离子标准溶液, 并稀释至刻度, 以上述试剂空白作参比, 于分光光度计上在波长 515 nm 处; 用 10 mm 比色皿测定抑制体系的吸光度.

2 结果与讨论

2.1 测定波长的选择

按上述实验方法, 测定了不同波长下抑制与非抑制反应体系的吸光度差值(图 1), 显然,

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在波长 515 nm 附近吸光度之差 ΔA 数值最大, 测定灵敏度最高, 本文选择测定波长为 515 nm.

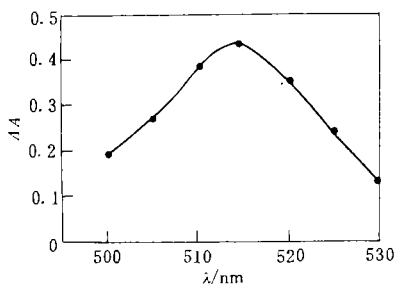


图1 吸收曲线

$$[\text{SAF}] = 8.0 \times 10^{-5} \text{ mol} \cdot \text{L}^{-1},$$

$$[\text{CTMAB}] = 8.0 \times 10^{-4} \text{ mol} \cdot \text{L}^{-1}$$

$$0.3 \mu\text{g} \cdot \text{ml}^{-1} \text{ Zr(IV)}, 0.08 \text{ mol} \cdot \text{L}^{-1} \text{ HCl}, 0.4 \mu\text{g} \cdot \text{ml}^{-1} \text{ F}^{-}$$

2.2 酸度的影响

考察了酸度对抑制-非抑制反应体系吸光度差值 ΔA 的影响, 结果表明, 在 $0.04-0.10 \text{ mol} \cdot \text{L}^{-1} \text{ HCl}$ 范围内, 体系酸度对吸光度差值的影响不显著, 本文选择实验中体系酸度为 $0.08 \text{ mol} \cdot \text{L}^{-1} \text{ HCl}$.

2.3 SAF 及 CTMAB 浓度的影响

SAF 和 CTMAB 浓度的影响如图 2 所示.

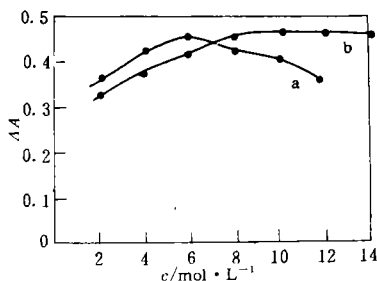


图2 SAF 和 CTMAB 浓度的影响

(a) $[\text{CTMAB}] = 8.0 \times 10^{-4} \text{ mol} \cdot \text{L}^{-1},$

$$[\text{HCl}] = 0.08 \text{ mol} \cdot \text{L}^{-1} \quad 0.3 \mu\text{g} \cdot \text{ml}^{-1} \text{ Zr(IV)},$$

$$0.4 \mu\text{g} \cdot \text{ml}^{-1} \text{ F}^{-}$$

(b) $[\text{SAF}] = 8.0 \times 10^{-5} \text{ mol} \cdot \text{L}^{-1},$

$$[\text{HCl}] = 0.08 \text{ mol} \cdot \text{L}^{-1}, 0.3 \mu\text{g} \cdot \text{ml}^{-1} \text{ Zr(IV)}, 0.4 \mu\text{g} \cdot \text{ml}^{-1} \text{ F}^{-}$$

从图 2 可以, 当 $[\text{SAF}] = 4.0-5.0 \times 10^{-5} \text{ mol} \cdot \text{L}^{-1}$ 时, ΔA 值稳定不变, 而当 $[\text{SAF}] = 1.2 \times 10^{-4} \text{ mol} \cdot \text{L}^{-1}$ 时, ΔA 值明显降低, 说明 SAF 浓度过大, 抑制了 F^{-} 与 Zr(IV) 的配位反应, 导致测定灵敏度降低. 图 2 结果表明, 当

$[\text{CTMAB}] < 6.0 \times 10^{-4} \text{ mol} \cdot \text{L}^{-1}$ 时, 吸光度差值随其浓度增大而增大; 而在 $6.0-1.4 \times 10^{-4} \text{ mol} \cdot \text{L}^{-1}$ 范围内, ΔA 值基本不随 CTMAB 的浓度而变化. 本文选择 SAF 及 CTMAB 的浓度分别为 $6.0 \times 10^{-5} \text{ mol} \cdot \text{L}^{-1}$ 和 $8.0 \times 10^{-4} \text{ mol} \cdot \text{L}^{-1}$.

2.4 Zr(IV)浓度的影响

图 3 结果表明, 随 Zr(IV) 浓度增大, 吸光度差值增大, 测定灵敏度提高, 当 Zr 含量为 $0.80-1.0 \mu\text{g} \cdot \text{ml}^{-1}$ 时, 体系吸光度之差达最大, 灵敏度最高; 继续增加 Zr(IV) 的浓度, 测定灵敏度呈下降趋势, 故本文选择 Zr 的浓度为 $1.0 \mu\text{g} \cdot \text{ml}^{-1}$.

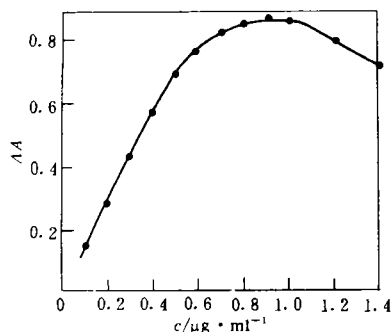


图3 Zr(IV)用量的影响

$$[\text{SAF}] = 8.0 \times 10^{-5} \text{ mol} \cdot \text{L}^{-1},$$

$$[\text{CTMAB}] = 8.0 \times 10^{-4} \text{ mol} \cdot \text{L}^{-1}$$

$$[\text{HCl}] = 0.08 \text{ mol} \cdot \text{L}^{-1}, 0.4 \mu\text{g} \cdot \text{ml}^{-1} \text{ F}^{-}$$

2.5 显色温度及显色时间的影响

本文显色体系在常温下的反应速度已足够快, 故选择显色反应在常温下进行. 另在向比色管中加入各有关试剂后, 反应立即进行并且趋于完全, 吸光度差 ΔA 趋于恒定并可稳定数小时不变. 因此, 反应时间长短对测定灵敏度基本没有影响. 本文选择在加入各试剂并充分混合后立即测定.

2.6 工作曲线

按文中所拟方法及条件, 考察了抑制及非抑制反应体系的吸光度差与氟离子浓度间的关系, 得到了下列线性回归方程:

$$\Delta A = 2.083 c_{\text{F}} (\mu\text{g} \cdot \text{ml}^{-1}) + 0.003$$

线性测定范围为 $0-0.48 \mu\text{g} \cdot \text{ml}^{-1} \text{ F}^{-}$; 按空白

测定标准偏差的 3 倍计, 方法检测限为 $2\text{ ng} \cdot \text{ml}^{-1}\text{F}^{-}$. 对含有 $0.20\text{ }\mu\text{g} \cdot \text{ml}\text{F}^{-}$ 的溶液平行测定了 11 次, 平均回收率为 96.8%; 另外, 加标回收试验表明, $0.20\text{ }\mu\text{g} \cdot \text{ml}^{-1}\text{F}^{-}$ 的加标回收率为 96%—104%.

2.7 干扰试验

试验了常见共存离子对测定 F^{-} 的干扰情况, 结果表明, 在本文条件下, 大量 K^{+} , Na^{+} , Cl^{-} , NH_4^{+} , Ca^{2+} , Mg^{2+} , Cu^{2+} , Zn^{2+} , Cr^{3+} , Cd^{2+} , Co^{2+} , Ni^{2+} 等对 F^{-} 的测定无干扰; 当 F^{-} 含量为 $0.30\text{ }\mu\text{g} \cdot \text{ml}^{-1}$ 时, 允许 2 倍量的 Fe^{3+} , VO_2^{+} 和 $\text{Cr}_2\text{O}_7^{2-}$ 存在, 超过限量时可加抗坏血酸还原消除干扰; Al^{3+} 的干扰可用蒸馏法去除; TiO^{2+} , WO_4^{2-} 和 MO_4^{2-} 对测定干扰严重, 但由于多数水样中这些离子的含量很少, 故不致于产生影响. 因此, 本法可望在不经分离的情况下, 直接测定水样中的氟离子.

2.8 样品测定

测定了土壤、蔬菜及不同水样中的氟含量. 样品处理方法如下:

土壤中可溶性氟: 将 10.0 g 过筛土样置于 250 ml 锥形瓶中, 加 20 ml 蒸馏水后于振荡器上充分振荡浸取 2.5 h , 过滤, 滤液稀释至 50 ml .

蔬菜: 将新鲜蔬菜样品用自来水和蒸馏水

反复洗涤数次并去除表面水分后, 称取适量充分搅碎, 加入适量水和 $1.0\text{ mol} \cdot \text{L}^{-1}\text{HCl}$, 在充分搅拌下浸取 2.5 h , 混合物过滤并将溶液适当稀释后待测.

水样经澄清处理后直接测定.

表 1 样品测定结果/ $\mu\text{g} \cdot \text{ml}^{-1}$

样品	测得量 ($n=5$)	加入量	回收率 /%	相对标准偏差 RSD/%
土壤样	3.743 ¹⁾	1.500	103.2	1.5
番薯叶	8.872 ¹⁾	1.500	98.1	1.2
海 水	0.440	1.000	98.5	1.7
雨 水	0.075	0.500	102.9	3.5
地表水	0.368	0.500	97.3	2.7

1) 浓度单位为 $\mu\text{g} \cdot \text{g}^{-1}$

用本法测定上述溶液中的氟含量, 并作了加入回收实验, 表 1 中给出了相应的测定结果.

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(上接第 63 页) 矿矿井废水经处理后就可达标排放, 处理 1 m^3 水的投药费用仅为 0.04 元. 目前, 鲍店煤矿对处理后的废水再经适量加药混凝微滤后回用于煤矿生产中.

3.2 用于处理油田含油废水

由于油田废水含油, 机械杂质等, 需经过

净化处理后才能回注地下. 以胜利油田现河首站废水为代表, 考察了 PAFC 用于处理油田含油污水的效果(见表 3), 表 3 表明, PAFC 的投量在 40 mg/L 以上时, 除油率及 SS 去除率都在 90% 以上, 可满足油田回注水的要求, 这说明 PAFC 是处理油田含油废水的高效净水剂.

表 3 PAFC 处理油田含油废水效果

PAFC 投量 / $\text{mg} \cdot \text{L}^{-1}$	含油量 / $\text{mg} \cdot \text{L}^{-1}$	SS / $\text{mg} \cdot \text{L}^{-1}$	浊度 /度	除油率 /%	SS 去除率 /%
0	140.0	72.4	150	0.0	0.0
20	20.5	15.3	20.6	85.4	79.4
30	15.2	8.4	12.5	89.1	88.7
40	9.1	4.5	8.8	92.5	93.9
50	5.4	2.3	7.7	96.1	96.9
60	4.3	1.8	6.9	96.9	97.6

08%, 90% and 90% respectively.

Key words: gangue, polyaluminum ferric chloride, preparation, structure, wastewater treatment.

Spectrophotometric Determination of Trace Fluoride Based on the Inhibitory Effect on the Formation of Trinary Complex. He Ronghuan and Xiu Huanhong (Chemistry Department of Yantai Teachers College, Yantai 264025); *Chin. J. Environ. Sci.*, 17(4), 1996, pp. 64–66

A kinetic method for the determination of trace fluoride was established based on the inhibitory effect of fluoride on the formation of trinary complex among Zr (IV), salicyl fluoronol (SAF) and cetyltrimethyl-ammonium bromide (CTMAB). The determination conditions were $[\text{SAF}] = 6.0 \times 10^{-5} \text{ mol} \cdot \text{L}^{-1}$, $[\text{CTMAB}] = 8.0 \times 10^{-4} \text{ mol} \cdot \text{L}^{-1}$, $[\text{Zr(IV)}] = 1.0 \mu\text{g} \cdot \text{ml}^{-1}$, $[\text{HCl}] = 0.08 \text{ mol} \cdot \text{L}^{-1}$, 25°C . The calibration graph was linear for $0.08 - 0.48 \mu\text{g} \cdot \text{ml}^{-1}$, and the detection limit was $2 \text{ ng} \cdot \text{ml}^{-1}$.

Key words: inhibitory on trinary complex formation, spectrophotometry, fluoride, determination.

Gas Chromatographic Determination of Chlorinated Pesticides and Polychlorinated Biphenyls in Sediment Using Ultrasonic Extraction and Steam Distillation Extraction (SDE). Zou Shichun and Zhang zhanxia et al. (Dept. of Chem, Zhongshan University, Guangzhou 510275); *Chin. J. Environ. Sci.*, 17(4), 1996, pp. 67–70

This paper described a new method for the extraction of trace organochlorinated pesticides (OCPs) and polychlorinated biphenyls (PCBs) in sediments using a combined technique of ultrasonic extraction (UE) with steam distillation extraction (SDE). The sediment sample was blended with water or n-hexane prior to the SDE step and the high recoveries for DDTs and PCBs, especially for PCBs, could be obtained by an auxiliary UE method. However, the recoveries for BHCs in sediment are unsatisfactory. Further experiments showed that the recoveries for β -, γ -BHC could be enhanced when a small amount of n-hexane was mixed with the sediment sample and was treated with the UE prior to the SDE step. It is possible that this combined technique is used in the extraction for semi-volatile compounds in other solid samples.

Key words: steam distillation extraction, ultrasonic bath, organochlorinated pesticides, polychlorinated biphenyls analysis.

A New PVC-Coated Carbon Rod Electrode for Anionic Surfactants and Its Application in Environmental Monitoring. Li Congrong (Chengdu Factory of Petroleum Chemistry, Chengdu 610083), Dan Dezhong et al. (Dept. of Environ. Sci. and Eng., Sichuan Union University, Chengdu 610065); *Chin. J. Environ. Sci.*, 17(4), 1996, pp. 71–72

A new anionic surfactant selective electrode prepared by coating a graphite rod with PVC compound containing triheptyl-dodecyl ammonium and dodecylbenzene sulfonate THDA-DBS is developed. The optimum coating membrane composition is THDA-DBS 5 mg, DBP 0.4 ml, NB0.3 ml and PVC 0.2 g. The electrode exhibits nernstain response to the surfactant anions over the concentration range 1×10^{-3} to $8.1 \times 10^{-7} \text{ mol/L}$ with a slope of 58 mV per decade. The detection limits are $5.6 \times 10^{-7} \text{ mol/L}$ for DBS and $7.4 \times 10^{-7} \text{ mol/L}$ for SDS. The electrode shows high stability and selectivity, and it is easy to make and store, and inexpensive. The electrode has been used successfully for the determination of anionic detergents in environmental water samples.

Key words: ion selective electrode, anionic surfactant, environmental monitoring, water sample.

Observation on Concentrations of Ammonia in Atmosphere by Diffusive Sampling. Chen Letian and Tong Yuqin (Research Center for Eco-Environmental Sciences, Chinese Academy of Sciences, Beijing 100085); *Chin. J. Environ. Sci.*, 17(4), 1996, pp. 73–74

Simple passive diffusion samplers were used for the determination of ammonia concentrations in atmosphere. The gas was collected by molecular diffusion on phosphoric-acid-impregnated paper and subsequently determined spectrophotometrically. In a survey of 3 apartments, concentrations of NH_3 indoors were higher than those outdoors. These preliminary data suggest that humans themselves may be a source of ammonia.

Key words: ammonia, diffusive sampling, environmental observation.

The Strategy, Countermeasures and Cost-Benefit Analysis of Industrial Wastewater Control in