

ENVIRONMENTAL SCIENCE

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## 西南典型碳酸盐岩高地质背景区农田重金属化学形 态、影响因素及回归模型

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摘要: 土壤重金属化学形态是决定重金属生物活性和生物毒性的重要因素, 是科学评价西南碳酸盐岩高地质背景区土壤重金 属生态风险的关键.为了探明碳酸盐岩高地质背景区土壤重金属化学形态分布情况,选择贵州省典型碳酸盐岩分布区,以第 二次全国土壤普查图斑为采样单元,在农田中采集土壤表层样品 309件,利用改进的 Tessier 七步顺序提取法,分析了 As、Cd、 Cu、Hg、Ni、Pb和Zn等7种重金属的水溶态(F1)、离子交换态(F2)、碳酸盐结合态(F3)、弱有机结合态(F4)、铁锰氧化物结 合态(F5)、强有机结合态(F6)和残渣态(F7)这7种化学形态.结果发现,土壤中重金属As、Cu、Hg、Ni、Pb和Zn残渣态比例 均超过50%,有效组分(F1~F3)比例均小于5%,潜在生物有效组分(F4~F6)比例低于45%,活性较低,生态风险不高.Cd的有 效组分和潜在生物有效组分占比分别为55.49%和29.37%,远高于其他重金属,基于土壤重金属形态的生态风险远小于基于土 壤总量的生态风险.逐步回归方程可以有效建立Cd、Cu和Pb生物有效组分与影响因素之间的关系.重金属全量和pH值是影响 碳酸盐岩高地质背景区土壤重金属化学形态的重要因子,受研究区长期土法炼锌活动和碳酸盐岩风化成土过程中重金属元素 倾向于在残渣态中富集的影响,土壤有机质(OM)和氧化物含量对土壤重金属化学形态影响相对较小.

关键词:土壤;重金属;化学形态;生态评价;影响因素;回归模型;高地质背景区

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## Chemical Speciation, Influencing Factors, and Regression Model of Heavy Metals in Farmland of Typical Carbonate Area with High Geological Background, Southwest China

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Abstract: The speciation of heavy metals in soil is an important factor determining their bioavailability and toxicity, and it is crucial for the scientific assessment of ecological risks posed by heavy metals in soils of typical carbonate areas with high geological background in southwest China. In order to investigate the distribution of speciation of heavy metals in soils of carbonate rock with high geological background, we selected a typical carbonate rock distribution area in Guizhou Province and used the second national soil survey plots as sampling units. A total of 309 topsoil samples were collected from farmland. The improved Tessier seven-step sequential extraction method was used to analyze the seven chemical forms of heavy metals: water-soluble (F1); exchangeable (F2); carbonate-bound (F3); weakly organic-bound (F4); iron-manganese oxide-bound (F5); strongly organic-bound (F6); and residual (F7) forms of arsenic (As), cadmium (Cd), copper (Cu), mercury (Hg), nickel (Ni), lead (Pb), and zinc (Zn). The study found that the residual forms of heavy metals As, Cu, Hg, Ni, Pb, and Zn in the soil accounted for more than 50%, the effective components (F1-F3) accounted for less than 5%, and the potential biological effective components (F4-F6) were less than 45%, indicating low reactivity and low ecological risk. The effective and potentially bioavailable components of Cd accounted for 55. 49% and 29. 37%, respectively, which were much higher than those of other heavy metals. The ecological risk based on the speciation of heavy metals in the soil was much lower than that based on the total content of heavy metals. The stepwise regression equations could effectively establish the relationship between the bioavailable and potentially bioavailable fractions of Cd, Cu, and Pb and their influencing factors. Total heavy metal contents and pH value were important factors influencing the speciation of heavy metals in soils of carbonate rock with high geological background areas. The enrichment of heavy metal elements in the residual fraction was influenced by long-term zinc smelling activities and the weathering of carbonate rocks into soil. Soil organic matter (OM) and oxide content had a relatively small influence on the speciation of heavy metals in the soil.

Key words: soil; heavy metals; chemical speciation; ecological assessment; controlling factors; regression model; high geological background areas

工业扩张、农业活动和矿山开采等人类活动引 起的土壤污染和粮食安全日益引起决策部门和科研 机构的重视[1~3].有研究表明,土壤中过量重金属积 累往往威胁人类健康,引发诸如发育受损、智力下 降、短期健忘症、认知障碍和心血管疾病等风 险<sup>[4,5]</sup>. 目前全球有超过1000万个污染场地,占地20 万 km<sup>2</sup>,其中 50% 以上的土地是由重金属污染引起

的[6,7]. 中国耕地也面临重金属污染的严峻考验, 在 国家层面先后实施了全国土地质量地球化学调查和

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国家土壤污染详查<sup>[8-12]</sup>,结果发现,中国每年新增 重金属污染的耕地约有1000万hm<sup>2[13,14]</sup>,耕地中度 污染以上的点位比例为2.5%,覆盖面积约2.3万 km<sup>2[15]</sup>.

上述国家尺度的土壤调查都是基于土壤重金属 全量,尽管土壤重金属的全量可以有效提供污染水 平信息,以确定阈值,制定适当的政策,然而,土 壤是由矿物、有机物、微生物和其他固体成分组成 的复杂生态系统,其数量、组成和表面特征深刻影 响着重金属的迁移转化[16,17].更为重要的是土壤重金 属化学形态才是决定重金属生物活性和生物毒性的 重要因素,等量重金属由于化学形态的不同表现出 的环境生物效应差异很大[18,19]. 目前国家尺度的调查 在土壤污染防控和土地资源精准管理方面做出了巨 大贡献,但是评价方法多以重金属全量为基础的经 验评价[8~12],如内梅罗指数法、地累积指数法和潜 在污染指数法等[20.21],也有学者用《土壤环境质量 农用地土壤污染风险管控标准》规定的筛选值和管 控值来评价生态风险<sup>[22,23]</sup>,以上方法忽视了我国幅 员辽阔,气候条件、地质背景和景观地貌区域性差 别较大的问题.如我国西南地区,重金属多是由碳 酸盐岩风化以及次生矿物形成[24.25],重金属含量很 高,但是其生物活性很低[26~29],采用总量很难科学 评价土壤重金属污染水平.因此,重金属化学形态 成为污染评价和风险预测的重要依据.长期以来, 围绕土壤酸碱度、有机质(OM)和土壤质地等理化指 标对重金属化学形态的影响开展了大量工作,研究 表明土壤重金属总量与其化学形态和生物效应间有 密切的关系<sup>[30~32]</sup>,但多数研究仅停留在定性描述, 缺乏影响因素的定量研究,且普遍存在样本量不 足,代表性不够的问题.

贵州省是我国重要的碳酸盐岩分布区,是特殊 的重金属高地质背景区<sup>[33]</sup>,重金属大多为地质成 因,生物有效性较低<sup>[34]</sup>,按照中国现行的土壤环境 质量标准,大部分耕地必须休耕或改变土地用途, 这将产生较大的社会影响和资源浪费.本研究选择 贵州省赫章县典型碳酸盐岩分布区农田表层土壤为 研究对象,分析了土壤中7种重金属化学形态,旨 在:①研究高地质背景区土壤重金属化学形态空间 分布规律;②探索影响土壤重金属形态分布的因 素;③建立高地质背景区土壤重金属生物有效组分 含量回归模型,定量描述主要影响因素的贡献.

#### 1 材料与方法

 研究区概况 研究区位于贵州省毕节市赫章县,北纬 26°46'12"~27°28'18",东经 104°10'28"~ 105°01'23",森林覆盖率高达59.03%,地形以山地 为主,碳酸盐岩分布广,属典型的喀斯特岩溶区. 气候属暖温带气候,年均气温在10~13.6℃之间,年 均降雨量介于786~1068 mm,雨量充沛,光照条件 较好,太阳辐射强度较高,季节变化不明显.全国 土地调查结果显示土地利用类型以耕地和林地为 主.研究区(图1)出露地层以碳酸盐岩、峨眉山玄武 岩和砂岩为主,309件土壤表层样品全部采自农田, 以较高的密度覆盖整个研究区和不同成土母质区.



#### 1.2 样品采集和分析

样品采集严格按照土地质量地球化学调查要求,采样点全部位于农田,采用梅花形布点法从5个点取等质量土壤混合组成1件分析样品,采样深度为0~20 cm,土壤样品去除石粒、杂草和动植物残体等非土壤成分,选择通风、整洁和无污染的场地自然风干,用橡皮锤敲打,全部过20目尼龙筛,均匀混合后四分法取样,装入乙烯样品瓶后送实验室分析.

本次测试指标为表层土壤中8种重金属(As、 Cd、Cr、Cu、Hg、Ni、Pb和Zn)的全量,采用"七 步提取法"定量分析了除Cr之外7种重金属的不同 化学形态.同时还分析了pH、Na<sub>2</sub>O、SiO<sub>2</sub>、Al<sub>2</sub>O<sub>3</sub>、 TFe<sub>2</sub>O<sub>3</sub>、CaO、K<sub>2</sub>O、MgO、OM和Mn的含量.所有 样品严格按照《生态地球化学评价样品分析技术要 求(试行)》(DD 2005-03)规范的方法测试,表层土 壤样品(全量)选用12种国家一级标准物质用选定分 析方法进行12次分析检验,pH值测定采用6种国家 一级标准物质进行分析,重金属形态样品选用4种 国家一级标准物质选定分析方法对每一个标准物质 进行 8次分析检验,分别统计各被测项目平均值与 标准值之间的对数差(ΔlgC)和相对标准偏差 (RSD%),合格率为100%.湖北省地质实验测试中 心完成了本次分析测试,抽取土壤外检样品122件, 形态外检样品51件送南京矿产资源监督检测中心进 行分析评价,合格率分别为100%和92.86%.各指标 的分析方法、检出限、准确度和精密度均符合有关 规范要求并通过了中国地质调查局质量监控中心验 收,结果真实可靠.

1.3 数据统计分析

点位设计及其成果图件利用 ArcGIS 10.2 完成, 地球化学统计参数,如最大值、最小值、中位值、 平均值、标准偏差和变异系数等利用 Excel 2020 完 成,斯皮尔曼(Spearman)相关系数和回归模型使用 SPSS 22.0(IBM, USA)软件完成.所有图件最后均通 过 CorelDRAW X8进行图形矢量处理.

1.4 土壤重金属形态生态风险评价等级

风险评价编码法(RAC)是根据元素生物有效性 含量占全量的百分比来评估环境风险<sup>[35]</sup>.本文形态 测量采用改进的Tessier七步顺序提取法,获取了重 金属水溶态(F1)、离子交换态(F2)、碳酸盐结合态 (F3)、弱有机结合态(F4)、铁锰氧化物结合态 (F5)、强有机结合态(F6)和残渣态(F7)这7种化学 形态.其中F1、F2和F3易被植物吸收,活性较大, 划分为生物有效组分;F4、F5和F6在一定环境下 易被植物吸收,划分为潜在生物有效组分;F7比较 稳定,不易被植物吸收<sup>[18,36]</sup>,为残渣态.RAC的计算 方法见下式.

RAC =  $(F1 + F2 + F3) / \omega \times 100\%$  (1) 式中,  $\omega$ 为元素全量(mg·kg<sup>-1</sup>), RAC值越高, 生物 有效性越高, 环境风险越大, RAC值越低, 生物有 效性越低, 环境风险越小.根据RAC值大小, 生态 风险划分为:极高风险(>50%)、高风险(30%~ 50%)、中等风险(10%~30%)、低风险(1%~10%)和 无风险(<1%).

#### 2 结果与分析

#### 2.1 表层土壤元素含量特征

研究区各元素的含量特征示于表 1,研究区表 层土壤中主量元素 $\omega(K_2O)$ 、 $\omega(Na_2O)$ 、 $\omega(SiO_2)$ 和  $\omega(CaO)$ 的平均值分别为 1.45%、0.55%、50.54% 和 0.98%,分别是中国土壤背景值的 0.65、0.39、0.78 和 0.45倍,为显著亏损;表层土壤中 Al<sub>2</sub>O<sub>3</sub>、TFe<sub>2</sub>O<sub>3</sub> 和 MgO 含量是中国土壤背景值的 1.33、3.06 和 1.48 倍,为相对富集;土壤 OM 平均含量为 2.65%,是中 国土壤背景值的1.47倍;研究区土壤pH值在4.33~8.09之间,平均值为6.04,多数为酸性土壤.

研究区表层土壤中8种重金属元素 $\omega$ (As)、  $\omega$ (Cd)、 $\omega$ (Cr)、 $\omega$ (Cu)、 $\omega$ (Hg)、 $\omega$ (Ni)、 $\omega$ (Pb)和  $\omega$ (Zn)的平均值分别为8.79、1.34、219.11、123.99、 0.13、80.02、32.19和173.30 mg·kg<sup>-1</sup>,分别是全国土 壤背景值的0.78、13.37、3.59、5.49、1.92、2.97、 1.24和2.34倍.除As之外,其余重金属元素均为显 著富集,特别是Cd和Cr,超过全国土壤背景值的5 倍以上.Cd、Cr、Cu和Hg的变异系数分别为69%、 46%、51%和51%,按照相关变异系数划分标准, 为高度变异,指示受人为或外来活动影响.

上述元素含量特征是两个原因引起的,一是受 水热条件和碳酸盐岩风化和成土过程的影响,可溶 性离子 Ca<sup>2+</sup>、Mg<sup>2+</sup>、Na<sup>+</sup>和 K<sup>+</sup>等从母岩中浸出淋失和 难溶或不溶物质如 Cd、Zn、Cr、As、Fe 和 Al 原地 残留或被铁锰氧化物吸收富集,反映了地质背景和 气候条件对土壤中元素含量的影响;二是本研究采 样点在农田,是人为扰动和自然环境交互影响的复 杂农田系统,受耕种过程中肥料、农药和农业用品 使用的影响,Cd、Cr 和 Cu 等农业活动标志性元素 含量升高,在土壤中高度富集,反映了农业活动对 土壤中元素含量的影响.

2.2 表层土壤重金属元素形态含量分布特征

重金属化学形态是影响重金属在农田系统中迁移转化的重要因素,研究重金属的形态可以更好地研究其环境地球化学行为和生物利用度<sup>[18]</sup>.土壤重金属元素(As、Cd、Cu、Hg、Ni、Pb和Zn)形态分布如表2所示.结果表明,土壤中重金属As(70.77%)、Cu(77.97%)、Hg(50.38%)、Ni(86.34%)、Pb(54.41%)和Zn(79.51%)多以残渣态存在,表明这些元素以原生矿物和次级硅酸盐矿物的形式在晶格中结合<sup>[40]</sup>,其化学性质比较稳定,不可能对生态系统产生影响<sup>[41]</sup>.除残渣态之外,Pb和Hg的潜在生物有效组分含量也较高,这与沉积物中的铁锰氧化物含量有关,Mn-Fe氢氧化物是金属的主要吸附剂<sup>[42]</sup>.部分重金属含量也与OM含量有关,与土壤中腐殖质物质具有很强的亲和力<sup>[43]</sup>.

生物有效组分(F1+F2+F3)被认为是评估生物利 用度和环境风险的重要因素<sup>[44]</sup>.土壤中各元素的生 物有效组分顺序如图2:Cd(55.49%)>Pb(4.88%)> Hg(4.76%)>Zn(3.28%)>Ni(1.85%)>Cu(1.01%)> As(0.67%).研究区除Cd外其他重金属生物活性均较 低,迁移能力较差.Cd的残渣态含量较低,仅为 15.14%,有效组分和潜在生物有效组分分别为 55.49%和29.37%,远远高于其他重金属,这与Cd

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Table 1 Geochemical statisticas of element concentrations in topsoil samples

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元素	最小值	最大值	平均值	中位数	标准偏差	变异系数	中国土壤背景值	贵州省土壤背景值
As	0.73	75.80	8.79	6.71	8.49	0.97	11.2	20
Cd	0.25	8.36	1.34	1.12	0.92	0.69	0.1	0.66
Cr	55.20	479.00	219.11	203.00	101.78	0.46	61	95.9
Cu	38.20	688.00	123.99	115.00	63.21	0.51	22.6	32
Hg	0.03	0.47	0.13	0.12	0.07	0.51	0.07	0.11
Ni	26.00	151.00	80.02	78.60	21.45	0.27	26.9	39.1
Pb	14.80	59.60	32.19	31.50	8.77	0.27	26	35.2
Zn	110.00	328.00	173.30	170.00	31.52	0.18	74.2	99.5
Mn	302.00	4 437.00	1 333.53	1 321.00	442.46	0.33	583	794
$K_2O$	0.36	4.15	1.45	1.25	0.60	0.41	2.24	1.88
$Na_2O$	0.05	2.90	0.55	0.19	0.64	1.15	1.4	0.1
$SiO_2$	37.44	76.69	50.54	49.98	6.00	0.12	65	66.4
$Al_2O_3$	7.38	27.36	16.64	16.73	2.73	0.16	12.5	10.88
$\mathrm{TFe_2O_3}$	5.35	20.09	12.86	12.75	2.62	0.20	4.2	6
MgO	0.44	5.48	1.93	1.26	1.33	0.69	1.3	1.18
CaO	0.14	5.24	0.98	0.80	0.69	0.70	2.16	0.87
OM	0.81	7.84	2.65	2.55	0.90	0.20	1.8	25 P
nН	4 33	8 09	6.04	5 90	1.02	0.17	67	62

1)样品数309件,氧化物和OM单位为%,pH和变异系数无量纲,其他单位均为mg·kg<sup>-1</sup>;中国土壤元素背景值SiO<sub>2</sub>引自文献[37],费州省土壤元素背景值SiO<sub>2</sub>引自文献[38],其他元素引自文献[39]





元素地球化学性质和矿业活动有关<sup>[34]</sup>.本研究结果 表明,除Cd以外,尽管Cr、Cu、Hg、Ni、Pb和Zn 均超过全国土壤背景值,含量较高,但是有效组分 均小于5%,难以被农作物吸收,生态风险较低.因 此,基于重金属全量作为生态风险评价的标准远高 于实际风险.

2.3 基于土壤重金属元素形态生态风险评价

土壤中7种重金属元素的RAC如表3所示,表 层土壤As、Cu和Ni有效态比例较低,供植物直接 吸收利用的部分较少,仅存在低生态风险等级以下 的土地;Hg、Pb和Zn仅存在中等生态风险等级以 下的土地,占比分别为10.68%、7.44%和0.65%, 比例较低,无高风险和极高风险的土地.研究区有 土法炼锌的历史,禁止几十年后,Pb和Zn的影响 已经较低,对农田影响较小,保持现状,定期监测 即可;Cd是风险最高的元素,其中极高风险、高风 险和中等风险的比例分别为65.37%、33.33%和 0.97%,需要开展风险区土壤Cd来源以及农作物安 全专项研究,加强这部分土地的监管(图3). 表3 表层土壤重金属元素生态风险等级分布<sup>1)</sup>

			Table 3 Cla	ss distribution o	of ecological r	isk for heavy m	etals in top so	il		
元素	RAC < 1%		$1\%{\leqslant}\mathrm{RAC} < 10\%$		$10\%{\leq}\mathrm{RAC}<30\%$		$30\%{\leq}\mathrm{RAC} < 50\%$		50%≤RAC	
	n	A/%	n	A/%	n	A/%	n	A/%	n	A/%
As	241	77.99	68	22.01	0	0.00	0	0.00	0	0.00
Cd	0	0.00	1	0.32	3	0.97	103	33.33	202	65.37
Cu	188	60.84	121	39.16	0	0.00	0	0.00	0	0.00
Hg	4	1.29	272	88.03	33	10.68	0	0.00	0	0.00
Ni	30	9.71	279	90.29	0	0.00	0	0.00	0	0.00
Pb	2	0.65	284	91.91	23	7.44	0	0.00	4	0.00
Zn	5	1.62	302	97.73	2	0.65	0	0.00	0	0.00

1)n表示某一风险水平的样本数,A表示某一风险级别的样本占样本总数的比例



3.1 土壤重金属存在形态的影响因素

土壤是一个复杂的生态系统,土壤重金属的形态是一个动态变化的过程,不仅受元素本身的地球化学性质的影响,还与土壤理化性质和根际环境关系密切<sup>[18.22,45]</sup>.本研究采用Spearman相关分析,研究了土壤 pH、OM、氧化物(SiO<sub>2</sub>、Al<sub>2</sub>O<sub>3</sub>、TFe<sub>2</sub>O<sub>3</sub>、MgO、CaO)和土壤全量与7种重金属不同化学形态的关系,相关分析结果如表4所示.

#### 3.1.1 土壤重金属全量对化学形态的影响

土壤中重金属各形态的含量与全量两者之间有 显著的正相关关系,自然成因的重金属主要赋存在 残渣态等相对稳定的相态中,人为成因的重金属主 要赋存在活性较大,易发生迁移的有效相态中<sup>[22]</sup>. 根据 Spearman 分析结果,研究区土壤残渣态含量与 全量呈明显的相关关系(表4),相关系数(*R*<sup>2</sup>值)范 围为 0.699~0.981,除 Cd 以外,相关系数均超过 0.834,为显著相关(图4).研究区Ni的水溶态(F1) 与总量之间有显著相关关系,其余重金属水溶态 (F1)与总量之间无明显的相关关系.Cu和Cd、Ni和 As的离子交换态(F2)与全量具有正相关关系,Cu和 Cd的离子交换态(F2)与全量相关系数均大于0.5, 呈显著正相关(表4).水溶态(F1)和离子交换态(F2) 具有较强的生物有效性,说明研究区Cu、Cd、Ni和 As这4种重金属的全量对生物有效性有较强的影响. 3.1.2 土壤理化指标对化学形态的影响

Spearman 相关分析结果显示,研究区pH值与 Pb和Zn的离子交换态(F2)具有显著负相关关系, 与As的离子交换态(F2)和碳酸盐结合态(F3)具有 显著正相关关系.pH值与Cd的各种形态关系都比 较紧密,与Cd的水溶态(F1)和离子交换态(F2)具 有显著负相关,与其它形态显著正相关(图5).说 明在酸性条件下,Pb、Zn和Cd生物活性较强,碱 性条件下,重金属生物活性较低.OM对重金属的 形态分布影响较小,与Pb弱有机结合态(F4)和铁 锰氧化物结合态(F5)具有显著正相关关系,与Zn、 Cd、As和Hg弱有机结合态(F4)具有显著正相关关系, 与Hb和Hg残渣态(F7)具有显著正相关关系,

#### 3.1.3 土壤氧化物含量对存在形态的影响

氧化物一般是影响土壤中重金属存在形态的关键因素之一<sup>[45,46]</sup>.研究区土壤中 TFe<sub>2</sub>O<sub>3</sub>含量与Cu强有机结合态(F6)和残渣态(F7)有显著相关关系,与其它重金属各形态均无较强的相关关系(图6);土壤中CaO与Cd的化学形态较为密切,与水溶态(F1)具有显著负相关,与碳酸盐结合态(F3)和强有机结合态(F6)具有显著正相关(图6).SiO<sub>2</sub>与Cu强有机结合态(F6)和残渣态(F7)具有显著负相关关系;Al<sub>2</sub>O<sub>3</sub>与As水溶态(F1)具有显著负相关;MgO与Pb的潜在生物有效组分和残渣态(F7)具有显著负相关关系,与Ni潜在生物有效组分和残渣态(F7)具有显著负相关关系,与Ni潜在生物有效组分和残渣态(F7)具有显著

表 4	土壤重金属各形态与元素含量及理化指标相关分析1)

Table 4 Correlations for chemical speciation, elements contents, and physicochemical properties

元素	形态	全量	$SiO_2$	$Al_2O_3$	$TFe_2O_3$	MgO	CaO	pН	Mn	ОМ
	F1	0.044	$0.125^{*}$	-0.466**	$-0.127^{*}$	0.322**	0.523**	0.385**	0.232**	0.062
	F2	0.522**	-0.305**	0.191**	$0.267^{**}$	-0.024	0.176**	0.085	0.155**	0.218**
	F3	0.561**	-0.324**	$0.170^{**}$	0.319**	0.185**	-0.004	-0.173**	0.105	-0.067
Cu	F4	$0.660^{**}$	-0.322**	0.229**	$0.340^{**}$	-0.228**	$0.154^{**}$	0.136*	$0.267^{**}$	$0.497^{**}$
	F5	0.773**	-0.452**	-0.022	0.426**	0.331**	0.465**	0.247**	0.443**	$0.134^{*}$
	F6	0.839**	$-0.548^{**}$	0.301**	0.516**	0.033	0.056	-0.062	$0.299^{**}$	0.225**
	F7	0.975**	-0.661**	0.395**	0.699**	-0.067	0.099	0.065	0.359**	$0.115^{*}$
	F1	-0.134*	0.244**	-0.351**	-0.227**	0.236**	0.063	-0.022	0.063	-0.026
	F2	$0.137^{*}$	0.076	0.315**	-0.105	-0.271**	-0.783**	-0.772**	-0.332**	0.074
	F3	$0.664^{**}$	0.259**	0.312**	-0.248**	-0.610**	-0.587**	-0.414**	-0.396**	$0.406^{**}$
Pb	F4	$0.845^{**}$	0.353**	$0.118^{*}$	-0.339**	$-0.589^{**}$	-0.265**	-0.096	-0.218**	0.681**
	F5	0.833**	0.077	0.053	0.011	-0.392**	$0.198^{**}$	0.352**	0.308**	0.673**
	F6	$0.578^{**}$	$0.167^{**}$	0.171**	$-0.147^{**}$	-0.566**	-0.190**	0.027	-0.085	0.413**
	F7	0.834**	0.074	0.303**	-0.061	-0.465**	-0.026	$0.114^{*}$	-0.130*	0.515**
	F1	-0.170**	0.156**	-0.147**	-0.174**	0.134*	-0.232**	-0.358**	-0.088	-0.107
	F2	-0.348**	-0.026	0.265**	0.027	-0.086	-0.795**	-0.904**	-0.266**	0.017
	F3	0.326**	$0.120^{*}$	-0.245**	-0.100	-0.009	0.369**	0.376**	0.137*	0.409**
Zn	F4	0.316**	$0.123^{*}$	0.094	-0.046	-0.416**	0.068	0.231**	-0.004	0.569**
	F5	$0.469^{**}$	-0.253**	-0.328**	0.259**	$0.583^{**}$	0.663**	0.415**	0.470**	-0.019
	F6	$0.427^{**}$	$-0.277^{**}$	-0.034	0.279**	0.234**	0.190**	0.057	0.329**	0.286**
	F7	$0.889^{**}$	-0.376**	0.153**	0.483**	-0.065	0.288**	0.369**	$0.484^{**}$	0.104
	F1	0.389**	0.269**	-0.426**	-0.260**	$0.268^{**}$	0.102	-0.028	0.006	-0.195**
1	F2	$0.278^{**}$	-0.074	0.080	0.084	$0.178^{**}$	-0.260**	-0.476**	-0.027	-0.127*
6	F3	0.512**	0.135*	-0.205**	-0.123*	$0.329^{**}$	-0.084	-0.282**	-0.194**	-0.297**
Ni	F4 7	-0.049	0.302**	0.042	-0.221**	-0.316**	-0.303**	-0.152**	-0.276**	0.165**
	F5	0.652**	-0.167**	-0.393**	0.172**	0.716**	0.470**	0.219**	0.327**	-0.346**
6	/F6	0.706**	-0.058	$-0.370^{**}$	0.007	$0.680^{**}$	0.255**	-0.011	$0.126^{*}$	-0.268**
Vg.	F7	0.969**	0.048	-0.360**	-0.035	0.638**	$0.182^{**}$	0.047	0.057	-0.446**
10	PP1 4	-0.045	0.107	0.088	-0.105	-0.073	-0.617**	-0.735**	-0.227**	0.049
- \v	F2	$0.589^{**}$	$0.223^{**}$	-0.078	-0.163**	-0.170**	$-0.274^{**}$	-0.285**	-0.004	$0.370^{**}$
- 1	F3	$0.852^{**}$	$0.158^{**}$	-0.234**	-0.123*	-0.147**	$0.544^{**}$	$0.609^{**}$	0.199**	$0.480^{**}$
Cd	F4	$0.842^{**}$	$0.202^{**}$	-0.102	-0.163**	-0.375**	$0.348^{**}$	0.481**	0.153**	$0.675^{**}$
	F5	$0.840^{**}$	0.005	-0.227**	0.029	-0.056	0.629**	$0.666^{**}$	0.432**	0.411**
	F6	0.719**	0.042	-0.354**	-0.009	$0.128^{*}$	0.737**	$0.740^{**}$	0.358**	0.186**
	F7	0.699**	-0.002	-0.099	0.119*	-0.319**	0.428**	0.576**	0.330**	0.291**
	F1	$0.124^{*}$	0.338**	-0.505**	-0.313**	$0.180^{**}$	0.415**	0.432**	0.107	0.023
	F2	0.305**	0.201**	-0.328**	-0.261**	-0.007	$0.505^{**}$	0.527**	0.079	0.258**
	F3	$0.464^{**}$	0.431**	-0.469**	-0.462**	0.023	0.499**	0.522**	-0.089	0.306**
As	F4	0.863**	0.252**	0.157**	-0.216**	-0.537**	-0.133*	$0.120^{*}$	-0.130*	0.545**
	F5	0.697**	0.521**	-0.197**	-0.473**	-0.266**	0.135*	0.286**	-0.137*	0.373**
	F6	0.431**	0.532**	-0.409**	-0.533**	-0.002	0.275**	0.304**	-0.163**	0.252**
	F7	0.981**	0.194**	0.046	-0.103	-0.466**	0.085	0.352**	0.108	0.437**
	F1	-0.066	$0.220^{**}$	-0.267**	-0.200**	0.173**	0.214**	0.212**	0.044	-0.025
	F2	0.005	0.076	0.032	-0.066	-0.059	-0.118*	-0.072	-0.076	-0.023
	F3	0.079	0.010	0.108	0.012	-0.107	-0.138*	-0.076	$-0.140^{*}$	0.087
Hg	F4	0.775**	$0.206^{**}$	0.159**	-0.203**	-0.508**	-0.283**	-0.113*	-0.106	0.567**
	F5	0.548**	-0.103	0.412**	0.133*	-0.357**	-0.369**	-0.190**	0.073	0.131*
	F6	0.625**	0.006	0.085	0.017	-0.231**	-0.025	$0.142^{*}$	0.242**	0.008
	F7	0.906**	0.128*	0.230**	$-0.112^{*}$	-0.619**	-0.088	0.099	-0.069	0.578**

1)\*\*表示在 0.01 水平(双侧)上显著相关,\*表示在 0.05 水平(双侧)上显著相关



Fig. 5 Scatter plot of chemical speciation of heavy metals and pH value in soil

影响较为复杂,与长期该地土法炼锌的活动有关. **3.2** 土壤重金属生物有效组分回归模型

土壤重金属化学形态受多种因素的影响和控制,如重金属之间的拮抗、协同作用以及上述讨论

中的土壤理化指标和土壤氧化物含量等,这些因素 之间存在多重共线性关系.因此,本研究采用逐步 回归(stepwise regression)分析方法,将土壤重金属 全量、土壤氧化物含量、pH 值和 OM 含量等指标



126	表 5 土壤重金属生物有效组分回归模型及主要影响因素 <sup>1)</sup> Table 5 Soil heavy metal bioavailable component regression model and influencing	g factors
元素	回归模型	$R^2$
Cd	生物有效组分_Cd = 0.676+0.343 Cd-0.86 pH+0.037 OM-0.011 Al <sub>2</sub> O <sub>3</sub> +0.45 K <sub>2</sub> O	0.867
Cu	生物有效组分_Cu = 1.781+0.031 Cu-0.337 TFe <sub>2</sub> O <sub>3</sub> +0.643 K <sub>2</sub> O-0.163 pH	0.564
Pb	生物有效组分_Pb = 5.528-0.742 pH+0.047 Pb-0.159 MgO-0.167 TFe <sub>2</sub> O <sub>3</sub>	0.650
Zn	生物有效组分_Zn = 8.70-1.221 pH+0.563 OM+1.248 K,O-0.79 MgO-0.011 Zn	0.471

生物有效组分\_Ni = 1.578-0.262 pH+0.268 K<sub>2</sub>O-0.105 MgO-0.015 Ni

生物有效组分\_As = 0.056+0.01 pH-0.03 Al<sub>2</sub>O<sub>3</sub>-0.003 TFe<sub>2</sub>O<sub>3</sub>+0.01 As

1)预测方程中重金属、氧化物和有机质(OM)均为全量; R<sup>2</sup>为决定系数, P为显著性

Cd生物有效组分含量预测方程预测精度较 高, R<sup>2</sup>值为0.867, P < 0.01, 自变量贡献降序排列 为: Cd > pH > OM > Al<sub>2</sub>O<sub>3</sub> > K<sub>2</sub>O, 其中主要影响变量 为 Cd 和 pH 值, 相对贡献率分别为 60.90% 和 17.04%. Pb生物有效组分含量预测方程预测精度为 可接受水平, R<sup>2</sup>值为0.650, P < 0.01, 自变量贡献 降序排列: pH > Pb > MgO > TFe<sub>2</sub>O<sub>3</sub>,其中主要影响 变量为pH、Pb和MgO,相对贡献率分别36.85%、 21.35%和 20.08%. Cu 生物有效组分含量预测方程 R<sup>2</sup> 值为0.564, P < 0.01, 自变量贡献降序排列: Cu >

Ni

As

TFe<sub>2</sub>O<sub>3</sub> > K<sub>2</sub>O > pH, 其中主要影响变量为Cu和 TFe<sub>2</sub>O<sub>3</sub>,相对贡献率分别 57.38% 和 26.24%. Zn 生物 有效组分含量预测方程R<sup>2</sup>值为0.471, P < 0.01, 自 变量贡献降序排列:  $pH > OM > K_2O > MgO > Zn, 其$ 中主要影响变量为 pH 和 OM, 相对贡献率分别 30.02% 和 20.44%. Ni 生物有效组分含量预测方程 R<sup>2</sup> 值为0.400, P < 0.01, 自变量贡献降序排列: Ni >  $pH > K_2O > MgO$ ,其中主要影响变量为Ni和pH,相 对贡献率分别 35.41% 和 30.30%. As 生物有效组分含 量预测方程R<sup>2</sup>值为0.342,精度较低.

0.400

0.342

< 0.01

< 0.01

< 0.01

< 0.01

#### 4 结论

(1)本研究清晰地说明了重金属化学形态是碳酸盐岩分布区农田生态风险评价的重要因素,基于重金属生物有效组分含量的生态风险远小于基于土壤总量的生态风险,土壤重金属生物有效组分可以更精确地表征重金属的生物活性和生物毒性,更好地服务于政府部门国土规划.

(2)本研究阐明了重金属全量是土壤重金属化 学形态的决定因素,具有显著的正相关关系,其次 pH值和氧化物含量也是影响重金属化学形态主要因 素.但本研究缺乏土壤矿物组成、根际环境和大气 环境等可能影响土壤重金属有效组分的环境因素指 标,今后随着研究深入,将会提供农田系统元素迁 移规律的成果.

(3)本研究建立的土壤重金属有效组分含量回 归方程解决了多种影响因素之间存在多重共线性关 系的问题,预测精度较高,为当前土壤金属元素有 效态提取方法众多,提取效果参差不齐,分析成本 较高等问题提供了较好的解决方案.

(4)本研究为高地质背景区生态风险评价提供 了新方法,研究结果表明仅以农田土壤重金属全量 评价土地质量状况,往往夸大了生态风险,基于土 壤重金属化学形态的生态风险评价为污染土壤修复 治理提供了新思路.

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Environmental Science (monthly)

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