

ENVIRONMENTAL SCIENCE

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# 基于特定源-风险评估模型的兰州黄河风情线绿地土 壤重金属污染优先控制源分析

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摘要:为甄别城市滨河公园景区绿地土壤重金属污染优先控制因子和污染源,以兰州市黄河风情线大景区为研究区,采集并测定 64 个绿地土壤样品重金属 As、Cd、Cr、Cu、Hg、Ni、Pb和Zn的含量;采用单因子污染指数法和污染负荷指数法量化重 金属污染程度,利用绝对因子得分-多元线性回归(APCS-MLR)模型对绿地土壤重金属进行源解析.并运用 APCS-MLR模型与 综合生态风险指数和人体健康风险评价模型相耦合方法,解析各污染源对生态风险和人体健康风险的贡献率.结果表明,除 Cr和Ni之外,As、Cd、Cu、Hg、Pb和Zn的含量均值高于兰州市土壤元素背景值,但所有元素含量均低于《土壤环境质量 建 设用地土壤污染风险管控标准》(GB 36600-2018)的筛选值.单因子污染指数结果显示,As、Cd、Cr、Cu、Ni、Pb和Zn为无污 染至轻微污染水平,而Hg属于轻度污染.污染负荷指数评价结果显示,绿地土壤总体上属于轻度污染水平.源解析表明,绿 地土壤重金属源自于交通源、自然-农业源和自然-工业源,贡献率分别为34.79%、23.12%和18.49%.特定源-综合生态风险指数结果表明,Cd和Hg为生态风险优先控制元素,自然-工业源为优先控制污染源;特定源-健康风险评价模型分析结果表明,As和Ni为人体健康优先控制元素,自然-农业源为优先控制污染源.

关键词:重金属; APCS-MLR模型; 源解析; 生态风险; 健康风险

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## Identification Priority Source of Heavy Metal Pollution in Greenspace Soils Based on Source-specific Ecological and Human Health Risk Analysis in the Yellow River Custom Tourist Line of Lanzhou

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Abstract: In order to identify the priority control factors and pollution sources for heavy metal contamination in greenspace soil from urban riverfront park areas, the Yellow River Custom Tourist Line in Lanzhou was selected as the research area. Sixty-four soil samples were collected and analyzed for the concentrations of As, Cd, Cr, Cu, Hg, Ni, Pb, and Zn. The single-factor pollution index and pollution load index were applied to assess heavy metal pollution. Then, the absolute principal component score-multiple linear regression (APCS-MLR) model was used for source apportionment of heavy metals in greenspace soils. Finally, source-specific ecological and human health risks were quantified by combining the APCS-MLR model and comprehensive ecological risk index, as well as the human health risk assessment model. The results showed that all the average values for As, Cd, Cu, Hg, Pb, and Zn concentrations, except for those of Cr and Ni, were higher than their soil background values in Lanzhou City, but all heavy metal contents were lower than the risk screening value for the Soil Environment Quality Risk Control Standard for Soil Contamination to Development Land (GB 36600-2018). The results of the single-factor pollution index revealed that As, Cd, Cr, Cu, Ni, Pb, and Zn were at levels of no contamination to low contamination, whereas Hg presented moderate contamination. The results of the pollution load index demonstrated that the overall level of heavy metal pollution in greenspace soils was mild. Source apportionment indicated that heavy metals in greenspace soils primarily originated from traffic sources, natural-agricultural sources, and natural- industrial sources, with contribution rates of 34.79%, 23.12%, and 18.49%, respectively. Source-specific ecological risk analysis showed that Cd and Hg were the priority control elements for ecological risk, and natural-agricultural sources were identified as the priority source for health risks in the study region.

Key words: heavy metals; APCS-MLR model; source apportionment; ecological risk; health risk

城市绿地是城市生态系统的重要组成部分,具 有为城市及其居民提供生态系统服务、增加游憩休 息空间、稳定生态安全格局和厚植教育文化等功 能<sup>[1-3]</sup>.绿地土壤是城市绿地建设的重要载体,也是 保障绿地土壤生物和功能多样性及促进植物、人类 和动物健康的基石<sup>[3-5]</sup>.然而,伴随着城市化进程的 快速推进,城市绿地面临的土壤环境污染问题日趋 严重,特别是铜(Cu)、锌(Zn)、镉(Cd)、铅(Pb)和 汞(Hg)等典型"城市重金属"元素已成为其主要污

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染物<sup>[3,4]</sup>.以上元素经由机动车轮胎的磨损、汽车尾 气和工业"三废"的排放及大气降尘等过程蓄积于 土壤之中,经长时间的累积从而出现不同程度的污 染现象<sup>[3,4]</sup>.由于重金属污染具有难降解性、不可逆 性、隐蔽性和生物毒性等特点,在严重影响城市绿 地植物和土壤动物及微生物健康生长的同时,也直 接或者间接地给人类健康带来潜在的威胁<sup>[1,3]</sup>.因 此,开展城市绿地土壤重金属污染状况调查、风险 评估和溯源解析研究,可为改善城市土壤环境质 量、促进生态城市建设及满足居民日益增长的美好 生活需求提供科技支撑,具有重要的理论与现实 意义.

土壤重金属优先污染源的风险管控是土壤重金 属污染精准防控的有效策略<sup>[6~8]</sup>,而重金属的溯源 分析是优先污染源确定的前提.一直以来,国内外 学者利用因子分析(factor analysis, FA)、聚类分析 (cluster analysis, CA) 和 主 成 分 分 析 (principal component analysis, PCA)等多元统计分析方法定性 识别了土壤重金属污染的主要来源类型[1,9],为推动 土壤重金属污染源头治理提供了直接的科学证据. 然而, 定性的污染源判别无法满足当前精准治污的 政策要求,具有一定的局限性.近年研究表明化学 质量平衡法(chemical mass balance, CMB)、同位素 法、主成分分析/因子分析-多元线性回归法 (principal component analysis/factor analysis-multiple linear regression, PCA/FA-MLR)、绝对因子得分-多 元线性回归法 (absolute principal component scoremultiple linear regression, APCS-MLR)和正定矩阵因 子分解法(positive matrix factorization, PMF)等受体模 型可有效量化各污染源对重金属元素的贡献 量[10~12],为土壤重金属污染精准防治和精细化管控 提供了强有力的科技支撑.然而,由于重金属污染 来源迥异,不同重金属毒性差异显著且其对生态环 境和人体健康的影响不同,进而会出现对重金属富 集贡献量大的污染源并未对生态环境和人体健康构 成污染风险的现象[6.13].为此,优先污染源的控制需 以各污染源的风险贡献量为基础,而非各污染源对 重金属总量的贡献量.近期研究发现通过源解析受 体模型与生态-人体健康风险评价模型相耦合的特定 源风险评估模型可计算出不同污染源中各重金属元 素的贡献率,从而来量化不同污染源对生态和人体 健康的风险<sup>[8,13,14]</sup>. 在有效地解决传统方法通过总量 角度来评估风险局限性的同时,还可以精确量化不 同重金属污染来源带来的潜在风险,可有效识别引 发高风险的具体污染源和需优先控制的风险源,从 而有助于污染风险的有效防控.

作为黄河流域生态保护和高质量发展的重要承 载地之一,兰州是黄河上游重要中心城市,黄河风 情线是兰州市最重要的滨水生态和文化休闲景观 带,是兰州的城市名片.然而,兰州作为一个典型 的河谷型工业城市,土壤污染现象较为严重[15~17]. 虽然已有研究对兰州黄河风情线周边绿地土壤重金 属污染状况进行了调查与评价[18],但对重金属污染 来源及其相对贡献率尚不清楚.为此,本文以兰州 黄河风情线绿地土壤为研究对象,在测定土壤中 As、Cd、Cr、Cu、Hg、Ni、Pb和Zn等元素含量的 基础上,分析重金属的污染特征;采用 APCS-MLR 模型定量解析重金属的污染源及其相对贡献,并结 合综合生态风险指数和人体健康风险评估模型来定 量解析不同污染源的生态和健康风险贡献率,从而 确定优先污染因子和污染源, 以期为兰州黄河风情 线绿地土壤重金属污染风险的精准管控和优先污染 源控制提供科学依据.

1 材料与方法

1.1 研究区概况

兰州市(35°58′~37°02′N, 102°58′~104°57′E) 位于甘肃中部黄河上游,是甘肃省省会城市,也是 黄河唯一穿城而过的省会城市.气候上属温带大陆 性半干旱气候,年平均气温在6~9℃,年平均降水 量为300mm左右,风力一般为1~4级<sup>[15]</sup>.作为中国 唯一的城市内黄河风情线,其核心段西起兰州市西 固区,东至兰州市城关区,占地面积为22.87 km<sup>2</sup>, 全长约20km.沿途建设有兰州黄河铁桥、黄河母亲 雕像、马拉松文化公园和水车博览园等游园广场、 城市主题公园和滨河健身步道,沿途建设有以乔灌 草植物、常绿草坪、常绿树和落叶树为主要绿化植 物的公共绿地,已成为中国最长的沿河开放式公园 和绿色生态长廊,是兰州市民和外来游客观景、休 闲、娱乐和健身的重要场所.近年来,随着城市工 业化和区域经济一体化进程的不断加快,工业源、 交通源和生活源排放的污染物日益增加[19].加之兰 州特殊气象条件形成的大气逆温层,阻碍了污染物 的扩散<sup>[20]</sup>,大量的重金属随大气降尘进入当地环 境,从而对生态和居民健康存在较大风险[21].

#### 1.2 样品采集与实验分析

根据黄河风情线沿途主题公园和游园广场绿地的分布情况及人流量的大小,在现场踏勘的基础上,于2020年7月连续7d以上晴朗无风的天气之后,采用梅花形布点法采集周边绿地土壤样品,每个土壤样品是由相距500m左右的5个挖取面积为25cm×25cm,深度为0~20cm的表层土壤样品混合

采集样品 64个,采样过程中严格遵照《土壤环境监测技术规范》(HJ/T 166-2004)的相关要求进行规范操作,避免相互交叉污染.



Fig. 1 Distribution of sampling sites for greenspace soils in the Yellow River Custom Tourist Line in Lanzhou

将采集的绿地土壤样品带回实验室, 平铺于干 净的纸上,掰成碎块并摊成厚约2cm的薄层,置于 阴凉通风处进行自然风干.之后弃去植物残渣、根 系和碎石等杂物,并用木棍压碎,过孔径为0.149 mm的尼龙筛,用于土壤样品重金属总量的测定.土 样的测定在中国科学院长春应用化学研究所完成. 经氢氟酸-硝酸-高氯酸微波消解法消解之后,使用 电感耦合等离子体质谱仪(Thermo X Series 2)测定样 品中的Cd、Cr、Cu、Ni、Pb和Zn含量,检出限分 别为: 0.02、0.4、0.2、1.0、2.0和2.0 μg·g<sup>-1</sup>. 在测定 土壤样品中As和Hg含量之前,先经硝酸-盐酸-水 溶液体系消解,后用原子荧光光谱仪(XGY-1011A) 测定,检出限分别为: 0.2 µg·g<sup>-1</sup>和 0.005 µg·g<sup>-1</sup>.为 确保数据的准确性,实验测定以20%的平行样、空 白样及土壤成分分析标准物质(GBW 07449)进行质 量控制,分析误差在5%以内.所有元素的回收率均 在90%~102%.

1.3 重金属污染评价方法

1.3.1 单因子污染指数法

单因子污染指数法(single factor pollution index, *P<sub>i</sub>*)是一种用于定量评价研究区土壤中单个重金属污染状况的方法<sup>[22]</sup>,其计算公式为:

$$P_i = C_i / S_i \tag{1}$$

式中, $C_i$ 为重金属i的实测含量, $mg \cdot kg^{-1}$ ; $S_i$ 为元素 i的兰州市土壤环境背景值<sup>[23]</sup>, $mg \cdot kg^{-1}$ . $P_i$ 的重金属 污染程度分级标准如表1所示.

1.3.2 污染负荷指数法

污染负荷指数法(pollution load index, PLI)是一 种用于定量评价研究区土壤多种重金属综合污染程 度的方法<sup>[24]</sup>,其计算公式为:

$$PLI = \sqrt[n]{P_1 \times P_2 \times \dots \times P_n}$$
(2)

式中,PLI为研究点位的污染物负荷指数,n为重金 属元素的数量.PLI的重金属污染程度分级标准如表 1所示.

1.4 重金属特定源-风险评价模型

重金属特定源-风险评价模型是在绝对因子得 分-多元线性回归(APCS-MLR)模型识别重金属污染 来源并定量分析污染源贡献率的基础上,结合综合 生态风险评价指数法和人体健康风险评价模型定量 解析各污染源对生态风险和人体健康风险的贡献率.

### 1.4.1 APCS-MLR模型

APCS-MLR模型是一种主成分分析法与多元线 性回归法相结合来定量解析污染物来源的受体模 型<sup>[25]</sup>,即在通过主成分分析(PCA)获取绝对主因子 得分(APCS)的基础上,再分别以APCS和各重金属 含量作为自变量和因变量进行多元线性回归分析, 之后利用回归系数计算各个因子所对应的污染源贡 献率<sup>[25]</sup>.具体计算步骤如下.

首先,利用公式(3)对重金属含量数据标准化:

$$X_{ij} = \frac{C_{ij} - C}{\sigma_i} \tag{3}$$

式中, $X_{ij}$ 为重金属含量标准化值,无量纲; $C_{ij}$ 为第j个样品中重金属i的含量, mg·kg<sup>-1</sup>; $\bar{C}$ 为重金属i的含量均值, mg·kg<sup>-1</sup>; $\sigma_i$ 为重金属i的标准偏差.

接着,引入含量为0的人工样本,计算各重金属0含量样本的因子分数,计算公式为:

$$X_{0i} = \frac{0 - \overline{C}_i}{\sigma_i} = -\frac{\overline{C}_i}{\sigma_i}$$
(4)

最后,将每个样品的因子得分与含量为0的人 工样本作减法,得到每个样本的APCS,并分别以 APCS和各重金属含量作为自变量和因变量进行多元 线性回归分析,根据公式计算源贡献率:

$$C_i = b_{j0} + \sum_{k=1}^{n} (b_{jk} \times APCS_{jk})$$
 (5)

式中, $b_{0}$ 为重金属j的多元线性回归常数; $b_{k}$ 为污染 源k对重金属j的回归系数, $b_{k}$ × APCS<sub>k</sub>为污染源k对  $C_{i}$ 的贡献率,%.

此外,需要注意的是在 APCS-MLR 模型计算过 程中,为防止贡献率可能会出现负值,从而影响污 染源分配的准确性,需采用绝对值计算污染源的贡 献率<sup>[26]</sup>.

#### 1.4.2 特定源-综合生态风险评价模型

特定源-综合生态风险评价模型是在 APCS-MLR 模型溯源结果的基础上,结合综合生态风险指数 (Nemerow integrated risk index, NIRI)法<sup>[14]</sup>定量解析 各污染源的生态风险.具体计算公式如下:

$$\operatorname{EIRI}_{ij}^{k} = \sqrt{\frac{(\operatorname{ER}_{ij\max}^{k})^{2} + (\operatorname{ER}_{ij\operatorname{average}}^{k})^{2}}{2}} \tag{6}$$

式中, EIRI<sup>*k*</sup><sub>*ij*</sub>为样品*i*中污染源*k*对多种元素的生态风险, ER<sup>*k*</sup><sub>*ijmax*</sub>和ER<sup>*k*</sup><sub>*jjavage*</sub>分别是同一样品中所有元素生态风险值的最大值和平均值.ER<sup>*k*</sup><sub>*ij*</sub>的计算公式如下所示:

$$\mathrm{ER}_{ij}^{k} = \frac{C_{ij}^{k}}{B_{i}} \times T_{r}^{i} \tag{7}$$

式中,  $ER_{ij}^{k}$ 为样品 i中污染源 k对元素 j的生态风险;  $B_{i}$ 为兰州市土壤元素背景值<sup>[23]</sup>;  $T_{r}^{i}$ 为重金属的毒性 响应系数, 无量纲, 本研究中 As、Cd、Cr、Cu、 Hg、Ni、Pb和 Zn的毒性响应系数别取值 10、30、 2、5、40、5、5和1<sup>[27]</sup>.  $C_{ij}^{k}$ 为样品 i中污染源 k对元素 j的质量贡献, mg·kg<sup>-1</sup>, 其计算公式为:

$$C_{ij}^{k} = C_{ij}^{k^*} \times C_i \tag{8}$$

式中, $C_{ij}^{k}$ 为样品i中污染源k对元素j的贡献值,无量纲, $C_i$ 为样品i中重金属的实测值, mg·kg<sup>-1</sup>.

基于 ER<sup>k</sup><sub>ij</sub>和 EIRI<sup>k</sup><sub>ij</sub>值的生态风险等级划分标准如表1 所示<sup>[14,27]</sup>.

	,	_		,	2 <b>1</b> · <b>1</b>		
		表1	Ⅰ 兰州黄河风情线绿地	!土壤重金属污	染和生态风险分级标准	/	~nF
	Tab	le 1 Classific	ation division standard of	heavy metal pol	llution and ecological risk i	n greenspace	812
			soils from the Yellow Ri	iver Custom Tou	rist Line in Lanzhou		
评价方法	数值范围	污染程度	评价方法 数值范围	污染程度	评价方法	数值范围	污染程度
	<1	无污染	( /<1	无污染	1521	<40	轻微风险
	1~2	轻微污染	1~2	轻度污染	I'V à	40~80	中等风险
$P_i$	2-3	轻度污染	PLI 2~3	中度污染	$\mathrm{ER}^k_{ij}$ 或 $\mathrm{EIRI}^k_{ij}$	80~160	较强风险
11	3~5	中度污染	3	重度污染	110	160~320	很强风险
31	>>5	重度污染	Solv P	6	10/23	>320	极强风险
	/		<b>¬</b> <i>µ</i> / <i>µ</i> · / <i>µ</i>		Z 08.1 1 3674		

1.4.3 特定源-健康风险评价模型

特定源-健康风险评价模型是基于 APCS-MLR 模型的溯源结果,利用美国环保署的人体健康风 险模型来量化不同污染来源对人体健康风险贡献 量的评价方法<sup>[28]</sup>.各污染源的健康风险计算公式 如下:

$$ADD_{ijing}^{k} = \frac{C_{ij}^{k} \times IngR \times ED \times EF}{BW \times AT} \times CF \qquad (9)$$

$$ADD_{ijinh}^{k} = \frac{C_{ij}^{k} \times InhR \times EF \times ED}{PEF \times BW \times AT}$$
(10)

$$ADD_{ijdermal}^{k} = \frac{C_{ij}^{k} \times SA \times AF \times ABS \times ED \times EF}{BW \times AF} \times CF$$

(11)

式中,  $ADD_{ijing}^{k}$ 、 $ADD_{ijinh}^{k}$ 和  $ADD_{ijdermal}^{k}$ 为在污染源 k 中 样品 i中j元素分别在手口摄入、呼吸吸入和皮肤接 触途径下的日均暴露量, mg·(kg·d)<sup>-1</sup>;  $C_{ij}^{k}$ 为样品 i中污染源 k对元素j的质量贡献, mg·kg<sup>-1</sup>. 其他参数 的定义与取值见表 2<sup>[29,30]</sup>.

特定源-非致癌健康风险计算公式如下所示:

$$\mathrm{HI}_{ij}^{k} = \sum \mathrm{HQ}_{i,j}^{k} = \sum \frac{\mathrm{ADD}_{ij,p}^{k}}{\mathrm{RfD}_{p}}$$
(12)

式中, HQ<sup>k</sup><sub>i</sub>和HI<sup>k</sup><sub>i</sub>分别为样品i中元素j在污染源k中

不同接触途径下的非致癌健康风险和总非致癌风 险,p为暴露途径,RfD为参考剂量值,mg·(kg·d)<sup>-1</sup>. 当 $HQ_{ij}^{k}$ 或 $HI_{ij}^{k} \le 1$ 时,认为重金属污染物不存在非致 癌风险;当 $HQ_{ij}^{k}$ 或 $HI_{ij}^{k} > 1$ 时,认为污染物存在非致 癌风险<sup>[31]</sup>.

特定源-致癌健康风险计算公式如下所示[29]:

$$CR_{ij}^{k} = \sum CR_{ij,p}^{k} = \sum ADD_{ij,p}^{k} \times SF_{i,p}$$
(13)

$$\mathrm{TCR}_{ij}^{k} = \sum \mathrm{CR}_{ij}^{k} \tag{14}$$

式中,  $CR_{ij}^{k}$ 为样品 *i* 中来自污染源 *k* 对元素 *j* 的致癌 风险值,  $TCR_{ij}^{k}$ 为总致癌风险值, SF 为致癌斜率因 子, mg·(kg·d)<sup>-1</sup>. 当  $CR_{ij}^{k}$ 或  $TCR_{ij}^{k} \leq 10^{-6}$ 时, 认为不 存在致癌风险; 当  $10^{-6} < CR_{ij}^{k}$ 或  $TCR_{ij}^{k} \leq 10^{-4}$ 时, 认 为具有潜在致癌风险; 当  $CR_{ij}^{k}$ 或  $TCR_{ij}^{k} > 10^{-4}$ 时, 认 认 为 存 在 致 癌 风 险<sup>[31]</sup>. RfD 和 SF 的 取 值 见 表 3<sup>[18,32,33]</sup>.

#### 1.5 数据分析方法

采用 SPSS 19.0 和 Microsoft Excel 2019 对样本数 据进行描述性统计分析,并利用 SPSS 19.0 软件进 行 PCA/APCS 受体模型拟合分析,运用 Origin 2020b 和 ArcGIS 10.7 进行图件的绘制、编辑与处理. 表2 健康风险评价模型的暴露参数及其参考值

	Table 2 Reference values for exposure p	parameter and its definition of hea	lth risk assessment model	
全粉	<b>今</b> 21	24 12-	数	:值
参奴	含义	- 単位	成人	儿童
IngR	手口摄入频率	$mg \cdot d^{-1}$	100	200
InhR	呼吸吸入频率	$m^3 \cdot d^{-1}$	14.5	7.5
ED	暴露年限	а	24	6
EF	暴露频率	$d \cdot a^{-1}$	180	180
BW	平均体重	kg	56.8	15.9
AT(致癌)	平均暴露时间(致癌)	d	ED×365	ED×365
AT(非致癌)	平均暴露时间(非致癌)	d	70×365	70×365
ABS	皮肤吸收因子	无量纲	0.01	0.01
SA	皮肤暴露表面积	$\mathrm{cm}^2$	5 075	2 448
AF	皮肤黏着度	$mg \cdot (cm^2 \cdot d)^{-1}$	0.07	0.2
CF	单位转换因子	无量纲	1×10 <sup>-6</sup>	$1 \times 10^{-6}$
PEF	悬浮颗粒沉降因子	$m^3 \cdot kg^{-1}$	1.36×10 <sup>9</sup>	1.36×10 <sup>9</sup>

表3 重	重金属不同	同暴露途径	的参考剂量(	RfD)和斜率	≤系数(SF)
------	-------	-------	--------	---------	---------

Table 3 Reference doses (RfD) and slope coefficient (SF) of different exposure pathways for heavy metals

壬人民		$RfD/mg \cdot (kg \cdot d)^{-1}$			SF/mg•(kg•d	)-1
里金周	手口摄入	呼吸吸入	皮肤接触	手口摄入	呼吸吸入	皮肤接触
As	3×10 <sup>-4</sup>	3×10 <sup>-4</sup>	1.23×10 <sup>-4</sup>	1.50	1.51×10	3.66
Cd	$1 \times 10^{-3}$	1×10 <sup>-3</sup>	1×10 <sup>-5</sup>	3.8×10 <sup>-1</sup>	6.3	3.8×10 <sup>-1</sup>
Cr	3×10 <sup>-3</sup>	2.86×10 <sup>-5</sup>	6×10 <sup>-5</sup>	5.01×10 <sup>-1</sup>	4.2×10	
Cu	$4 \times 10^{-2}$	4.02×10 <sup>-2</sup>	$1.2 \times 10^{-2}$	1. 20	- /	WZZ/
Hg	3×10 <sup>-4</sup>	8.57×10 <sup>-5</sup>	2.10×10 <sup>-5</sup>	10 - Q V.	1 -	B)
Ni	$2 \times 10^{-2}$	2.06×10 <sup>-2</sup>	5.4×10 <sup>-3</sup>	VI City d	9.01×10 <sup>-1</sup>	4.25×10
Pb	$3.5 \times 10^{-3}$	3.52×10 <sup>-3</sup>	5.25×10 <sup>-4</sup>	1.1+1	—	C-0
Zn	3×10 <sup>-1</sup>	3×10 <sup>-1</sup>	6×10 <sup>-2</sup>	(6)-2)	—	A 3 /
1)"一"表示	无相关数据,下同	11/10	15	154		S

**2.1** 兰州黄河风情线绿地土壤重金属含量描述性统计

兰州黄河风情线绿地土壤重金属测定结果如表 4所示.从中可知,重金属含量平均值由高到低依次 为:Zn>Cr>Ni>Pb>Cu>As>Cd>Hg.其中Hg、Cd、Zn、 Pb、Cu和As含量平均值分别是兰州市土壤元素背 景值<sup>[23]</sup>的2.37、1.48、1.40、1.22、1.14和1.06倍, 表明绿地土壤存在一定程度的As、Cu、Zn、Cd、 Hg和Pb的富集超标.然而,研究区所有绿地土壤样 点的重金属含量均低于《土壤环境质量建设用地土 壤污染风险管控标准(试行)》(GB 36600-2018)中第 二类建设用地(绿地与广场用地)的风险筛选值<sup>[34]</sup>, 说明绿地土壤质量健康安全,为安全无污染风险

重金属变异系数(coefficient of variance, CV)是 一种量化研究区域样点重金属含量变异特征和离散 程度的指标.重金属CV值越大,说明重金属含量受 人类活动影响越显著<sup>[35]</sup>.根据CV值的大小,分为低 度变异(<15%)、中度变异(15%<CV<36%)和高度变 异(>36%)<sup>[35]</sup>.由表4可知,研究区绿地土壤重金属 的CV值由大到小依次为:Hg>Cd>Zn>Pb>Cu>Ni>As >Cr.其中As和Cr为低度变异,Cu、Ni、Pb和Zn属 于中度变异,Cd和Hg表现为高度变异特征(表2), 说明除了As、Cr和Ni含量在研究区绿地土壤分布较 为均匀之外,其余元素均受到了人类活动的干扰, 特别是Cd和Hg元素的空间含量差异显著,这或与 不同绿地土壤样点所处地理位置的人类活动影响强 度差异有关.

此外,与兰州市大气降尘重金属含量均值相 比<sup>[21,36]</sup>(表5),研究区绿地土壤重金属含量均值 均表现为较低的特征,这可能与黄河风情线种植 的不同植物种群对道路扬尘和大气降尘所携带重 金属具有缓冲和滞留等作用有关<sup>[18]</sup>.与同为国内 西北地区的银川市<sup>[37]</sup>和阿克苏市<sup>[38]</sup>的绿地土壤研 究相比,研究区Cr含量均值相对较高,除As外, 其余重金属含量均值也要高于银川市,说明兰州 市工业活动和交通排放等污染相比于银川市和阿 克苏市较为严重,另一方面也可能与当地逆温现

表4 兰州黄河风情线绿地土壤重金属描述性统计结果

	Table 4 Descriptive statistics of heavy	metals in g	reenspace so	ils from the	Yellow Rive	r Custom Tou	rist Line in	Lanzhou	
项目	类型	As	Cd	Cr	Cu	Hg	Ni	Pb	Zn
	最大值/mg·kg <sup>-1</sup>	14.37	0.88	81.88	40.77	0.43	38.47	53.13	169.95
纽山上婶	最小值/mg·kg <sup>-1</sup>	8.23	0.02	42.43	15.87	0.01	11.72	17.18	38.13
继地工場	平均值/mg·kg <sup>-1</sup>	11.10	0.26	59.69	24.89	0.07	28.80	26.38	77.78
	变异系数/%	13.43	53.91	11.90	19.29	103.06	16.73	26.49	31.81
会老店	兰州市土壤元素背景值 <sup>[23]</sup> /mg·kg <sup>-1</sup>	10.46	0.175	63.85	21.92	0.028	30.34	21.7	55.73
<b>爹</b> ′写'阻	第二类建设用地风险筛选值 <sup>[34]</sup> /mg·kg <sup>-1</sup>	60	65	—	18 000	38	900	800	—

象对污染物的抑制扩散有关<sup>[15]</sup>. 与其他地区同类 研究相比,研究区Zn的含量均值高于南京市绿 地土壤<sup>[4]</sup>, As、Cd和Ni的含量均值也高于北京 市绿地土壤<sup>[39]</sup>,然而,相较于齐齐哈尔市<sup>[40]</sup>和 上海市<sup>[41]</sup>的绿地土壤重金属含量均值,研究区重 金属的含量均值均相对较低.这可能与不同区域 的城市产业结构差异和土壤背景值的地域性 有关<sup>[42]</sup>.

#### 表 5 兰州黄河风情线绿地土壤重金属和兰州市大气降尘和国内其他城市绿地土壤重金属的对比/mg·kg<sup>-1</sup>

Table 5 Comparison of heavy metals in greenspace soils in the Yellow River Custom Tourist Line and atmospheric

dust in Lanzhou and in greenspace soils of other cities in China/mg·kg<sup>-1</sup>

				,	, I			0 0			
项目	城市	区域	As	Cd	Cr	Cu	Hg	Ni	Pb	Zn	文献
十左欧小	兰州市	西北		2.85	43.96	32.37	—	33.61	133.54	406.32	[36]
人乀陲主	兰州市	西北	—	3.89	95.61	84.25	-	43.13	133.78	379.84	[21]
	兰州市	西北	11.5	0.33	61.00	26.45	0.09	30.39	28.33	91.53	本研究
	北京市	华北	7.98	0.25 🌽	25	—	3.76	25.11	38.42	-//	[39]
	上海市	东部	- 0	0.4	62.6	46.6	+¢	33.0	53.1	240.0	[41]
绿地土壤	南京市	东部	-//	0.62	87.29	36.89	(- V	- 1	46.30	80.37	[4]
$\sim$	银川市	西北	11.69	0.13	57.12	19.61	0.04	24.11	22.22	51.54	[37]
11	阿克苏市	西北	-\ 0	MH &	33.58	89.66	"+ "i	1-2	79.29	106.12	[38]
91	齐齐哈尔市	东北	57.96	SON	267.74	151.38	0.61	5-0	156.5	103.92	[40]

兰州黄河风情线绿地土壤重金属污染评价 2.2兰州黄河风情线绿地土壤 Pi和 PLI 评价结果显 示,8种重金属的Pi平均值由大到小依次为:Hg (2.37) > Cd (1.47) > Zn (1.40) > Pb (1.25) > Cu(1.14) > As(1.06) > Ni(0.95) > Cr(0.93), 见图 2. 由  $P_i$ 平均值可知, Hg为轻度污染(2 <  $P_i$  ≤ 3), Cd、Zn、 Pb、Cu和As属于轻微污染( $1 < P_i \leq 2$ ), Ni和Cr为无 污染(P<sub>i</sub>≤1).具体而言,绿地土壤中Hg污染最为广 泛, 分别有占比为 4.69%、 20.31%、 20.31% 和 20.31%的样点处于重度、中度、轻度和轻微污染状 态,仍有34.38%的样点为无污染;其次是处于无-重度污染水平的Cd,分别有占比为1.56%、1.56%、 18.75% 和 46.88% 的样点处于重度、中度、轻度和轻 微污染状态,还有31.25%的样点为无污染(图2); 接着是Zn,整体为无-中度污染,分别有1.56%、 3.13%、81.25%和14.06%处于中度、轻度、轻微和 无污染状态;再者为无-轻度污染的Pb(图2),分别 有 3.13% 和 71.88% 的样点为轻度和轻微污染,还有 25.00%的样点为无污染; Cu、As、Ni和Cr均属于 无-轻微污染(图2),其中分别有占比为70.31%、 71.88%、40.63%和20.31%的样点为轻微污染.

研究区绿地土壤重金属的PLI值介于0.69~2.20 之间,平均值为1.20,总体上为轻度污染(图2).其 中处于中度、轻度和无污染的样点占比分别为 28.13%、71.88%和28.12%,说明研究区绝大部分绿 地土壤为轻度污染.

2.3 兰州黄河风情线绿地土壤重金属污染源解析

采用 APCS-MLR 模型对研究区绿地土壤重金属 污染进行定量源解析,将原始数据导入 SPSS 26.0软 件进行标准化处理后,进行 Kaiser-Meyer-Olkin 值 (KMO = 0.831)和 Bartlett's球体检验(P<0.001),结果 表明该数据适用于 PCA 分析.接着以主成分特征值 大于1和前 n个主成分累计解释总方差大于76%的 原则共提取了3个主成分因子,分别解释了总方差 的54.00%、12.12%和10.28%(表 6).接着将 PCA 分 析的因子得分转化为绝对主因子得分(APCS),再将 各 APCS 与各重金属含量做多元线性回归,可得 As、 Cd、Cr、Cu、Hg、Ni、Pb和 Zn 的拟合度 R<sup>2</sup>分别为 0.776、0.834、0.805、0.678、0.562、0.903、0.863 和 0.690,除 Hg 的拟合度较低外,其余元素均大于 0.65, 且各重金属元素预测含量与实测含量比值均 接近于1,说明多元线性回归方程的拟合效果较好,





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表 6 黄河风情线绿地土壤重金属含量因子分析	〒结 果
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Table 6 Results of heavy metal factor analysis for greenspace soils from the Yellow River Custom Tourist Line in Lanzhou

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项目		PCA1	PCA2	PCA3	APCS1	APCS2	APCS3
$\cap$	As	0.783	0.031	0.403	0.329	0.726	0.374
919	Cd	0.854	-0.137	-0.294	0.853	0.272	0.179
	Cr	0.45	0.769	0.105	0.078	0.082	0.89
EAR	Cu	0.803	-0.176	0.038	0.621	0.517	0.157
里金周	Hg	0.602	0.447	-0.008	0.342	0.178	0.643
23 17	R	0.681	-0.319	0.58	0.232	0.921	0.032
( P	РЬ	0.839	-0.055	-0.396	0.884	0.159	0.238
	Zn	0.774	-0.152	-0.26	0.777	0.261	0.137
特征值		4.32	0.969	0.822	2.783	1.849	1.479
方差贡献率/%		54.005	12.117	10.277	34.79	23.118	18.49
累计贡献率/%		54.005	66.122	76.399	34.79	57.909	76.399

APCS-MLR模型的源解析结果整体较好.绿地土壤重金属的污染源贡献率图谱见图3.

APCS1 中 Pb、Cd、Zn 和 Cu 的载荷较高,贡献 率分别为 88.4%、85.3%、77.7% 和 62.1% [图 3(a)]. 其中 Cd 属于高度变异,Pb、Cu 和 Zn 为中度变异, 且前述分析表明绿地土壤存在一定程度的 Cu、Zn、 Cd 和 Pb 的富集超标,说明 APCS1 可能受人类活动 影响较大.有研究表明,机动车引擎、刹车片、镀 锌部件和轮胎的磨损与腐蚀以及燃料燃烧和汽车尾 气的排放等交通活动会向周边环境释放 Cd、Pb、Cu 和 Zn 元素<sup>[43,44]</sup>.兰州黄河风情线坐落于交通动脉的 南北滨河路旁,交通繁忙、车流如织,部分核心景 区和景点处在交通瓶颈地段,随着兰州市机动车保 有量的快速增加,交通流量、道路扬尘密度和尾气 排放量也随之剧增,道路交通活动产生的重金属颗粒经大气沉降和空气粉尘吸附作用,最终富集于周边绿地土壤中.因此,APCS1受道路交通的影响,可视为交通源.

APCS2载荷较高的元素为Ni、As和Cu,贡献率 分别为92.1%、72.6%和51.7%[图3(a)].其中Cu属 于中度变异,As和Ni为低度变异,Ni的含量均值略 低于兰州市背景值,而As含量均值却略高于兰州市 背景值(表4).说明APCS2受到了一定的人类活动影 响.有研究表明Ni属于铁族元素,易与土壤中氧化 物结合,其来源与成土母质关联密切<sup>[43]</sup>.而农田土 壤中杀虫剂和除草剂制品的使用、氮肥和磷肥等复 合肥及有机肥的施用会引起As和Ni含量的增加<sup>[45]</sup>. 兰州黄河风情线是由滨河路绿色长廊沿线景点串联 而成的滨河风景区,景区管委会不定期地对沿线绿地植物进行喷施农药和施肥等绿化养护管理工作.虽然景区使用低毒、低残留农药和生物农药进行养护工作,但长期的施肥和杀虫除草剂的使用势必会引起绿地土壤中As含量的累积.因此,APCS2是自然-农业源.

APCS3的主要载荷元素为Cr和Hg,贡献率为 89.0%和64.3%[图3(a)].Cr的含量均值低于兰州市 背景值,且Cr为低度变异,而Hg属于高度变异, 表明APCS3受到了一定的人类活动影响.有研究发 现土壤中Cr元素主要受成土母质和地球化学作用的 影响,且其分布多与自然因素有关<sup>[43]</sup>.兰州市耕地 土壤重金属溯源研究也认为兰州土壤Cr元素源自成 土母质<sup>[46]</sup>. 而土壤中 Hg 的累积与燃煤排放、有色 金属冶炼和水泥生产等工业活动有关<sup>[43,47]</sup>. 研究 区位于黄河上游最大的工业城市——兰州,石油、 机械、电力和冶金等工业为兰州市的支柱产业, 以上工业活动产生的烟尘、粉尘和废气等也会通 过风力作用或雨水径流向周边环境扩散、沉降并 最终蓄积于土壤中.除此之外,位于兰州以西的 河西走廊地区的矿产开发、金属冶炼和燃煤发电 等工业活动丰富而密集,加之河西走廊地区常年 以西北风为主,工业活动产生的粉尘颗粒进入大 气,在风力搬运作用下,进行长距离传输,最终 通过大气干湿沉降进入土壤<sup>[18]</sup>. 因此, APCS3 代表 了自然-工业源.





综上所述,交通源对研究区绿地土壤重金属累 计的贡献率最高,为34.79%;其次为自然-农业源, 贡献率为23.12%;自然-工业源的贡献率最小,为 18.49%[图3(b)].

**2.4** 兰州黄河风情线绿地土壤重金属生态风险与特定源-综合生态风险评价

研究区绿地土壤重金属单项生态风险指数(*E*<sub>r</sub>) 和综合生态风险指数(NIRI)评价结果显示,*E*<sub>r</sub>的平 均值由高到低依次为:Hg(94.61)>Cd(44.18)>As (10.61)>Pb(6.08)>Cu(5.68)>Ni(4.75)>Cr(1.87)>Zn (1.40)[图4(a)],除Cd和Hg分别存在中等和较强 生态风险之外,其余重金属均为轻微生态风险.此 外,绿地土壤重金属的NIRI值为16.18~433.89[图4 (a)],说明存在轻微至极高的生态风险,平均值为 69.60,为中等生态风险.

基于特定源-综合生态风险评价模型的结果如图 4(b)所示.绿地土壤重金属的3种污染源对研究区综 合生态风险的贡献率由高到低依次为:自然-工业源 (40.39%)>交通源(38.66%)>自然-农业源(20.94%). 由此可见,自然-工业源对生态风险的贡献率最高, 这与APCS-MLR源解析结果存在差异.由APCS-MLR 分析结果可知,具有高载荷 Cr和Hg元素的自然-工 业源是对重金属贡献最低的污染源,但却是对综合 生态风险贡献最高的污染源,其次为具有高载荷 Pb、 Cd、Zn和Cu元素的交通源.这既与Cd和Hg是研究 区绿地土壤主要的潜在生态危害元素有关,也与重



图 4 兰州黄河风情线绿地土壤重金属生态风险评价

Fig. 4 Ecological risk assessment of heavy metals from greenspace soils in the Yellow River Custom Tourist Line in Lanzhou

金属元素的毒性高低有关.由不同污染源的重金属对综合生态风险的贡献值分析可知,Cd和Hg的总风险占比超过90%,说明Cd和Hg是污染源生态风险的主要贡献元素.此外,Hg和Cd元素具有较高的毒性,远远高于其他污染元素的毒性系数,从而具有较高的生态风险<sup>[6,48]</sup>.这也印证了具有高贡献率的污染源并不一定具有高的生态风险<sup>[6,48]</sup>.综上,Cd和Hg为研究区绿地土壤生态风险优先控制污染元素,自然-工业源为优先控制污染源.

2.5 兰州黄河风情线绿地土壤重金属特定源-人体 健康风险评价

特定源-人体健康风险评价模型的结果如表7所 示.就非致癌风险而言,不同污染源对成人和儿童 的总HI值分别为5.43×10<sup>-2</sup>和8.55×10<sup>-2</sup>,均小于阈值 1, 表明不同污染源下的绿地土壤重金属对人体健 康不存在非致癌健康风险.对于单个重金属元素而 言, 各污染源情境下的不同重金属元素对成人和儿 童的HI值也均小于1,表明研究区绿地土壤重金属 对成人和儿童不存在非致癌健康风险.对于致癌风 险而言,不同污染源对成人和儿童的 TCR 值分别为 2.17×10<sup>-4</sup> 和 1.38×10<sup>-3</sup>,均高于致癌风险量级水平 (1.0×10<sup>-4</sup>),说明存在致癌风险,且儿童的风险要高 于成人,这与儿童的体型、生理特征和行为习惯相 关<sup>[48]</sup>. 从单个重金属元素来看, 自然-农业源情境下 的不同重金属元素对成人的TCR值超过阈值1×10-4, 而3种污染源情境下的不同重金属元素对儿童的 TCR值均超过阈值1×10<sup>-4</sup>.表明在自然-农业源下的 绿地土壤重金属暴露对成人存在较高致癌健康风 险,而在各污染源情境下的绿地土壤重金属暴露对 儿童均有较高致癌风险,其中Ni和Cr对成人存在显 著致癌风险, Ni、Cr和As对儿童存在显著致癌风 险.然而,由前述分析可知,研究区绿地土壤所有 重金属含量均低于《土壤环境质量 建设用地土壤污

染风险管控标准(试行)》(GB 36600-2018)中第二类 建设用地(绿地与广场用地)的风险筛选值[34],意味 着重金属对人体健康的风险可以忽略,而致癌风险 评价模型结果却显示存在显著致癌性,这是由于致 癌风险评价模型的结果一方面取决于土壤重金属的 含量和生物有效性[49],另一方面也与评价模型中评 估基础参数数据有密切的关系[50].由于本研究基于 土壤重金属的总量来量化对人体健康的直接影响, 并未考虑绿地植物对重金属的转归和人体对重金属 的拮抗作用等间接因素,所以评价结果有一定的放 大效应,但仍需引起关注并加强预警和防范.此外, 本研究结果不仅证实了基于重金属总量的健康风险 评价结果具有放大性,而且也说明了致癌风险模型 中的评估基础参数在建设用地健康风险评价过程中 需要调整,即未来还需进一步开展不同土地利用条 件下和不同土壤类型情境下全面、系统和连续的暴 露参数调查,不断完善和细化我国的健康风险评估 暴露参数.

此外,3种污染源对成人和儿童的非致癌风险 贡献率变化趋势一致(图5),均为:自然-农业源> 交通源>自然-工业源.其中贡献较高的自然-农业源 具有的载荷元素为As和Ni,相较于其他元素,As 和Ni显然具有较低的参考剂量(RfD)和较高的斜率 系数(SF),更易于产生健康风险<sup>[29]</sup>.此外,具有较 高毒性的As和Ni也不是污染较重的元素(图2),却 是最高的健康风险贡献元素.这与不同重金属健康 风险暴露参数不同和毒性系数的强弱及健康风险叠 加累积效应等有关<sup>[51]</sup>.此现象与石文静等<sup>[7]</sup>的研究结 果相似,也进一步证实了污染程度严重的元素不一 定就具有较高的健康风险<sup>[7,51]</sup>.

由上可知,研究区绿地土壤重金属对成人和儿 童不具有非致癌健康风险,但是As、Cr和Ni却对成 人和儿童构成不同程度的致癌健康风险.为此,As 表 7 不同来源重金属对成人和儿童的特定源-健康风险评价

	Т	able 7 Sourc	e-specific health	risks of heavy me	etals from diffe	erent sources f	or adults and chil	dren	
米山	西口		成	人			儿	童	
尖型	坝目	交通源	自然-农业源	自然-工业源	总和	交通源	自然-农业源	自然-工业源	总和
	As	8.98×10 <sup>-3</sup>	9.98×10 <sup>-3</sup>	$1.47 \times 10^{-3}$	2.04×10 <sup>-2</sup>	1.56×10 <sup>-2</sup>	$1.74 \times 10^{-2}$	2.56×10 <sup>-3</sup>	3.56×10 <sup>-2</sup>
	Cd	3.56×10 <sup>-4</sup>	$2.64 \times 10^{-4}$	2.65×10 <sup>-5</sup>	6.46×10 <sup>-4</sup>	4.81×10 <sup>-4</sup>	$3.57 \times 10^{-4}$	3.58×10 <sup>-5</sup>	$8.74 \times 10^{-4}$
	$\mathbf{Cr}$	$1.19 \times 10^{-2}$	$1.38 \times 10^{-2}$	1.12×10 <sup>-3</sup>	$2.68 \times 10^{-2}$	$1.70 \times 10^{-2}$	$1.96 \times 10^{-2}$	1.59×10 <sup>-3</sup>	3.83×10 <sup>-2</sup>
	Cu	$1.62 \times 10^{-4}$	$1.72 \times 10^{-4}$	2.95×10 <sup>-5</sup>	3.64×10 <sup>-4</sup>	$2.80 \times 10^{-4}$	$2.98 \times 10^{-4}$	5.09×10 <sup>-5</sup>	$6.28 \times 10^{-4}$
非致癌风险	Hg	$1.02 \times 10^{-4}$	7.71×10 <sup>-5</sup>	4.74×10 <sup>-5</sup>	$2.26 \times 10^{-4}$	1.63×10 <sup>-4</sup>	$1.23 \times 10^{-4}$	7.58×10 <sup>-5</sup>	$3.62 \times 10^{-4}$
	Ni	3.68×10 <sup>-4</sup>	4.06×10 <sup>-4</sup>	8.09×10 <sup>-5</sup>	$8.55 \times 10^{-4}$	6.34×10 <sup>-4</sup>	6.99×10 <sup>-4</sup>	$1.39 \times 10^{-4}$	$1.47 \times 10^{-3}$
	Pb	2.25×10 <sup>-3</sup>	2.23×10 <sup>-3</sup>	3.05×10 <sup>-4</sup>	$4.78 \times 10^{-3}$	3.78×10 <sup>-3</sup>	3.74×10 <sup>-3</sup>	5.13×10 <sup>-4</sup>	$8.04 \times 10^{-3}$
	Zn	$8.24 \times 10^{-5}$	6.93×10 <sup>-5</sup>	1.17×10 <sup>-5</sup>	1.63×10 <sup>-4</sup>	$1.40 \times 10^{-4}$	$1.18 \times 10^{-4}$	2.00×10 <sup>-5</sup>	$2.78 \times 10^{-4}$
	总 HI	2.42×10 <sup>-2</sup>	2.70×10 <sup>-2</sup>	3.09×10 <sup>-3</sup>	5.43×10 <sup>-2</sup>	3.82×10 <sup>-2</sup>	4.24×10 <sup>-2</sup>	4.99×10 <sup>-3</sup>	8.55×10 <sup>-2</sup>
	As	1.18×10 <sup>-5</sup>	1.31×10 <sup>-5</sup>	1.93×10 <sup>-6</sup>	2.68×10 <sup>-5</sup>	8.22×10 <sup>-5</sup>	9.13×10 <sup>-5</sup>	1.34×10 <sup>-5</sup>	$1.87 \times 10^{-4}$
	Cd	8.98×10 <sup>-8</sup>	6.67×10 <sup>-8</sup>	6.68×10 <sup>-9</sup>	1.63×10 <sup>-7</sup>	6.34×10 <sup>-7</sup>	4.71×10 <sup>-7</sup>	4.72×10 <sup>-8</sup>	1.15×10 <sup>-6</sup>
致癌风险	$\mathbf{Cr}$	2.16×10 <sup>-5</sup>	2.49×10 <sup>-5</sup>	2.02×10 <sup>-6</sup>	4.85×10 <sup>-5</sup>	$1.48 \times 10^{-4}$	$1.70 \times 10^{-4}$	1.38×10 <sup>-5</sup>	3.31×10 <sup>-4</sup>
	Ni	6.09×10 <sup>-5</sup>	6.72×10 <sup>-5</sup>	1.34×10 <sup>-5</sup>	$1.41 \times 10^{-4}$	3.71×10 <sup>-4</sup>	4.10×10 <sup>-4</sup>	8.16×10 <sup>-5</sup>	8.63×10 <sup>-4</sup>
	TCR	9.44×10 <sup>-5</sup>	$1.05 \times 10^{-4}$	1.73×10 <sup>-5</sup>	2.17×10 <sup>-4</sup>	6.02×10 <sup>-4</sup>	6.72×10 <sup>-4</sup>	$1.09 \times 10^{-4}$	1.38×10 <sup>-3</sup>



Fig. 5 Proportion of source-specific health risks from different sources for adults and children

和 Ni 为人体健康风险优先控制污染元素,自然-农业源为优先控制污染源.

#### 3 结论

(1)研究区绿地土壤重金属除 Cr 和 Ni 的含量均 值较低外,其余 As、Cd、Cu、Hg、Pb 和 Zn 的含量 均值均高于兰州市土壤元素背景值,存在一定程度 的累积超标现象.

(2)单因子污染指数评价结果显示,研究区绿 地土壤 Cr和Ni为清洁无污染,Cd、Zn、Pb、Cu、 As和Hg存在不同程度的污染,需关注Hg污染.污 染负荷指数评价结果显示,研究区整体上处于无-中 度污染水平,以轻度污染为主.

(3) APCS-MLR 源解析表明,研究区绿地土壤重 金属 Pb、Cd和Zn主要受交通源的影响,贡献率分 别为88.4%、85.3%和77.7%; Ni和As受自然-农业 源的影响,贡献率分别为92.1%和72.6%;Cu分别 受交通源和农业源的影响,贡献率为62.1%和 51.7%;Cr和Hg受自然-工业源的影响,贡献率为 89.0%和64.3%.

(4)特定源-综合生态风险指数表明,自然-工业 源对研究区生态风险贡献最大,为优先控制污染源, Cd和Hg为生态风险优先控制污染元素;特定源-健康 风险评估表明,研究区重金属对人群不构成非致癌健 康风险,但As、Cr和Ni对人群存在不同程度较高致 癌健康风险.此外,自然-农业源为人体健康风险优先 控制污染源,As和Ni为优先控制污染元素.

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Environmental Science (monthly)

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Formation Mechanism , Structural Characteristics of Ultrafine Mineral Particles, and Their Environmental Effects Research Progress in Application of Biochar-immobilized Bacteria Composites in Environmental Remediation Effect of Microplastics on Ammonia Nirogen Adsorption by Zeolite in Water and Its Mechanism Mechanism of Ultraviolet Aging Effect on the Adsorption of Ciprofloxacin by Nano-biochar <sup>-</sup> Response of Phytoplankton Communities and Environmental Factors Under the Influence of Land Use in the Wuding River Basin Bacterial Community Diversity and Functional Gene Abundance of Culturable Bacteria in the Wetland of Poyang Lake Bacterial Community Structure and Its Relationship with Heavy Metals in Sediments of Diannong River Impacts of Treated Wastewater on Bacterial and Fungal Microbial Communities in Receiving Rivers Metagenomics Reveals the Characteristics and Functions of Bacterial Community in the Advanced Wastewater Treatment Process Spatio-temporal Characteristics of Habitat Quality and Natural-human Driven Mechanism in Dabie Mountain Area Typerspectral Inversion of Soil Organic Matter Content Based on Continuous Wavelet Transform, SHAP, and XCBoost Effects of Straw Retention , Film Mulching, and Nitrogen Input on Soil Quality in Dryland Wheat Field Effects of Different Soil Salinities on N <sub>2</sub> O Emission ; A Meta-analysis Massessment and Prediction of Carbon Storage Based on Land Use/Land Cover Dynamics in the Guangdong-Hong Kong-Macao Greater Bis Simulation of Temporal and Spatial Changes in Ecosystem Carbon Storage in Funiu Mountains Based on InVEST Model Relationship Between Microbial Nutrient Limitation and Soil Organic Carbon Fraction During Shelterbelts Construction Characteristics and Driving Forces of Organic Carbon Mineralization in Brown Soil with Long-term Straw Returning Effects of Bicchar Combined with Different Types of Nitrogen Fertilizers on Denitrification Bacteria Community in Vegetable Soil Investigation of Seil Microbial Characteritse During Stand Development in <i>Pin</i>	<ul> <li><sup></sup>QU Si-tong, SHAN Su-jie, WANG Chong-ming, et al. (2160)</li> <li>LIU Zhen-hai, ZHANG Zhan-hua, YUAN Yu-xin, et al. (2171)</li> <li><sup></sup>SUN Shu-yu, HUANG Meng-xin, KONG Qiang, et al. (2185)</li> <li><sup></sup>LIAN Jian-jun, XIE Shi-ting, WU Pei, et al. (2195)</li> <li><sup></sup>CUO Shan-song, HU En, DING Yi-tong, et al. (2211)</li> <li><sup></sup>YU Jiang, WANG Chun, LONG Yong, et al. (2223)</li> <li><sup></sup>LIU Shuang-yu, MENG Jun-jie, QIU Xiao-cong, et al. (2223)</li> <li><sup></sup>LIU Shuang-yu, MENG Jun-jie, QIU Xiao-cong, et al. (2223)</li> <li><sup></sup>CUO You-shun, YU Zhong, HAO Wen-bin, et al. (2246)</li> <li><sup></sup>HU Jian-shuang, WANG Yan, ZHOU Zheng, et al. (2259)</li> <li>HENG Ya-ping, ZHANG Jun-hua, TIAN Hui-wen, et al. (2268)</li> <li><sup></sup>YE Miao, ZHU Lin, LIU Xu-dong, et al. (2292)</li> <li><sup></sup>WANG Chang-yuan, MA Xiao-chi, GUO De-jie, et al. (2304)</li> <li><sup></sup>HUANG Yi-hua, SHE Dong-li, SHI Zhen-qi, et al. (2313)</li> <li>ay Area <sup></sup>ZHENG Hui-ling, ZHENG Hui-feng (2321)</li> <li><sup></sup>ZHANG Zhe, SHI Zhen-qin, ZHU Wen-bo, et al. (2332)</li> <li><sup></sup>ZHANG Zhe, SHI Zhen-qin, ZHU Wen-bo, et al. (2333)</li> <li><sup></sup>ZHANG Xin-yue, XIAO Qi-tao, XIE Hui, et al. (2363)</li> <li><sup></sup>ZHANG Xin-yue, XIAO Qi-tao, XIE Hui, et al. (2373)</li> <li><sup></sup>ZHANG Xin-yue, XIAO Qi-tao, XIE Hui, et al. (2344)</li> <li><sup></sup>ZHANG Xin-yue, XIAO Qi-tao, XIE Hui, et al. (2424)</li> <li><sup></sup>ZHANG Xin-yue, XIAO Qi-tao, XIE Hui, et al. (2424)</li> <li><sup></sup>ZHANG Xin-yue, LI Xing-ren, GAO Shang-jie, et al. (2417)</li> <li>Analysis in the Yellow River Custom Tourist Line of Lanzhou</li> <li><sup></sup>ZHANG Hai-lin, ZHANG Yu, WANG Ding, et al. (2428)</li> <li>Southwest China</li> <li><sup></sup>ZHANG Shi-li, ZHANG Yu, WANG Ding, et al. (24240)</li> <li><sup></sup>LIAO Ze-yuan, LI Jie-qin, SHEN Zhi-jie, et al. (2424)</li> <li><sup></sup>WEI Hong-bin, LUO Ming, XIANG Lei, et al. (2474)</li> <li><sup></sup>WEI Hong-bin, LUO Ming, XIANG Lei, et al. (2474)</li> <li><sup></sup>WEI Hong-bin, LUO Ming, XIANG Lei, et al. (2474)</li> </ul>