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中国城市PM25和PM10时空分布特征和影响因素分析

李江苏,段良荣*,张天娇

(河南大学黄河文明与可持续发展研究中心暨黄河文明省部共建协同创新中心, 开封 475001)

摘要: PM_{25} 和 PM_{10} 浓度超标引发的空气质量问题严重影响公众健康,研究 PM_{25} 和 PM_{10} 浓度对制定有效的污染防控和治理措施具有重要意义. 运用时空分析法,分析 2018 年季度 PM_{25} 和 PM_{10} 浓度时空分布,并用 GWR 探究浓度差异的原因. 结果表明:① PM_{25} 和 PM_{10} 的浓度均呈冬春高、夏秋低的季节性规律;四季污染物浓度在胡焕庸线两侧存在显著差异,该线以东地区高浓度聚集在京津冀地区,该线以西地区高浓度聚集在新疆中南部.② PM_{25} 和 PM_{10} 浓度的 Moran's I 在四季均为正,且均在冬季增至最大值; PM_{25} 和 PM_{10} 的分布格局基本一致,"高-高"类和"低-低"类集中分布现象明显.③各因素对 PM_{25} 和 PM_{10} 浓度的影响存在较大空间异质性. 温度和坡度对 PM_{25} 和 PM_{10} 浓度的影响均表现为南部地区呈负相关,北部地区呈正相关. 降水和风速对大部分地区 PM_{25} 和 PM_{10} 浓度的影响呈负相关,且夏季负相关的范围均较小. 人类活动强度、二产占比和能源消耗总量对地区 PM_{25} 和 PM_{10} 浓度的影响在大部分地区呈正相关,各季度的正相关高值区也存在差异.

关键词: PM₂₅浓度; PM₁₀浓度; 时空分布; 地理加权回归(GWR); 影响因素

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Analysis of Spatio-temporal Distribution Characteristics and Influencing Factors of PM_{2.5} and PM₁₀ in Chinese Cities

LI Jiang-su, DUAN Liang-rong*, ZHANG Tian-jiao

(Key Research Institute of Yellow River Civilization and Sustainable Development & Collaborative Innovation Center of Yellow River Civilization Provincial Co-construction, Henan University, Kaifeng 475001, China)

Abstract: In recent years, the air quality problems caused by extreme haze events have become increasingly serious in China, especially those caused by fine particulate matter ($PM_{2,5}$), which has become the main component of haze. The air quality problems caused by excessive concentrations of $PM_{2,5}$ and PM_{10} have seriously affected the public health of Chinese cities. Studying the concentrations of $PM_{2,5}$ and PM_{10} is of great significance for formulating effective pollution prevention and control measures. This study selected seven pollutant concentration impact factors, including four natural factors (temperature, precipitation, wind speed, and slope) and three socio-economic factors (human activity intensity index, the proportion of secondary industry, and total energy consumption). This study used the spatio-temporal analysis method to analyze the spatio-temporal distribution of $PM_{2,5}$ and PM_{10} concentrations throughout the year and quarter in 2018 and used GWR to explore the factors that caused the concentration difference. The results showed that: ① the concentrations of $PM_{2,5}$ and PM_{10} were higher in winter and spring than in summer and autumn. Pollutant concentrations differed significantly on both sides of the Hu Huanyong line in all four seasons, with high concentrations in the east gathering in Beijing, Tianjim, and Hebei and concentrations in the west gathering in south-central Xinjiang. ② The Moran indexes of $PM_{2,5}$ and PM_{10} concentrations were positive in all four seasons and reached the maximum in winter; the distribution pattern of $PM_{2,5}$ and PM_{10} concentrations showed great spatial heterogeneity. The influence of temperature and slope on $PM_{2,5}$ and PM_{10} concentrations was negative in the south and positive in the north, and their distribution was not identical. Precipitation and wind speed had a negative correlation effect on $PM_{2,5}$ and PM_{10} concentrations were positively correlated high value areas from quarter to quar

Key words: concentration of PM2, 5; concentration of PM10; spatio-temporal distribution; geographically weighted regression (GWR); influencing factors

伴随着我国经济和城市化快速发展,大气环境面临着人口增长和经济发展的双重压力,环境问题已经成为社会公众关注的焦点^[1,2]. 对大气污染物的关注主要集中于 PM₂₅、PM₁₀、SO₂、NO₂、O₃和 CO₂等^[3,4],特别是由 PM₂₅和 PM₁₀浓度超标引发的空气质量恶化,成为了公众和社会舆论关注的热点问题^[4,5]. 根据空气动力学,将大气颗粒物分为总悬浮颗粒物(TSP)、可吸入颗粒物(PM₁₀)和细颗粒物(PM₂₅),由于 PM₂₅和 PM₁₀粒径小、吸附性强,对人体心血管系统产生严重危害,同时对肺癌发生率和死亡率有直接关系^[6]. 根据《危险的呼吸 2:大气 PM₂₅对中国城市公众健康效应研究》^[7],2013年大气 PM₂₅

污染导致全国31座省会或直辖市中25.7万人超额死亡,空气污染已成为威胁全球人类健康的第四大因素. 习近平总书记在中国共产党第二十次全国代表大会中指出,深入推进环境污染防治,持续深入打好蓝天、碧水和净土保卫战,基本消除重污染天气. 在此背景下,研究全国地级市 PM₁₀浓度

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作者简介:李江苏(1983~),男,博士,副教授,主要研究方向为城市

发展与大气环境遥感,E-mail; lijs. 09b@igsnrr. ac. cn * 通信作者,E-mail; 104753211575@henu. edu. cn 的时空分布特征,分析其影响因素并对其空间异质 性进行解释,对各区域有效制定污染防控和环境治 理措施具有较大的借鉴意义,对城市的生态文明建 设具有重要的指导意义.

国内外学者对PM25和PM10的研究,主要从PM25 和 PM₁₀浓度的时空演化特征、相关性研究和影响因 素等方面进行探究.在时间尺度上,已有研究得出 污染物浓度 "U" 型逐月变化规律和周期性 "U-脉 冲"型逐日变化规律[8,9];也有学者利用2013年后 公开的站点数据估算其他年份污染物浓度, 研究近 10年污染物浓度逐年增减趋势[9,10]. 在空间尺度上, 不少学者选择某一代表性城市为研究对象[11];也有 学者以城市群[12]为研究对象,如京津冀[13]、长三 角[14,15]和辽中南[16,17]等城市群.在研究方法上,使用 多元回归[18]、空间计量[19,20]、地理探测器[18,21]和地 理加权回归[21-23]. 大多研究采用空间计量模型, 假 定影响因素对PM25和PM10在不同空间位置的影响一 致,忽略了局部影响的空间差异.然而不同区域大 气污染物具有明显的差异,不同的自然条件和社会 经济因素对其影响存在空间差异.故本研究拟采用 地理加权的方法, 分析不同区域受各因素影响的区 域差异.

综上所述, 大部分学者多研究单一城市或区域 单一污染物,部分研究将PM25和PM10进行比较,发 现 PM25和 PM10表现出很强的相关性[24,25]. 研究方法也 多采用统计方法,忽略了空间差异.鉴于此, 运用空间自相关和地理加权回归(GWR),对全国范 围内的PM25和PM10的季节变化及其影响因素的空间 异质性进行研究.已有研究表明自然因素中的气温、 降水、风速、坡度、气压和相对湿度等^[26]对 PM₂₅和 PM₁₀浓度产生较大影响,社会经济因子中人口密度、 人均 GDP、能源消耗和第二产业占比等[27]因素对 PM₂,有显著的影响.基于文献的梳理、数据精度和 数据可获得性,本文选取了污染物的影响因子,自 然因子为:气温(X1)、降水(X2)、风速(X3)和坡度 (X4); 社会经济因子的选取, 不局限于单一视角, 一方面从土地利用视角,借鉴徐勇[28]的人类活动强 度指数,以建设用地当量为度量单位,计算各区域 的人类活动强度(X5);另一方面,考虑到工业发展 和取暖对污染的影响,选取第二产业占比(X6)和能 源消耗总量(X7). 以 2018年季度 PM25和 PM10浓度为 变量,定量分析自然因素和社会因素对城市大气污 染物浓度的影响,揭示影响因素的空间异质性,以 期为PM25和PM10污染治理提供科学依据.

1 材料与方法

1.1 数据来源

本研究涉及的数据有: ①2018年全国 363 个地 级市城市的PM25和PM10季度数据,用城市边界统计 城市内所有监测站点值, 计算年均值和四季均值, 用24h数据计算日均值、用日均值计算月均值,进 而计算四季均值,四季时段划分为:春(2018年3~ 5月)、夏(2018年6~8月)、秋(2018年9~11月)和 冬(2018年12月至2019年2月),数据来源于中国空 气质量在线监测分析平台(https://www.aqistudy. cn/). ②气温、降水和风速的季度数据,数据处理方 法与大气污染数据处理方法一致,对于少数城市其 边界内无气象站点,选择距离该城市最近的站点提 取该城市气象数据,该数据来源于美国国家气候数 据中心 (National Climatic Data Center) 的开放数据 国 DEM 数据,空间分辨率为 30 m×30 m,用于计算 坡度. ④2018年中国土地利用/土地覆盖遥感监测数 据库(LUCC), 空间分辨率为30m×30m, 用于计算 人类活动强度.数据③和④来源于中国科学院资源 环境科学与数据中心(http://www.resdc.cn/). ⑤全国 263个城市的第二产业占 GDP 比重和能源消耗数据 均来源于2019年《中国城市统计年鉴》. 在上述数 据中,数据①~④包含中国香港和澳门;数据⑤暂 缺中国香港、澳门和台湾数据.在影响因素分析中, 均对变量和自变量数据进行了标准化.

1.2 研究方法

1.2.1 人类活动强度

人类活动强度(human activity intensity of land surface, HAILS)是根据土地利用类型来判断人类活动对生态环境干扰程度的指标.不同的人类活动强度对大气污染的影响程度不同,本文借鉴徐勇^[28]的人类活动强度测度方法,测度中国城市人类活动强度.利用2018年中国大陆的土地利用类型数据,以建设用地当量(construction land equivalent, CLE)为度量单位,将所有土地利用类型按照其对应的建设用地当量折算系数(conversion index of construction land equivalent, CI)换算成建设用地当量(表1),区域内建设用地当量面积占区域总面积的百分比即为该区域的人类活动强度.计算公式为:

$$HAILS = \frac{S_{CLE}}{S} \times 100\% \tag{1}$$

$$S_{\text{CLE}} = \sum_{i=1}^{n} (\text{SL}_{i} \cdot \text{CI}_{i})$$
 (2)

式中,HAILS为人类活动强度; S_{CLE} 为区域内建设用

表 1 不同土地利用类型的建设用地当量折算系数

Table 1	Conversion	indox of	Construction	land aguiva	lant for differen	nt land use types

类型	土地利用类型	特征标志说明	CI_i
耕地	旱地/水田	表层自然覆被改变,种植1年生作物	0.200
++ 116	有林地/灌林地/疏林地	表层自然覆被未改变且未被利用	0.000
林地	未成林造林地/苗圃/各类园地	表层自然覆被改变,种植多年生植物	0.133
草地	天然/改良牧草地	表层自然覆被未改变但被利用	0.067
建设用地	城镇用地/农村居民点/独立工矿/交通用地/特殊用地	表层有人工隔层,水分、养分、空气和热量交换阻滞	1.000
44.4	河渠/湖泊/滩涂/滩地/永久性冰川雪地	表层自然覆被未改变且未被利用	0.000
水域	水库坑塘	表层自然覆被改变,空气和热量交换阻滞	0.600
未利用地	沙地/戈壁/盐碱地/沼泽/裸土地	表层自然覆被未改变且未被利用	0.000

地当量面积;S为区域总面积;SL为第i种土地利用类型的面积;CI为第i种土地利用类型的建设用地当量折算系数;n为土地利用类型的总数.

1.2.2 空间自相关方法

空间自相关(spatial autocorrelation)用于反映某一区域中的某种现象与邻近区域同一现象的相关程度,揭示现象之间潜在的相互依赖性.通常分为全局空间自相关与局部空间自相关两大类.计算公式为:

Moran's
$$I = \frac{\sum_{i=1}^{N} \sum_{j=1}^{N} \mathbf{W}_{ij} (y_i - \bar{y}) (y_j - \bar{y})}{S^2 \sum_{i=1}^{N} \sum_{j=1}^{N} \mathbf{W}_{ij}}$$
 (3)

式中, y_i 为城市i大气污染程度;y和 S^2 分别为城市大气污染程度平均值和方差;N为城市总数; W_i 为空间权重矩阵。Moran's I取值范围为(-1,1),若其值高于0,表明大气污染程度之间具有空间正相关性。另外,采用标准统计量Z来检验 Moran's I的显著性水平,计算公式为:

$$Z(\text{Moran's } I) = \frac{\text{Moran's } I - E(\text{Moran's } I)}{\sqrt{\text{VAR}}}$$
 (4)

$$E(\text{Moran's }I) = -\frac{1}{n-1} \tag{5}$$

1.2.3 普通最小二乘法

普通最小二乘(ordinary least squares, OLS)模型是用来解释因变量 Y_i 与自变量 X_i 之间关系的多元线性函数、计算公式为:

$$Y_i = \beta_0 + \sum_i \beta_i X_i + \varepsilon_i \tag{6}$$

式中, β_0 为常数项; β_i 为回归系数; ϵ_i 为随机误差项.

1.2.4 地理加权回归

地理加权回归(geographically weighted regression, GWR)是 Fotheringham 等在传统最小二乘法(ordinary least squares, OLS)模型的基础上将数据地理位置加入回归参数中,同时考虑相邻点的空间权重,允许局部参数估计的地学统计方法^[29].本研究采用 GWR

揭示大气污染物 $(PM_{25} \pi PM_{10})$ 影响因素的空间异质性, 计算公式为:

$$Y_i = \beta_0(\mu_i, \theta_i) + \sum_k \beta_k(\mu_i, \theta_i) X_{ik} + \varepsilon_i$$
 (7)
式中, Y_i 为 i 城市的因变量解释值, (μ_i, θ_i) 为城市 i 的地理坐标; X_{ik} 为城市 i 的自变量解释值,本文中包括 4 个自然因子和 3 个人文因子; $\beta_k(\mu_i, \theta_i)$ 为研究单元 i 区域质心 (μ_i, θ_i) 处的回归参数; ε_i 为随机误差项.

2 结果与分析

2.1 中国 PM_{2.5}和 PM₁₀的时空变化特征

2.1.1 PM, 和 PM 10季节变化特征

基于中国《环境空气质量标准》(GB 3095-2012)中污染物浓度标准,将PM25和PM10浓度值划 分为5个区间(表2和表3),分析研究时段内各区间 浓度的城市数量.结果表明:①中国PM25浓度总体 呈冬春高、夏秋低的季节性规律,具体表现为:冬 季>春季>秋季>夏季,ρ(PM₂₅)均值从大到小依次 为 55.20、41.03、32.79 和 24.63 μg·m⁻³, 这与已有研 究的结果基本一致[30]. 根据二类环境空气质量功能 区(居住区、商业交通居民混合区、文化区、工业 区和农村地区)的浓度标准,这类区域 $\rho(PM_{25})$ 应低 于35 µg·m³; 按此浓度标准, 冬春两季均不符合浓 度标准, 仅夏秋两季符合浓度标准; 四季符合浓度 标准的城市数量分别占 36.09%、89.81%、59.50% 和 30.03%. ② PM10浓度表现为冬春高, 夏季低, 春夏 秋冬的 $\rho(PM_{10})$ 均值分别为91.08、50.01、62.90和 88.13 μg·m⁻³, 这与已有研究的结果较为一致^[31]. 根 据二类环境空气质量功能区的浓度标准, $\rho(PM_{10})$ 应 低于70 µg·m⁻³,按此浓度标准,冬春两季不符合浓 度标准, 仅夏秋两季符合浓度标准; 四季符合浓度 标准的城市数量分别占 40.50%、87.61%、71.07% 和 39.12%.

2.1.2 PM, 和 PM 应 包 变 化 特征

中国季度 PM25和 PM10浓度空间格局演化呈现的

表 2 PM_{2.5}季均浓度变化特征及其浓度区间的城市数量

Table 2 Characteristics of the change in quarterly average PM_{2.5} concentration and the number of cities in its concentration range

季节		$PM_{2.5}$,各浓度区间城市数	量/个		ρ(PM _{2.5})均值
	< 15	15 ~ 35	36 ~ 75	76 ~ 115	> 115	ρ(PM _{2.5})均值 /μg·m ⁻³
春季	6	125	227	2	0	41.03
夏季	42	284	36	1	0	24.63
秋季	29	187	147	0	0	32.79
冬季	13	96	174	65	15	55.20

表 3 PM₁₀季均浓度变化特征及其浓度区间的城市数量

Table 3 Characteristics of the change in the quarterly average PM₁₀ concentration and the number of cities in its concentration range

季节		PM	10各浓度区间城市数量	量/个		ρ(PM ₁₀)均值 /μg·m ⁻³
表 h	< 40	41 ~ 70	71 ~ 150	151 ~ 300	> 300	$/\mu g \cdot m^{-3}$
春季	15	141	185	17	5	91.08
夏季	125	197	38	2	1	50.01
秋季	67	188	100	3	0	62.90
冬季	39	100	176	41	1	88.13

特征为: ①胡焕庸线以东地区, 两季 $\rho(PM_{25})$ 高于 35 μg·m⁻³的区域主要分布在华北平原、东北平原 和长江中下游平原, 其中冬季华北平原地区 ρ (PM_{2.5}) 高于 75 μ g·m⁻³; 夏季 ρ (PM_{2.5}) 高于 35 μg·m-3的区域范围缩小至京津冀地区; 秋季范围 有所扩大, 在山东、河南、江苏、湖北和安徽等 区域也有分布(图1). ②胡焕庸线以西地区, 四季 ρ(PM_{2.5})均高于35 μg·m⁻³的地区主要分布在新疆的 和田、阿克苏和喀什等地区, 其中冬春两季出现 ρ(PM_{2.5})高于115 μg·m⁻³的地区. ③PM₁₀浓度季节分 布格局和 PM25大体一致, 年均值低于二级浓度标 准的区域主要分布在呼伦贝尔、塔河、黑河、青 藏高原和秦岭淮河以南的大片区域; 而年均值高 于二级浓度标准的地区主要分布在黄河中下游、 京津冀、新疆中南部和华中地区,以上地区颗粒 物污染较为严重. ④PMio浓度较高地区在冬春两季 分布范围最大, 夏秋两季缩减至京津冀和新疆 南域.

2.1.3 PM25和PM10空间相关性特征

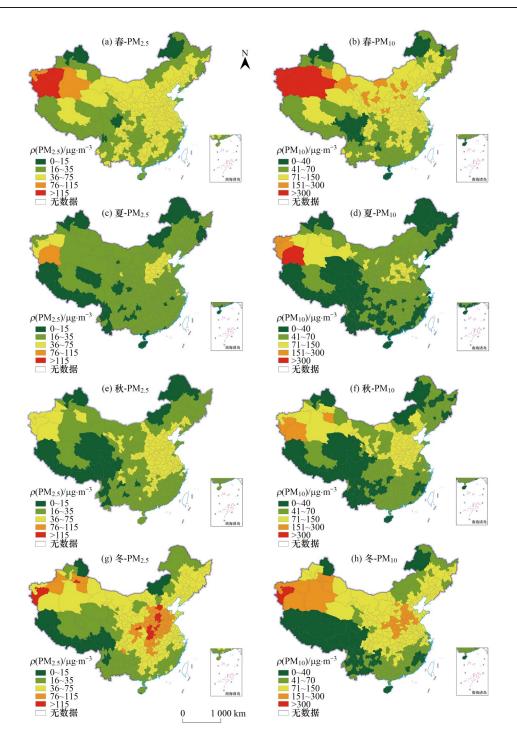
中国城市 PM₂₅和 PM₁₀浓度的 Moran's *I*在 2018年四季均为正,且通过 1%的显著性检验(表 4),表明城市间存在显著的全局空间正相关,这与已有研究的结果一致^[32].而且 Moran's *I*均在冬季增至最大,说

明中国城市的 PM_{2.5}和 PM₁₀浓度差距在冬季达到最大.可见,城市 PM_{2.5}和 PM₁₀浓度具有全局自相关性,由于全局空间自相关分析无法反映局部空间异质性,故采用 LISA 集聚图进行局部相关性分析.

根据四季各城市 PM25和 PM10浓度的集聚情况, 将我国城市划分成4类聚集区,在显著性水平为 0.01的条件下,大多数城市污染物浓度集聚类型为 "高-高"和"低-低"两类,说明我国PM25和PM10 浓度空间集聚特征明显. PM25和 PM10的分布格局基 本一致,"高-高"类主要分布在京津冀、黄河中下 游和新疆部分地区,其中在春夏两季PM10"高-高" 类范围更大,包含了巴彦淖尔、阿拉善盟和鄂尔多 斯等地区."低-低"类主要分布在长江流域以南, 包括东南沿海、川西、两广和云南, 以及东北部的 呼伦贝尔和齐齐哈尔.集聚类型受季节影响明显, PM10的"高-高"类地区由春夏到秋冬南移较为明 显; PM25的"高-高"类地区呈轻微南移."低-高" 和"高-低"两类集聚特征不明显, PM25 "低-高" 类地区分散布局于山东半岛、黄土高原、内蒙古和 新疆地区; PM10 "低-高" 类地区分布于阿里地区, 秋冬季节范围扩大至乌兰察布、张家口、鄂尔多斯 和阿拉善盟. "高-低"类城市数量极少, PM25四季 分别有6、10、3和1个城市呈现"高-低"集聚,

表 4 PM_{2.5}和 PM₁₀ Moran's I 计算结果 Table 4 PM_{2.5} and PM₁₀ Moran's I calculation results

番目		PM	I _{2.5}			PM	1 ₁₀	
项目	春季	夏季	秋季	冬季	春季	夏季	秋季	冬季
Moran's I	0.29	0.44	0.69	0.75	0.18	0.19	0.59	0.61
Z 值	22.26	32.77	50.39	53.87	13.82	16.28	43.11	44.29
P值	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00



基于自然资源部审图号为 GS(2020)4632的标准地图制作,底图无修改,下同

图 1 2018年各季度 PM_{2.5}和 PM₁₀浓度分布

Fig. 1 ~ PM $_{\!2.5}$ and PM $_{\!10}$ concentration distribution by quarter in 2018

PM₁₀四季分别有 2、7、2 和 4 个城市呈现"高-低"集聚(表 5). 总体来看,大部分城市属于"高-高"

和"低-低"类,表明大部分城市污染浓度在空间上 呈集聚特征,即PM₂₅和PM₁₀两者浓度较高或较低的

表 5 $PM_{2.5}$ 和 PM_{10} 浓度集聚各类型城市数量

Table 5 Number of cities with $PM_{2.5}$ and PM_{10} concentration clusters of various types

				2. 3 10		7.1		
季节		PM	2.5			PM	1 10	
→ h	H-H	H-L	L-H	L-L	Н-Н	H-L	L-H	L-L
春	90	6	12	106	77	2	2	154
夏	100	10	19	115	78	7	7	151
秋	107	3	16	93	155	2	13	146
冬	111	1	14	120	104	4	11	128

城市在空间上集中分布.

2.2 影响因素分析

2.2.1 OLS和 GWR 模型对比分析

对比 GWR 和 OLS 模型运算结果,各季度 PM₂₅和 PM₁₀浓度的影响因素均呈现 GWR 模型拟合效果优于 OLS(表 6). GWR 模型的调整 R²高于 OLS,AICc和-2倍的对数似然值低于 OLS. 各季度 GWR 模型的拟合效果存在显著差异,其中 PM₂₅在秋冬季节拟合效果较好,调整 R²高于 0.75,夏季调整 R²为 0.59,拟合效果差于其他季节. PM₁₀调整 R²在四季均高于 0.72,也表现出春秋冬三季的拟合效果优于夏季. 夏季拟合效果差的原因可能在于: 受数据制约,部分对夏季影响较大的指标未能纳入回归模型,该类指标如太阳高度角大小. 夏季太阳直射北半球,太阳高度角在四季中最大,太阳直射导致强烈的光化学反应会使部分一次污染物生成大量的二次颗粒,且 夏季逆温弱、边界层不稳定等特点会造成污染物不同程度的扩散^[23,33].

表 6 OLS和GWR模型输出结果对比

Table 6 Comparison of output results between OLS and GWR models

		- F X	/ /
因变量	模型输出结果	OLS	GWR
	AICc	-200.325	-354.856
春-PM _{2.5}	-2倍对数似然值	-219.036	-478.623
91	调整 R ²	0.267	0.653
1.1	AICc	-259.471	-394.659
夏-PM _{2.5}	-2倍对数似然值	-278.182	-516.566
01/	调整 R ²	0.198	0.591
10 h	AICc	-226.787	-395.233
秋-PM _{2.5}	-2倍对数似然值	-245.499	-570.179
\//	调整 R ²	0.390	0.751
41	AICe	-142.276	-371.954
冬-PM _{2.5}	-2倍对数似然值	-160.987	-543.112
	调整 R^2	0.318	0.778
	AICc	-332.129	-627.775
春-PM ₁₀	-2倍对数似然值	-350.841	-800.839
	调整 R^2	0.229	0.805
	AICe	-131.336	-355.206
夏-PM ₁₀	-2倍对数似然值	-150.047	-510.289
	调整 R^2	0.189	0.723
	AICc	-306.876	-541.031
秋-PM ₁₀	-2倍对数似然值	-325.588	-768.892
	调整 R^2	0.361	0.818
	AICe	-341.635	-595.848
冬-PM ₁₀	-2倍对数似然值	-360.347	-764.226
	调整 R ²	0.412	0.825

2.2.2 GWR模型影响因素的空间异质性分析

本文采用 ArcGIS 对 PM₂₅和 PM₁₀浓度的影响因素 进行地理加权回归分析,并将回归结果进行可视 化,进而分析各影响因素时空差异.

(1)温度 温度(X1)对 PM₂₅和 PM₁₀浓度影响均 呈两极分化的空间格局,总体上呈南负北正.四季 存在明显差距,负效应范围大小排序:冬季>秋 季>春季>夏季, 其中夏季的拟合效果远不如其他 季节(表6). 负影响的高值区由春季珠三角向夏秋冬 季逐渐北移到长三角和山东半岛地区;正效应范围 大小排序: 夏季 > 春季 > 秋季 > 冬季, 在此排序下 正效应强度逐渐增大,高值区由东北部地区逐渐过 渡到京津冀地区.夏季温度对PM25负效应最强的地 区集中在以南阳和驻马店为中心的中部地区, 以及 东南沿海的宁德、福州和温州等地区, 负相关系数 达到-0.20, 意味着温度每增加1, PM25浓度可降低 0.20; 正效应最强的地区集中分布东北部地区, 其 中黑河、哈尔滨、牡丹江和呼和浩特等正相关系数 均高于0.40, 意味着温度每增加1, PM25可增高 0.40. 夏季温度对 PM10 负效应最强的地区也分布在南 阳和驻马店地区,连云港和盐城也有分布,负相关 系数均达到-0.50.

温度对中国污染物浓度的影响呈南负北正的空间差异^[34],造成这种分布格局的原因与地区气温特点密切相关,受经纬度和季风的影响,北方地区年均气温普遍低于南方地区,南方地区气温越高对流活动越频繁,其引发的对流雨对污染物有清除作用.北方地区正效应主要受冬季特殊气象条件影响,冷空气活动是影响大气污染物累积和消散的主要气象条件之一.在温度较高的晴朗天,冷空气活动较弱,容易产生静温或逆温现象^[35],进而导致大气污染积聚,空气污染加重;但是冬季东北地区受到来自蒙古北部寒冷空气的低温影响,冷空气活动强,不易发生逆温现象^[36],污染物浓度会相对降低.

(2)降雨 降雨(X2)对 PM₂₅和 PM₁₀浓度的影响 在大部分地区呈现负效应, PM10负效应范围更大. 降雨对 PM₂₅的负效应范围在春夏秋冬分别是 95.81%、61.98%、70.34%和64.26%;对PM10负效应 范围在四季分别是90.87%、71.86%、68.82%和 80.99%, 表明充沛的降水对我国大部分地区的PM25 和 PM10浓度有降低作用,与已有研究的结论一致[37]. 不同季节之间具有明显的空间差异, 负影响的高值 区由中西部地区向东部地区偏移.在春夏秋三季中, 黄河中上游地区污染物浓度受降水影响呈显著负效 应,且负相关程度较强,酒泉、金昌和张掖等地区 的 PM₂₅和 PM₁₀负相关系数分别达到 0.30 和 0.50,表 明降水量每增长1, PM25浓度可降低0.30, PM10浓度 可降低 0.50. 然而, 冬季黄河中上游地区降水对污染 物浓度呈正效应,正相关系数达到0.70,表明降水 量每增长1, PM25和PM10浓度均可增加0.70.

降水可以捕获大气中的颗粒物和气溶胶粒子,使之从大气中清除,同时对地表扬尘有一定的抑制作用. 秋季华北地区降雨较多,导致山东半岛及京津冀地区的污染物浓度降低. 然而,冬季甘肃和宁夏等大部分地区降水对 PM₂₅的影响呈现正效应,主要是因为冬季甘肃、宁夏等地风速较大,其北部阿拉善盟和巴彦淖尔的平均风速可达 3.64 级,而全国均值仅为 2.54 级,加之以上地区地表植被覆盖率低,地表细小颗粒物易被扬起,造成空气中细小颗粒物浓度增大.

(3)风速 风速(X3)对 PM₂₅和 PM₁₀浓度的影响主要呈负相关,风速对 PM₂₅的影响范围在春夏秋冬分别是 70.72%、73.76%、57.80% 和 69.96%;风速对 PM₁₀的影响范围在春夏秋冬分别是 72.62%、70.34%、60.08% 和 79.47%. 负相关影响范围主要分布在东北三省及东南沿海地区. 负相关显著的东北部地区相关系数最高可达到-0.60,意味着风速每增大1,PM₂₅和 PM₁₀的浓度可降低 0.60. 然而,在四季,风速对陕西、甘肃等地区污染物浓度呈正效应^[23];在夏秋两季,风速对四川盆地也呈现正效应.

风速对 PM_{2.5}和 PM₁₀浓度主要呈负效应,这与已有研究的结果一致^[37].北方地区的负影响程度高于南方地区,由于北方地势平坦,为大气污染物提供了良好的扩散条件,风速的增大有利于污染物的扩散,从而改善空气质量,尤其对相关系数最高的东北平原更为显著.但风速对陕西、甘肃等地区的污染物浓度均具有正效应,原因是该地区地处黄土高原,土质较为疏松,当风速达到一定程度时,虽然有利于空气中细颗粒的扩散,但同时也会卷起地面粉尘,加重了大气污染^[23].夏秋两季风速对四川盆地的污染物呈正相关,原因是该地区地形封闭,年均风速低且气流难以扩散,易导致污染物聚集;且夏季气温偏高时,光化学反应速率加快,增加硫酸盐和硝酸盐颗粒,污染物二次来源增多.

(4)坡度 坡度(X4)对 PM₂₅和 PM₁₀浓度的影响存在显著的空间差异.东南部大部分地区呈现负相关,即坡度越大,PM₂₅和 PM₁₀浓度越小;东北部地区呈正相关,即坡度越大,PM₂₅和 PM₁₀浓度越大.不同季节存在显著差异,春季负影响范围最大,冬季最小.由春季到冬季,负影响的高值区逐渐由长三角地区南移到珠三角地区;正影响的高值区逐渐由京津冀地区北移到东北地区.夏秋两季坡度对污染物负影响强度远高于春冬两季,秋季 PM₂₅的负影响高值区主要集中在长三角地区,负相关系数达到-1.20,意味着坡度每增加 1,污染物浓度将下降1.20.

西南大部分地区坡度对污染物浓度影响呈显著负效应.四川、云南和重庆等地空气污染主要集中在坡度较小的盆地,盆地外围山脉对气流形成阻挡效应,导致气流无法快速扩散,加重了空气污染程度.然而,东北平原和华北平原,地形平坦,坡度小,有利于污染物扩散,所以局部地区的污染物浓度较低.黄河"几"字湾北部的鄂尔多斯和榆林,其地形对PM₂₅和PM₁₀浓度呈显著正效应,原因是太行山、阴山和贺兰山将以上地区包围,受高大山脉的阻挡,导致气流在此处难以扩散,加剧污染程度.

(5)人类活动强度 人类活动强度(X5)对地区 PM₂₅和 PM₁₀浓度的影响在大部分地区呈现正相关,即人类活动强度越大,PM₂₅和 PM₁₀浓度越大(图 2). 在春夏秋冬季 PM₂₅正相关影响范围占比分别是 78.33%、68.06%、70.72%和 86.70%,PM₁₀正相关影响范围分别是 65.02%、67.30%、70.72%和 100%.人类活动强度对 PM₂₅和 PM₁₀浓度影响呈负相关的地区主要分布在长三角和珠三角地区. 从春到冬正相关的范围逐渐扩大;正效应高值区逐渐向北移动. PM₂₅浓度正效应高值区主要集中在中原城市群、山东半岛城市群、京津冀城市群、长江中游和西南地区. PM₁₀与 PM₂₅浓度正效应高值区分布格局基本相似,也主要集中在中原城市群和东北部等地区.

本文借鉴徐勇等。28]提出的人类活动强度计算方 法,从土地利用类型的视角来判断人类活动对大气 环境的干扰程度.人类活动强度对PM25和PM10浓度 的影响较大,这与已有研究中其他社会因素对大气 污染产生较大影响一致[30, 37]. 西南地区的宜宾、攀 枝花和重庆等地区工业较为发达.2018年官宾市工 业增加值比上年增长9.80%,其中火电比上年增长 了27.30%; 2018年重庆电力、热力、燃气及水生产 和供应业增长8%,黑色金属冶炼和压延加工业增 长43%.可见,西南地区工业快速发展尤其是火电和 黑色金属冶炼等高耗能产业的快速增长加大该地区 的污染物浓度. 中原城市群和山东半岛城市群的人 口密集、工业生产规模大,生活生产排放的污染源 较多,导致人类活动与污染物呈正相关.负值区域 主要分布在东南沿海地区,该地区受东南季风的影 响,季风带来海洋水汽,形成丰富的降水,对污染 物有冲刷作用,气象因素对东南沿海的影响程度远 高于人文因素.

(6)二产占比 二产占比(X6)对 PM₂₅和 PM₁₀浓度的影响呈正相关,即二产占比越大,PM₂₅和 PM₁₀浓度越大.不同季节二产占比对污染物浓度的影响存在显著差异.PM₂₅的正效应高值区,由春夏向秋冬季,从西南地区转变为京津冀地区;PM₁₀的正效应

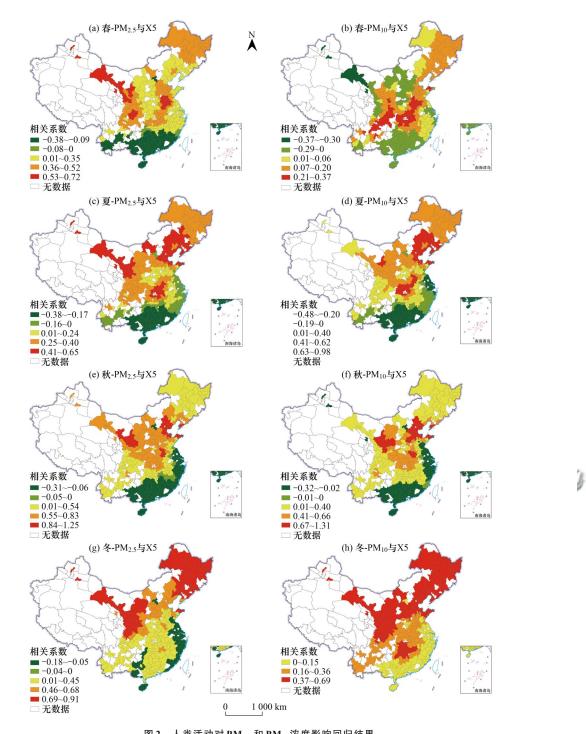


图 2 人类活动对 PM_{2.5}和 PM₁₀浓度影响回归结果

Fig. 2 $\,$ Regression results of the effect of human activities on $PM_{2.5}$ and PM_{10} concentrations

高值区,夏季主要集聚在西南地区,春秋冬三季主要分布在京津冀地区(图 3). 夏季西南地区 PM₂₅和 PM₁₀浓度的正相关系数可达到 0.25 和 0.34,表示二产占比每增加 1,污染物浓度将分别上升 0.25 和 0.34.

二产占比与污染物浓度成正比这一结论与已有研究结果一致^[37-39]. 工业生产过程直接或间接向大气环境排放 PM₂₅. PM₂₅的直接排放源中,贡献较大的工业部门主要为冶金、建材和化工,特别是炼焦、钢铁、水泥和砖瓦等行业. 以上工业部门排放

PM₂₅的多少与其工艺水平和管理水平密切相关. PM₂₅也会由硫和氮的氧化物转化而成,以上污染物主要来自化石燃料和垃圾的燃烧. PM₁₀一方面来自污染源的直接排放,如烟囱与车辆、工业生产中材料的破碎碾磨;另一方面来自硫氧化物、氮氧化物和挥发性有机化合物等互相作用形成的细小颗粒物.工业比重提高,"三废"及烟粉尘等污染物加重,导致地区的PM₂₅和PM₁₀浓度上升.

(7)能源消耗总量 能源消耗总量(X7)对 PM_{2.5}

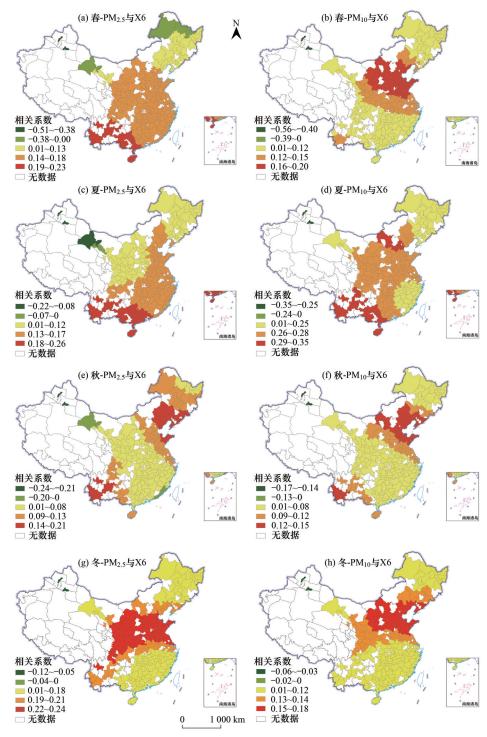


图 3 二产占比对 $PM_{2.5}$ 和 PM_{10} 浓度影响回归结果

Fig. 3 Regression results of the effect of the proportion of secondary production on $PM_{2.5}$ and PM_{10} concentrations

和 PM₁₀浓度的影响在大部分地区呈正相关,即能源消耗总量越大, PM₂₅和 PM₁₀浓度越大(图 4). 春夏秋三季污染物受能源消耗总量的正向影响更显著,春夏秋冬 PM₂₅正相关影响范围占比分别是 72.24%、86.69%、65.78% 和 38.03%, PM₁₀正相关影响范围分别是 51.71%、74.91%、58.17% 和 25.48%. 秋冬两季京津冀、山东半岛的 PM₂₅和 PM₁₀浓度回归系数呈负值,这与京津冀地区交通限行、发展新能源车及大

气污染防治等政策影响有关.

秋冬两季京津冀、山东半岛的 PM₂₅和 PM₁₀浓度 回归系数呈负值,这与2017年以来这些地区的污染 防治政策密切相关.自2017年北京陆续淘汰国一、 国二和国三机动车,积极发展新能源车,优化交通 结构;实施了《〈京津冀及周边地区2017-2018年秋 冬季大气污染综合治理攻坚行动方案〉北京市细化 落实方案》;采取"煤改电、气"的方式调整能源

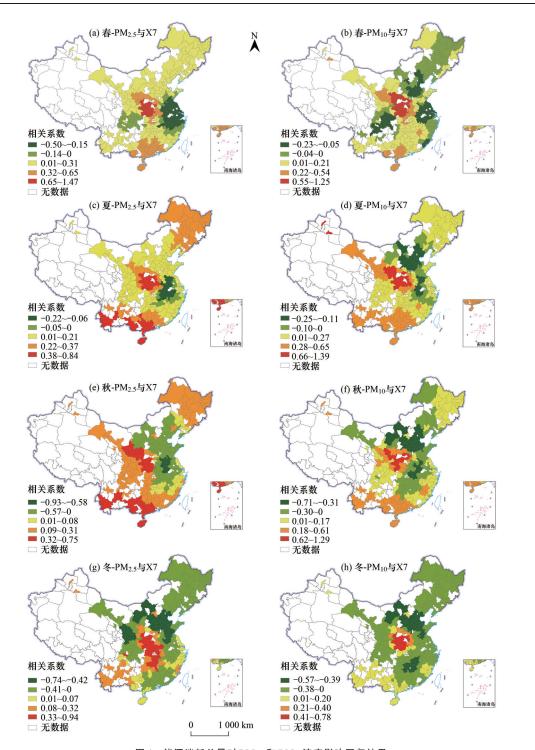


图 4 能源消耗总量对 PM_{2.5}和 PM₁₀浓度影响回归结果

Fig. 4 $\,$ Regression results of the effect of total energy consumption on $PM_{2.5}$ and PM_{10} concentrations

结构,降低能源消耗对大气环境的影响;2017年,将北京、天津及河北、山西、山东和河南的26个城市("2+26"城市)确定为京津冀大气污染传输通道,作为区域大气污染治理的重点;京津冀、山东半岛和东北老工业基地工业发展历史悠久,集中处理工业污染效果良好,可有效降低能源消耗对大气环境的影响.东南沿海部分地区污染物浓度相关系数也呈负值,但该地区的模型拟合效果较差,说明能源消耗不是该地区大气环境污染的主要影响因素.

3 讨论

大气污染物浓度受自然环境和人类社会经济活动共同影响,不同季节、不同地区各类自然环境和人类社会经济活动对大气污染物浓度的作用方向、影响程度和范围不尽相同.研究中需要从自然地理环境因素和人文经济社会因素综合构建影响因素体系.已有研究更多关注自然要素对大气污染物浓度的影响,这些要素主要包括气温、降水、风速和坡

度等.本研究除了上述自然要素之外,一方面观测基于土地利用类型的人类活动强度对大气污染物浓度的影响,另一方面观测工业发展、能源消耗对大气污染物浓度的影响.事实上,量化人类社会经济活动的指标有很多,诸如人口密度、经济发展水平和城镇化水平等.本研究尝试把上述所有表征人类社会经济活动的指标纳入回归模型,然而人口密度、经济发展水平和城镇化水平等指标拟合效果差.这说明区域大气污染浓度的主要社会因子并非人口密度、经济发展水平和城镇化率,举例来说,两个人口密度或经济发展水平和城镇化率,举例来说,两个人口密度或经济发展水平相近的地区,却存在大气污染浓度差距,这种差距的成因可能主要来自于自然要素或人类社会经济活动中的工业发展、能源消耗和人类活动强度差距.

本研究发现具体影响因子在不同时段、不同地区的影响程度不尽相同,由于本研究的时段较短(仅为一年四季),今后的研究有待拉长时间尺度,充分考虑各因子的时间演进规律和地域差异,为PM₂₅和PM₁₀污染防治提供因时因地制宜的防治措施.本研究发现,京津冀地区、中原城市群等人口密集区域受二产占比影响较大,未来PM₂₅和PM₁₀污染治理应优化产业结构,促进产业转型升级.甘肃和宁夏的部分城市,受地表扬尘的影响,需在重点时节做好提前防控,如风速较大时,限制部分工业材料碾磨.对于北方冬季易发生逆温天气的区域,需在逆温天气时减少工业生产活动.

4 结论

- (1) PM_{2.5}和 PM₁₀的浓度均呈现冬春高、夏秋低的季节性规律. PM₁₀的季节性空间差异明显,冬季黄河中下游地区、京津冀、新疆中南部和华中地区浓度较高;春季浓度较高的地区主要位于新疆、甘肃和宁夏等地;夏季 PM₁₀浓度最低,尤其是长江中下游和东南沿海地区的颗粒物浓度较低.
- (2)PM_{2.5}和PM₁₀浓度的 Moran's *I*在四季均为正, 且均在冬季增至峰值,说明城市间两类污染物浓度 差距在冬季最大.PM_{2.5}和PM₁₀浓度的分布格局基本 一致,"高-高"类主要分布在华北、京津冀、黄河 中下游和新疆部分地区,"低-低"类主要分布在长 江以南、西南地区、东北部的呼伦贝尔和齐齐哈尔 等,"高-高"类和"低-低"类集中分布现象明显.
- (3)用GWR、OLS对PM₂₅和PM₁₀浓度影响因素分析发现,GWR模型拟合效果优于OLS,PM₁₀和PM₂₅拟合效果基本一致,均表现为春秋冬三季的拟合效果好,夏季效果差,模型的显著性水平依次是冬>秋>春>夏.

(4)四季各影响因子对 PM₂₅、PM₁₀浓度变化存在较大空间异质性.自然因素中温度和坡度对 PM₂₅、PM₁₀浓度的影响基本一致,既存在正影响也存在负影响,呈南北分化的格局.降水和风速对大部分地区 PM₂₅和 PM₁₀浓度呈负相关.人类活动强度、二产占比和能源消耗总量对 PM₂₅和 PM₁₀浓度的影响在大部分地区呈正相关.

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