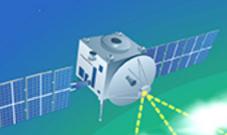


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ENVIRONMENTAL SCIENCE

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PM_{2.5}和O₃污染协同防控区的遥感精细划定与分析 李沈鑫,邹滨,张凤英,刘宁,薛琛昊,刘婧



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PM_{2.5}



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海南岛半干旱区农用地土壤重金属富集因素、健康风险及来源识别

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摘要:为了解海南岛半干旱区农业土壤中重金属富集因素和污染状况,在感城镇采集 1818 件表层土壤样品,测定其重金属含量和化学组成.采用相关分析、地累积指数 $(I_{\rm geo})$ 、综合生态风险指数 (RI)、危害指数 (HI)、致癌风险指数 (CR) 和正定矩阵因子分析 (PMF) 开展重金属风险评价和来源识别. 结果显示,重金属 $\omega(As)$ 、 $\omega(Cd)$ 、 $\omega(Cr)$ 、 $\omega(Cu)$ 、 $\omega(Hg)$ 、 $\omega(Ni)$ 、 $\omega(Pb)$ 和 $\omega(Zn)$ 的平均值分别为 22. 7、0. 128、33. 4、14. 5、0. 032、9. 32、32. 5 和 43. 3 mg·kg⁻¹,除 Zn 外,均高于海南岛土壤背景值. 相关分析表明,重金属富集与土壤中 Fe、Mn、Al 和有机质含量密切相关. $I_{\rm geo}$ 结果表明,研究区农业土壤主要受到 As的污染,其次为 Cd 和 Cu; RI 结果显示,高风险以上的样品占比为 29. 4%,其中 As 是潜在生态风险的主要贡献者;健康风险评估结果显示,As、Cr 和 Ni 对儿童存在致癌风险,需要引起注意. 基于 PMF 模型,确定了研究区重金属的 4 种主要来源,其中Hg 主要来自工业排放;As 主要来自农业活动;Ni、Cu、Cr 和 Zn 主要来自与成土母质密切相关的自然来源;Pb 和 Gd 主要来自农业活动和机动车尾气的混合源. 研究表明 PMF 模型与相关分析相结合,能够有效识别土壤重金属来源.

关键词:半干旱区; 土壤重金属; 生态风险; 健康风险; 正定矩阵因子分析(PMF)

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Enrichment Factors, Health Risk, and Source Identification of Heavy Metals in Agricultural Soils in Semi-arid Region of Hainan Island

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Abstract: To understand the enrichment factors and pollution levels of heavy metals in agricultural soils in the semi-arid region of Hainan island, 1818 surface soil samples were collected in Gancheng Town and analyzed for their heavy metal contents and physicochemical composition. Correlation analysis was used to determine the heavy metal enrichment factors. The geo-accumulation index $(I_{\rm geo})$, comprehensive ecological risk index (RI), and hazard index (HI), as well as carcinogenic risk (CR), were used to assess the degree of pollution and health risk. Positive matrix factorization (PMF) was used to determine the primary sources of pollution and priority sources. The average values of heavy metal contents in the topsoil were 22.7, 0.128, 33.4, 14.5, 0.032, 9.32, 32.5, and 43.3 mg·kg⁻¹ for As, Cd, Cr, Cu, Hg, Ni, Pb, and Zn, respectively. With the exception of Zn, the concentrations of other heavy metals in the topsoil were higher than the soil background values of Hainan, showing different degrees of heavy metal accumulation effect. The $I_{\rm geo}$ revealed that the major pollutant element in soils was As, followed by Cd and Cu. The RI showed that the proportion of soil samples that were high-risk level or worse was 29.4% of the total number of samples, among which As was the major source of risk. The health risk assessment results indicated that As, Cr, and Ni exposure presented carcinogenic risk for children with high CR values. Based on PMF, four major sources of heavy metals were identified in the study area. Hg was derived mainly from industrial sources, and As was closely associated with agricultural activities. Ni, Cu, Cr, and Zn were related to soil parent materials. Pb and Cd were associated with agricultural activities and traffic emissions. The PMF models combined with correlation analysis were useful for estimating the source apportionment of heavy metals in soils.

Key words; semi-arid region; soil heavy metal; ecological risk; health risk; positive matrix factorization (PMF)

土壤是人类生存的重要物质基础,随着人口的快速增长,特别是城市化和工业化的发展,土壤污染问题越发严峻[1].由于农用地土壤可以通过食物链直接影响公众健康,因此,土壤污染,特别是重金属污染,引起了全世界的广泛关注[2,3].我国第一次耕地调查结果显示,土壤点位超标率高达19.4%,其中重金属是主要污染物[4].而在诸多干旱-半干旱农业区,由于长期使用地下水乃至废水灌溉,使得土壤重金属污染问题更为显著[5,6].重金属具有难降解、高毒性和生物累积性等特点,可进一步造成水体、大气和农作物的污染,进而危害生态系统和人类健

康^[7~10],因此,了解干旱-半干旱区农用地土壤重金 属富集因素,并开展重金属污染评价十分必要.

为了解重金属累积引起的环境风险,前人提出多种评价体系. 其中地累积指数 $(I_{geo})^{[11]}$ 、浓度富集指数 $(CEF)^{[12]}$ 和生态风险指数 $(RI)^{[13]}$ 通过与对应重金属的地球化学背景值、基准值或风险阈值等

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对比来确定土壤污染情况,而危害系数(HI)和致癌风险(CR)^[14,15]通过计算居民重金属摄入量来评估健康风险.然而,仅评价重金属的风险状况,无法准确识别重金属来源,对重金属污染防控的导向性较差.为此,聚类分析(CA)^[16]、主成分分析(PCA)^[17]、相关分析^[18]和正定矩阵因子分析(positive matrix factorization,PMF)^[19]等方法被用于辨别污染物来源.其中,PMF模型是一种利用相关矩阵对变量进行降维分析的受体模型,能够有效识别污染源和定量获得非负源贡献率而广泛应用于土壤重金属污染的朔源^[19,20].

海南岛主体气候温润、雨量充沛,而位于海南岛西南部的东方市,由于年蒸发量远大于降雨量,被划分为半干旱区而区别于海南岛其它地区^[21].长期高强度的人工灌溉、施肥和除虫等农业活动可能会导致土壤重金属的累积^[5,22-25],伴随着脱贫攻坚任务的完成,农用土地集约化程度进一步提高,也可能增加农用地土壤重金属生态风险.尽管有学者对海南岛内农业土壤进行了综合评价^[26,27],但是缺少在半干旱区重金属污染的研究,对区内土壤环境污染现状认识不足.为此,本次工作通过系统采集区域内土壤样品,分析东方市农业区土壤中8种重金属元素(As、Cd、Cr、Cu、Hg、Ni、Pb和Zn)的污染情况和主要富集因素,评价重金属生态风险和健康风险,并利用PMF模型识别污染源,以期为海南岛半干旱

区农用地土壤重金属污染防控提供科学依据.

1 研究区概况

调查区位于海南省东方市感城镇(108°36′~108°53′E,18°47′~18°55′N),调查面积182 km². 地势东高西低,东部地貌以低矮丘陵为主,西部为临海平原区.区内土壤类型以燥红壤为主,同时发育少量砖红壤、水稻土和冲积土.土地利用方式主要为园地,其次为耕地,其中园地以种植芒果为主,多分布在低矮丘陵区,耕地多分布在河流两岸,以种植水稻、花生和蔬菜为主.研究区属热带季风海洋性气候区,平均降雨量900 mm·a⁻¹,蒸发量2596 mm·a⁻¹,日照时间为2558 h·a⁻¹.国内日照时长仅次于西藏,被誉为热带瓜果基地.

2 材料与方法

2.1 采样原则

本次工作参考东方市农用地分布进行采样点布设,采样点位置需距离道路 50 m以上,采样深度为0~20 cm,采样密度约 10点·km⁻²(图 1),共采集土壤样品 1818 件. 原始样品由预布点周围 10~15 m范围内 3个子样等量混合而成,质量约 1.5 kg. 每件样品在去除地表落叶、碎石块等杂物后装入干净样品袋中,于阴凉处风干后过 10 目尼龙筛,缩分后称取 200 g 用于测试.



图 1 研究区地理位置和采样点示意

Fig. 1 Location of the study and sampling sites

2.2 样品分析与质量监控

土壤样品的分析测试在中国地质科学院地球物理地球化学勘查研究所测试中心完成. 样品前处理、分析检测过程和质量监控参考文献[28]. 根据不同的测试指标,选取合适的分析方法,其中全氮(TN)的测定采用氧化燃烧-气相色谱法(OC-GC),Al₂O₃、TFe₂O₃、CaO、SiO₂、全磷(TP)和全钾(TK)的测定采用 X 射线荧光光谱法(XRF),As 和 Hg 的测定采

用原子荧光光谱法(AFS),MgO、Na₂O、Zn 和 Cr 的 测定采用等离子体发射光谱法(ICP-OES),Cd、Cu、Pb 和 Ni 的测定采用电感耦合等离子体质谱法 (ICP-MS),pH 的测定采用电位法,有机质采用高 频燃烧-红外碳硫仪进行测定. As、Cd、Cr、Cu、Hg、Ni、Pb 和 Zn 的检出限分别为 0.5、0.02、3、1、0.0005、1、1.5 和 2 mg·kg⁻¹,Al₂O₃、TFe₂O₃、CaO 和 SiO₂ 的检出限是 0.05%,Na₂O 和有机质的

检出限是 0.1%, TN、TP 和 TK 的检出限分别是 0.02、0.01 和 0.4 g·kg⁻¹. pH 测定采用 GpH-1、GpH-4、GpH-7 和 GpH-10 这 4 种国家一级标准物质进行分析, 其余指标采用 4 种国家一级标准物质(GSS-3a、GSS-7a、GSS-39、GSS-44) 和重复分析样进行质量监控. 结果显示, 8 种重金属和对应的化学指标报出率、相对标准差和加标回收率均优于国家标准^[29].

2.3 评价模型

2.3.1 地累积指数

地累积指数 (I_{geo})^[11]最初用于沉积物污染识别,随后广泛应用于土壤污染划分,其计算公式为:

$$I_{\text{geo}} = \log_2 \left[C_i / (1.5 \times C_b) \right]$$
 (1) 式中, C_i 为土壤中重金属的实测值; C_b 为对应重金属的土壤背景值,本文选用海南岛表层土壤背景值。100 进行计算。根据 I_{geo} 值可以将污染划分为 7 个等级[11],分别为:无污染($I_{\text{geo}} \leq 0$)、无污染至中度污染($0 < I_{\text{geo}} \leq 1$)、中度污染($1 < I_{\text{geo}} \leq 2$)、中度污染至重度污染($2 < I_{\text{geo}} \leq 3$)、重度污染($3 < I_{\text{geo}} \leq 4$)、重度污染至极度污染($4 < I_{\text{geo}} \leq 5$)和极度污染($I_{\text{geo}} > 5$).

2.3.2 潜在生态风险指数

潜在生态风险指数(RI)^[13]可用于评估重金属对生态系统的潜在危害. 计算方法如下:

$$EI_{i} = (C_{i}/C_{b}) \times CF^{i}$$

$$RI = \sum_{i=1}^{n} EI_{i}$$
(3)

式中, EI_i 为重金属 i 的潜在生态风险指数, CF^i 为该重金属的毒性响应因子(即 As = 10、Cd = 30、Cr = 2、Cu = 5、Hg = 40、Ni = 5、Pb = 5、Zn = 1) $[^{223}$. EI 值可分为 5 个等级 $[^{133}]$:低水平(EI < 40)、中等水平($40 \le EI$ < 80)、高水平($80 \le EI$ < 160)、极高水平($160 \le EI$ < 320)和危险水平(EI ≥ 320)生态风险. RI 为综合潜在生态风险指数,为各重金属潜在生态风险指数之和.该指标分为低风险(EI < 150)、中风险(EI ≥ 300)、强风险(EI ≥ 300)和极高风险(EI ≥ 300)。

2.3.3 非致癌风险

HI 指数用于评价人体摄入重金属所导致的非致癌风险. 通过计算经口、皮肤和呼吸摄入重金属量与参考摄入量对比来计算风险. 这 3 种摄入途径的计算方法为:

$$ADD_{\text{ingest}} = \frac{C_i \times \text{IngR} \times \text{EF} \times \text{ED}}{\text{BW} \times \text{AT}} \times 10^{-6} \quad (4)$$

$$ADD_{\text{dermal}} = \frac{G_i \times \text{SA} \times \text{AF} \times \text{ABS} \times \text{EF} \times \text{ED}}{\text{BW} \times \text{AT}} \times 10^{-6} \quad (5)$$

$$ADD_{\text{inhalation}} = \frac{C_i \times \text{APM} \times \text{InhR} \times \text{EF} \times \text{ED}}{\text{BW} \times \text{AT}} \times 10^{-6} \quad (6)$$

式中, ADD_{ingest} 、 ADD_{dermal} 和 $ADD_{inhalation}$ 分别为消化吸收、皮肤接触和呼吸摄入的重金属量, C_i 为实测土壤中重金属含量 $(mg \cdot kg^{-1})$. 其余参数的含义和参考值见表 1.

表 1 重金属暴露指标 [2,31,32]

Table 1 Exposure parameters for the heavy metals

	Tubic 1	Exposure parameters for the ne	avy metais	
指标	含义	单位	成人	儿童
IngR	土壤摄人率	mg•d ⁻¹	100	200
EF	暴露频率	d•a⁻¹	350	350
ED	暴露年限	a	26	6
BW	平均体重	kg	70	15
AT	平均暴露时间	d	$ED \times 365$	$ED \times 365$
SA	暴露皮肤表面积	cm^2	5 700	2 800
AF	皮肤黏附因子	$mg \cdot (cm^2 \cdot d)^{-1}$	0. 2	0. 2
ABS	皮肤吸收因子	_	0.001	0.001
APM	单位体积微粒量	mg⋅m ⁻³	0.0651	0.0651
InhR	日空气吸入量	$m^3 \cdot d^{-1}$	14. 5	7.5

非致癌风险计算方法为:

$$HQ = \frac{ADD}{RfD} \tag{7}$$

$$HI = \sum HQ \tag{8}$$

式中,HQ 为重金属不同暴露途径的危害指数; RfD 为此重金属对应暴露途径的参考计量,具体参考值 见表 2. HI 用于表征重金属的非致癌风险,当 HI 值 小于 1,表示非致癌风险水平为可接受范围,反之则

为不可接受范围.

2.3.4 致癌风险

致癌风险用于表征个体由于重金属暴露所导致 的终身致癌指数^[14],其计算方法为:

$$CR = \sum ADD \times SF$$
 (9)

式中,CR 为重金属的致癌风险,SF 为对应重金属不同暴露途径下的致癌转换因子,具体参考值见表 2.

衣4 里亚属多方儿里州以照科学囚丁	表 2	重金属参考计量和致癌斜率因子	2,31,3
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Table 2	Reference	dose	and	slope	factor	of	heavy	metals

重金属	参考	計量 RfD/mg・(kg・c	d) -1	致	癌因子 SF/(kg·d)・	mg ⁻¹
里並馮	经口	皮肤	呼吸	经口	皮肤	呼吸
As	3.0×10^{-4}	1. 23 × 10 ⁻⁴	4. 29 × 10 ⁻⁶	1. 5	1. 5	1. 51 × 10 ¹
Cd	1.0×10^{-3}	2.5×10^{-5}	2.86×10^{-6}	_	_	6. 3
Cr	3.0×10^{-3}	3×10^{-5}	2.86×10^{-5}	5.01×10^{-1}	2.0×10^{1}	4.2×10^{1}
Cu	4.0×10^{-2}	1.2×10^{-2}	_	_	_	_
Hg	3.0×10^{-4}	2. 14×10^{-5}	_	_	_	_
Ni	2.0×10^{-2}	5.4×10^{-3}	9.0×10^{-5}	1.7	4.25×10^{1}	8. 4×10^{-1}
Pb	1. 4×10^{-3}	5. 24×10^{-4}	_	8. 5×10^{-3}	_	4.2×10^{-2}
Zn	3.0×10^{-1}	6. 0×10^{-2}	_	_	_	_

其中 CR ≥ 1 × 10⁻⁴表示致癌风险为不可接受水平; CR < 1 × 10⁻⁴则表示致癌风险为可接受水平^[14].

2.4 数据处理

数据处理使用 Office Excel 2016,进行土壤重金属含量的均值、最大值、最小值、标准差和变异系数等基本描述性统计分析. 相关分析和箱形图采用Origin 2021 进行运算和制作. 土壤重金属元素的来源及其贡献率采用 EPA PMF 5.0 模拟. 采用 Arcgis 10.7 对各影响因子进行空间分析,通过普通克里格法对因子进行空间插值.

3 结果与讨论

3.1 土壤理化性质和重金属含量

研究区土壤 pH 值为 3.55 ~ 8.66, 平均值为 5.72(表 3),土壤 pH < 5.0 的样品有 396 件,占总调查样品数量的 21.8%.一般情况下,由于水热条件不同,干旱区土壤 pH 要高于湿润区^[33],而研究区土壤 pH 较低可能与大量施肥或者富含硫化物矿物氧化造成的土壤酸化有关.海南岛主体水热充足,而研究区干旱缺水,土壤淋溶速率低,可溶性离子Na⁺、Ca²⁺和 K⁺易沉积,因此土壤中 Na₂O、CaO 和TK 水平高于海南表层土壤整体水平,而 Al₂O₃、TFe₂O₃ 和 MgO 含量相对偏低(表 3),反映了干旱气候条件对土壤化学成分的影响.此外,土壤中 TN、TP 和 TK 含量相对偏高(表 3),也反映了农业活动对土壤化学组成的影响.

研究区土壤中 8 种重金属相关统计参数显示 (表 3),重金属 ω (As)、 ω (Cd)、 ω (Cr)、 ω (Cu)、 ω (Hg)、 ω (Ni)、 ω (Pb) 和 ω (Zn) 的平均值分别为 22.7、0.128、33.4、14.5、0.032、9.32、32.5 和 43.3 mg·kg⁻¹,除 Zn 外,均高于海南岛表层土壤背景值(表 3),表明重金属存在不同程度的累积效应.对比《土壤环境质量农用地土壤污染风险管控标准》(GB 15618-2018)规定的筛选值,As、Cd、Cu 和 Pb 的超标率分别为 13.6%、6.5%、2.9% 和 2.1%,其余 4 种重金属超标率均小于 0.6%.8 种重金属

的变异系数范围为 70.8%~256.5%,由低到高排序为: Cr < Pb < Ni < Cu < Zn < Hg < As < Cd,均属于高度变异,尤其是 As、Cd 和 Hg 的变异系数均超过147%,表现出更强的离散程度,表明它们在土壤中的含量变化大,受区域背景和人为因素等空间差异影响大.

3.2 土壤化学组成对重金属富集的影响

通过 Pearson 相关系数分析土壤中氧化物、TN、 TP、TK、有机质和 pH 对重金属富集的影响,结果 见表 4. 通常情况下, 土壤中 Al,O, 与 SiO, 比值越 高,土壤黏粒含量越高,土壤质地越细[36]. 表 4 显 示,研究区土壤中 Al,O, 与 8 种重金属呈显著正相 关,SiO₂ 与8种重金属呈显著负相关,重金属随着土 壤 Al,O,增加而升高,随着 SiO,增加而降低,表明土 壤重金属富集与土壤黏粒和土壤质地密切相关,这 与前人的研究结果一致[37]. 土壤中 Al₂O₃ 一定程度 上也反映出黏土矿物的含量,伴随黏土矿物和有机 质的增加,土壤对重金属的吸附量也会增加.研究区 土壤中 TN、TP 和 TK 与 8 种重金属呈显著正相关 关系(表4),有研究表明,施用不同剂量有机肥,会 导致土壤重金属含量增加,且随有机肥施用量的增 加而增大[38]. 研究区土壤 TN 和 8 种重金属, TK 和 Pb 均具有较高的相关系数(r > 0.530, P < 0.01), TP 和 8 种重金属也呈显著正相关关系(P < 0.01), 表明土壤中 TN、TP、TK 的升高能促进重金属的 富集.

土壤中 TFe_2O_3 和 Mn 与 Cr、Ni、Cu 和 Zn 等元素也具有较高的相关系数(表 4 和图 2),表明这些重金属容易被富含 Fe、Mn 的矿物吸附或以类质同象进入矿物晶格中,这是引起土壤重金属富集的主要因素[37]. 例如在中等酸性条件下, As 能以 $(AsO_4)^{3-}$ 及其共轭酸的形式被吸附在含铁矿物表面,或形成络合物形式与 $MnFe_2O_4$ 相结合[37]. 相比之下,土壤 pH 与重金属的相关系数较小(表 4),表明研究区 pH 对重金属富集程度影响较弱. 红壤相

对于其它土壤发生类型,通常具有较高的黏土矿物 和铁锰结核,这些矿物不仅富含 Al、Fe、Mn 元素, 重金属元素也异常富集[33]. 研究区 8 种重金属元素 与 Al₂O₃、TFe₂O₃、Mn 和有机质均为显著正相关关 系(表4),因此推测,研究区土壤含 Fe、Mn、Al 矿 物和有机质是重金属富集的主要因素.

表 3 研究区土壤重金属及其它理化组成基本统计参数1)

Table 3	Basic statistics of heavy	metals and other	physicochemical	components in soils of the study are	а

项目	pН	As	Cd	Cr	Cu	Hg	Ni	Pb	Zn	Mn
最小值	3. 55	0.660	0.002	0.008	1. 258	0. 005	1. 082	4. 745	4. 395	31.0
最大值	8. 66	1161.3	7. 821	201. 1	169. 9	1. 273	125.8	423.4	1 077. 3	5 019. 1
平均值	5. 72	22.7	0. 128	33.4	14. 5	0.032	9. 32	32. 5	43.3	366. 2
中位值	5. 64	8.66	0.069	26.0	9.64	0.024	6.07	30. 5	33.0	247.6
变异系数	16. 5	227.8	256. 5	78. 2	108.0	147. 4	106. 0	70.8	108.9	104. 1
琼中[26]	_	3. 30	0.063	49. 9	9. 93	0. 035	13. 4	33.4	60. 1	
海南土壤背景[30]	_	1. 34	0.04	27.5	6. 1	0.02	7. 24	24. 4	44. 4	163.5
全国表层土壤背景[34]	_	11	0.097	61	23	0.065	27	26	74	585
	< 5.5	40	0.3	150	50	1. 3	60	70	200	_
筛选值(GB 15618-2018) ^[35]	5. 5 ~ 6. 5	40	0.3	150	50	1.8	70	90	200	_
师是国(GD 13016-2016)。	6. 5 – 7. 5	30	0.3	200	100	2.4	100	120	250	_
	>7.5	25	0.6	250	100	3.4	190	170	300	1
项目	TN	TP	TK	有机质	Al_2O_3	CaO	Na ₂ O	$\mathrm{TFe_2O_3}$	MgO	SiO_2
最小值	0.06	0.07	0.98	0.78	1.41	0.04	n. d.	0.12	0.02	50.16
最大值	2.41	3.39	41.60	24.98	21.21	5.05	7.03	14.71	4.58	96.70
平均值	0.74	0.37	22.92	7.17	9.56	0.19	0.42	2.39	0.30	77.73
中位值	0.71	0.32	24.28	6.63	9.50	0.16	0.37	1.88	0.23	77.61
变异系数	49.1	61.8	45.6	50.3	41.5	99.2	109.3	77.1	87.9	9.9
琼中[26]	_	1 11/4) = y /		_	(A)	all -	1/2	_	7/
海南土壤背景[30]	0.20	0.197	19.75	_	18.19	0.11	0.04	3.05	0.31	65.86
全国表层土壤背景[34]	- ,) 40	19.10	2-	12.5	2.2	1.5	4.2	1.3	65
61	_	7#/	F - 1	_	_	1 1/2	100	1-	_ ~	\
筛选值(GB 15618-2018) ^[35]	_	1 1	1-4/	4	_	12	f - 1	1-	_	- A
лижент (ОН 15010-2010)		(*//	1-4	+\	_	11- 1	_	\J-	_	_
VA VILA I VE	5	1-9"	14-L	19	_	_	_	-9	_	

1) n=1 818; "—"表示没有相关数据; 其中 pH 值无量纲,重金属单位为 $mg \cdot kg^{-1}$, TN、TP、TK 和有机质单位为 $g \cdot kg^{-1}$, 其余单位为%

表 4 半干旱区土壤重金属与理化组成相关系数1)

Table 4 Pearson correlation coefficients between soil pH, chemical properties, and heavy metals

	As	Cd	Cr	Cu	Hg	Ni	Pb	Zn
TN	0. 554 **	0. 552 **	0. 726 **	0. 643 **	0. 619 **	0. 693 **	0. 530 **	0. 698 **
TP	0. 338 **	0. 383 **	0. 469 **	0. 510 **	0. 352 **	0. 407 **	0. 371 **	0. 513 **
TK	0. 275 **	0. 268 **	0. 326 **	0. 205 **	0. 243 **	0. 211 **	0. 682 **	0. 395 **
Mn	0. 629 **	0. 646 **	0. 704 **	0. 676 **	0. 418 **	0. 753 **	0. 399 **	0. 711 **
有机质	0. 564 **	0. 548 **	0. 723 **	0. 650 **	0. 614 **	0. 704 **	0. 478 **	0. 680 **
pН	0. 039 *	0. 159 **	0.016	0. 051 *	0. 107 **	0. 070 **	0.014	0. 104 **
$\mathrm{Al_2O_3}$	0. 615 **	0. 527 **	0. 785 **	0. 640 **	0. 478 **	0. 716 **	0. 720 **	0. 754 **
CaO	0. 327 **	0. 473 **	0. 419 **	0. 409 **	0. 396 **	0. 417 **	0. 466 **	0. 521 **
$\mathrm{TFe_2O_3}$	0. 736 **	0. 623 **	0. 961 **	0. 849 **	0. 511 **	0. 912 **	0. 526 **	0. 833 **
MgO	0. 695 **	0. 618 **	0. 898 **	0. 792 **	0. 530 **	0. 881 **	0. 549 **	0. 852 **
Na_2O	0. 252 **	0. 258 **	0. 309 **	0. 146 **	0. 206 **	0. 189 **	0. 577 **	0. 329 **
${\rm SiO}_2$	-0.638 **	-0.558 **	-0. 839 **	-0.735 **	-0.480 **	-0.803 **	-0.580 **	-0.764 **

1) n=1 818;除了 pH 值,均进行对数处理; ** 表示 P < 0.01, *表示 P < 0.05

3.3 土壤重金属污染程度

以海南岛表层土壤重金属含量为背景值,计算地 累积指数[图3(a)],结果表明,研究区土壤主要受到 As 污染,其次为 Cd 和 Cu,其余重金属污染程度较 低. 其中 As 元素中度污染、中度至重度污染和重度污 染级别以上的比例分别为 28.4%、25.7% 和 28.5%; Cd 元素中度污染、中度至重度污染和重度污染级别 以上的比例分别为 14.5%、5.2% 和 3.4%; Cu 元素 中度污染、中度至重度污染和重度污染级别以上的 比例分别为14.1%、5.0%和1.5%.

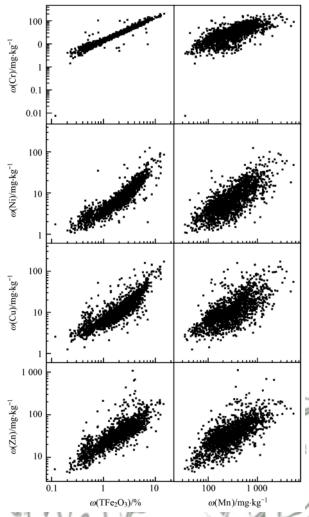


图 2 土壤 TFe₂O₃ 和 Mn 与 Cr、Ni、Cu 和 Zn 含量散点图 Fig. 2 Scatter diagrams of TFe₂O₃, Mn, Cr, Ni, Cu, and Zn in soils

潜在生态风险结果显示[图 3(b)], EI 中位值由大到小依次为: As (64.6) > Cd (51.9) > Hg (47.0) > Cu(7.9) > Pb(6.2) > Ni(4.2) > Cr(1.9) > Zn(0.74), 平均值降序排序与中位值一致,表明As、Cd和Hg是研究区重要的风险元素. 其中, As、Cd和Hg中等生态风险级别以上的比例分别为73.3%、66.9%和61.3%, 其余5种重金属在置信区间内均表现为低生态风险. 29.4%的土壤点位 RI值大于300,表现为强生态风险级别以上. 这部分土

壤中 As 是主要贡献者,平均贡献率为 50.1%.因此,研究区内土壤存在 As 污染,这也导致大量的土壤样品存在较强的重金属潜在生态风险,需要引起关注.

3.4 土壤重金属健康风险评价

基于经口、皮肤和呼吸摄入重金属的健康风险 评价模型通常用于场地(建设用地)土壤的风险评 估,但近些年广泛运用于矿区[14]、工业区[39]和农用 地[22,31]土壤的健康风险评价. 研究区土壤非致癌风 险评价结果显示[图 3(c)],成人 HI 值在置信区间 内小于1,表明重金属对成人不存在非致癌风险.对 于儿童,少量样品点 As 的 HI 值大于 1,需要引起注 意. 致癌风险结果表明,重金属对成人的致癌风险处 于可接受水平. 对于儿童, As、Cr 和 Ni 的致癌风险 (CR)平均值大于1×10⁻⁴,存在致癌风险,Cd和Pb 为可接受水平[图 3(d)]. 由于儿童特殊的生理和 行为特征,他们对单位重量污染物的作用更敏感,在 游戏和散步过程中更容易摄入含重金属的土壤微粒 而增加暴露风险[40]. 此外,致癌风险模型是基于人 体经消化吸收、皮肤吸收和呼吸摄入重金属微粒所 建立的风险模型,其结果直接取决于土壤重金属含 量,而风险模型中各系数的参考值却能够很大程度 上影响模型结果. 研究区土壤中 As、Cr 和 Ni 的中 位值均低于全国土壤背景值(表3),对儿童却存在 不可接受的致癌风险,这表明致癌风险模型中相关 参数在农用地健康风险评价过程中需要调整,或者 说全世界范围内土壤中 As、Cr 和 Ni 对儿童普遍存 在不可接受的致癌风险.

3.5 土壤污染源识别与分布

重金属的相关性分析能够初步判断它们是否具有同源性,为重金属来源提供信息.本文采用Pearson 相关系数表示重金属间的相关性,结果见表5. Cr、Cu、Ni、Zn和Cd这5种重金属间具有较高的相关系数,Pb和Cd也具有较高的相关系数,Pb和Cd也具有较高的相关系数,为为显著正相关关系,表明它们可能有相似的来源.而As、Hg和其它重金属相关系数相对较小,表明As和Hg各自具有不同的来源.

表 5 土壤重金属含量相关性分析1)

Pearson correlation matrix for the heavy metal concentrations in soil samples As Cd Cr Cu Hg Ni Ph Cd 0.341 * 0. 457 ** Cr0. 272 ** Cu 0.426 ** 0. 263 ** 0.826 ** Hg 0.358 ** 0.220 ** 0. 164 ** 0.165 ** 0.454 ** 0. 257 ** 0. 888 ** 0.771 ** 0. 142 ** Ni Pb 0.402 ** 0.609 ** 0. 254 ** 0. 193 ** 0. 241 ** 0. 155 ** Zn 0.617 ** 0.690 ** 0.528 ** 0.479 ** 0.515 ** 0.483 ** 0.605 **

1) ** 表示 P < 0.01

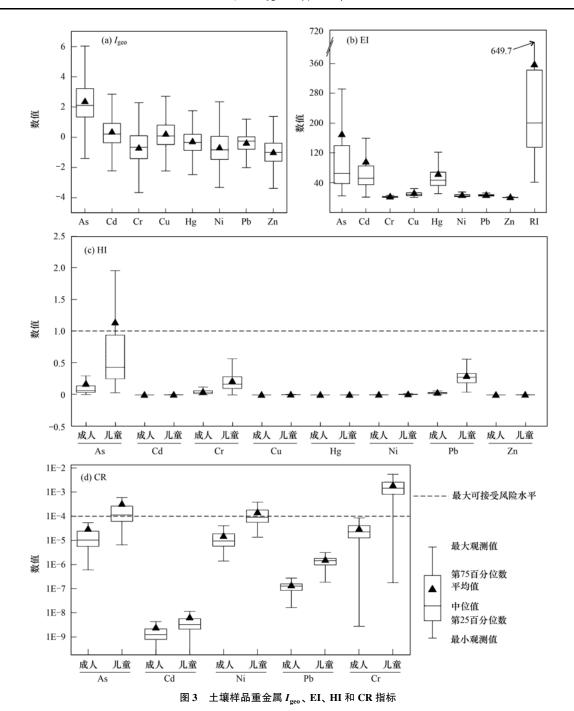


Fig. 3 The I_{geo} , EI, HI, and CR values of heavy metals in soil samples

PMF 模型是美国环保署基于主成分分析法建立的受体模型,是识别和定量分析土壤中重金属主要来源的有效工具. Wang 等^[2]介绍了该模型的原理和操作方法. 在本文中,8种重金属的信号强度均设置为强,信噪比(S/N)大于2.0,说明研究中的变量是可靠的. 将因子数设置在3~5之间,使用随机的初始点运算20次,对比不同结果,选取最小最稳定的Q值,以获得最佳因子个数. 当选取4个因子时,绝大部分样品残差位于-3~3之间,As、Cr、Cu、Hg、Ni和Pb的r²均大于0.78,表明PMF模型的结果是合理的. 东方市半干旱区土壤样品中重金属的来源及其贡献如图4和图5所示. 各因子的贡

献率大小依次为:因子 3 (45.2%) > 因子 4 (22.9%) > 因子 1 (16.4%) > 因子 2 (15.5%).因子 1 中 Hg 占据绝对优势载荷,贡献率为 87.0%,其次为 Cd(18.3%).因子 2 中 As 的载荷量最大,贡献率为 83.6%,其它重金属贡献率较低.因子 3 为污染源的主要控制因子,主要由 Ni (94.4%)、Cu (86.5%)、Cr (81.7%)、Zn (55.6%)和 Cd (33.8%)构成.因子 4 以 Pb (89.6%)的载荷量最大,其次为 Cd,贡献率为 34.1%.

为了更好地识别重金属的来源与分布,通过普通克里格法对 PMF 各因子得分进行空间插值,获取各因子的空间分布(图 6). 因子 1 中以 Hg 为主要载

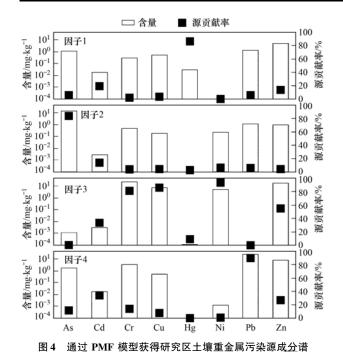


Fig. 4 Source profiles and source contributions of soil heavy metals in the study area from PMF

荷,有研究表明,全球约60%~80%的汞来自人为排放^[41].工业活动过程中如燃煤发电、石油生产和废弃物燃烧等均会增加 Hg 的排放,并通过大气干湿沉降进入农业土壤^[3,22,42].因此,有研究者将土壤中Hg 的累积与人类城市化进程相关联^[43].如图6所示,因子1的高载荷点多分布在西侧沿海平原一带,此处人口相对密集、存在一定的工业基础.在研究区的中部,发现小面积的点源污染,结合土壤中Hg的高变异性(CV=147%),因子1应表示与工业活动有关的人为排放源.

因子 2 以 As 的载荷量最大,贡献率为 83.6%. 工业排放^[22]和农业活动^[2,3]均会导致土壤中 As 的富集,其中后者在农用地中被广泛报道.图 6 显示, 因子 2 高载荷点大面积分布在研究区东侧低矮丘陵一带. 经过实际调查, 东侧土壤由于长期干旱缺少降雨, 多发育燥红壤, 加之昼夜温差较大, 土地以种植芒果为主, 几乎不存在工业基地. 据报道, 海南岛芒果园年平均施肥量为 695. 32 kg·hm^{-2[44]}, 长期的施肥和除虫活动易导致 As 的累积^[2,3]. 此外, 由于区域性的缺水环境, 果树更需要高强度的人工灌溉, 这也可能导致重金属的累积. 因此, 因子 2 应表示农业活动来源.

因子 3 为污染源的主要控制因子,主要由 Ni (94.4%)、Cu(86.5%)、Cr(81.7%)、Zn(55.6%)和 Cd(33.8%)构成.有研究表明,土壤中的 Cr 和 Ni 可以来自农业活动、工业活动和成土母质及其风化过程^[3,45,46],其中,大范围内 Cr、Ni 和 Cu 的来源多与成土母质相关^[22,47,48].相比于 As 和 Cd,本研究土壤中 Cr、Ni、Cu 和 Zn 的含量更接近背景值(表 3),同时受控于土壤中含 Fe、Mn、Al 矿物,表明它们主要来自成土母质.此外,因子 3 贡献率在空间分布图中具有丘陵区高、平原区低的特点(图 6),也表明这些重金属受到区域背景影响.结合本文研究,可以将因子 3 定义为与成土母质和成土过程中产生大量含 Fe、Mn、Al 矿物有关的自然来源.

因子 4 以 Pb (89.6%)的载荷量最大,其次为 Cd,贡献率为 34.1%.研究区气候干旱,为增加农作物产量,常采用地膜覆盖的方式来保持土壤中的水份. Pb 常作为热稳定剂添加到地膜中,由于缺乏回收措施,会导致土壤的 Pb 污染^[2].同时,肥料也是我国农业土壤中 Pb 的重要来源^[22],赵文等^[49]对海南商品有机肥分析结果显示,肥料中的 Pb 要高于海南岛土壤背景,而 Hg、Cr 和 Ni 则相反.此外,汽车尾气曾是土壤中 Pb 的主要来源^[15,23],尽管 2000 年以后,我国禁止生产和销售含有四乙基铅的汽油,但

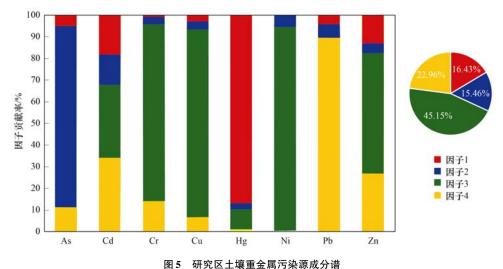


Fig. 5 Factor profiles of heavy metal sources identified by PMF

由于 Pb 的半衰期长达数百年,土壤中 Pb 的残留量仍然较高. 因此,研究区土壤中的 Pb 可以来自农业活动和机动车尾气. Cd 在因子 4 中也有较大的载荷,Cd 被认为与农业活动密切相关,化肥、杀虫剂的使用和污水灌溉都会增加 Cd 的含量^[2,22]. 图 6 显示,因子 4 的高载荷点多分布在研究区东侧种植园,

农产品收获季节,会有大量高排放厢式货车来往于该区域,为土壤重金属提供了来源. 因此,因子 4 应该表示农业生产和机动车尾气的混合源. PMF 与相关分析在重金属来源分类具有相似的结果,但 PMF 能够有效量化污染物来源. 因此,将 PMF 与相关分析结合,能够有效识别重金属来源.

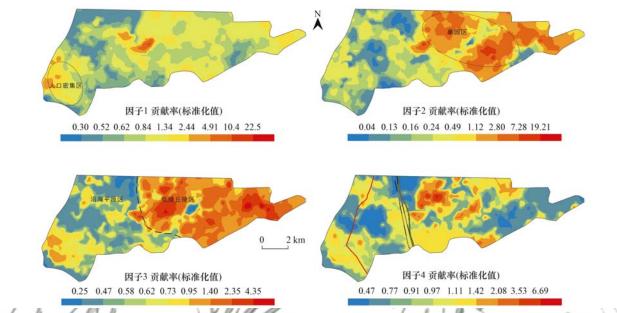


图 6 PMF 影响因子空间分布

Fig. 6 Kriging interpolation of PMF-factors

4 结论

- (1)在海南岛半干旱区,农用地土壤中除 Zn 之外,As、Cd、Cr、Cu、Hg、Ni 和 Pb 平均含量均高于海南岛土壤背景值,这与土壤中富含 Fe、Mn、Al 和有机质密切相关.
- (2)依据国家土壤污染风险筛选值,研究区农用地土壤中 As、Cd、Cu 和 Pb 的超标率分别为13.6%、6.5%、2.9%和2.1%,Cr、Ni、Zn 和 Hg 的超标率均小于0.6%. I_{geo} 结果表明,研究区农业土壤主要受到 As 的污染,其次为 Cd 和 Cu,中度污染以上样品分别占 82.7%、23.1%和20.6%; EI 指数表明,土壤中 As、Cd 和 Hg 为主要风险元素; RI 指数显示,29.4%的样品为强生态风险级别以上,其中 As 是其主要贡献者. 因此,在海南岛半干旱区开展农业活动,应注意 As 的污染防控.
- (3)人体健康风险评估表明,研究区土壤重金属对成人不会产生非致癌风险和致癌风险.重金属对儿童不存在非致癌风险,但 As、Cr 和 Ni 存在一定的致癌风险.因此,儿童在农用地嬉戏时,应该注意土壤污染物的防范.
- (4) PMF 模型源识别结果显示, 研究区重金属 主要有4种来源. Hg 主要来源于工业排放; As 主要

来源于农业生产活动; Ni、Cu、Cr 和 Zn 来自于成土母质和成土过程中产生大量含 Fe、Mn 和 Al 矿物有关的自然来源; Pb 和 Cd 来自农业生产和机动车尾气的混合源. 其中自然来源为土壤重金属的主要来源.

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