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## 猪粪堆肥过程中金霉素去除及重金属形态变化

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摘要: 抗生素和重金属的广泛使用导致了畜禽粪便中抗生素和重金属的大量残留, 堆肥工艺不但可降解残留的抗生素, 也能固化重金属. 本文以养殖场猪粪为对象, 利用中试好氧堆肥反应装置, 研究不同金霉素浓度时的猪粪堆肥特性及去除情况[0 mg·kg<sup>-1</sup> (CK)、10 mg·kg<sup>-1</sup> (T1)和50 mg·kg<sup>-1</sup> (T2)], 同时开展堆肥过程重金属形态变化的研究. 结果表明, 堆肥结束后, CK 组金霉素没有检出, T1和T2组抗生素降解率分别达到96.31%和97.32%, 金霉素降解过程符合一级动力学模型. 堆肥可以使重金属固化, Cu、Zn 元素的生物可利用态(可交换态、可还原态)逐渐转化为生物毒性低的可氧化态与残渣态, Cu、Zn 明显钝化. 相关性分析表明金霉素的去除与生物可利用态 Cu、Zn 呈现显著的正相关性.

关键词:猪粪;好氧堆肥;金霉素;重金属;生物可利用态

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# Removal of Chlortetracycline and Morphological Changes in Heavy Metals in Swine Manure Using the Composting Process

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Abstract: The widespread use of antibiotics and heavy metals in livestock farms results in large residues of antibiotics and heavy metals in the livestock manure. Composting technology can biodegrade residual antibiotics and solidify heavy metals. A pilot composting reactor was used to analyze the characteristics of chlortetracycline (CTC) removal at different antibiotic concentrations [0 mg·kg<sup>-1</sup> (CK), 10 mg·kg<sup>-1</sup> (T1), and 50 mg·kg<sup>-1</sup> (T2)]. Moreover, the morphological changes in heavy metals during the composting process were analyzed. After composting, no chlortetracycline was detected in the CK group and the antibiotics degradation rates of T1 and T2 groups reached 96.31% and 97.32%, respectively. The chlortetracycline degradation fits the pseudo-first-order kinetics model. Heavy metals can be solidified during the composting; thus, the bioavailable state of Cu and Zn (exchangeable state, reducible state) changed into the oxidation state and residues with apparent passivation formed. The correlation analysis showed that the removal of CTC showed strong positive correlations with the biological available Cu and Zn.

Key words: swine manure; aerobic composting; chlorotetracycline (CTC); heavy metals; bio-available forms

近年来,随着规模化集约化畜禽养殖业的迅猛发展,养殖粪便的产量急速增加.据统计,我国大型畜禽养殖场中猪粪年产量达到2.5亿 t<sup>[1,2]</sup>.为了减少畜禽患病的几率,畜禽养殖饲料中通常会添加抗生素药物.据调查发现,我国每年生产的抗生素有52%(162000 t)用于畜禽养殖<sup>[3]</sup>.同时为了缩短畜禽饲养周期,常在饲料中添加重金属<sup>[4]</sup>.但添加的大部分抗生素和重金属不被畜禽吸收,随粪便和尿液排出<sup>[5]</sup>.含抗生素和重金属残留的粪便利用后,会引发环境污染和促进耐药基因传播<sup>[6]</sup>,随着抗生素使用量的增长,微生物的抗药性和重金属抗性也不断提高<sup>[7]</sup>.

因此,如何去除粪便中残留的抗生素和重金属成为研究的热点,有研究证明:堆肥处理可以对残留的抗生素进行降解同时能固化重金属<sup>[8~14]</sup>. 抗生素的去除受到多种因素的影响,如温度、pH、含水

率、重金属浓度、微生物、TOC、C/N的影响,但是最主要的因素还是温度、重金属浓度以及微生物的影响<sup>[15-17]</sup>. 众多实验开展粪便堆肥过程抗生素降解研究,如沈颖等<sup>[18]</sup>通过正交试验发现温度是影响土霉素和四环素降解的主要影响因素,Arikan等<sup>[19]</sup>进行牛粪堆肥实验,经过 38 d 后 99.8% 的金霉素得到降解. 同时开展粪便堆肥过程抗生素降解及重金属形态变化的研究鲜见报道. 本文利用中试装置,分析粪便堆肥过程抗生素降解及重金属形态变化,并开展抗生素的去除与生物可利用态 Cu、Zn 相关性的研究,探索养殖粪便安全回田的技术对策.

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#### 1 材料与方法

#### 1.1 实验材料

实验所用猪粪取自黑龙江省哈尔滨市呼兰区某大型养猪场哺乳期母猪粪便,木屑取自于哈尔滨某

大型木材厂. 堆肥原料的基本理化性质见表 1.

#### 1.2 实验装置及设计

本实验采用中试规模,实验装置如图 1 所示, 反应器有效容积为 320 L.

共设3组实验,猪粪和木屑按照2:1(质量比)

表 1 堆肥原料基本理化性质

Table 1 Characteristics parameters of materials

原料	含水率/%	有机质(DW)/%	总碳/%	总氮/%	C/N	pH
猪粪	69. 9	92. 35	38. 65	3. 35	11. 535 82	6. 520
木屑	8. 6	78. 25	46. 51	1.35	34. 455	_

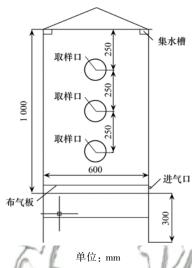


图1 实验装置示意

Fig. 1 Schematic diagram of experimental process

进行混合, 堆肥初期含水率控制在 65%. 1 号为对照组, 不投加金霉素; 2 号金霉素投加量(以DW 计, 下同) 为 10 mg·kg<sup>-1</sup>; 3 号投加 50 mg·kg<sup>-1</sup>的金霉素. 堆肥周期 60 d, 采用静态强制通风, 曝气量为 0.5 L·(kg·min) <sup>-1</sup>, 每 10 d 翻堆一次, 40 d 后 3 组反应器停止供气, 在通风条件下进行静置发酵.

#### 1.3 常规指标分析测定方法

常规指标的检测方法:温度检测为实时监测装

置;取5g样品加25 mL水,振荡1h,然后3500 r·min⁻¹离心15 min,测定pH(便携式pH/mV/℃计,HI8424C);同时取1 mL上清液稀释50倍后,进行水溶性碳氮含量的测定(TOC/TN-VCPN,岛津公司);种子发芽指数的测定,按照样品·蒸馏水(1:10)浸提,振荡1h,3500 r·min⁻¹离心15 min,选取20颗饱满的白菜种子,每组设3个平行样,30℃生活培养箱中培养72h,以蒸馏水作为空白对照,测定根长及种子发芽率.

种子发芽指数(GI) = 样品处理的发芽率 × 样品处理平均根长 空白的发芽率 × 空白平均根长

#### 1.4 抗生素和重金属的检测

抗生素检测:通过固相萃取小柱对金霉素进行固相萃取,取 0.5 g 样品,加入 McIlvaine-Na<sub>2</sub>EDTA 浸提,利用甲醇洗脱,氮气吹干后定容,过 0.22 μm 滤膜,然后利用 LC-MS/MS (2777 C-ACQYITY UPLC-Xevo<sup>TM</sup> TQ MS,美国 waters 公司)进行金霉素的检测 $^{[6,9,11,18-20]}$ . 固相萃取方法的金霉素加标回收率为 97.94% ~110.63% 符合要求 $^{[21,22]}$ .

重金属的检测为 Cu、Zn 不同形态含量,采用改进的 BCR 提取法,提取方法见表 2,利用 ICP-MS (Optima-5300DV,美国 PE 公司)进行检测.

表 2 BCR 法提取重金属步骤

Table 2 BCR extraction process of heavy metals

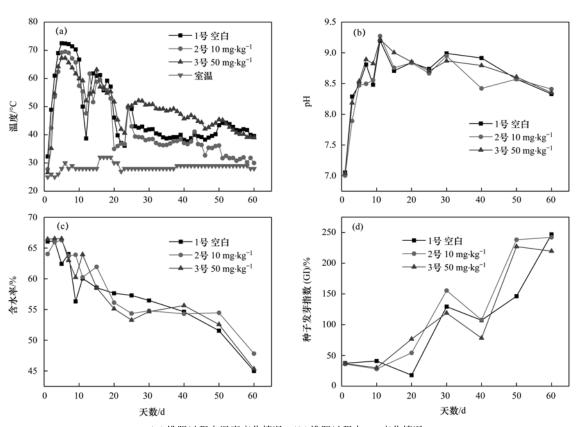
步骤	重金属形态	提取剂	提取方法
1	可交换态	30 mL 0. 11 mol⋅L <sup>-1</sup> CH <sub>3</sub> COOH	25℃ 水浴振荡 16 h, 离心 20 min
2	可还原态	30 mL 0.5 mol·L $^{-1}$ NH $_2$ OH·HCl, pH 2	25℃ 水浴振荡 16 h, 离心 20 min
3	可氧化态	$10~\text{mL}$ 8. 8 $\text{mol}\cdot\text{L}^{-1}\text{H}_2\text{O}_2,\text{pH}$ 2. 5	室温轻微振荡 1 h, 85℃ 水浴振荡 1 h
		$10~\text{mL}$ 8. 8 $\text{mol}\cdot\text{L}^{-1}\text{H}_2\text{O}_2$ , $\text{pH}$ 2. 5	85℃水浴振荡 1 h
		30 mL 1.0 mol ${}^{\textstyle \cdot} L^{-1}$ $CH_3 COONH_4$ , pH 2	25℃水浴振荡 16 h, 离心 20 min
4	残渣态	$\mathrm{HNO_3}\text{-}\mathrm{HClO_4}$	消解

#### 2 结果与讨论

#### 2.1 堆肥过程理化性质的变化

堆肥过程中温度的变化见图 2(a), 堆肥初期有机物质分解释放大量热量,3 组实验高温期持续的时间均在 20 d 左右,最高温度分别达到 72.5、69.5、67.2  $\mathbb{C}$ . 其中,1号升温最快,第 3 d 达到了60  $\mathbb{C}$ . 堆肥 20 d 后,3号的温度明显持续高于1号与2号,可能是由于在堆肥初期微生物活性受到高剂量的金霉素抑制,随着金霉素残留浓度的降低,继续进行有机物分解,使得温度维持在 45  $\mathbb{C}$  左右.pH 变化见图 2(b),堆肥初期因为含氮有机物分解

产生铵态氮使 pH 值升高, 在堆肥第 11 d 达到最高值. 随后, 大量的 N 元素以 NH<sub>3</sub> 的形式释放, 使 pH 值逐渐降低, 最终稳定在 8.5 左右. 含水率变化见图 2(c), 在堆肥初期有机物分解产生大量热的同时带走大量水分, 使得含水率下降. 第 10 d 进行翻堆时, 对含水率进行控制, 维持在 60% 左右. 40 d 后, 含水率迅速降低, 60 d 后达到 45% 左右. 种子发芽指数(GI)是评价堆肥腐熟程度的一个重要指标. 图 2(d)为堆肥过程中的 GI 变化情况. 在堆肥初期(0~10 d), GI 在 50% 左右, 第 30 d 均达到 100% 以上, 60 d 可以达到 250%, 表明堆肥已经达到腐熟.



(a) 堆肥过程中温度变化情况; (b) 堆肥过程中 pH 变化情况;

(c) 堆肥过程中含水率变化情况;(d)堆肥过程中种子发芽指数变化情况

#### 图 2 堆肥过程理化性质的变化情况

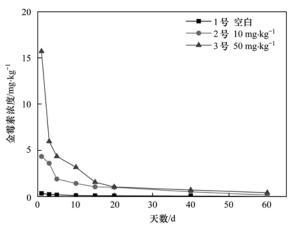
Fig. 2 Changes in temperature, pH, moisture and GI during composting

#### 2.2 堆肥过程中金霉素的降解情况

作为制约畜禽粪便资源化的主要因素, 抗生素的降解率是堆肥过程中必须重视的指标, 猪粪中检测的金霉素含量(以 DW 计, 下同)为 0.6 mg·kg<sup>-1</sup>,与 Arikan 等<sup>[19]</sup>的检测结果 0.208 mg·kg<sup>-1</sup>相近,但是与张树清等<sup>[23]</sup>的测定结果 15.66 mg·kg<sup>-1</sup>存在有很大的差距,主要是由于本实验猪粪选取为哺乳期的母猪粪便,其饲养过程中添加的抗生素含量低.

堆肥过程中金霉素的降解变化情况见图 3,金霉素降解主要在升温及高温期,主要途径为水解、分解和微生物降解,受到温度、pH、微生物活性等因素影响.最终,1号没有检测出金霉素,2号金霉素剩余量为 0.16 mg·kg<sup>-1</sup>、3号金霉素剩余量为 0.42 mg·kg<sup>-1</sup>(见图 3).

由图 4 得出, 堆肥初期, 金霉素迅速降解, 在第 10 d 可分别达到 68.15%、74.73%、79.78%,



#### 图 3 堆肥过程中金霉素的降解情况

Fig. 3 Changes in the concentrations of chlortetracycline during composting

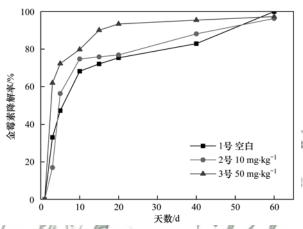


图 4 堆肥过程中金霉素的降解率变化情况

Fig. 4 Degradation rate of CTC during composting

最终降解率分别达到 100%、96.31%、97.32%.2 号实验结果相对于 Wu 等[11] 报道的金霉素在 3.1 mg·kg<sup>-1</sup>时的降解率 74% 更高,金霉素降解率达到 96.31%,其中 Wu 实验中的温度最高为 60℃,在本实验中,堆肥温度可以依靠自身达到 70℃以上,并且维持 3~5 d.可以看出,提高高温期温度可以促进金霉素的分解和水解有利于去除金霉素.3号 50 mg·kg<sup>-1</sup>的金霉素降解率达到了 97.32%与 Hu 等[24] 报道的初始含量为 60 mg·kg<sup>-1</sup>时金霉素降解率为 97% 相近.胡真虎实验中温度为 62℃,3号反应器中温度最高为 68℃,持续 2 d,然后逐渐降低,

在堆肥周期前40 d都维持在一个较高的温度,与胡真虎实验现象相近.通过对3组实验结果的分析,可以得出,提高高温期温度可以促进金霉素的水解及分解从而提高金霉素的去除效率.

#### 2.3 一阶动力学模型研究金霉素降解情况

堆肥过程中金霉素的降解情况符合一级动力学 模型,图5为 $\ln(c/c_0)$ 与堆肥时间t之间的关系, 由此可以得出两者之间的线性关系(见表3). 在堆 肥过程 20 d 内, 金霉素降解符合一级动力学模型, 其拟合度  $R^2$  达到0.8876、0.8675、0.8383、半衰 期为 7.8、6.86、4.94 d. 此实验结果较吴晓峰[11] 等研究结果更短,其金霉素半衰期为8.25 d,证明 堆肥效果较其更加. 2号与3号的金霉素降解半衰 期不同,可能是由于添加金霉素,对堆肥反应产生 的影响不同. 金霉素的去除和微生物活性以及温度 三者之间互相影响, 微生物呼吸作用分解有机物使 堆体温度升高,促进抗生素的分解和水解,但是温 度过高或者抗生素浓度过高也会对微生物的活性产 生抑制作用,从而反作用于温度和抗生素的降解. 在本实验过程中, 1号第3d可以达到60℃以上, 高温期最高可以达到70℃以上,使得微生物的活性 受到了抑制,而2号、3号反应器的升温速度较1 号慢, 第 3 d 温度尚未达到 55℃, 最高温度低于 70℃, 使得微生物的相关酶的活性受温度的抑制较 1号小, 金霉素降解速率较快, 半衰期较短.

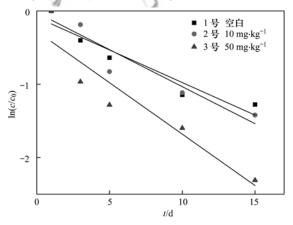


图 5 堆肥过程中  $\ln(c/c_0)$  与时间(t) 之间的关系

Fig. 5 Plot of natural logarithm of the concentration/starting concentration (  $c/c_0$  ) versus time ( t ) of CTC

#### 表 3 金霉素降解的线性拟合方程、半衰期、R<sup>2</sup>

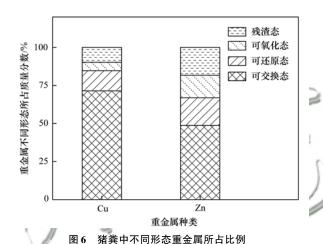
Table 3  $\,$  Linear fitting equation, half-life, and  $R^2$  of the degradation of CTC

编号	方程	斜率	$R^2$	半衰期/d
1	$\ln(c/c_0) = -0.08891t - 0.08743$	-0.08891	0. 887 6	7. 80
2	$\ln(c/c_0) = -0.10105t - 0.0225$	-0.10105	0. 867 5	6. 86
3	$\ln(c/c_0) = -0.14033t - 0.27697$	-0. 140 33	0. 838 3	4. 94

#### 2.4 堆肥过程中重金属的不同形态的变化情况

畜禽粪便中的重金属是制约其资源化的另外一因素. 通常来说, 堆肥对重金属的总量影响不大, 但是对重金属的形态变化有重要作用, 可以对重金属进行固化<sup>[23~30]</sup>. BCR 连续提取法将重金属分为 4 种形态: 可交换态、可还原态、可氧化态、残渣态. 其中可交换态与可还原态容易被吸收存在一定生物毒性, 因此将可交换态与可还原态之和称为生物可利用态. Cu、Zn 不同形态所占比例见图 6.

从图 6 可以看出, 猪粪中 Cu、Zn 主要以可还原态与可交换态存在, 其含量的多少直接影响生物毒性高低. 猪粪中重金属含量最高的是 Cu、Zn, 含量达到 110.51 mg·kg<sup>-1</sup>和2 736.52 mg·kg<sup>-1</sup>. 我国对于 Cu、Zn 含量的要求还没有相应的标准, 但按照德国标准(Cu 100 mg·kg<sup>-1</sup>、Zn 400 mg·kg<sup>-1</sup>)来看, Cu、Zn 含量存在明显的超标, 其中生物可利用



. 6 Proportion of different forms of heavy metals in swine manure

态占有绝大部分,如猪粪未经堆肥处理还田存在着 风险.

在堆肥初始阶段, Cu、Zn 的生物可利用态占很大比例,但是经过堆肥处理后明显地减少,可交换态与可还原态的重金属在堆肥过程中逐渐转变为可氧化态和残渣态.堆肥处理后使得生物可利用态Cu(可交换态和可还原态之和)从初始的65.74%、68.34%、67.26%变为原来的16.73%、17.59%、19.47%,同时对于Zn 的钝化作用也十分明显,从初始76.82%、76.57%、77.89%减少到26.09%、30.76%、37.19%.由此可见堆肥处理可以使重金属明显地固化,使得其生物可利用形态的重金属含量减少,继而减少重金属的毒害和还田过程中的土壤重金属污染的风险.

#### 2.5 相关性分析

如表 4 所示,金霉素的剩余量与含水率、pH、以及生物可利用形态 Cu、Zn 呈现出明显的相关性. Ramaswamy 等<sup>[31]</sup>认为 pH、温度以及微生物酶活性在降解过程中起重要作用. Bao 等<sup>[32]</sup>研究发现抗生素的去除效果受到温度、pH、重金属,微生物,TOC、TN 等因素的影响. 金霉素的剩余量与生物可利用态 Cu、Zn 的含量呈正相关(P<0.05),随着堆肥反应的进行,金霉素和生物可利用态重金属同步降低. 实验结果中金霉素剩余量 1 号 < 2 号 < 3 号,生物可利用态 Cu、Zn 含量 1 号 < 2 号 < 3 号,生物可利用态比例高,金霉素剩余量多. 同时与 pH 成负相关(P<0.01). 此外,实验证明温度也是影响重金属去除效果的重要因素,提高高温期温度有利于促进金霉素水解和分解,提高金霉素的降解率.

表 4 金霉素剩余量与温度、含水率、pH、GI 以及生物可利用态 Cu、Zn 的相关系数

Table 4	Correlations between	CTC	temperature	moietura	nН	CI	and bioavailable Cu and Zn

				·,,	,, p, o,		
	温度	含水率	рН	GI	金霉素剩余量	生物可利用态 Cu	生物可利用态 Zn
温度	1	0. 693 * *	0. 326	-0.357	-0.011	-0.513	-0.863
含水率		1	-0.246	-0.625 *	0. 652 *	0. 976 *	0. 725
pH			1	0.032	810 * *	-0.843	-0.962 *
GI				1	-0.445	- 0. 843	-0.454
金霉素剩余量					1	0. 966 *	0. 956 *
生物可利用态 Cu						1	0. 855
生物可利用态 Zn							1

1) \* 表示相关性在 0.05 以上显著(双尾); \* \* 表示相关性在 0.01 以上显著(双尾)

#### 3 结论

(1)堆肥过程可以有效地去除猪粪中残留的金霉素和固化重金属. 最终对照组没有检测出金霉素,添加10 mg·kg<sup>-1</sup>和50 mg·kg<sup>-1</sup>金霉素的2号、

- 3号堆肥处理中金霉素降解率达到96.31%、97.32%.,仍检测到有少量的残余. 堆肥能够显著降低生物可利用态重金属比例,明显地钝化Cu、Zn.
  - (2)提高高温期温度有利于提高金霉素的去除

- 率,高温促进了金霉素的水解和分解,利用一阶动力学模型研究金霉素降解过程, $R^2$ 均达0.83以上.
- (3)金霉素的降解主要受温度,重金属、pH 和 微生物活性的影响. 经过相关性分析,发现金霉素的去除与生物可利用态 Cu、Zn 呈现显著的正相关性,生物可利用态 Cu、Zn 比例高,金霉素剩余量高. 同时,金霉素的去除与含水率、pH 亦呈现显著的相关性.

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