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# 基于 LDI 的土地利用类型与湿地水质的相关性:以苏州太湖三山岛国家湿地公园为例

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摘要:以苏州太湖三山岛国家湿地公园为例,探索不同时空条件下土地利用类型整体与水质之间的相关关系. 使用主成分分析计算出每个采样点的水质综合指数,基于 GIS 空间分析技术计算表征土地利用类型综合效应的景观开发强度指数 (landscape development intensity,LDI),运用 Pearson 相关分析探讨水质与土地利用类型之间的整体相关性. 结果表明,水质随着湿地公园空间分布不同而有较大的差异,研究区西部水质较好,东部水质较差;建设用地与水质综合指数存在明显的正相关,自然水体则与水质综合指数呈显著的负相关;缓冲区半径为100、150、200、250、300、350、400、450 和500 m的 LDI 综合指数与水质综合指数的 Pearson 相关系数分别为 0.641、0.678、0.691、0.685、0.691、0.695、0.680、0.653 和 0.649(P < 0.01),它们之间存在显著而稳定的相关性,可以体现出多种土地利用类型对湿地水质的整体影响,部分克服了单项土地利用类型表达不完整、难以解释水质变化的不足.

关键词:湿地公园; 土地利用类型; 水质; 主成分分析; 景观开发强度指数

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## Correlation Between LDI-based Land Use Types and Water Quality in Sanshan Island of Taihu Lake National Wetland Park, Suzhou

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**Abstract**: Sanshan island of Taihu Lake National Wetland Park in Suzhou was taken as a case study to explore the relationship between land use types and water quality under different spatial and temporal conditions. Firstly, principal component analysis was used to calculate the comprehensive index of water quality for a given sampling site. Secondly, landscape development intensity index (LDI), which can represent combined effects of land use types, was calculated based on GIS spatial analysis technology. Finally, overall correlation between water quality and land use types was obtained by using Pearson correlation analysis. The results showed that the water quality varied with the spatial distribution of the wetland park. Totally, water quality of west region was good and water quality of east region was poor; Built-up land and water quality integrated index exhibited obvious positive correlation. And natural water and water quality index was significantly negatively correlated; By building relationship of water quality index and LDI index within 100, 150, 200, 250, 300, 350, 400, 450 and 500 m radius buffer, Pearson's r values between them were 0.641, 0.678, 0.691, 0.685, 0.691, 0.695, 0.680, 0.653 and 0.649 respectively (P < 0.01). These statistics indicated obvious and stable overall correlation between land use types and water quality. This can reflect a variety of land use types' comprehensive effects on wetland water quality, and partly overcome the weakness of incomplete and difficult explanation for water quality changes with single type of land use. **Key words**; wetland park; land use type; water quality; principle component analysis; landscape development intensity index

作为水陆交界处的自然过渡区,湿地是自然界非常重要的人类生存环境和生态景观之一,在调节气候、调蓄洪水、净化水质、维持生物多样性等方面发挥着十分重要的作用[1]. 但是,近年来伴随着中国经济的高速发展,城区全面扩张,人口急剧增加,湿地受到越来越严重的破坏. 湿地面积大量减少,水污染程度日益加重,生物多样性逐渐消失,湿地生态功能严重退化[2]. 为此,国内许多城市建立了湿地公园,使其在承担一般公园游憩休闲功能的同时,也承担着类似自然保护小区的生态功能. 希望借此探索出一条协调城市化发展过程中湿地保护

与科学利用的发展模式.

一方面,湿地土地利用类型的变化会影响湿地公园内部原有的景观结构,改变湿地公园的生态过程,影响着湿地公园的健康状态.另一方面,水作为湿地生态系统中最重要的特征之一,是湿地物质循环和能量流动的核心载体.湿地水质的健康状态也影响着湿地公园的健康状况.因此,研究土地利用

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类型和湿地水质之间的关系,可以为湿地公园水土 资源可持续利用、湿地保护与管理提供科学依 据[3]. 国外学者基于 GIS 与遥感等技术,使用统计 分析方法初步探明了土地利用类型与水质之间存 在着紧密联系,土地利用类型的构成方式可以显 著影响到水质[4~6]. 湿地土地利用类型与水质化 学性质之间,湿地水体中的氮、磷含量与周围林地 面积以及与湿地的距离之间均存在明显相关 性[7]. 国内学者则在太湖、新安江、九龙江、洱海 和艾比湖等流域探讨了土地利用结构与水质之间 的关系[8~15]. 有研究表明城市建设用地、未利用 地、林地、水体等土地利用类型均与水质存在明 显的相关关系. 由于不同尺度的流域生态过程及 水质均存在明显的时空差异,这导致湿地生态系 统中土地利用与湿地水质之间的关系错综复杂. 上述研究侧重于土地利用类型构成和土地利用空 间格局与湿地水质之间的响应关系,大多将单个 的土地利用类型与单个水质监测指标作相关性分 析. 一些学者建立了能够衡量多种土地利用类型 之间复合关系的综合指标[16,17],但较少研究在时 空差异条件下土地利用类型综合指标与湿地水质 之间的联系.

本研究以苏州太湖三山岛国家湿地公园为例, 通过构建基于主成分分析的水质综合指数与代表土 地利用类型综合效应的景观开发强度指数,揭示在 不同时空条件下多种土地利用类型对湿地水质的综 合影响与相关性,以期为湿地公园水污染防治和土 地利用优化方案提供理论依据和技术基础,并为湿 地保护和管理活动提供科学指导.

#### 1 材料与方法

#### 1.1 研究区概况

苏州太湖三山岛国家湿地公园位于苏州城西太湖之中的三山岛区域,总面积 6.252 km²,依托三山岛、泽山岛等岛屿及毗连太湖水域构建而成,是全国第一个以村级和岛屿形式创建的国家级湿地公园. 地理坐标介于东经 120°16′45″~120°17′55″,北纬 31°01′25″~31°02′30″. 该湿地公园是在退渔还湿的基础上,通过恢复三山岛原有的湖滨带而形成的湿地. 受太湖小气候环境影响,三山岛水热资源丰富,气候温和宜人,空气湿润. 年平均气温 16℃,年平均降雨为1 100~1 140 mm,年平均湿度 76%,具有良好的人居环境. 研究范围包括三山岛、泽山岛及毗连的太湖水域,如图 1 所示.

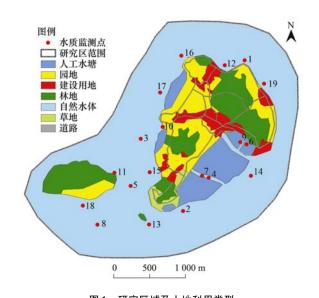


图 1 研究区域及土地利用类型

Fig. 1 Location and land use types of the study area

#### 1.2 数据来源与预处理

根据苏州太湖三山岛国家湿地公园周边土地利用类型结构及生态系统特征的空间差异,在三山岛、泽山岛周边水域布设了19个水质采样点(图1),分别在2014年7月1日、2014年9月16日、2014年12月16日和2015年4月16日进行了4期的水质采样工作.样品采集后带回实验室,参考已有相关文献[10-15],选择总悬浮固体(TSS)、溶解氧(DO)、化学需氧量(COD)、氨氮和硝态氮这5项代表性水质指标,依据《地表水和污水监测技术规范(HJ/T91-2002)》进行测定.测定时为保证结果的准确性,对每个采样点的样品做3份平行检测,取3份平行检测结果的平均值作为最终的水质指标.

此外,利用已有的土地利用现状图和遥感影像图,结合野外实地调查结果,得到如图1所示的7类湿地公园土地利用类型,包括人工水塘、园地、建设用地、林地、自然水体、草地和道路,其中建设用地主要指居住用地与商服用地.

#### 1.3 基于主成分分析的水质综合指数

水质评价作为生态环境质量评估的重要内容之一,涉及到众多的影响因子,在综合评价过程中很难从众多的影响因子中提取重要信息<sup>[18]</sup>.主成分分析方法(PCA法)的基本思想是认为在众多相关因子之间必然存在着起支配作用的共同因子.它可以从许多变量中筛选出相互独立的主要因子,在保留原始重要信息的同时,使主要因子比原始变量具有某些更优越的特性.由于 PCA 法具有提取主要因子、简化数据结构、避免主观选择等特点,可被应

用于水质综合评价中<sup>[19,20]</sup>,从众多的水质监测指标中提取真正反映水质优劣的主要信息. PCA 法的数学模型为:

$$\begin{cases} F_{1} = a_{11}Z_{1} + a_{12}Z_{2} + \dots + a_{1m}Z_{m} \\ F_{2} = a_{21}Z_{1} + a_{22}Z_{2} + \dots + a_{2m}Z_{m} \\ \vdots \\ F_{m} = a_{m1}Z_{1} + a_{m2}Z_{2} + \dots + a_{mm}Z_{m} \end{cases}$$
(1)

式中,m 为变量因子个数;  $F_1$ ,  $F_2$ , …,  $F_m$  为提取出的主成分;  $a_{11}$ , …,  $a_{mm}$  为标准化处理变量 Z 协方差矩阵特征值对应的特征向量.

组合上述主成分,可得到如下的综合评价函数 F:

$$F = \frac{\lambda_1}{\lambda_s} F_1 + \frac{\lambda_2}{\lambda_s} F_2 + \dots + \frac{\lambda_p}{\lambda_s} F_p \tag{2}$$

式中, $\lambda_1$ ,  $\lambda_2$ , …,  $\lambda_p$  为 Z 矩阵的特征值, $\lambda_S = \lambda_1 + \lambda_2 + \dots + \lambda_n$ ; p 为主成分提取个数.

本研究中,共使用了 TSS、DO、COD、氨氮和硝态氮这 5 项水质监测指标. 对这些指标进行 KMO和 Bartlett 检验,可得 KMO的值为 0.712, Bartlett 的显著性概率 P 值 <0.05, 说明这些指标之间具有较

强的相关性,适用于主成分分析. 考虑到 TSS、COD、氨氮和硝态氮的数值均与水质呈反比,只有 DO 指标随着水质的变好而增加. 因此取所有采样点 DO 指标中的最大值减去各采样点的 DO 指标原始值,使其数值也与水质呈反比. 经过处理后,所有的指标均随着水质的提高而变小.

将每期监测数据的 5 项水质指标作为变量因子构成原始变量矩阵,使用式(3)对其进行标准化处理得到均值为 0、标准差为 1 的标准化数据,以消除原始不同数据量纲和数量级差异的影响. 在此基础上计算标准化数据的相关系数矩阵,并求出特征根和方差所占比例,结果如表 1 所示.

$$Z_{ij} = \frac{x_{ij} - \bar{x}_{j}}{S_{j}}$$

$$i = 1, 2, \dots, n; j = 1, 2, \dots, t$$
(3)

式中, $Z_{ij}$ 表示第 i 个采样点的第 j 个指标的标准化值; $x_{ij}$ 表示第 i 个采样点的第 j 个指标的原始测量值; $\bar{x}_j$  表示所有采样点第 j 个指标的平均值; $S_j$  表示第 j 个指标的标准差. n 为采样点个数,t 为水质监测指标个数. 本文中 n 为 19 , t 为 5 .

表 1 主成分分析的特征值与方差

Table 1 Eigenvalues and percent of total variance by using PCA analysis

| 期数 | 项目    |          |          | 主成分       | 主成分        |             |  |
|----|-------|----------|----------|-----------|------------|-------------|--|
|    | 坝目    | 1        | 2        | 3         | 4          | 5           |  |
| 1  | 特征值   | 2.67262  | 1.04008  | 0.895892  | 0.260975   | 0.130433    |  |
| 1  | 方差比/% | 53.4524  | 20.8016  | 17.9178   | 5.219 50   | 2.608 67    |  |
| 2  | 特征值   | 2.19020  | 1.564 07 | 0.805 126 | 0.392965   | 0.047 641 9 |  |
| 2  | 方差比/% | 43.8040  | 31.2814  | 16. 102 5 | 7.85930    | 0.952838    |  |
| 3  | 特征值   | 1.95288  | 1.493 17 | 0.894920  | 0.532 144  | 10.6429     |  |
| 3  | 方差比/% | 39.0577  | 29.8633  | 17.8984   | 10.6429    | 2.537 70    |  |
| 4  | 特征值   | 3.026 56 | 0.987416 | 0.651 114 | 0. 244 486 | 0.0904219   |  |
| 4  | 方差比/% | 60.5312  | 19.7483  | 13.0223   | 4.88972    | 1.808 44    |  |

从表 1 可知,前两个主成分的特征值几乎都大于 1,前 3 个主成分方差贡献率的总和分别占总方差的 92.2%、91.2%、86.8%和 93.3%,均大于 85%.因此可令 p=3,取前 3 个主成分来代替原来 5 个变量因子,据此确定因子载荷.以第一期观测数据为例,主成分因子载荷矩阵如表 2 所示.

表 2 主成分载荷矩阵

Table 2 Loading matrix of principal component

|      | U         | 1 1       | 1         |
|------|-----------|-----------|-----------|
| 水质指标 | 主成分 $F_1$ | 主成分 $F_2$ | 主成分 $F_3$ |
| TSS  | -0.2298   | -0.7898   | 0. 564 7  |
| DO   | 0. 876 7  | -0.1940   | -0.3023   |
| COD  | 0. 415 1  | 0. 581 7  | 0. 691 0  |
| 氨氮   | 0. 939 8  | -0.1083   | 0. 083 6  |
| 硝态氮  | 0. 892 1  | -0.1693   | 0. 032 9  |

根据式(1)和式(2),主成分 $F_1$ 、 $F_2$ 和 $F_3$ 是水质监测指标的线性函数,综合评价函数F又是主成分 $F_1$ 、 $F_2$ 和 $F_3$ 的线性函数.分别计算4期的水质综合评价函数,最后取平均值得到每个采样点水质的综合指数,将其作为衡量水质优劣的依据.

以第一期观测数据为例,设标准化后的指标变量分别为 $ZX_1$ 、 $ZX_2$ 、 $ZX_3$ 、 $ZX_4$ 、 $ZX_5$ ,则主成分表达式为:  $F_1 = -0.2298ZX_1 + 0.8767ZX_2 + 0.4151ZX_3 + 0.9398ZX_4 + 0.8921ZX_5$ 

$$F_2 = -0.789 \text{ 8ZX}_1 - 0.194 \text{ 0ZX}_2 + 0.581 \text{ 7ZX}_3 - 0.108 \text{ 3ZX}_4 - 0.169 \text{ 3ZX}_5$$

$$F_3 = 0.5647 \text{ ZX}_1 - 0.302 3\text{ZX}_2 + 0.691 0\text{ZX}_3 + 0.083 6\text{ZX}_4 + 0.032 9\text{ZX}_5$$
 (4)

以各主成分对应的方差贡献率为权重建立水质 综合函数:

$$F = 0.579F_1 + 0.226F_2 + 0.195F_3$$
 (5)

由于经过处理的 5 项水质监测指标均随着水质的提高而变小,所以 F 值越小表示水质越好, F 值越大则表示水质越差.

#### 1.4 土地利用类型的 LDI 综合指数

湿地相关的生物群落健康状况通常与人类活动强度紧密相关,人类活动越强烈,对湿地生态系统健康的干扰也越严重.作为人类活动的具体表现形式,土地利用可以反映出邻近湿地的健康状况<sup>[21]</sup>. Brown等<sup>[22]</sup>提出了景观开发强度方法(landscape development intensity,LDI).该方法结合土地利用类型与单位面积单位能耗的开发强度,测算不同土地利用类型所对应的 LDI 系数. LDI 系数代表了单位面积利用不可再生能量的定量测量值,即不可更新能源能值.它可以用来表示土地利用类型所消耗的能值,能值消耗越多则对应的 LDI 系数越大.最后累积研究范围内不同土地利用类型及其 LDI 系数,得到 LDI 综合指数:

$$LDI_{index} = \sum LU_i(\%) \times LDI_i$$
 (6)

式中, $LDI_{index}$ 为湿地某区域的 LDI 综合指数; $LU_i$  (%)为第i 种土地利用类型的面积占该区域总面积的百分比; $LDI_i$  为第i 种土地利用类型所对应的LDI 系数.

参考文献[23],本研究中人工水塘、园地、建设用地、林地、自然水体、草地和道路对应的 LDI 系数如表 3 所示. LDI 系数处于 1~10 之间,1 代表湿地完全为自然健康状态,10 代表湿地被人类高度开发利用而极度退化. LDI 系数越大,表明湿地受人类干扰强度越大,相应的湿地健康状况越差. 由于 LDI 系数是根据当地土地利用类型对应的不可更新能源能值估算得到,考虑到能值计算的过程比较复杂,并且受不同国家不同地区经济发展水平的影响而有所差异. 因此在今后的研究中,须根据苏州研究区域实际情况精确测定 LDI 系数,以提高本文方法的适用性.

LDI 方法的实质是核算不同土地利用类型的不可更新能源能值. 能值以太阳能焦耳[J·(hm²·a)<sup>-1</sup>]作为衡量单位,表示每年每公顷在能量转换中可利用的生产产品或服务所需要的能量. LDI 方法使用湿地生态系统人为的能量输入来表示人类活动对湿地自然环境的整体干扰程度,通过人类对湿地区域土地利用的影响来间接评价湿地

健康<sup>[24,25]</sup>. 该方法考虑了土地利用类型的构成和综合效应,可以更全面客观地反映土地利用类型对湿地水质的影响.

#### 表 3 湿地区域土地利用类型与对应的 LDI 系数

Table 3 Land use types and corresponding LDI

coefficients within wetland

| 序号 | 土地利用类型 | 说明        | LDI 系数 |
|----|--------|-----------|--------|
| 1  | 人工水塘   | 人工开挖水塘    | 1. 83  |
| 2  | 园地     | 成片的果园农用地  | 3. 68  |
| 3  | 建设用地   | 住宅用地和商服用地 | 8.66   |
| 4  | 林地     | 天然林和人工林   | 1.58   |
| 5  | 自然水体   | 开放的自然湖泊水体 | 1.00   |
| 6  | 草地     | 成片的草地     | 2.77   |
| 7  | 道路     | 主干道公路用地   | 7. 81  |

#### 1.5 水质与土地利用类型的关系

土地利用与水质的关系往往随着空间位置改变表现出局部变化的特征.即使在同一研究区域的不同位置,同一土地利用类型对水质的影响在大小、方向、距离上均可能表现出不同<sup>[26]</sup>.由于传统的LDI方法仅考虑了土地利用类型的结构组成而没有考虑距离衰减因素造成的影响,本研究采用反距离权重LDI方法<sup>[27]</sup>,充分考虑距离在土地利用类型对水质采样点的影响作用.也就是说,同一土地利用类型距离采样点位置越近,其对水质的影响也越大.改进的计算公式如下:

$$LDI_{index} = \sum \lambda_i \times LU_i(\%) \times LDI_i$$
 (7)

式中, $\lambda_i = d_i^{-2} / \sum_{i=1}^n d_i^{-2}$ ,且满足  $\sum_{i=1}^n \lambda_i = 1$ .  $\lambda_i$  是土地利用类型对应的反距离权重系数, $d_i$  是土地利用类型斑块重心位置至对应采样点的距离,n 是土地利用类型的总数.

缓冲区大小的设定直接影响着参与分析的土地利用类型面积和反距离权重系数,决定了 LDI 综合指数的大小. 在实际应用中,需根据生态学意义和具体评价对象等因素综合确定该范围<sup>[28,29]</sup>. 本研究在野外调查的基础上,以采样点为中心,分别设置不同半径的圆形缓冲区,如图 2 所示. 利用 ArcGIS 10.1 软件叠置分析功能的交集操作工具和归纳统计工具,分别统计出各个采样点对应的缓冲区范围内土地利用类型面积及面积百分比,并计算各土地利用类型斑块重心位置到采样点之间的距离. 从而揭示不同土地利用类型结构对不同空间位置水质采样点的影响,探明水质与该点缓冲区内土地利用类型整体结构的关系.

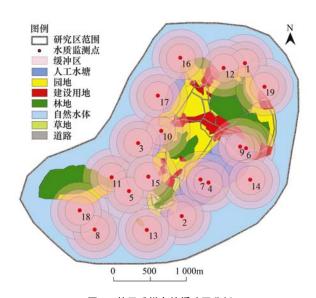


图 2 基于采样点的缓冲区分析

Fig. 2 Buffering analysis based on sampling sites

#### 2 结果与讨论

#### 2.1 水质空间分布特征

使用主成分分析方法求出每一期水质监测数据对应的水质综合函数 F,将不同季节 4 期数据求出的 F 再取其算术平均值,以减小季节的影响. 最后将其线性拉伸至 1~10 区间,得到每个采样点对应的水质综合指数,结果如图 3 和表 4 所示. 结果表明,研究区域西部的 5、8、11、18 号采样点,远离人口相对密集、旅游设施较多的三山岛主岛,受到人为干扰较少,水质较好. 相反,东北部的 1、12、19 号采样点位于旅游核心区域,附近有东泊和山东码头,还有大量的农家乐,人为干扰因素较多,所以水

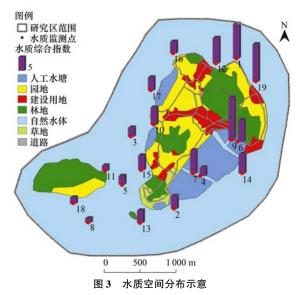


Fig. 3 Spatial distribution map of water quality

质较差. 东南部的9号采样点处于三山岛主要的污水处理厂和垃圾处理厂附近,水质也比较差. 南部区域有两条快艇通道,但由于该区域建立了人工修复湿地,所以该区域的水质整体处于居中水平.

#### 2.2 土地利用类型结构

缓冲区的合理设置决定了 LDI 综合指数大小, 也影响着土地利用类型与水质之间的关系[30]. 本 研究以50 m 为间隔,分别设置半径为100、150、 200、250、300、350、400、450 和 500 m 的缓冲区并 计算相应的 LDI 综合指数,计算结果如表 4 所示. 由于采样点都位于水面上,以采样点为中心建立的 圆形缓冲区中,自然水体所占的比例相对其它土地 利用类型要高. 下面以 400 m 为半径的缓冲区分析 为例来说明,所有采样点对应的土地利用类型结构 如图 4 所示,不同土地利用类型结构差异明显. 其 中顶端小圆点、矩形框下边缘、矩形框上边缘分别 代表不同土地利用类型面积所占百分比的最大值、 最小值和平均值. 总体上自然水体面积占比最高, 平均为60.4%,8号采样点处达到最大的99.1%, 3、5和13号采样点也都超过了85%;人工水塘面 积占比排名第二,平均为21.5%,4号点处达到最高 的 65.1%; 林地面积占比平均为 10.7%, 6 号点处 为最高的 24.9%; 园地面积占比平均为 10.5%, 10 号点处达到最高的25.8%;建设用地面积占比平均 为 4.6%, 9 号点处为最高的 15.7%; 草地面积占比 平均为 3.7%, 2 号点处为最高的 7.8%; 道路面积 占比平均为1.9%,9号点处为最高的4.5%.

对于每个采样点来说,对应缓冲区内土地利用 类型面积占比如图 5 所示,据此计算出的 LDI 综合 指数如表 4 所示.

#### 2.3 水质与土地利用类型相关性

#### 2.3.1 单项分析

以半径400 m的缓冲区为例,将各单项的土地

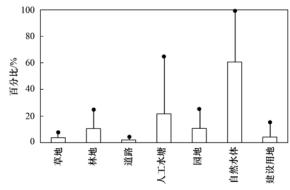


图 4 不同土地利用类型所占的面积百分比

Fig. 4 Area proportion of different land use types

| 表 4 | 采样点的 | 水质综合 | :指数与 | LDI 综 | 合指数 <sup>1)</sup> |
|-----|------|------|------|-------|-------------------|
|     |      |      |      |       |                   |

| Table 4 | Water quality | integrated | index and | I LDI | index for | r the sampling sites |
|---------|---------------|------------|-----------|-------|-----------|----------------------|
|         |               |            |           |       |           |                      |

|      |            |            |            | 1 , 0      |            |            | 1 0        |            |            |            |
|------|------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|
| 采样点号 | 水质综合<br>指数 | L100<br>指数 | L150<br>指数 | L200<br>指数 | L250<br>指数 | L300<br>指数 | L350<br>指数 | L400<br>指数 | L450<br>指数 | L500<br>指数 |
| 1    | 8. 05      | 1. 59      | 1. 72      | 1.68       | 1. 65      | 1. 56      | 1. 61      | 1. 52      | 1. 45      | 1. 39      |
| 2    | 3.01       | 1. 22      | 1. 13      | 1. 28      | 1.31       | 1.42       | 1.50       | 1.46       | 1. 21      | 1. 43      |
| 3    | 2.07       | 1.00       | 1.00       | 1.00       | 1.02       | 1.05       | 1. 32      | 1. 23      | 1. 16      | 1. 11      |
| 4    | 1.95       | 2. 10      | 1.99       | 1.71       | 1.85       | 1.70       | 1. 87      | 1. 85      | 1.75       | 1.82       |
| 5    | 2.01       | 1.00       | 1.00       | 1.00       | 1.06       | 1.09       | 1. 14      | 1. 25      | 1. 23      | 1.42       |
| 6    | 5. 20      | 2.88       | 2.99       | 2.91       | 3. 15      | 2.96       | 2. 77      | 2.73       | 2. 84      | 2. 75      |
| 7    | 5. 44      | 1.76       | 1.84       | 1.76       | 2. 22      | 2. 11      | 2. 31      | 2. 28      | 2. 34      | 2. 28      |
| 8    | 1.00       | 1.00       | 1.00       | 1.00       | 1.00       | 1.00       | 1. 02      | 1.02       | 1.02       | 1.01       |
| 9    | 10.00      | 2. 94      | 2. 92      | 2. 79      | 3.43       | 3.01       | 3. 22      | 3.02       | 2. 83      | 2. 92      |
| 10   | 4. 02      | 2. 84      | 2.77       | 2.61       | 2. 64      | 2. 56      | 2. 43      | 2.46       | 2. 53      | 2. 59      |
| 11   | 1. 24      | 1.55       | 1.46       | 1.31       | 1.44       | 1. 36      | 1. 42      | 1. 34      | 1. 27      | 1. 18      |
| 12   | 5. 91      | 2.77       | 2. 57      | 2. 62      | 2. 78      | 2. 68      | 2. 47      | 2.49       | 2.38       | 2. 27      |
| 13   | 2. 85      | 1.31       | 1. 22      | 1.03       | 1.04       | 1.07       | 1. 24      | 1. 15      | 1.04       | 1.03       |
| 14   | 4. 62      | 1.00       | 1. 21      | 1.09       | 1. 13      | 1. 18      | 1. 45      | 1. 39      | 1. 27      | 1. 36      |
| 15   | 3.43       | 1.73       | 1.71       | 1.50       | 2. 33      | 2.05       | 2. 27      | 2. 20      | 2. 32      | 2. 51      |
| 16   | 2.72       | 1.00       | 1. 59      | 1.58       | 1.67       | 1. 55      | 1. 44      | 1.63       | 1.76       | 1. 67      |
| 17   | 3. 07      | 1.00       | 1. 13      | 1.16       | 1.43       | 1. 39      | 1. 73      | 1. 67      | 1.53       | 1. 45      |
| 18   | 1.47       | 1.35       | 1.42       | 1.48       | 1.75       | 1.60       | 1. 67      | 1. 52      | 1.38       | 1.44       |
| 19   | 8.76       | 2. 62      | 2.56       | 2.49       | 2.61       | 2. 51      | 2. 22      | 2. 24      | 2.41       | 2. 52      |

1)L100 指数、L150 指数、L200 指数、L250 指数、L300 指数、L350 指数、L400 指数、L450 指数和 L500 指数分别对应 100、150、200、250、300、350、400、450 和 500 m 半径的缓冲区而得到的 LDI 综合指数

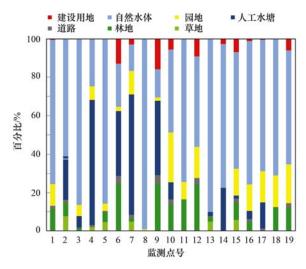


图 5 采样点缓冲区内土地利用类型面积占比

Fig. 5 Area proportion of land use types within buffer of the sampling sites

利用类型与水质综合指数作相关分析,可发现草地、林地、道路、人工水塘、园地、自然水体和建设用地在缓冲区内所占面积比和水质综合指数之间的Pearson相关系数分别是 - 0.068、0.485、0.703、0.174、0.051、-0.466、0.642. 其中道路和建设用地与水质综合指数呈明显的正相关,即道路和建设用地面积占比越大,水质综合指数越高,水质越差.主要原因可能是建设用地中以旅游设施和农家乐为

主,另外道路的修建也改变了原有土地利用类型的结构,加剧了人为的干扰,并通过影响流域水文循环、水土流失及污染物迁移转化,影响着水环境质量,导致水质退化. 因此可将道路与建设用地看作水质下降的主导因素. 自然水体则与水质综合指数呈明显的负相关,自然水体面积占比越大,水质综合指数越小,水质越高. 应该是由于随着自然水体面积的增加,对水质污染物的稀释作用也更加明显,因此可将自然水体看作水质提高的主导因素.

此外,人工水塘、园地与水质综合指数存在弱的正相关性,林地与水质综合指数则存在较强的正相关性,这与一般的分析结果不完全相符. 部分原因可能是三山岛国家湿地公园的经济结构以休闲旅游业为主,果树、渔业为副. 水质污染物的来源以生活污染物为主,这导致上述土地利用类型对水质的影响相对较小. 另一部分原因可能是多种土地利用类型是以综合作用的方式影响着湿地水质,单凭其中的某一项很难给出明确的解释,需要使用综合分析方法进一步明晰土地利用类型整体对水质的影响.

#### 2.3.2 综合分析

基于水质综合指数与 LDI 综合指数,使用 Pearson 相关对湿地公园水质与土地利用类型进行 整体分析,结果如图 6 所示. 其中水质综合指数与

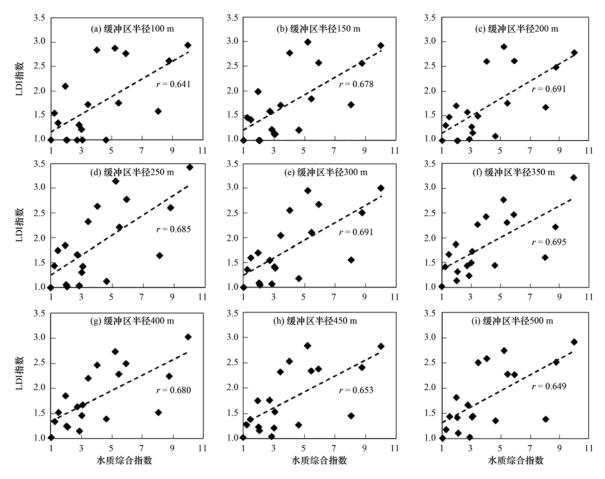


图 6 水质综合指数与 LDI 综合指数的相关分析

Fig. 6 Correlation analysis between water quality integrated index and LDI integrated index

缓冲区半径为 100、150、200、250、300、350、400、450 和 500 m 的 LDI 综合指数的 Pearson 相关系数分别为 0.641、0.678、0.691、0.685、0.691、0.695、0.680、0.653 和 0.649(P < 0.01),它们之间均呈现很强的正相关. 这较好解释了土地利用类型与湿地水质之间的整体相关关系,避免了单项土地利用类型表达不完整、难以解释水质变化的不足.

根据 r 系数曲线变化特征,半径为 350 m 缓冲区对应的 LDI 综合指数与水质综合指数的相关性最大,这与文献[27]的结论基本相符. 当缓冲区半径小于 150 m 时,由于采样点均位于水面上,土地利用类型中自然水体占据绝对大的比例,其它土地利用类型的种类与面积均较少,只能主要显示自然水体与水质之间的相关性而难以揭示多种土地利用类型对水质的综合影响;当缓冲区半径大于 450 m 时,距离采样点较远而对其影响较小的土地利用类型也被包含入缓冲区而参与 LDI 综合指数的计算,削弱了土地利用类型对湿地水质之间的总体相关性.

#### 3 结论

- (1)使用主成分分析计算水质综合指数,结果显示水质随着湿地公园空间分布的不同而有较大的差异.研究区西部的泽山岛没有居民点,植被密集,因此该区域整体上水质较好;研究区的东部位于旅游核心区域,附近景点众多,存在大量的农家乐,较易产生较多的生活污水,导致该区域的水质较差;其它区域为非核心旅游区域,受到少量的人为干扰,整体水质处于居中水平.
- (2)直接将各单项土地利用类型与水质综合指数作 Pearson 相关分析,结果表明道路和建设用地与水质综合指数呈明显的正相关,自然水体则与水质综合指数呈明显的负相关. 但是人工水塘、园地、林地与水质综合指数存在的相关性较难直接做出合理的解释.
- (3)利用水质综合指数与 LDI 综合指数,可以 反映出土地利用类型整体对湿地公园水质之间的相 关性,体现出多种土地利用类型对湿地水质的综合

影响. 此外,在计算土地利用类型所占面积比的同时也考虑了该土地利用类型斑块与采样点之间距离远近的影响,更精确地得出水质与土地利用类型之间的相关关系.

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