## 新焼 様 (HUANJING KEXUE)

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## $CO_2$ 气氛热解与硝酸改性的生物炭 $Pb^{2+}$ 吸附性能对比

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摘要:酸改性处理常被用于生物炭的改性过程,但也存在酸消耗量大、废液处理难和成本高等问题.利用热解过程直接改性提高 生物炭重金属的去除效果、降低改性成本,是未来实现改性生物炭广泛使用的重要前提.为评估 CO<sub>2</sub>气氛热解法在生物炭制备和 应用方面的优势和潜力,对 CO<sub>2</sub>气氛热解与 HNO<sub>3</sub>改性生物炭对 Pb<sup>2+</sup>的去除性能进行对比分析.采用元素分析、傅里叶红外光谱 (FTIR)和X射线光电子能谱(XPS)对生物炭的元素组成和结构特性进行表征.结果表明,500℃热解条件下,HNO<sub>3</sub>改性的生物炭 产生了较多的 C=O和O=C-O-等羧基类官能团,并引入了-NO<sub>2(my)</sub>和-NO<sub>2(sm)</sub>基团,提高了生物炭的表面活性和络合能力. CO<sub>2</sub>气氛制备生物炭中含有较多的金属碳酸盐,可通过离子交换和共沉淀作用吸附去除 Pb<sup>2+</sup>.此外,CO<sub>2</sub>改性生物炭具有较大的 比表面积和更优的微孔结构,有利于 Pb<sup>2+</sup>传质扩散,促进表面吸附,W500CO<sub>2</sub>和W700CO<sub>2</sub>对 Pb<sup>2+</sup>的吸附容量提升显著,达到 60.14 mg·g<sup>-1</sup>和 71.69 mg·g<sup>-1</sup>.而 HNO<sub>3</sub>改性生物炭 W500N<sub>2</sub>-A 和 W700N<sub>2</sub>-A 对 Pb<sup>2+</sup>的最大吸附容量较低,分别为 42.26 mg·g<sup>-1</sup>和 68.3 mg·g<sup>-1</sup>.CO<sub>2</sub>改性生物炭吸附容量优于 HNO<sub>3</sub>改性生物炭,是由于比表面积和官能团的双重作用.因此,CO<sub>2</sub>气氛直接热解相比于 HNO<sub>3</sub>改性生物炭具有成本低、环境友好和重金属去除效率高等优点,是一种值得推广应用的生物炭改性方法.

关键词: 生物炭; 二氧化碳(CO<sub>2</sub>); HNO<sub>3</sub>改性; 重金属; 吸附; 矿物

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### Comparison of Pb<sup>2+</sup> Adsorption Properties of Biochars Modified Through CO<sub>2</sub> Atmosphere Pyrolysis and Nitric Acid

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**Abstract:** Acid modification has been widely used to modify the structural properties of biochars. However, acid modification led to the large consumption of acid, increased difficulty of waste effluent disposal, and a high application cost. To evaluate the advantages and application potential of biochars prepared under  $CO_2$ , utilizing pyrolysis to directly modify biochars to improve heavy metal removal efficiency and reduce production cost, would be an important prerequisite for the broad application of biochars. The sorption performance of  $Pb^{2+}$  with  $CO_2$ -modified biochars was compared with that of  $HNO_3$ -modified biochar. The elemental compositions and structural properties of biochars were characterized through elemental analysis, Fourier transform infrared spectroscopy, and X-ray photoelectron spectroscopy. The results revealed that for biochars produced at 500 °C,  $HNO_3$  modified biochars,  $CO_2$ -modified biochars had the larger specific surface area and better microporous structures, which were beneficial to the diffusion of  $Pb^{2+}$  and  $TI. 69 \text{ mg} \cdot \text{g}^{-1}$ . By contrast,  $HNO_3$ -modified biochars  $W500N_2$ -A and  $W700N_2$ -A showed the lower  $Pb^{2+}$  sorption capacity of  $W500CO_2$  and  $W700CO_2$ , which were 60. 14 mg  $\cdot \text{g}^{-1}$  and 71. 69 mg  $\cdot \text{g}^{-1}$ . By contrast,  $HNO_3$ -modified biochars  $W500N_2$ -A and  $W700N_2$ -A showed the lower  $Pb^{2+}$  sorption capacity of  $CO_2$ -modified biochars. Consequently, the  $CO_2$ -modified biochar had the advantages of low cost, environmental friendlines, and high heavy metal removal efficiency, which is a modification method worthy of promotion and application. **Key words**; biochars; carbon dioxide( $CO_2$ );  $HNO_3$ -modification; heavy metal; sorption; heavy metal; sorption; mineral

生物炭由生物质原料在限氧或缺氧条件下热解 制备生成<sup>[1]</sup>,因其表面具多孔结构,化学结构稳定,且 碳含量高,具有较大的阳离子交换容量和比表面积, 被广泛应用于污染物吸附、降解和土壤固碳等环境 处理领域<sup>[2]</sup>.但生物炭的吸附能力也受到一些因素的 限制,为扩大生物炭在环境、能源等领域的应用,生 物炭改性备受关注<sup>[3]</sup>.改性生物炭能够提高土壤固碳 能力,修复土壤重金属污染,具有广阔的市场前景. 生物炭的多孔结构、表面含氧官能团特性受生物质 类型<sup>[4]</sup>、热解温度<sup>[5]</sup>和改性剂<sup>[6]</sup>等的影响.硝酸 (HNO<sub>3</sub>)改性方法简单有效,能提升生物炭表面官能 团含量<sup>[7]</sup>及比表面积.HNO<sub>3</sub>改性过程中较高的改性 温度和较大的HNO<sub>3</sub>用量可使污泥基生物炭表面形 成更多的羧基和 C—N=O结构<sup>[8]</sup>.生物炭表面官能 团的增加,可促进其对重金属的静电吸引、离子交换 及表面螯合过程,有效提高对重金属 Pb<sup>2+</sup>、Cd<sup>2+</sup>和 Cr<sup>3+</sup>

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等的去除效率<sup>[9]</sup>.此外,生物炭中的矿物盐,如碳酸盐、硅酸盐和磷酸盐等与重金属共存时,其所含的Ca<sup>2+</sup>和Mg<sup>2+</sup>等金属离子,能和重金属离子发生离子交换反应,形成稳定的矿物沉淀<sup>[10]</sup>.然而,尽管HNO,改性能在一定程度上提高生物炭的比表面积和表面含氧官能团数量,但也会溶解生物炭的比表面积和表面含氧官能团数量,但也会溶解生物炭中的矿物灰分,降低生物炭对重金属的去除效率.此外,酸改性过程伴随着HNO<sub>3</sub>的大量消耗<sup>[11]</sup>,增加了改性生物炭的生产成本和环境风险.因此通过在热解过程直接改性提高生物炭对重金属的去除效率以及降低改性生物炭的成本,是实现改性生物炭广泛应用的重要前提.

N<sub>2</sub>气氛制备的生物炭常常存在孔坍塌现象,而 CO,气氛制备的生物炭孔结构则更清晰发达,孔壁也 更厚实<sup>[12]</sup>.另外,CO2气氛下,当生物质热解温度超过 700℃,CO2会与生物炭上的石墨结构发生反应,生成 CO气体(即 boudouard reaction 过程)<sup>[13]</sup>. 而热解环境 下, 生物质脱氢与CO反应(CO+3H<sub>2</sub>→ CH<sub>4</sub>+ H<sub>2</sub>O)<sup>[14]</sup>、热腐蚀作用(CO<sub>2</sub>+C → CO)<sup>[15]</sup>及挥发性有 机碳(VOCs)的热裂解反应(CO<sub>2</sub>+VOCs →→CO+ H<sub>2</sub>)<sup>[16]</sup>,均有利于生物炭的孔隙形成,增大生物炭的比 表面积. 若生物质中含有碱金属,热解过程中CO,还 能与碱金属离子(如Ca<sup>2+</sup>和Mg<sup>2+</sup>等)形成碳酸盐矿物, 增强生物炭对重金属的去除效果[17]. 然而,目前针对 CO2气氛与HNO3改性后生物炭的结构特征,及其对 Pb2\*的吸附性能、去除机制的对比研究还非常缺乏. 通过CO2气氛下热解改性生物炭,使用常被视为工业 废气的CO<sub>2</sub>,进行二次利用,可助力国家"双碳"战略, 实现碳中和[18].因此,本研究通过对比CO,气氛与 HNO,改性生物炭的结构性质差异,经等温吸附实验 测定了两种生物炭对 Pb2+的去除效率,分析了生物炭 比表面积、矿物含量和含氧官能团对 Pb2+吸附机制的 影响,以期明确CO2气氛改性相比传统HNO3改性的 优势及其应用前景.

#### 1 材料与方法

1.1 改性生物炭的制备

柳木生物质取自Czylok, Poland;硝酸(HNO<sub>3</sub>)、氢 氟酸(HF)、过氧化氢(H<sub>2</sub>O<sub>2</sub>)、氢氧化钠(NaOH)和溴 化钾(KBr)购自中国阿拉丁试剂有限公司;硝酸铅 [Pb(NO<sub>3</sub>)<sub>2</sub>]购自广东汕头市西陇化工厂.

采用柳木为生物质原料,将原材料磨碎烘干后 过 2 mm 筛网,放入马弗炉,在升温速率为 7 °C · min<sup>-1</sup> 的 N<sub>2</sub> 或 CO<sub>2</sub> 气氛 (3.0 L · min<sup>-1</sup>)条件下,于 500 °C 或 700 °C 热解 3 h 制备生物炭,标记为W500N<sub>2</sub>、 W500CO<sub>2</sub>、W700N<sub>2</sub>和W700CO<sub>2</sub>.对热解后的生物炭进 行改性,将生物炭与含量为 65%的 HNO<sub>3</sub>溶液混合, 在 60℃条件下混合反应 3 h,并于 80℃烘箱内干燥 4 h. 改性后的生物炭标记为 W500N<sub>2</sub>-A、W500CO<sub>2</sub>-A、W700N<sub>2</sub>-A和W700CO<sub>2</sub>-A.将原始生物炭及改性生物炭用去离子水清洗,直到 pH 值恒定不变为止,放入 65℃烘箱烘干备用.

#### 1.2 改性生物炭的表征

通过元素分析仪(UNICUBE元素分析仪,德国) 测定生物炭的元素含量;生物炭表面官能团采用傅 里叶红外光谱仪(FTIR, Variance 640-IR,美国)进行 测定;生物炭表面元素分布采用X射线光电子能谱法 (XPS, Escalab 250Xi 激发光源, Thermo Fisher Scientific,美国)进行分析;生物炭重金属元素分析采 用电感耦合等离子体质谱法(ICP-MS, NexION 350X, PerkinElmer,美国)检测;生物炭比表面积和孔 径等表面性质采用比表面积分析仪(BET, ASAP 2020,美国麦克)测定;生物炭表面微观结构和形貌特 征通过扫描电子显微镜(SEM, Regulus 8100, 日本)进 行表征.

#### 1.3 等温吸附实验

通过等温吸附实验测定改性生物炭对Pb<sup>2+</sup>的吸 附性能.由于溶液pH会影响Pb<sup>2+</sup>的存在形态,进而影 响生物炭对Pb<sup>2+</sup>的去除<sup>[19]</sup>,结合pH对Pb<sup>2+</sup>形态的影响 及生物炭表面的质子化现象,选择pH=5.8开展吸附 实验.分别称取8mg不同改性生物炭于10mL 安瓿 瓶中,加入去离子水调节溶液pH,将不同性质的改性 生物炭(pH范围为4.3~9.71)与水振荡摇匀后,通过 浓度为0.01mol·L<sup>-1</sup>的HNO<sub>3</sub>和0.01mol·L<sup>-1</sup>的NaOH 进行调节.在调好的生物炭水溶液中加入质量浓度 为5~300mg·L<sup>-1</sup>的Pb<sup>2+</sup>溶液[Pb<sup>2+</sup>原液用Pb(NO<sub>3</sub>)<sub>2</sub>固 体配制],确保最终反应体积为8mL.放入25℃恒温 摇床(150r·min<sup>-1</sup>)振荡24h,取上清液过0.45 μm水 系微孔滤膜,用原子吸收分光光度计(Hitachi Z-2000 Series,日本)测定滤液中Pb<sup>2+</sup>的平衡浓度.每批实验 重复3次.

生物炭对 Pb<sup>2\*</sup>的平衡吸附容量通过公式(1)<sup>[20]</sup> 求得:

$$Q_e = \frac{(c_0 - c_e) \times V}{m} \tag{1}$$

式中, $Q_e$ 为平衡吸附容量, $mg \cdot g^{-1}$ ;  $c_0$ 为溶液初始质量 浓度, $mg \cdot L^{-1}$ ;  $c_e$ 为溶液平衡质量浓度, $mg \cdot L^{-1}$ ; V为溶 液体积,L; m为生物炭投加量,g.

为进一步分析改性生物炭对重金属 Pb<sup>2+</sup>的吸附 机制,采用 Langmuir 和 Freundlich 等温吸附模型对 Pb<sup>2+</sup>的吸附过程进行拟合.Langmuir等温吸附模型假 定材料吸附污染物为单分子层吸附,通过公式(2)<sup>[21]</sup> 求得:

$$Q_{\rm e} = \frac{Q_{\rm m} \times K_{\rm L} \times c_{\rm e}}{1 + c_{\rm e} \times K_{\rm L}} \tag{2}$$

式中, $Q_m$ 为最大吸附容量, $mg \cdot g^{-1}$ ; $c_e$ 和 $Q_e$ 分别为Pb<sup>2+</sup> 的平衡质量浓度和平衡吸附容量, $mg \cdot L^{-1}$ 和 $mg \cdot g^{-1}$ ; $K_L$ 为吸附常数.

Freundlich等温吸附模型假设吸附剂表面为非均 质表面,吸附位点分布不均匀,吸附剂对污染物的吸 附属多层吸附,通过公式(3)<sup>[22]</sup>求得:

$$Q_{e} = K_{\rm F} \times c_{e}^{1/n} \tag{3}$$

式中,n为吸附平衡常数, $K_{\rm F}$ 为吸附常数, $c_{\rm e}$ 和 $Q_{\rm e}$ 分别为 Pb<sup>2+</sup>的平衡质量浓度和平衡吸附容量,mg·L<sup>-1</sup>和 mg·g<sup>-1</sup>.

1.4 生物炭消解实验

称量 50 mg改性生物炭于消解罐中, 留出 2 个 消解罐(无生物炭)做空白对照, 每批实验重复 3 次. 在消解罐中加入 5 mL HNO<sub>3</sub>, 1 mL HF, 2 mL H<sub>2</sub>O<sub>2</sub>, 放入赶酸器预消解 20 min, 待冷却后补加 2 mL HNO<sub>3</sub>; 在微波消解仪中消解 180 min. 待消解后, 采用 150 ℃赶酸器赶酸至消解罐中溶液剩 1 mL, 剩 余溶液倒入 50 mL 容量瓶定容.取 10 mL 溶液过 0.45 μm 水系微孔滤膜, 采用 ICP-MS 检测溶液中的 重金属含量.

#### 2 结果与分析

#### 2.1 生物炭的性质特征

生物炭元素组成如表1所示.CO,气氛制备的生 物炭产生了较多灰分,W500CO2及W700CO2的灰分 含量高达11%,而W500N2及W700N2的灰分含量均 小于5%.通过消解实验测定生物炭金属含量,结果 显示, CO2气氛下制备的生物炭中Ca2+含量较高.生 物炭灰分可能来源于热解过程中CO,与生物炭中的 金属离子反应生成了碳酸盐矿物[23].在CO,气氛条件 下,生物炭灰分的增加也导致CO2气氛制备的生物 炭C的含量(73.70%和75.88%)均低于N。气氛条件 下制备的生物炭的C含量(80.97%和85.13%).但随 着热解温度升高,CO,热解制备的生物炭灰分含量 并未显著提升,这可能是由于生物炭中有机碳的进 一步分解以及生成的部分碳酸盐矿物受热分解导 致<sup>[24]</sup> HNO<sub>3</sub>改性后 CO<sub>2</sub>热解炭灰分含量又恢复到较 低水平,这是因为HNO3环境下溶解了Ca2+和Zn2+形 成的碳酸盐矿物<sup>[25]</sup>.以上结果均证实,CO,热解制备 生物炭中灰分为碳酸盐矿物.

$\cap$ /	a 8			表し以	性生物灰兀	<u>家宮重分</u>	M S	124			(1)		
610	10		Table	Table 1 Element content analysis of modified biochar						(			
<b>生物</b> 岩	18	ω/%		Nº V	2	原子比	10.	ω(灰分) _		$\omega^{1}/\mathrm{mg}\cdot\mathrm{g}^{-1}$	3 1		
110 K	-C	Н	0/	N	H/C	0/C	O+N/C	1%	Ca <sup>2+</sup>	$Zn^{2+}$	Cu <sup>2+</sup>		
W500N <sub>2</sub>	80.97	2.44	11.69	1.27	0.36	0.11	0.12	3.63	3.701	0.22	0.015		
W500N <sub>2</sub> -A	72.51	2.52	22.13	3.61	0.42	0.23	0.27	2)	2.068	0.15	0.009 2		
W700N <sub>2</sub>	85.13	1.31	10.76	1.17	0.18	0.09	0.11	1.63	2.221	0.16	0.012		
$W700N_2$ -A	77.96	1.41	19.54	1.93	0.22	0.19	0.21	_	0.791	0.12	0.064		
W500CO <sub>2</sub>	73.70	2.43	11.83	0.95	0.39	0.12	0.13	11.09	9.81	0.051	0.019		
W500CO <sub>2</sub> -A	72.18	2.44	24.42	3.94	0.41	0.26	0.30	—	2.97	0.030	0.013		
W700CO <sub>2</sub>	75.88	0.93	10.68	0.94	0.15	0.11	0.12	11.57	7.52	0.21	0.075		
W700COA	75.25	1 54	21.22	1 71	0.24	0.21	0.23	0.28	0.041	0.11	0.015		

1)离子含量采用ICP-MS获得; 2)"一"表示未检测到灰分

经 HNO<sub>3</sub>改性生物炭 O 含量提升近一倍, O/C 和 H/C 均升高, O/C 分别从 0.09~0.12 升高至 0.19~ 0.26, H/C 分别从 0.15~0.39 升高至 0.22~0.42.这与 Li等<sup>[26]</sup>通过 HNO<sub>3</sub>(2 mol·L<sup>-1</sup>)改性竹源生物炭, 在表 面成功引入羧基、内酯基、酚基和羰基等含氧基团的 研究结果一致.这说明酸改性破坏了芳香性结构,并 在断键位置引入了羧基等含氧官能团, 提高了生物 炭的亲水性<sup>[27]</sup>.此外, 酸后改性 CO<sub>2</sub>热解炭, O 含量进 一步提升, 这可能是低温热解过程中, 生物质中 Ca<sup>2+</sup> 抑制了 C—H 断裂, 饱和烃键被保留, CO<sub>2</sub>气氛热解炭 缩合度更低, 使得酸性氧化更易引入表面含氧官 能团<sup>[28]</sup>.

通过 SEM 对生物炭表面微观结构和形貌特征

进行识别,如图1所示.N<sub>2</sub>和CO<sub>2</sub>气氛制备的生物炭 有明显差异:前者侧壁较光滑[图1(b)~1(d)],孔 隙结构不均匀;后者表面粗糙,孔隙多且均匀[图1 (f)~1(h)].而HNO<sub>3</sub>改性生物炭侧壁和孔隙结构 均出现坍塌,形成更多的微介孔结构<sup>[29]</sup>.表2中的 BET结果显示,W700CO<sub>2</sub>的比表面积高达494.6 m<sup>2</sup>·g<sup>-1</sup>,相比于同一温度下N<sub>2</sub>气氛制备的生物炭提 升了182.5倍,平均孔径也由16.89 nm降至3.86 nm.这是由于CO<sub>2</sub>的热腐蚀作用(CO<sub>2</sub>+C→CO)以 及VOCs的热裂解(CO<sub>2</sub>+VOCs→CO+H<sub>2</sub>),导致了 比表面积的显著增加和孔径的明显缩小,为孔隙形 成提供了有利的条件<sup>[30]</sup>.而500℃未达到CO<sub>2</sub>与生 物炭中有机物的反应温度,对比表面积、微孔形成





Fig. 2 XPS spectral analysis of modified biochars

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的贡献度不高<sup>[31]</sup>. HNO<sub>3</sub>改性后导致 W700CO<sub>2</sub>的比 表面积降至 366.14 m<sup>2</sup>·g<sup>-1</sup>,下降为原来的 74%,这可 能是浓 HNO<sub>3</sub>的强氧化性腐蚀生物炭表面造成孔隙 壁塌陷导致.

Table 2 Analysis of specific surface area and pore size of modified biochars								
生物炭	比表面积(BET)/m <sup>2</sup> ·g <sup>-1</sup>	BJH孔径/nm	BJH 孔体积/cm <sup>3</sup> ·g <sup>-1</sup>					
W500N <sub>2</sub>	2.89	6.32	0.023					
$W500N_2$ -A	2.27	7.35	0.072					
W700N <sub>2</sub>	2.71	16.89	0.011					
$W700N_2$ -A	3.43	17.78	0.016					
W500CO <sub>2</sub>	1.06	6.83	0.032					
W500CO <sub>2</sub> -A	2.71	6.83	0.058					
W700CO2	494.6	3.86	0.045					
W700C0A	366.14	3.75	0.039					

表 2 改性生物炭比表面积与孔径分析

XPS 能谱图 2(a)显示生物炭主要由 C 1*s* 和 O 1*s* 组成,此外还含有极少量的 N 1*s* 和 Ca 2*p*. C 1*s* 的高分 辨率 XPS 光谱如图 2(b)所示,含 C 化学键主要包括 C—C/C=C 键 (284.78 eV)<sup>[32]</sup>, C—O/C—OH 键 (285.92 eV)<sup>[33]</sup>,及 O—C=O 键<sup>[34]</sup>(288.93 eV).相比 于 W500N<sub>2</sub>,经 CO<sub>2</sub>和 HNO<sub>3</sub>改性后,C—C/C=C 键的 占 比 均 下 降 (C—C/C=C 从 72.37% 下 降 至 62.84%),C—O/C—OH 键和 O—C=O 键的占比均出 现上升(C—O/C—OH 从 20.92% 提升至 27.13%, O—C=O 键从 6.70% 提升至 10.03%),这表明酸改 性和 CO<sub>2</sub>热解破坏了 C—C/C=C 键,并在断键处形成 了含氧官能团.

O1s的高分辨率XPS光谱如图2(c)所示, 含O化 学键主要包括C--O键(531.50~532.0eV)<sup>[35]</sup>和C==O 键(532.5~533.0eV)<sup>[36]</sup>.根据图谱分析,CO<sub>2</sub>气氛热解 制备的生物炭中,W500CO<sub>2</sub>的C==O官能团原子百分 比低于同温度下N<sub>2</sub>气氛下制备的生物炭(W500N<sub>2</sub>), 而W700CO<sub>2</sub>的C==O官能团原子百分比高于W700N<sub>2</sub> 样品.这可能是因为,在低温条件下,生物炭中的矿 物质起到了保护作用,阻碍了C--O官能团向C=O官 能团的转化;而在高温条件下,部分矿物质被分解,保 护作用减弱,C--O官能团更容易被氧化为C=O官能 团.这表明高温条件下CO<sub>2</sub>气氛对生物炭中C=O键的 形成有促进作用.在强酸的作用下,W500N<sub>2</sub>、 百分比分别降低了 19.01%、4.05%、6.02% 和 9.32%,而C—O官能团原子百分比则相应增加.这可 能是因为强酸使得C—O键断裂,形成了C—O键<sup>[37]</sup>.

FTIR 光谱揭示了各生物炭的表面所含官能团差 异,在3439~3446 cm<sup>-1</sup>的伸缩振动峰为酚羟基、醇羟 基及水分子羟基; 2913 cm<sup>-1</sup>的吸收峰为烷烃 C-H<sup>[38]</sup>;在1520 cm<sup>-1</sup>与1343 cm<sup>-1</sup>的吸收峰是 -NO<sub>2(asy)</sub>和-NO<sub>2(sym</sub>),在1056 cm<sup>-1</sup>处的尖峰属于C-O-C和C-C拉伸<sup>[39]</sup>;871 cm<sup>-1</sup>处吸收峰为芳香族C 一H面外弯曲振动峰<sup>[40]</sup>;位于1616~1618 cm<sup>-1</sup>的强吸 收峰属于羰基、羧基中的C=O拉伸振动[41],1384 cm<sup>-1</sup>处的吸收峰为酯基、羧基的0=C-0-拉伸振 动<sup>[42]</sup>. 图 3 显示, HNO3 改性后, 500°C下 N2和 CO2下制 备的生物炭上0=C-0-和C=0、-NO<sub>2(av)</sub>和-NO<sub>2(sym</sub>)的官能团峰值增强,说明酸改性引入了羧基类 官能团,碳基质中氢离子被HNO。中硝基亲电取代生 成了硝基团<sup>[43]</sup>.在700℃下,生物质在CO2气氛中热裂 解制备的生物炭C-O-C的拉伸峰明显高于同温N, 气氛制备生物炭,这是由于700℃下CO2分子能够与 生物炭表面碳基质发生反应插入氧原子,提高生物 炭表面含氧官能团的数量.

2.2 等温吸附实验结果

采用 Langmuir 和 Freundlich 吸附等温方程对 Pb<sup>2+</sup>吸附进行拟合分析,如图4所示,拟合参数列 于表3.

生物炭 一	L	angmuir吸附等温曲	线	Freundlich吸附等温曲线			
	$K_{\rm L}/{\rm L}\cdot{\rm mg}^{-1}$	$Q_{\rm e}/{ m mg} \cdot { m g}^{-1}$	$R^2$	$K_{\mathrm{F}}/\mathrm{mg}^{1-(1/n)} \cdot \mathrm{L}^{1/n} \cdot \mathrm{g}^{-1}$	1/n	$R^2$	
W500N <sub>2</sub>	0.987	18.35	0.807 1	9.58	0.18	0.770 8	
$W500N_2$ -A	0.355	42.26	0.859 5	16.34	0.24	0.896 7	
W700N <sub>2</sub>	4.20	15.41	0.802 7	11.95	0.071	0.631 3	
$W700N_2$ -A	0.121	68.35	0.932 6	11.47	0.46	0.939 6	
W500CO2	0.019 4	60.14	0.926 1	3.42	0.54	0.957 0	
W500CO <sub>2</sub> -A	0.070 8	44.74	0.877 1	6.94	0.41	0.924 6	
W700CO <sub>2</sub>	18.207	71.69	0.797 1	62.36	0.045	0.997 7	
W700CO <sub>2</sub> -A	0.338	80.75	0.931 6	35.52	0.21	0.959 5	

表 3 改性生物炭 Pb<sup>2+</sup>吸附等温曲线拟合参数 Table 3 Modified biochar Pb<sup>2+</sup> adsorption isotherm curve fitting parameters



Fig. 4 Pb<sup>2+</sup> adsorption isotherm of modified biochars

由图4可知,酸改性能够提高生物炭对重金属 的吸附能力.HNO,改性后,W500N,-A、W700N,-A和 W700CO<sub>2</sub>-A的最大吸附容量分别从18.35、15.41 和 71.69 mg·g<sup>-1</sup>提升至 42.26、68.35 和 80.75 mg· g-1. 然而,相比于 N,气氛制备的生物炭在 HNO,改性 后吸附量提升了 2~3 倍, CO, 气氛制备的生物炭在 HNO,改性后吸附量仅提升了14%,提升效果有限. 另一方面,经CO,气氛制备的W500CO,和W700CO,, 其最大吸附容量分别达到 60.14 mg·g<sup>-1</sup>和 71.69 mg· g<sup>-1</sup>,显著高于在 N,气氛下制备的 W500N,和 W700N, (18.35 mg·g<sup>-1</sup>和15.41 mg·g<sup>-1</sup>),分别是N<sub>2</sub>气氛制备 生物炭的3.2倍和4.6倍.然而,经过HNO。改性后 W500CO,-A的最大吸附容量由 60.14 mg·g<sup>-1</sup>下降为 44.74 mg·g<sup>-1</sup>, 与 W500N<sub>2</sub>-A (42.26 mg·g<sup>-1</sup>)相似, 而 W700CO,-A的最大吸附容量仅略高于W700CO,,为 80.75 mg $\cdot$ g<sup>-1</sup>.

由表3可知,W500N<sub>2</sub>和W700N<sub>2</sub>生物炭对Pb<sup>2+</sup>的 吸附更符合Langmuir单层等温吸附模型,而酸改性的 W500N<sub>2</sub>-A和W700N<sub>2</sub>-A生物炭则更适用于Freundlich 多层等温吸附模型.此外,CO<sub>2</sub>热解炭无论是否经过 HNO<sub>3</sub>改性,都表现出较高的Freundlich多层等温吸附 模型拟合度.

#### 3 讨论

改性生物炭对 Pb2+的吸附结果表明,酸改性能 增强生物炭对重金属的吸附效果.酸改性后生物 炭上含有的含氧官能团能与Pb2+发生络合,导致吸 附 Pb2+后 pH 值显著降低[10]. 此外,生物炭吸附 Pb2+ 后,一OH和C=O组分会发生显著减少[44].本研 究中,通过生物炭 Pb<sup>2+</sup>吸附实验前后的 FTIR 图谱 对比,发现吸附重金属后在1616~1618 cm<sup>-1</sup>处的 C=O以及1384 cm<sup>-1</sup>处的酯基、羧基的0=C-O 一信号出现不同程度的降低(图 5),这表明官能团 的表面络合在 Pb2+吸附中发挥了重要作用.此外, 通过 BET 分析发现,除 W700CO,外,其余 3 种未经 HNO,改性的生物炭比表面积均不超过3m<sup>2</sup>·g<sup>-1</sup>;但 W500CO,较低的表面积却有较高的Pb<sup>2+</sup>吸附容量, 这说明比表面积不是影响吸附效果的唯一因素. 表1中元素分析及ICP-MS结果表明,W500CO。的灰 分中的主要成分为碳酸盐矿物,碳酸盐中的金属 离子能与溶液中的 Pb<sup>2+</sup>发生阳离子交换及共沉淀 反应,为Pb2+的去除创造有利条件.但是经HNO,改



图5 吸附后改性生物炭红外光谱(FTIR)分析

 $Fig. \ 5 \quad Infrared \ spectrum \ (\ FTIR \ ) \ analysis \ of \ modified \ biochars \ after \ adsorption$ 

性后 W500CO<sub>2</sub>-A 的平衡吸附容量由 60.14 mg·g<sup>-1</sup>下 降为 44.74 mg·g<sup>-1</sup>(表 3),这可能是 HNO<sub>3</sub>环境下 Ca<sup>2+</sup>、Zn<sup>2+</sup>等金属元素发生浸出而被水洗去除,进而 降低了 Pb<sup>2+</sup> 与碳酸盐的共沉淀作用<sup>[45]</sup>.虽然 W700CO<sub>2</sub>-A 的金属盐矿物也被去除,但高温条件 下,CO<sub>2</sub>的热腐蚀作用(CO<sub>2</sub>+C→CO)以及 VOCs 热裂解(CO<sub>2</sub>+VOCs→CO+H<sub>2</sub>)增加了生物炭的比 表面积和孔隙结构,为 Pb<sup>2+</sup>提供了更多的吸附位 点,弥补了灰分减少的影响,Pb<sup>2+</sup>的吸附容量稍有 提升.

综上所述,采用CO<sub>2</sub>、HNO<sub>3</sub>改性的生物炭是一种 高效的Pb<sup>2+</sup>吸附剂.改性后的生物炭表面官能团含量 增加,具有较大的比表面积和优化的孔隙结构,有利 于Pb<sup>2+</sup>在生物炭表面发生络合、离子交换和共沉淀等 吸附反应.对中温生物炭而言,CO<sub>2</sub>直接热解改性生 物炭对Pb<sup>2+</sup>的去除效果明显高于N<sub>2</sub>再经酸改性的生 物炭;而高温条件下,CO<sub>2</sub>生物炭的孔隙度丰富,显著 增加了生物炭对Pb<sup>2+</sup>的去除效果.因此,利用CO<sub>2</sub>直 接热解低成本生物质制备高效的环境修复材料,能 作为高成本HNO<sub>3</sub>改性生物炭的替代性方法,值得进 一步推广应用.

### 4 结论

(1)500 ℃ CO<sub>2</sub>气氛热解和 HNO<sub>3</sub>改性都能增加生物炭表面官能团含量,并有效促进 Pb<sup>2+</sup>的络合去除, 其最大吸附容量分别从 18.35 mg·g<sup>-1</sup>提升至 60.14 mg·g<sup>-1</sup>和 42.26 mg·g<sup>-1</sup>.

(2)热解过程能促进 CO<sub>2</sub>与生物炭中所含金属反应生成碳酸盐矿物,有利于 Pb<sup>2\*</sup>离子交换作用及共沉 淀反应的发生,吸附效率提升明显.

(3)700℃ CO<sub>2</sub>气氛下制备的生物炭比表面积达 到494.6 m<sup>2</sup>·g<sup>-1</sup>,比 N<sub>2</sub>气氛下制备的生物炭具有更大 的比表面积和更优的微孔结构,改善了 Pb<sup>2+</sup>传质扩散 过程,最大吸附容量由 15.41 mg·g<sup>-1</sup>提升至 71.69 mg·g<sup>-1</sup>.

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Environmental Science (monthly)

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