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太湖水体和沉积物中有机磷酸酯的时空分布和风险评估

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摘要:为探究有机磷酸酯(OPEs)在太湖水环境中的时空分布特征和生态风险水平,利用固相萃取和超高效液相色谱-串联质谱方法测定太湖水体和沉积物中的单体和低聚 OPEs,其中单体 OPEs 包括烷基 OPEs、氯代烷基 OPEs 和芳香基 OPEs. 结果发现,OPEs 在太湖水体和沉积物的空间分布差异不显著;水体和沉积物赋存的 OPEs 皆有明显的季节差异性,夏季 OPEs 的赋存含量平均值(水体中为 752.7 ng·L⁻¹,沉积物中为 124.0 ng·g⁻¹)显著高于冬季(水体中为 498.5 ng·L⁻¹,沉积物中为 54.5 ng·g⁻¹),其中单体 OPEs 的季节差异性尤为明显;不同类型 OPEs 在水体和沉积物中的赋存浓度排序均为:氯代烷基 OPEs > 烷基 OPEs > 芳香基 OPEs «低聚 OPEs; OPEs 在水体和沉积物间的伪分配系数(K_a)在夏季(0.05~4.17 L·g⁻¹)高于冬季(0.02~3.47 L·g⁻¹),且与其辛醇水分配系数显著正相关.基于 OPEs 在太湖水体中的中位浓度,采用风险商值(RQ)法评估太湖水环境 OPEs 的生态风险,发现水体中的 OPEs 整体上呈现中等生态风险(RQ 在冬、夏两季分别约为 0.34 和 0.35);冬夏两季不同类型 OPEs 的生态风险排序与 OPEs 的赋存浓度相对应;单体 OPEs 在夏季的生态风险高于冬季;虽然评估结果表明低聚 OPEs 的生态风险较低,并且低于单体 OPEs,但目前有关低聚 OPEs 的毒理研究还鲜见报道,亟需开展深入研究以提供数据支持.

关键词:有机磷酸酯(OPEs);太湖;地表水;沉积物;生态风险

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Spatiotemporal Occurrence of Organophosphate Esters in the Surface Water and Sediment of Taihu Lake and Relevant Risk Assessment

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Abstract: Aiming to explore the spatiotemporal occurrence of organophosphate esters (OPEs) in the aquatic environment of Taihu Lake and to assess the relevant ecological risk, monomeric and oligomeric OPEs in the surface water and sediment of Taihu Lake were determined using solid-phase extraction and ultra-high performance liquid chromatography-tandem mass spectrometry. The target monomeric OPEs included chlorinated OPEs, alkyl OPEs, and aryl OPEs. There was no significant difference in the spatial distribution of OPEs in water and sediment between the pollutant-impacted and less impacted zones. The average concentrations of OPEs in summer and winter were, respectively, 752.7 and 498.5 ng^*L^{-1} in water and 124.0 and 54.5 ng^*g^{-1} in sediment, indicating an obvious seasonal difference, especially for the monomeric OPEs. The OPEs levels in both the water and sediment were ranked in the order of chlorinated OPEs > alkyl OPEs > aryl OPEs oligomeric OPEs. The pseudo-partitioning coefficients (K_d) of OPEs between the sediment and surface water of Taihu Lake were much higher in summer (0.05-4.17 L^*g^{-1}) compared to those in winter (0.02-3.47 L^*g^{-1}) and were significantly positively correlated with the lg K_{ov} values of OPEs. Based on the median concentrations of OPEs in the water of Taihu Lake, the ecological risk of OPEs was assessed by risk quotient (RQ) values, which indicated a medium risk at an RQ level of 0.34 during winter and 0.35 during summer. The assessment results showed that the risk ranking of OPEs was consistent with that of their concentrations, and the monomeric OPEs posed a higher ecological risk in summer compared to that in winter. Although the ecological risk of oligomeric OPEs in this study was not serious and was lower than that of monomeric OPEs, it is an urgent requirement to conduct ecotoxicology studies on oligomeric OPEs in the future since available data is highly limited at present.

Key words: organophosphate esters (OPEs); Taihu Lake; surface water; sediment; ecological risk

有 机 磷 系 阻 燃 剂(organophosphate flame retardants, OPFRs)作为溴代阻燃剂的替代品,在全球的需求增长迅速,近些年已跻身于第一大类阻燃剂行列^[1,2].中国是 OPFRs 生产大国,拥有 201 家生产厂,约占全球 55%的份额^[3].广泛应用的 OPFRs 主要为有机磷酸酯(organophosphate esters, OPEs),OPEs 按取代基的不同可划分为烷基 OPEs、氯代烷基OPEs和芳香基 OPEs. 该阻燃剂通过物理混合而非化学结合的方式添加到电子设备、纺织品、塑料、新能源汽车、电缆和家具等产品中^[4],因而容易经挥发和磨损释放到环境中^[2].截至目前,OPEs 在大气、室

内、地表水、污水、土壤等环境中皆有检出^[5-9]. 然而,有关 OPEs 在环境中赋存特征的研究仍然存在很大空白. 例如:大多数研究仅关注单体 OPEs 而忽略了新兴低聚 OPEs 在环境中的赋存状况^[10];不同季节条件下环境中 OPEs 的赋存差异还鲜有报道. 鉴于OPEs 的潜在危害,如引发斑马鱼甲状腺功能障碍^[11]、抑制鲢鱼生长发育^[12]、影响孩童骨骼健康^[13]、

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境污染与修复,E-mail: zcn9809@163.com * 通信作者,E-mail: liyifei@jiangnan.edu.cn 增强直肠癌细胞活性^[14]等,非常有必要针对 OPEs 的环境赋存特征开展更广泛、更深入的研究,进而为 OPEs 生态环境风险的评估以及相关治理技术的探索提供支持.

地表水环境与人类息息相关,具有生物多样性 保护、饮用水供给、农业灌溉、水产养殖和休闲娱乐 等重要功能[15,16]. 近几年, OPEs 在地表水环境中的赋 存情况日益引起关注.曹渺等[17]、李栋等[18]和 Lv 等[19]分别对黄河、长江和鄱阳湖中赋存的单体 OPEs 开展了相关研究,发现氯代OPEs在水体中的浓度显 著高于其他 OPEs,并且此类 OPEs 的潜在风险高.除 单体 OPEs 外, 地表水环境中的低聚 OPEs 也不容忽 视,黄河[20]、长江[21]和莱州湾[21]等地皆检出低聚 OPEs,并且有研究发现太湖沉积物中双酚 A-双(二 苯基磷酸酯)(BDP)的浓度与单体OPEs的浓度相 当[10]. 在地表水环境的水体和沉积物中, Zhao等[10]和 Liu 等[22]的研究发现 OPEs 的赋存状况存在明显差异, 这一现象一方面可能归因于水体是污染物的载体而 沉积物是污染物的"汇"和"源",另一方面可能与 OPEs 的物化特性(如 $\lg K_{ow}$ 值)有关^[23]. 此外,有学者 针对不同季节下 OPEs 的赋存情况,以黄河[20]和长 江[24]为模型开展了研究,发现 OPEs 的赋存有显著性

季节差异.因而,针对典型地表水环境,同时调研水体和沉积物中单体和低聚 OPEs 的赋存特征,并分析季节的影响有重要意义.

太湖是中国第三大淡水湖泊,是重要的饮用水水源地和水产养殖基地^[25],毗邻苏州、无锡和湖州等经济发达地区,是中国污染较为严重的大型湖泊之一^[26,27].本研究选择太湖作为地表水环境模型,分析了12种单体OPEs和3种低聚OPEs在冬夏两季太湖水体和沉积物中的污染水平和分布特征,并采用风险商值法评估了不同类型OPEs在冬夏两季间的风险差异.本研究既可以补充不同季节太湖水环境中OPEs赋存差异的研究空白,也可以为太湖流域的OPEs防控和环境管理提供依据,还可以为地表水环境中OPEs风险评估模型的建立提供数据支持.

1 材料与方法

1.1 标样与试剂

15种目标 OPEs 的标准品(纯度 ≥ 99%)购自德国 Dr. Ehremstorfer 公司和加拿大 Toronto Research Chemicals公司,其名称及理化性质如表1所示。实验分析所用的乙腈(色谱级)、二氯甲烷(色谱级)和甲醇(色谱级)购自国药集团化学试剂公司(北京).

表 1 15种目标 OPEs 的名称及理化性质

1)	Table 1 Names and pi	nysicocnemicai pi	opernes of the 1	5 target OFEs	1	-	1 1
分类	名称	缩写	CAS号	分子式	相对分 子质量	$\lg K_{_{\mathrm{ow}}}$	25℃时溶解度 / mg·L ⁻¹
VB VI	磷酸三丁酯	ТВР	126-73-8	$C_{12}H_{27}O_4P$	266.3	4	2.8×10 ²
lester P	磷酸三乙酯	TEP	78-40-0	$\mathrm{C_6H_{15}O_4P}$	182.1	0.8	4.9×10^{5}
烷基 OPEs	磷酸三丙酯	TPrP	513-08-6	$\mathrm{C_9H_{21}O_4P}$	224.1	2.3	4.0×10^{5}
	磷酸三(丁氧基乙基)酯	TBEP	78-51-3	${\rm C_{18}H_{39}O_{7}P}$	398.5	3.8	1.1×10^{3}
	磷酸三辛酯	TEHP	78-42-2	$\mathrm{C_{24}H_{51}O_4P}$	434.6	4.2	2.0×10 ⁻⁶
← /\ \\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\	磷酸三(2-氯乙基)酯	TCEP	115-96-8	$C_6H_{12}C_{13}O_4P$	285.5	1.8	7.8×10 ⁴
氯代烷基 OPEs	三-(2-氯异丙基)磷酸酯	TCPP	13674-84-5	${\rm C_9H_{18}C_{13}O_4P}$	327.6	2.6	1.6×10^{3}
	磷酸三(1,3-二氯异丙基)酯	TDCP	13674-87-8	${\rm C_9H_{15}C_{16}O_4P}$	430.9	3.7	7.0
	磷酸三甲苯酯	TCrP	1330-78-5	$\mathrm{C_{21}H_{21}O_4}$	368.4	6.3	0.1
芳香基	磷酸三间甲苯酯	TMPP	563-04-2	$\mathrm{C_{21}H_{21}O_4P}$	368.4	5.4	8.1×10^{-7}
OPEs	磷酸三苯酯	TPhP	115-86-6	$(C_6H_5)_3PO_4$	326.3	4.6	1.9
	2-乙基己基二苯基磷酸酯	EHDPP	1241-94-7	${\rm C}_{20}{\rm H}_{27}{\rm O}_4{\rm P}$	362.4	5.7	1.9
/代 形文	双酚 A-双(二苯基磷酸酯)	BDP	5945-33-5	${\rm C_{39}H_{34}O_8P_2}$	692.6	8.2	1.9×10^{-1}
低聚 OPEs	间苯二酚四苯基二磷酸酯	RDP	57583-54-7	${\rm C_{30}H_{24}O_8P_2}$	574.5	6.4	7.4×10^{-2}
	2,2-双氯甲基-三亚甲基-双[双(2-氯乙基)磷酸酯	i] V6	38051-10-4	$\rm C_{13} H_{24} C_{16} O_8 P_2$	583.0	1.9	58.0

1.2 研究区域和样品采集

太湖坐落在经济发达、人口密集的中国长三角地区,其水域面积为2338km²,水深为1.9 m^[28].太湖具有明显的进水和出水区域,其北部及西北部为主要的进水区,出水主要通过东南部的太浦河流出^[29,30].太湖周边的常住人口及企业在北部及西北部尤为密集,因而有更多的污染物从这些区域人

湖^[31-33]. 基于太湖进出水方向及污染物入湖特点,本研究将太湖划分为污染物高影响区和低影响区,采样点分布如图1所示.

样品采集工作于2022年1月(冬季)和8月(夏季)开展,采样点共计17个,依次用L1~L17表示,具体位置如图1和表2所示.在每个采样点,表层水样使用不锈钢采水器采集约1200mL(水面下0~0.5

m),避光保存于棕色玻璃瓶中;表层沉积物样品使用不锈钢抓泥器采集约500g(深度0~10cm),放入装有铝箔的塑封袋内.水样和沉积物样品皆冷藏运回实验室,水样在24h内完成预处理,沉积物样品放入-80℃冰箱内以备冷冻干燥.

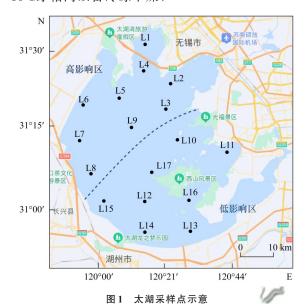


Fig. 1 Sampling locations in Taihu Lake

表 2 太湖采样点的地理坐标信息

Table 2 Geographic coordinates of sampling locations in Taihu Lake

/ 1	/ 4 1	1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1
采样点	经度(E)/(°)	纬度(N)/(°)
~L1	120.172 767	31.468 611
L2	120.246 161	31.378 889
13/	120.229 464	31.310 278
VB 141	120.162 222	31.396 619
L5	120.096 944	31.333 333
L6	119.963 889	31.311 111
L7	119.958 363	31.216 667
L8	119.984 127	31.139 818
L9	120.103 333	31.225 833
L10	120.269 722	31.232 778
L11	120.439 255	31.189 404
L12	120.150 575	31.062 778
L13	120.299 982	30.977 494
L14	120.134 667	30.993 056
L15	120.028 696	31.060 364
L16	120.311 624	31.053 876
L17	120.189 722	31.134 722

1.3 样品预处理

水样先后用 1.2 μm 和 0.7 μm 的玻璃纤维滤膜 (直径 47 mm, Whatman, 英国)过滤,过滤后的水样量取 1 L,用盐酸调 pH 为 3. 然后,采用固相萃取法对水样进行净化和浓缩,所用的固相萃取小柱为 ENVI-18 (6 mL,500 mg, Supelclean,美国),固相萃取步骤如下:用 15 mL二氯甲烷、15 mL乙腈和 15 mL超纯水(pH = 3)对小柱进行活化;以每秒 1~2滴的流速上

样;上样完毕后用 10 mL 含 10% 甲醇的超纯水淋洗;淋洗完成后负压抽干 30 min;干燥结束后,使用 12 mL的二氯甲烷和乙腈混合溶剂(体积比为 1:3)进行洗脱;收集洗脱液氮吹至干,最后用乙腈定容至500 μL.

沉积物样品经冷冻干燥后,研磨并过100目筛,取1g样品置于30mL玻璃离心管内,采用超声离心法提取样品中的目标OPEs,具体步骤如下:在样品中加入10mL乙腈,涡旋振荡1min,超声30min,离心20min(3000r·min⁻¹),上述步骤重复3次,合并上清液至棕色玻璃瓶内,加入超纯水(pH=3)定容至500mL. 然后,参考水样固相萃取的步骤对定容后的样品进行固相萃取,氮吹至干后用乙腈定容1mL.

1.4 目标 OPEs 分析

目标 OPEs 分析所用的仪器为超高效液相色谱-三重四级杆串联质谱仪 (UPLC - MS/MS, Waters ACQUITY UPLC Xevo TQ,美国),配置的色谱柱型号为 Waters ACQUITY BEH C18(2.1 mm×100 mm, 1.7 μ m). 色谱分离的柱温为 40° C,进样量为 5 μ L,以 5 mmol·L⁻¹乙酸铵为流动相 A,以乙腈为流动相 B,流动相流速为 0.4 mL·min⁻¹,相关的梯度洗脱程序如下:0~0.25 min, 10% B; 0.25~3 min, 10%~90% B; 3~4 min, 90% B; 4~4.01 min, 90%~10% B; 4.01~5 min, 10% B. 质谱分析的条件为:离子源温度 150%,脱溶剂温度 500%,锥孔反吹气流量 50 L·h⁻¹. 脱溶剂气流量 900 L·h⁻¹,碰撞气流量 0.15 mL·min⁻¹. 所有目标OPEs 在电喷雾正离子模式下分析,相关的离子对、驻留时间和碰撞能信息如表 3 所示.

表3 目标OPEs的质谱信息

Table 3 Mass spectrometry conditions for the target OPEs

化合物名称	母离子(m/z)	子离子 ¹⁾ (m/z)	驻留时间 /min	碰撞能 /eV
TEP	182.9	98.79	0.027	18
TPrP	225.28	98.85	0.022	16
TBP	267.32	98.84	0.22	18
TCEP	285.18	98.90	0.027	24
TCPP	329.00	99.00	0.022	30
TPhP	327.29	215.05	0.022	28
EHDPP	362.90	250.86	0.058	20
TMPP	368.90	164.96	0.029	48
TCrP	368.97	164.99	0.029	50
TBEP	399.40	299.33	0.029	12
TDCP	430.83	98.82	0.022	24
TEHP	435.17	98.78	0.022	14
RDP	574.84	151.74	0.029	66
V6	582.65	360.77	0.022	20
BDP	693.40	367.19	0.078	34

1)子离子为定量子离子

1.5 质量控制与保障(QA/QC)

OPEs的定量分析方法为外标法,为有效实现质 量控制与保障,采用自来水和空白沉积物样品制备 目标 OPEs 的标准溶液以避免基质效应. 具体方法 为:首先按照样品的预处理方法(1.3节)对自来水和 空白沉积物进行预处理,然后在氮吹至干复溶后的 乙腈溶液中加入15种目标OPEs的标准品,配制浓度 为 1、5、10、50、100 和 200 μg·L⁻¹的标准溶液. 空白 样品中皆未检出 OPEs,15种 OPEs 的标准曲线线性相 关系数 R^2 均大于 0.99. 水样 OPEs 检出限(LOD, S/N = 3)为 0.01~0.34 ng·L⁻¹,方法定量限(LOO,S/N = 10)

$$RQ_{\text{MEC/PNEC}} = \sum_{i=1}^{n} \frac{\text{MEC}_{i}}{\min\left(\text{EC50}_{i,\text{algae}}, \text{EC50}_{i,\text{crustacean}}, \text{EC50}_{i,\text{fish}}\right)} \times AF$$

式中,MEC表示太湖水体中目标 OPEs 的实际浓度 (μg·L⁻¹); PNEC 为预测无效应浓度(μg·L⁻¹),通过查 找最为敏感的水生生物的半数致死浓度,再按照计 算公式得到;MEC表示检测的n类混合物中的第i个 组分的实际最大环境浓度;EC50,是指n类混合物中 第 i 个组分的半致死浓度(包括藻类、甲壳类、鱼类 的半致死浓度); RQ_{MEC/PNEC}表示基于 MEC/PNEC 总和 的风险商;AF为评价因子,采用欧盟水框架指令的 推荐值 10~1 000. 当 RQ<0.1 时,生态风险低;当 0.1 ≤ RQ < 1.0 时, 生态风险中等; 当 RQ ≥ 1.0 时, 生 态风险高.

1.7 数据处理

采用 Origin 2021 和 Excel 2020 软件处理数据并 作图,采用IBM SPSS Statistics 25分析不同类型 OPEs 的季节差异性以及高、低影响区间的差异性(独立样 本 T 检验).

2 结果与讨论

2.1 太湖水体中 OPEs 的时空分布特征

在冬夏两季,太湖水体中OPEs在各采样点的浓 度分布如图 2 所示. 由图 2 可知,15 种目标 OPEs 除 RDP在部分采样点检出外(冬季检出率为35.3%,夏 季检出率为41.2%),其余OPEs在冬夏两季的检出率 均为 100%. 与 2014 年[10] 和 2019 年[25] 太湖水体中 OPEs的检出率相比,本研究发现有6种OPEs(TBEP、 TEHP、EHDPP、TMPP、BDP和RDP)的检出率有较 明显增加,预示着太湖周边 OPEs 的使用及排放正在 逐年增多.

从空间分布来看,在高影响区, $\rho(\sum_{i,j} OPEs)$ 范 围为 390.5~624.4 ng·L⁻¹(冬)和 662.81~1 002.20 $\operatorname{ng·L}^{-1}(夏)$;在低影响区, $\rho\left(\sum_{15} \operatorname{OPEs}\right)$ 范围为 388. 10 ~ 679. 18 ng·L⁻¹(冬)和 315. 79~1 161. 39 ng·L⁻¹(夏).

为 0.03 ~ 1.14 ng·L⁻¹; 沉积物 OPEs 检出限(LOD, S/ N = 3)为 0.09 ~ 0.25 ng·g⁻¹,方法定量限(LOQ,S/N = 10)为0.12~0.68 ng·g-1.在每批样品分析时,同时检 测一个加标样品(质量控制样品),用于分析回收率, 以保障检测质量.水样的加标回收率为64.8%~ 100.3%, 沉积物的加标回收率为61.0%~99.7%. 实 验过程用到的所有玻璃容器均经过自来水和超纯水 清洗3遍,最后超声烘干以避免环境污染.

1.6 生态风险评估

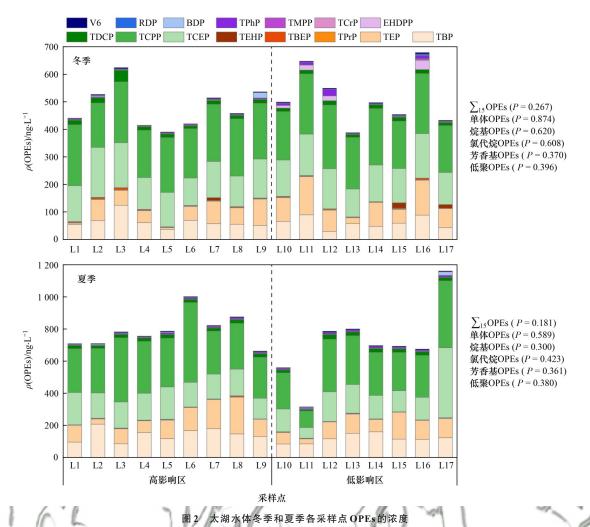
采用风险商值(risk quotient, RQ)法对太湖水体 中 OPEs 的风险进行评估, 计算公式如下所示[21,24]:

$$\frac{\text{MEC}_i}{\text{EC50}_i} \times \text{AF}$$

通过独立样本 T 检验分析发现,在冬夏两季, $\sum_{i,j}$ OPEs在两个区域之间皆没有显著性差异(P >0.05). 张文萍等[25]研究发现,太湖西北部 OPEs 的浓 度水平较高,但从整个太湖来看OPEs在水体中的浓 度无明显差异,与本研究的发现基本一致,庄园[34]的 研究表明,太湖水体中的OPEs同样在西北部污染较 为严重,并将太湖水体中OPEs的来源归因于入湖河 流的携带.

基于太湖水体中OPEs空间分布差异性小的特 点,本研究从整个太湖的角度分析了不同类型OPEs 在冬夏两季的赋存浓度(图3). 在冬季,太湖水体中 氯代烷基 OPEs 的浓度最高,尤其是 TCEP(99.71~ 181.00 ng·L⁻¹)和 TCPP(161.42 ~ 231.02 ng·L⁻¹);烷 基 OPEs 中的 TBP 和 TEP 的浓度也较高,其浓度范围 分别为 28.85 ~ 124.29 ng·L⁻¹和 5.49 ~ 138.95 ng·L-1; 芳香基OPEs 和低聚OPEs 的浓度相对较低, 最 高浓度皆低于50 ng·L-1,其中浓度较高的芳香基 OPEs有 EHDPP(0.70~33.40 ng·L⁻¹)和 TPhP(1.76~ 24. 05 ng·L⁻¹),浓度较高的低聚 OPEs 为 BDP(0. 30~ 19.44 ng·L-1). 在夏季,不同类型 OPEs 的浓度排序和 冬季基本一致:氯代烷基OPEs>烷基OPEs>芳香基 OPEs≈低聚 OPEs. 但是,值得注意的是夏季 ∑₁₁₅ OPEs 的浓度明显高于冬季,其中单体 OPEs 中除了芳香基 OPEs之外,其余单体 OPEs(氯代烷基 OPEs和烷基 OPEs)的浓度在夏季显著高于冬季,另外,虽然低聚 OPEs的浓度在两个季节之间不存在显著差异,但在 夏季的浓度更高(表4). 夏季浓度较高的 OPEs 包括 TCEP (66. 33 \sim 436. 00 ng·L⁻¹) \ TCPP (102. 98 \sim 496. 75 $\operatorname{ng} \cdot L^{-1}$) TBP (84. 13 ~ 207. 71 $\operatorname{ng} \cdot L^{-1}$) TEP (31.45 ~ 229.25 ng·L⁻¹)、和 BDP (0.60 ~ 21.92 $ng \cdot L^{-1}$).

张文萍等[25]调研了太湖流域中13种单体OPEs,



2 Concentrations of OPEs in the water of Taihu Lake at each sampling location in winter and summer

研究结果发现氯代烷基OPEs的含量最高,与本研究 的发现一致. 曹渺等[17]针对黄河流域14种单体OPEs 开展研究,其中赋存浓度最高的也为氯代烷基OPEs. 在中国胶州湾[7]和日本东京湾[35]同样发现氯代烷基 OPEs的浓度最高.上述现象表明,氯代烷基OPEs可 能是目前地表水体中占比最大的 OPEs,这可能与氯 代烷基的产量大和应用广泛相关,加之氯代烷基在 环境中的半衰期较长(8.6~21.3d),且不易被降 解[36],从而导致了氯代烷基 OPEs 在水体中的高赋存 现象. OPEs在夏季太湖水体中的浓度显著高于冬季, 这一现象在李栋等[18]和 Shi 等[37]的研究中也有类似 发现,该现象可能与气候条件有密切关系.有学者发 现强光照和高温更易使 OPEs 扩散至大气中,进而在 降雨和大气沉降的作用下汇集至河流湖泊中[38]. 太 湖处于亚热带地区,夏季光照强度大、温度高,OPEs 较容易从环境中被释放至大气和灰尘中,加之夏季 雨量充沛,OPEs易通过降雨作用及地表径流(或入湖 河流)进入湖泊中富集,因此可能导致了夏季太湖水 体中 OPEs 的高赋存.

表 4 太湖水体冬夏两季不同类型 OPEs 的显著性差异分析 1)

Table 4 Significant difference analysis for the concentrations of OPE in the water of Taihu Lake between the winter and summer

类型	冬季浓度平均 值/ng·L ⁻¹	夏季浓度平均 值/ng·L ⁻¹	P值
\sum_{15} OPEs	498.54	752.74	0.000
单体 OPEs	492.61	743.51	0.000
烷基 OPEs	135.86	246.54	0.000
氯代烷基 OPEs	345.49	488.06	0.001
芳香基 OPEs	11.27	8.91	0.525
低聚 OPEs	5.93	9.23	0.570

1)P值 < 0.05表示存在显著性差异

2.2 太湖沉积物中 OPEs 时空分布特征

不同季节太湖沉积物中OPEs 在各采样点的含量分布如图 4 所示. 冬季除 V6 未检出外,其余 14 种OPEs 均有检出,其中 TPrP、TEHP和 RDP的检出率低于 100%,分别为 17.7%、76.5%和 52.9%,其余 11 种OPEs 的检出率皆为 100%. 夏季沉积物中 OPEs 均被检出,除了 TPrP、RDP和 V6之外(检出率分别为23.5%、94.1%和 5.9%),其余 12 种 OPEs 的检出率皆为 100%. 与文献报道数据相比较,沉积物中 OPEs

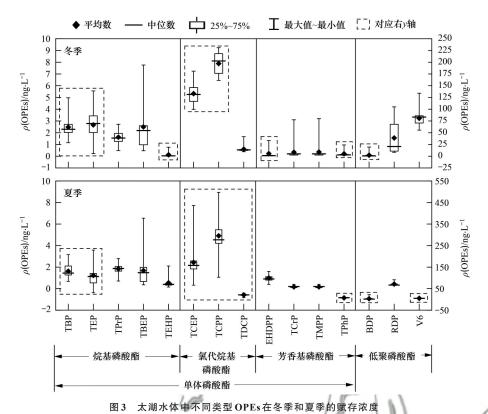


Fig. 3 Concentrations of OPEs in the water of Taihu Lake in winter and summe

的检出率较 2014年[10]和 2019年[39]均有提高

在空间分布上, $\omega(\sum_{15} \text{OPEs})$ 在高影响区的范围 (以 dry weight, dw 计)为 45. 30 ~ 117. 00 ng·g⁻¹(冬)和 69. 70 ~ 227. 40 ng·g⁻¹(夏),在低影响区为 47. 80 ~ 104. 80 ng·g⁻¹(冬)和 80. 50 ~ 176. 40 ng·g⁻¹(夏). 对不同区域的 OPEs 进行显著性差异分析发现,所有 OPEs 在区域间不存在明显差异(P > 0.05),并且不同类型 OPEs 在高、低影响区之间也没有差异性(P > 0.05),这与水体的规律基本一致。严小菊^[40]和张文萍^[41]的研究结果表明太湖沉积物中 OPEs 不存在显著区域分布特征,与本研究的结果一致。

太湖沉积物中 OPEs 的赋存规律与水体基本一致,即: 氯代烷基 OPEs > 烷基 OPEs > 芳香基 OPEs \approx 低聚 OPEs,具体情况如图 5 所示. 在冬季,氯代烷基 OPEs 中 TCEP 和 TCPP 含量较高,范围分别为 5. 10 \approx 46. 40 ng·g⁻¹ 和 19. 20 \approx 36. 50 ng·g⁻¹; 烷基 OPEs 中 TBP含量较高,为 7. 10 \approx 18. 30 ng·g⁻¹; 芳香基 OPEs 和 低聚 OPEs 虽然整体含量均较低,但值得注意的是 ω (BDP)的最大值可达 50. 10 ng·g⁻¹. 在夏季,太湖沉积物中 \sum_{15} OPEs 含量显著高于冬季,其中单体 OPEs 尤为明显. 另外,虽然低聚 OPEs 在沉积物中的含量不存在明显的季节差异 (P=0.074)但其在夏季的含量更高(表 5). 夏季含量较高的 OPEs 包括 TBP(2. 30 \approx 49. 30 ng·g⁻¹)、TCEP(1. 80 \approx 119. 10 ng·g⁻¹)、TCPP

(19.80~84.80 ng·g¹)、TPhP(7.00~16.70 ng·g¹)和BDP(1.10~75.20 ng·g¹). 地表水环境的沉积物和水体之间存在污染物的吸附和解吸过程^[38],有研究报道沉积物对污染物的吸附能力与水体的pH和水温呈正相关关系^[42]. 本研究现场监测的水体pH在冬夏两季分别为7.90和9.03,水温在冬夏两季分别为10.38℃和30.64℃因此推测太湖沉积物中OPEs在夏季的高赋存可能与水体的pH以及水温有关.

基于太湖沉积物中各 OPEs 的中位含量可知,TBP、TCPP和TCEP是太湖沉积物赋存的主要 OPEs,与长江^[20]和胶州湾^[7]沉积物中的主要 OPEs 一致. TCEP和TCPP是典型的两种氯代烷基 OPEs,其产量大、应用广泛、半衰期长和不易降解,因而较容易富集在地表水环境中^[36]. TBP作为一种添加型阻燃剂,除常用于阻燃剂和增塑剂外,同时还被用在润滑油和液压油中^[43],较容易通过挥发作用释放至环境中^[44]. 此外,虽然低聚 OPEs 中 BDP的含量低于上述3种单体 OPEs,但其在沉积物中的含量不容忽视(含量最大值可达 75. 20 ng·g⁻¹),Zhao等^[10]也有类似的发现. 鉴于 BDP的高产量(全国 OPEs 产量第 4)^[3]及其潜在代谢产物双酚 A^[45]的风险,该污染物同样需要引起高度重视.

2.3 OPEs在水体和沉积物中的分配

为了解 OPEs 在水体和沉积物中的分配情况,本研究对冬夏两季 OPEs 的伪分配系数[pseudo-

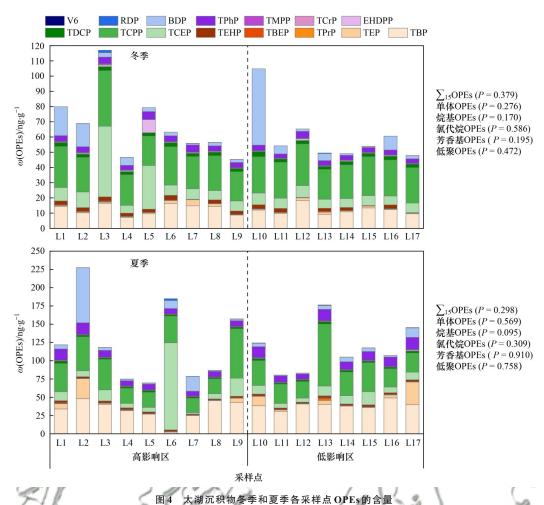


Fig. 4 Concentrations of OPEs in the sediment of Taihu Lake at each sampling location in winter and summer

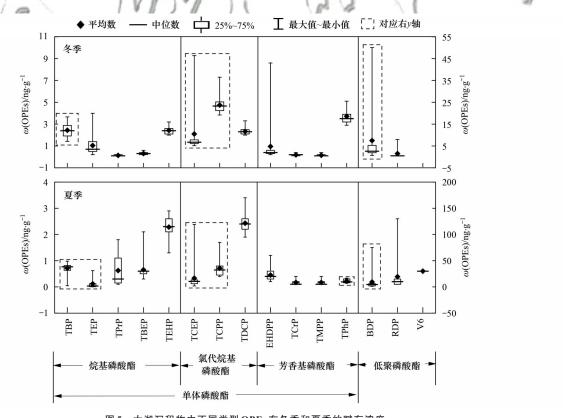


图 5 太湖沉积物中不同类型 OPEs 在冬季和夏季的赋存浓度

 $Fig. \ 5 \quad Concentrations \ of \ OPEs \ in \ the \ sediment \ of \ Taihu \ Lake \ in \ winter \ and \ summer$

表 5 太湖沉积物冬夏两季不同类型 OPEs 的显著性差异分析 ¹⁾
Table 5 Significant difference analysis for the concentrations of OPE in the sediment of Taihu Lake between the winter and summer

类型	冬季含量平均 值/ng·g ⁻¹	夏季含量平均 值/ng·g ⁻¹	P值
\sum_{15} OPEs	54.53	124.01	0.000
单体 OPEs	49.23	110.26	0.000
烷基OPEs	15.42	48.71	0.000
氯代烷基 OPEs	29.86	46.50	0.000
芳香基 OPEs	3.95	15.04	0.000
低聚 OPEs	5.30	13.76	0.074

1)P值 < 0.05表示存在显著性差异

partitioning coefficient, $K_a = \%$ 和物中污染物含量 $(ng \cdot g^{-1})/x$ 体中污染物浓度 $(ng \cdot L^{-1})$]^[20]进行了分析 (表 6). 由表 6可知, OPEs 在冬季的 K_a 值 $(L \cdot g^{-1}, F \cap D)$ 范围为 $0.02(TEP) \sim 3.47(BDP)$, 在夏季的 K_a 值范围为 $0.05(TEP) \sim 4.17(TEHP)$, BDP和 TEHP的 K_a 值分别在冬夏两季的所有 OPEs 中最高,表明这两种 OPEs 较容易被沉积物吸附,这与其在沉积物中的赋存情况相符(图 5). 在冬夏两季,15种目标 OPEs 的 K_a 值和 $lg K_{ow}$ 皆呈现显著的正相关关系(表 6),表明 $lg K_{ow}$ 越高的 OPEs 更容易被沉积物吸附,因而亲脂性高的BDP和 TEHP在沉积物中呈现出较高的含量,由此可知,OPEs 的物化性质可能是影响其在水体和沉积物之间分配的主要因素之一.

此外, \sum_{15} OPEs 在夏季的 $K_{\rm d}$ 值(0.87 ${\bf L} \cdot {\bf g}^{-1}$) 是冬表6 OPEs 在太湖沉积物和水体间的伪分配系数($K_{\rm d}$) 及与 ${\bf Ig} K_{\rm nu}$ 的相关性 $^{1)}$

Table 6 Pseudo-partitioning coefficients of OPEs between the sediment and surface water of Taihu Lake and their correlations with $\lg K_{\rm em}$

			ow ow	
类型	$K_{ m d}/{ m L}$ •	g ⁻¹	与 $\lg K_{ow}$	的相关性
天至	冬季	夏季	冬季	夏季
TBP	0.20	0.27		
TEP	0.02	0.05		
TPrP	0.06	0.34		
TBEP	0.13	0.39		
TEHP	0.80	4.17		
TCEP	0.08	0.10		
TCPP	0.12	0.12		
TDCP	0.16	0.12	P = 0.000	P = 0.000
EHDPP	0.18	0.45	$R^2 = 0.36$	$R^2 = 0.72$
TCrP	0.53	1.04		
TMPP	0.41	1.00		
TPhP	0.72	1.45		
BDP	3.47	2.60		
RDP	0.19	0.90		
V6	_	0.11		
\sum_{15} OPEs	0.50	0.87		

1)P值 < 0.05表示存在显著性差异,"一"表示无有效数据

季(0.50 L·g⁻¹)的1.74倍,表明OPEs在夏季更容易富集在沉积物中.虽然在夏季沉积物更容易吸附水体中的OPEs,但是水体中的OPEs在夏季明显高于冬季,主要原因可能是在这个季节水体污染物的输入量更大,沉积物吸附的量无法抵消输入量的影响,因而虽然沉积物吸附能力在夏季增强,但整体上水体中的OPEs浓度还是呈现增大趋势.

2.4 太湖水环境 OPEs 的生态风险评估

鉴于太湖水体具有水产养殖、饮用水供给和农业灌溉等重要功能,本研究重点针对水体中OPEs的赋存开展太湖水环境生态风险评估,但这并不表明太湖沉积物中的OPEs不存在生态风险,有关内容需要进一步开展研究.根据太湖水体中OPEs的浓度计算所得的风险商值如图6所示,整体上太湖水体中的 \sum_{15} OPEs具有中等生态风险(冬季RQ=0.34;夏季RQ=0.35),其中单体OPEs的生态风险中等,低聚OPEs的生态风险较低(冬季RQ=0.08;夏季RQ=0.04).在单体OPEs中,氯代烷基OPEs的生态风险最高,为中等程度,而烷基OPEs和芳香基OPEs为低生态风险.张文萍等[25]和吕佳佩等[46] 也对太湖水体中的OPEs进行了生态风险评估,发现氯代烷基OPEs为低生态风险,低于本研究的评估结果,预示着OPEs的风险可能正在逐年加剧.

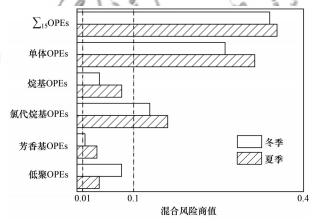


图 6 不同季节太湖水环境 OPEs 的风险商值

Fig. 6 Risk quotient values of OPEs in the water environment of Taihu Lake in different seasons

除低聚 OPEs 之外, OPEs 在夏季的生态风险皆高于冬季, 并且冬夏两季 OPEs 的生态风险排序整体上皆呈现为: 氯代烷基 OPEs > 烷基 OPEs > 芳香基 OPEs≈低聚 OPEs, 这与 OPEs 在水体中的赋存规律基本一致, 表明 OPEs 的浓度是影响 OPEs 生态风险的一个重要因素.单体 OPEs 中的氯代烷基 OPEs 是目前开展毒理研究较多的一类 OPEs [47], 已有研究发现其会对鱼类的生长发育产生危害[48], 并且还具有神经毒性[49]、基因和细胞毒性[50], 因而地表水体中的氯代

烷基OPEs亟需探索相关的控制措施. 低聚OPEs在本研究中的生态风险较低,但并不表明此类OPEs不需要引起重视,目前与其有关的毒理研究非常有限,需要深入开展相关研究以为生态风险评估提供更多的数据支持.

3 结论

- (1)太湖水体中的∑₁₅OPEs在高、低影响区之间的赋存浓度皆不存在显著性差异.水体中不同类型的OPEs在夏季的赋存浓度均高于冬季,其中单体OPEs的季节差异性显著.OPEs在太湖水体中的赋存浓度排序为:氯代烷基OPEs>烷基OPEs>芳香基OPEs≈低聚OPEs.
- (2)太湖沉积物中的 \sum_{15} OPEs 在高、低影响区 之间的赋存含量无明显差异. 沉积物中的单体 OPEs 在冬夏两季差异明显且在夏季含量更高,虽然低聚 OPEs 不存在显著性季节差异,但在夏季的赋存含量 更高. OPEs 在太湖沉积物中的赋存含量排序与水体 基本一致
- (3) OPEs 在水体和沉积物中的伪分配系数(K_d) 在夏季更高,表明 OPEs 在夏季更容易被沉积物吸附,但不能忽略夏季污染物排放通量的影响. OPEs 的 K_d 值与其 $\lg K_{ow}$ 显著正相关,表明亲脂性较高的 OPEs 更容易被沉积物吸附.
- (4)风险商值结果表明∑₁₅OPEs具有中等生态风险,不同类型的单体OPEs在夏季的生态风险皆高于冬季.整体上冬夏两季OPEs的生态风险排序为: 氯代烷基OPEs > 烷基OPEs > 芳香基OPEs≈低聚OPEs. 低聚OPEs 的生态风险虽然较低,但因其相关毒理数据的匮乏亟需开展深入研究.

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