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农田土壤除草剂污染的修复技术研究进展

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(1.中国科学院沈阳应用生态研究所污染生态与环境工程重点实验室,沈阳 110016; 2.中国科学院大学,北京 100049) 摘要:除草剂主要用于保护农作物免受杂草的侵害,是现代农业生产中用量最多的一类农药.然而,随着全球粮食需求的增加,除草剂的用量逐年增大,药效也不断增强,导致除草剂在农田土壤中出现累积、迁移转化和毒害作用等问题.为了降低除草剂给土壤-作物系统带来的生态风险,根据除草剂污染特征和区域农业生产规律研发绿色低碳修复技术,是目前生态环境领域需要关注的问题.基于此,整理了近年来关于农田土壤除草剂污染治理的相关报道,重点分析了修复技术的研究进展和应用案例,并对除草剂修复领域未来的发展动态进行了展望.目前应用于农田除草剂的修复技术主要有基于微生物修复、酶修复和植物修复的生物修复技术,以及基于生物炭基材料的吸附固定技术.其中,生物修复技术的发展相对成熟,已经应用于实际农田除草剂的修复治理工作,并形成了成功的修复案例.为了提升对农田土壤除草剂污染的修复效果,修复技术逐渐从单一模式向物理化学-生物多技术耦合模式发展,以充分发挥多技术集成应用的协同作用.

关键词:除草剂;农田土壤;污染;修复技术;生物修复

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Research Progress on the Remediation Technology of Herbicide Contamination in Agricultural Soils

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Abstract: Asthe most-used pesticides in the agricultural production process, herbicides are mainly applied to protect crops from weeds. However, with the increased global demand for food, the dosage of herbicides is rising annually, and the efficacy of herbicides is getting stronger, which can cause some environmental issues including the accumulation, migration and transformation, and toxic effects of herbicides in agricultural soils. According to the characteristics of herbicide contamination and regional agricultural production, developing green and low-carbon technologies to reduce the ecological risks of herbicides to the soil-crop systems is a current concern in the ecological environment field. In this paper, relevant studies in recent years on herbicide pollution management in agricultural soils were identified and reviewed, the research progress and application eases of remediation technologies for herbicide pollution was analyzed and demonstrated, and future research and development tendency regarding the remediation of herbicides pollution was also prospected. Current remediation technologies for herbicides mainly include bioremediation technologies (e. g., microbial remediation, enzyme remediation, and phytoremediation), adsorption, and immobilization technologies (e. g., biochar-based materials). The bioremediation technologieswere rather mature and had been applied to the herbicide-contaminated soil in fields. Additionally, many successful bioremediation cases have been reported. Moreover, in order to enhance the remediation effect on herbicide pollution in agriculture soils, remediation technologies have been gradually developed from a single model to a coupled model with physical, chemical, and biological technology, which can maximize the synergy of the multi-technology application.

Key words: herbicide; agricultural soils; pollution; remediation technology; bioremediation

农药是为了消灭、预防或控制有害的真菌、杂草或动物害虫而生产合成的物质,根据目标生物的不同,可以分为除草剂、杀虫剂和杀菌剂等[1].以上农药被广泛用于农业生产过程中,保护作物免受各种害虫和杂草的侵害,为人类粮食供给做出了巨大贡献.在各类农药产品中,除草剂占全球产量的比例最高,已经成为现代农业实践中不可或缺的一部分.据统计,2019 年除草剂的销售额约占全球农药市场份额的52% [2.3].除草剂在集约农业中的使用,保证了农作物的产量和质量,但随着全世界日益增长的粮食需求,除草剂的种类和用量都在持续增加.中国除草剂年使用量在1995 年时仅为4.6万t,2015 年就达到了100多万t.2014 年累计使用除草剂的作物面积达到了1.07亿 hm²,相当于同年作物总面积的65% 左右[4].2000 年农田的除草剂用量约为

 $0.41 \text{ kg} \cdot \text{hm}^{-2}$,而这一数值在 2015 年时增加了近 15 倍,农田除草剂用量达到了 $6.08 \text{ kg} \cdot \text{hm}^{-2[5,6]}$.

大量且频繁的使用除草剂,会使其在农田土壤中残留,并通过土壤、水和空气等介质迁移扩散.有研究表明,喷洒的除草剂对杂草靶向性不到1%,99%以上的除草剂分布在周围的环境中,最终汇入土壤,并残留在土壤中^[7,8].在现有报道中乙草胺和莠去津等是目前我国农田土壤中检出率和检出浓度较高的除草剂. Geng 等^[9]研究发现吉林省迁安和公主岭地区农田土壤中莠去津的检出率为97%,检出

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浓度为 11.1 µg·kg⁻¹; 而且土壤中残留的除草剂会 向地下水中迁移,上述地区地下水中莠去津的检出 率为89%, 检出浓度为106.8 ng·L-1. Sun 等[10] 在松 花江流域河岸带采集了189个农田、草地、林地和 裸地的土壤样品,并分析了土壤中莠去津和乙草胺 的残留情况,结果表明97%的土壤样品中检测到除 草剂残留,其中乙草胺和莠去津在农田土壤中的平 均残留量分别为 26. 10 μg·kg⁻¹和 11. 28 μg·kg⁻¹. 农田系统中除草剂的残留与区域土壤性质、作物类 型、气候条件和农业生产模式等多种因素密切相 关. 棉花田常用除草剂氟乐灵在我国几个主要产棉 区土壤中的残留存在显著差异,有研究表明河北省 的氟乐灵检出率最高(75%),其次是新疆(60%)和 山东(40%),含量平均值(以 dw 计)大小为:新疆 $(5.98 \, \mu g \cdot kg^{-1}) > 河北 (5.06 \, \mu g \cdot kg^{-1}) > 山东$ $(3.19 \ \mu g \cdot kg^{-1})^{[11]}$.

农田土壤中除草剂的残留会给农业生态系统带 来不可逆的生态效应和健康风险[12]. 首先,除草剂 长期大量地施用会使种植系统过度依赖除草剂,使 农田杂草种群迅速更迭,加速杂草除草剂抗性的演 变[13]. 根据"国际抗除草剂杂草调查",全球已经报 道了486 例独特的抗除草剂杂草生物型,其中包括 147 例双子叶杂草和106 例单子叶杂草[14]. 其次,土 壤中的除草剂残留会显著影响土壤微生物群落结构 和功能,抑制酶的活性[15,16]. Carpio 等[17]研究了绿 麦隆和氟噻草胺复合吡氟酰草胺这两种商用配方除 草剂对田间条件下土壤微生物活性、生物量和结构 的影响,结果表明除草剂的施用使土壤中微生物总 量急剧下降,土壤微生物群落结构发生了明显的变 化. 除草剂残留还会降低土壤中氮磷等养分的有效 性,从而减少后茬作物对养分的吸收和积累,不利于 后茬作物的生长. Guan 等[18] 研究发现长期使用乙 草胺降低了红壤旱地中花生和大豆幼苗的出苗率、 叶片数、茎粗和株高,明显抑制了作物幼苗的生长. 更为重要的是,土壤中残留的除草剂还可以通过土 壤-作物系统进入食物链,对人类健康构成潜在威 胁[19,20].

综上所述,农田土壤中的除草剂污染普遍存在,对生态系统和人类健康构成潜在威胁,如何去除农田土壤中的除草剂残留并降低其危害成为了目前需要研究的重点,本文从可实际应用的和具有应用前景的技术类型进行了综述.其中,微生物修复、酶修复、植物修复和生物炭修复技术比较成熟,已经开展了大量的实验,并有研究将相关技术应用到实际农田土壤除草剂污染的修复中去,得到了不错的修复效果.而近几年新型的植物-微生物和材料-微生

物等联合修复技术,仍处于起步阶段,针对农田土壤除草剂修复的研究还较少,基本上是在实验室条件下开展的模拟实验.但联合修复技术可以突破单一修复技术存在的一些限制,提升修复效果,是一种非常具有研究价值的修复技术^[21].因此本文将重点综述目前可实际应用的技术类型,并对具有应用潜力技术的作用机制、修复效果和应用可行性进行相关的阐述.

1 微生物修复技术

微生物修复技术是指通过增加微生物的代谢活性来去除环境中有机污染物的方法,主要包括生物刺激和生物强化^[22,23].生物刺激是通过添加有限的营养物质来刺激本地微生物的降解,生物强化则是接种外源降解微生物菌剂来加速土壤中污染物的降解,其中生物强化是目前应用较多的微生物修复方式.在修复过程中,微生物对除草剂的作用方式主要有两种,一种是微生物通过酶促反应直接作用于除草剂,即化合物通过一定途径进入微生物体内,然后在各种酶的作用下,最终将除草剂完全分解为无毒或毒性较小的小分子化合物;另一种是微生物通过矿化作用和共代谢作用等非酶促反应间接作用于除草剂^[24~27].

随着微生物筛选鉴定和分析检测技术的不断发 展,关于对除草剂降解菌株的相关报道越来越多,许 多具有除草剂降解能力的细菌和真菌被从环境中分 离出来. 例如, 从污水厂污泥和土壤中分离出了 Paracoccus sp. Y3B-1, Pseudomonas oleovorans LCa2、Achromobacter sp. D-12 和 Serratia ureilytica AS-1 等可降解除草剂的细菌^[28~37]. 还有一些真菌 也被鉴定出具有降解除草剂的能力,如 Umbelopsis isabellina, Aspergillus niger, Hypholoma dispersum ECS-705 和 Paecilomyces marquandii 等^[38~41]. 以上 菌株对不同种类除草剂的去除效果如表 1 所示. 在 修复过程中,这些除草剂降解菌株发挥了较为显著 的作用. Chen 等[42]研究了菌株 Paenarthrobacter sp. W11 对莠去津的去除效果,结果表明 50 mg·kg⁻¹的 莠去津污染土壤培养 49 d 后, Paenarthrobacter sp. W11 使土壤中莠去津的降解率从 75.7% 增加到了 96.0%,显著促进了土壤中莠去津的降解,减轻了莠 去津残留对小麦生长的毒害作用,并且矿化莠去津 产生的 NH; 作为氮源增加了土壤中微生物的数量. Zheng 等^[43] 研究表明菌株 Catellibacterium caeni sp. nov DCA-1T 对丁草胺具有较好的降解效果,在培养 基条件下,84 h 内对 50 mg·L-1 丁草胺的降解率达 到了81.2%,并且对土壤中丁草胺的降解也有显著

的促进作用. Dwivedi 等^[44] 研究发现菌株 *Stenotrophomonas acidaminiphila* JS-1 能以丁草胺为唯一碳源和能源,可将土壤中1 000 mg·kg⁻¹的丁草胺在 20d 内完全去除.

近年来,为了提高微生物对除草剂的降解效果,研究人员针对目标除草剂的特点,将具有不同功能的菌株有机结合,发挥菌群对除草剂的协同降解作用. Jiang 等^[45]研究了莠去津降解菌 Arthrobacter sp. DNS10 和解磷菌 Enterobacter sp. P1 对莠去津的降解特性,结果表明混合菌株对 100 mg·L⁻¹莠去津的降解率为 99.2%,而单一菌株 Arthrobacter sp. DNS10 的降解率仅为 38.6%,这种现象是两个菌株之间的代谢产物交换造成的,菌株 Arthrobacter sp. DNS10 降解莠去津产生的代谢产物乙胺和异丙胺可以作为氮源支撑 Enterobacter sp. P1 的生长, Enterobacter sp. P1 释放的某些次生代谢物也能促进 Arthrobacter sp. DNS10 对莠去津的降解.

目前,也有研究将微生物添加到实际农田中进行原位修复.曹博^[46]用莠去津降解菌剂 DNS10 来消减玉米田间土壤中的莠去津,莠去津按推荐用量施用,结果发现菌剂的添加可以有效增加莠去津在农田土壤中的降解速率,并能增加玉米产量. 王歆鑫^[47]将由莠去津降解菌 DNS10、溶磷菌 P1、植物

促生菌 JD37 组成的固定化复合菌剂 SP I 1 添加到以莠去津为除草剂的玉米田中,在莠去津正常施用的情况下,复合菌剂的添加降低了玉米拔节期土壤中莠去津的残留浓度,并对土壤中的有机质、TN、碱解氮和 TP 等含量有一定程度的提升.

虽然微生物修复具有范围广、操作简单、成本 低和环境友好等优点,但同时也受很多因素的制约, 包括内部因素和外部环境因素. 内部因素主要包括 微生物种类和除草剂自身. 首先, 微生物的种类会直 接影响除草剂的降解转化,不同种类的微生物或同 一种类的不同菌株对除草剂的降解能力和适应性也 有所差异. 其次是除草剂的种类、性质和结构等因 素影响着微生物降解的效率,除草剂作为人工合成 的生物异源有机物,往往对微生物的降解表现出抵 抗力,这让微生物对除草剂的自然降解速度远远达 不到环境和人类的需求[48]. 另外, 微生物的生存受 很多外部环境因素的影响,例如,温度、湿度、盐 度、pH、氧气和营养等都会影响微生物的生长[49]. 并且,外源微生物添加到土壤中会和土著微生物产 生竞争关系,这可能导致接种微生物的数量迅速减 少,影响修复效果[50].因此在微生物修复过程中,高 修复效率、高环境耐受性微生物的筛选、适宜的降 解方法和降解环境是需要不断研究的问题.

表 1 可降解除草剂的微生物

Table 1	Herbici	de-deg	rading	micro	bes
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散生物种类	菌株	除草剂降解效果	文献
	Paracoccus sp. Y3B-1	培养 3 d,对 100 mg·L ⁻¹ 乙草胺、丙草胺和丁草胺的降解率可达 86.7%、69.1%和65.5%	[28]
	Pseudomonas oleovorans LCa2	培养 $7d$,对 $7.6\mathrm{mg}\cdot\mathrm{L}^{-1}$ 乙草胺的降解率为 98.03% ,并且对乙草胺的耐受浓度可达 $200\mathrm{mg}\cdot\mathrm{L}^{-1}$	[29]
	Paracoccus sp. Strain FLY-8	能降解并利用 $100 \text{ mg} \cdot \text{L}^{-1}$ 的甲草胺、乙草胺、丁草胺、异丙草胺、异丙甲草胺和丙草胺作为碳源生长	[30]
	Achromobacter sp. D-12	培养 5 d,对 10 mg·L ⁻¹ 乙草胺的降解率为 95%	[31]
细菌	Serratia ureilytica AS-1	培养 10 d, 能完全降解 500 mg·L-1的丁草胺	[32]
	Sphingomonas sp. DC-6	培养 48 h,对 100 mg·L $^{-1}$ 丁草胺、乙草胺和甲草胺的降解率分别为 76.7%、93.6% 和 98.6%	[33]
	Bacillus sp. CY	培养 30 d,对 50 mg·L-1丙酯草醚的降解率为 65%	[34]
	Hansschlegelia sp. CHL1	培养 4 d,对 50 mg·L-1氯嘧磺隆的降解率超过 95%	[35]
	Ammoniphilus sp. JF	培养 24 h, 可完全降解 100 mg·L-1的丁草胺	[37]
	Pseudomonas putida G3	培养 360 h, 可完全降解 700 mg·L-1的丁草胺	[36]
	Umbelopsis isabellina	培养 5 d,对 25 mg·L ⁻¹ 2,4-二氯苯氧乙酸(2,4-D)的降解率为 98%	[38]
	Aspergillus niger	培养7 d, 可将 20 mg·L-1 乙草胺完全降解	[39]
真菌	Hypholoma dispersum ECS-705	培养 12 d,对 100 mg·L ⁻¹ 百草枯的降解率达 70.7%	[40]
	Paecilomyces marquandii	培养 7 d, 对 50 mg·L $^{-1}$ 甲草胺降解率为 96.7%, 对 100 mg·L $^{-1}$ 甲草胺的降解率为 79.9%	[41]

2 酶修复技术

微生物降解除草剂的潜力来源于微生物中存在的功能酶. 例如, 假单胞菌 *Pseudomonas* sp. strain ADP 在降解莠去津的过程中, *AtzA、AtzB* 和 *AtzC* 这

3 种酶共同作用把莠去津逐步转化成羟基莠去津、 N-异丙基三聚氰胺、三聚氰酸和异丙胺,最终被分 解为 CO₂ 和 NH₃^[49]. 相较于微生物降解,酶的直接 作用效率高得多,因此研究人员将从微生物细胞中 分离出来的功能酶应用于除草剂污染修复^[51]. Pizzul 等 $^{[52]}$ 通过体外实验研究发现漆酶、纯锰过氧化物酶对除草剂草甘膦均有一定的降解潜力,漆酶在 $MnSO_4$ 、Tween 80 和 ABTS [2,2'-azino-bis (3-ethylbenzthiazoline- 6-sulphonic acid)] 存在的情况下,24 h 可将 0. 06 mmol· L^{-1} 的草甘膦降解 90. 1%, 纯锰过氧化物酶在 $MnSO_4$ 和 Tween 80 存在的情况下,24 h 可将 0. 06 mmol· L^{-1} 的草甘膦完全降解. Scott 等 $^{[53]}$ 发现莠去津脱氯酶和三嗪水解酶可以将莠去津代谢成氯化物、 NH_3 和 CO_2 .

在实际应用中,游离酶存在着稳定性差和失活 快等缺点[54,55]. 为了克服这些缺点,可以通过吸附、 包封、封装、交联、共价结合和附着等方法将酶固 定在载体上,来增加其稳定性、回收率和可重复使用 性[56]. Yu 等[57] 用交联明胶(GLT) 和壳聚糖(CTS) 作为酯酶 SulE 的固定化载体,并通过微观实验评测 了 GLT/CTS-SulE 对甲基苯磺隆和甲磺隆污染土壤 的修复效果,结果表明 GLT/CTS-SulE 相较于单独 的酶表现出更好的 pH 和温度适应性以及更高的除 草剂降解率,说明这是一种对甲基苯磺隆和甲磺隆 污染土壤进行原位强化修复的可行方法. Yang [58] 等 利用菌株 Hansschlegelia sp. CHL1 表达了一种可以 初级降解磺酰脲类除草剂的酯酶 SulE,将其包埋在 一种环境友好、生物相容性好和可生物降解的交联 聚谷氨酸/明胶水凝胶体系(CPE)中,并以氯嘧磺隆 为研究对象,研究了 CPE-SulE 和游离 SulE 在土壤 和水中的降解能力,结果发现,CPE-SulE 和游离 SulE 相比, 热稳定性、pH 稳定性和可重复使用性明 显提高,同时也提高了氯嘧磺隆的降解效率.

目前,已有研究将酶修复技术应用到田间土壤中来验证其修复效果. Yu 等^[59]利用简单纯化的酶和金属有机框架低成本合成用于除草剂污染土壤修复的纳米生物催化剂,通过进一步亲和层纯化了芳氧苯氧丙酸酯除草剂水解酶 QpeH,通过仿生矿化将其嵌入了两种类型的沸石咪唑酯骨架材料(ZIFs)中,形成了十字花状形状的 QpeH@ ZIF-10 和菱形十二面体的 QpeH@ ZIF-8 两种复合材料,并应用于清除西瓜田中的精喹禾灵(推荐用量施用),结果表明,添加 QpeH@ ZIF-10 和 QpeH@ ZIF-8 复合材料分别清除了土壤中88%和84%的精喹禾灵,并且发现复合材料的使用还可以恢复土壤中的细菌群落.

以上研究说明酶修复技术用来修复除草剂污染 土壤是可行的方法,但酶在环境中的脆弱性和不可 重复利用性限制其在农田土壤中的应用,因此常用 固定酶方法来增加其可使用性.在酶固定化方法中, 固定化载体是一个非常关键的因素,它需要满足以 下条件:良好的稳定性和机械强度,利于酶的回收和 重复利用;较大的比表面积或多孔结构,来增加酶的固定化量;易于修饰和涂布,方便酶的固定;成本低,来源丰富,环境友好等^[54,60].因此,在酶修复技术的实际应用中,固定化载体的研究显得尤为重要.

3 植物修复技术

利用植物修复环境中污染物的方法被称为植物 修复,这是一种被动的且低成本的就地修复方法,在 保持土壤现有的活性、物理结构和肥力的同时,对 污染土壤进行修复[61,62]. 植物主要通过植物积累、 植物稳定、植物降解、植物挥发和根际降解等过程 来修复污染土壤[63]. 植物积累是指植物通过其根系 吸收土壤中的除草剂,并将其储存在其体内的过程. 植物挥发则是指除草剂从土壤中吸收后通过植物的 蒸腾作用挥发的过程. 植物稳定是指通过根系吸附 等形式来固定土壤中的除草剂,降低其移动性和生 物有效性. 植物降解,又称植物转化,指在植物体内 酶的作用下将除草剂转化为无毒或低毒性物质的过 程[64]. 根际降解,又称植物刺激,根系分泌物和植物 根部释放的酶可以改变除草剂的生物有效性以及刺 激根际细菌和真菌的活动,帮助降解除草剂[65].因 此理想的修复植物应该具有生长迅速、场地适应性 强、根系发达和积累能力强等特征.

植物修复技术的首要工作就是筛选出对除草剂 同时具有耐受能力和降解能力的植物品类. 目前报 道较多的除草剂修复植物种类主要有禾木科和豆 科. 在以往的研究中,黑麦草和狼尾草等禾本科牧草 被证明对除草剂具有较好的去除能力[66,67]. Mimmo 等[68]研究了黑麦草对特丁津的去除效果,结果表明 尽管除草剂对黑麦草地上部的长度和重量产生负面 影响,但黑麦草还是能去除30%~40%的特丁津. Lin 等[69] 通过盆栽试验发现狼尾草的种植使莠去津 的降解率由 15.22% 提高到了 51.46%. 豆科植物由 于根部常有固氮作用的根瘤,并且还是优良的绿肥 和饲料作物,也常被用于土壤除草剂的去除. Souto 等[70]研究发现洋刀豆、大豆、箭筈豌豆、百脉根和 白三叶草组合的种植促进了土壤中咪唑乙烟酸、甲 咪唑烟酸和灭草烟的微生物降解,这3种除草剂的 平均降解率分别达到了 91%、92% 和 93%. Teófilo 等[71]研究发现大托叶猪屎豆和白羽扇豆对环嗪酮 污染土壤具有一定的修复潜力,可分别去除 2.5% 和 8.6% 的环嗪酮.

植物因其自身的生长规律和生理特性各不相同,对不同类型除草剂的耐受能力和富集能力也有所差异.以旱田中检出率和检出浓度最为突出的除

草剂莠去津为例,不同禾本科牧草对其吸收和降解 的能力差异显著. Khrunyk 等[72]研究了柳枝稷、大 蓝茎和黄印度草对莠去津的去除能力,结果表明柳 枝稷和大蓝茎是比较有修复潜力的植物,水培条件 下对莠去津的去除率为40%,沙培条件下为20%~ 33%,而黄印度草则表现出对莠去津的低抗性和低 吸收量.同时,相同植物对不同除草剂的修复效果也 存在差异. Teófilo 等[71]研究了黑麦草和狼尾草对敌 草隆、环嗪酮和甲嘧磺隆的去除效果,结果表明黑 麦草和狼尾草对环嗪酮的修复潜力较大,分别去除 了土壤中7.3%和1.1%的环嗪酮,而对敌草隆和甲 嘧磺隆的吸收率均低于 1%. Mendes 等[73]研究了大 托叶猪屎豆、洋刀豆、黧豆和白羽扇豆对除草剂快 杀稗和丁噻隆的修复效果,结果表明4种植物分别 吸收了4.48%、22.49%、16.71%和15.00%的丁噻 隆,1.75%、13.44%、6.20%和10.02%的快杀稗, 其中对丁噻隆的修复效果要优于快杀稗.

由于除草剂主要应用于农业生产过程中,来保 证农作物的生长免受杂草的危害[71],因此有学者研 究了农作物本身对土壤除草剂残留的耐受性和去除 能力. 例如,虞云龙等[74]研究了棉花、水稻、小麦和 玉米种植过程中作物根际土壤和非根际土壤丁草胺 浓度的差异,发现作物根系显著促进了丁草胺的降 解,使其降解半衰期缩短了26.6%~57.2%.有研究 发现不同作物对除草剂的吸收能力还有差异. Sánchez 等[75]研究了大麦和玉米对莠去津的修复效 果,结果表明大麦和玉米对莠去津均有一定的吸收 和解毒作用,但玉米对莠去津的吸收能力要大于大 麦,能够积累土壤中38.4%的莠去津.作物生物量 和叶表面积的不同可能是造成这一差异的主要原 因,较大的生物量往往能吸收更多的除草剂,而较大 的叶表面积代表着更高的蒸腾速率,这可能会增加 作物对除草剂的吸收[75,76].

随着转基因技术的发展,相继出现了将特定外源基因导入植物以提高植物对除草剂的耐受性和降解效率的相关研究.沙特阿拉伯 Azab 的研究团队对特定基因导入植物植株来修复除草剂污染的方面进行了一系列研究,结果表明转 CYP1A2 基因拟南芥对苯脲类除草剂利谷隆具有更高的活体代谢和解毒能力,人类 CYP1A2 基因的过表达增加了拟南芥对利谷隆的修复能力和耐受性;此外,表达人体细胞色素 P450-1A2 转基因海棠对除草剂异丙隆也表现出较强的耐受性,即使高剂量异丙隆处理下,转基因海棠依然能旺盛生长,并且能够代谢异丙隆^[77,78]. Siminszky 等^[79]研究也发现同时表达 CYP71A10 和大豆 P450 还原酶基因的烟草植株相比单独表达

CYP71A10 的植株,对苯脲型除草剂利谷隆、伏草隆和绿麦隆的降解能力提高了 20%~23%. 转基因植物对除草剂的代谢为提高植物修复能力提供了高效、环保的手段,转基因植物修复是一种非常有潜力的修复方式.

目前,也有一些研究对实际田间条件下植物修复的效果进行了验证. Madalão 等^[80]在田间条件下用洋刀豆来修复甲磺草胺,甲磺草胺用量为1000g·hm⁻²的有效成分,约为1.4倍最大商业剂量,并以高粱为指示植物来评估其修复效果,结果表明洋刀豆修复过污染土壤中高粱的干物质产量、植物数量、生产力和高度等指标与未污染土壤相当,而未修复污染土壤中高粱的干物质量等指标均有所下降,这说明洋刀豆可以修复土壤中的甲磺草胺污染,减少其对作物的危害. Shimazu 等^[81]研究表明在稻田条件下,携带 CYP2B6 基因的转基因水稻植株能代谢除草剂异丙甲草胺,显著降低异丙甲草胺在植物和土壤中的残留.

植物修复具有成本低、环境友好和无二次污染 等优点. 但同时植物修复在应用过程中也存在一些 局限性. 例如,植物修复通常缺乏对除草剂全面降解 的能力; 植物修复应用在除草剂修复时,除草剂本 身会对植物造成一定的危害,其中一些可能是永久 性损伤,会降低植物修复的有效性,并对植物修复系 统的稳定性构成威胁; 当除草剂污染向土壤剖面深 处延伸时,植物的根部难以接近,会导致植物修复效 率低; 修复场地的环境和气候条件会影响植物的生 长情况,很大程度上决定了植物修复的效率:不同 除草剂的性质也会影响植物修复的效率, 当植物修 复一些难降解和生物可利用性低的除草剂时,往往 得不到理想的效果: 并且不同的植物对除草剂的修 复效果也存在差异[82,83]. 因此,在植物修复过程中, 要根据污染地区的实际情况来选择适当的植物的品 种和适宜的修复方法,以达到预期的修复效果.

4 生物炭修复技术

吸附是除草剂进入土壤后发生的第一个过程,会影响除草剂的生物利用度和对非目标物的毒性^[84].因此,用环境友好的材料来吸附土壤中的除草剂被认为是一种经济有效的修复方法.生物炭是一种在限氧条件下热解生物质得到的碳质材料,具有较大的比表面积和发达的孔隙结构,可以有效地吸附土壤中的除草剂^[21].近几年有关生物炭在除草剂修复中的应用被广泛研究.首先,生物炭可以通过吸附、固定除草剂和其代谢物,控制除草剂在土壤中的迁移和生物有效性,从而降低人类接触除草剂

和环境污染的风险^[85,86]. 其次,生物炭的添加可以改善土壤的通气条件,为土壤中的微生物提供碳源、无机营养和微量元素,增强了微生物的活动和新陈代谢,有利于土壤微生物对除草剂降解过程的发生^[87~90]. 生物炭作为土壤改良剂还可以提高土壤肥力,增加作物产量,恢复退化的土壤^[91,92]. 目前生物炭的原材料主要有作物秸秆、动物粪便和木质材料^[93].

水稻、玉米、小麦和豆科等植物秸秆是大量且 易于获取和可再生的潜在资源,常被制成生物炭用 来修复除草剂污染土壤. Jiao 等[94] 研究了具有高比 表面积的茶秆生物炭和石墨化茶秆生物炭对莠去津 的吸附固定性能,结果表明茶秆生物炭和石墨化茶 秆生物炭改良的土壤中莠去津的释放显著减少了 3.5%~36.0%. 添加到土壤中的生物炭在吸附除草 剂,限制其迁移的同时,还会对土壤中的微生物群落 产生影响. Meng 等[95] 研究了添加小麦秸秆制成生 物炭对氟磺胺草醚胁迫下小麦幼苗的生长、生理特 性和根际微生物群落的影响,生物炭在土壤中的添 加量为1%、2%和4%,结果表明生物炭可以显著降 低小麦对氟磺胺草醚的吸收,减轻其对小麦幼苗的 毒害作用,还增加了小麦幼苗根际有益细菌和真菌 群落的丰富度和多样性,其中2%生物炭添加量得 到的效果最好,这说明在修复过程中需要确定合适 的生物炭添加量以达到最佳的修复效果. Yang 等[%]研究了小麦秸秆制成生物炭对黑土中莠去津 吸附和微生物降解的影响,结果表明生物炭的添加 明显增强了莠去津的去除,激活了土壤功能微生物 的降解,并且提高了土壤中有机质的吸附.而含有大 量N、P元素和其他营养物质的畜禽粪便也是制备 生物炭的优良原料[97]. Zhang 等[98] 以猪粪为原料, 分别在350℃和700℃的热解温度下制得BC350和 BC700 的生物炭样品,用来测试对莠去津的吸附性 能,结果发现两种材料均对莠去津有较好的吸附性 能,但 BC700 的吸附性能要优于 BC350,这是因为 在较高的热解温度下,生物炭的灰分比较高,这说明 生物炭的性质会影响其对除草剂的吸附性能. Martin 等[99]研究发现了用家禽粪便制成生物炭改 良的土壤对除草剂敌草隆和莠去津的吸附量是未改 良土壤的2~5倍,但观察到土壤中生物炭的老化会 降低它们的吸附性能,经过32个月的老化生物炭处 理的土壤和对照土壤有着相似的吸附-解吸性能,而 新改良的土壤对除草剂的吸附增加,解吸滞后性较 大. 也有一些研究用木质材料制成的生物炭来修复 除草剂污染土壤. Wang 等[100]评估了用木屑和木炭 制成的生物炭对特丁津在森林土壤上吸附的影响,

结果显示生物炭的添加增强了土壤对特丁津的吸附. Sopeña 等[101]报道了木炭制成的生物炭改性土壤对异丙隆的吸附容量为非改性土壤的 5 倍,生物炭的添加大大增加了土壤的吸附性能.

已有研究将生物炭修复应用到实际田间土壤中.解钩^[102]研究了水稻壳制成的生物炭对自然条件下田间土壤中莠去津和乙氧氟草醚环境行为的影响,结果表明生物炭的添加使除草剂在农田中的消散速率加快,莠去津的半衰期从119 d 缩短到52 d,乙氧氟草醚的半衰期从260 d 缩短到了102 d,对莠去津和乙氧氟草醚淋溶现象的阻控率为81.1%和28.4%,对横向扩散的阻控率为89%和40%.并且对莠去津和乙氧氟草醚污染土壤进行了田间修复试验,结果表明生物炭的添加减少了大豆对莠去津37%~90%的吸收量和玉米对乙氧氟草醚50%~86%的吸收量,减轻了除草剂污染对大豆和玉米的药害,提高了农产品的产量和质量.

上述研究证明了生物炭的添加能够吸附土壤中 的除草剂,减少作物吸收,增强微生物降解过程,并 且可以改良土壤,在农田土壤除草剂的修复中具有 广阔的前景. 但生物炭修复在应用中还存在一些问 题.一方面生物炭的吸附作用降低了土壤中除草剂 的生物利用度,减缓了生物降解速度,另一方面生物 炭又可以通过提供碳源和其他营养物质增加微生物 的活性,加速除草剂的微生物降解过程. 因此关于生 物炭对土壤中除草剂命运的影响和其机制还需进一 步研究. 并且生物炭在制造过程中会产生如生物油 和灰烬等副产品,热解时会留在生物炭的表面,而这 些物质进入土壤后是否会对土壤造成影响,还没有 得到充分的了解[103~105]. 在实际应用过程中还需要 考虑制备条件、添加量和修复时间等限制因素对生 物炭修复效率造成的影响以及成本控制[89,92]. 因 此,如何将生物炭安全和高效地应用到农田土壤除 草剂污染的修复中,还需要进行不断的研究.

5 联合修复技术

为了弥补单一修复技术在土壤除草剂降解方面的不足,研究人员将不同的修复技术进行有机耦合,以达到对土壤除草剂的强化处理效果.

目前研究的热点是植物-微生物联修复技术应用于除草剂的去除,该技术结合了植物修复和微生物修复的优势,充分发挥了植物和微生物在修复过程中的作用,并且成本低,能耗低,可大面积应用,适用于农田土壤面源除草剂污染的修复. 植物-微生物联合修复技术是利用植物在土壤中构成一个特异的根际系统. 首先,植物根系可以为微生物的生长提供

有利场所,植物在生长过程中会分泌一些氨基酸、糖和小分子酸等物质,为根际中的微生物提供生长必须的碳源和氮源,促进微生物的生长代谢活动,加速除草剂的微生物降解过程.与此同时,根际土壤微生物对除草剂的降解矿化作用可以改变一些难利用或高毒性除草剂的形态,增强植物对除草剂诱导的氧化损伤的抗性,减少除草剂对植物的毒害作用,还可以增加植物对除草剂的吸收,增强植物修复效果^[21,106,107].植物根系和根际的微生物的协同作用加速了除草剂的降解.

James 等[108]用菖蒲、芦苇和香蒲与从它们根面 分离的菌株 A. calamus-Pseudomonas sp. strain ACB 联合修复莠去津,结果表明植物-假单胞菌的组合对 5 mg·L⁻¹ 莠去津的去除率分别为 91%、72% 和 86%,对 10 mg·L⁻¹莠去津的去除率分别为 87%、 64%和80%,显著高于单独植物或微生物的处理. Zhang 等[109] 从东北黑土中筛选出 1 株莠去津降解 菌株 Arthrobacter sp. strain DNS10 和狼尾草联合修 复莠去津,结果表明植物-微生物联合作用对莠去津 的降解率为98.10%,而单独菌株或植物的降解率只 有 87. 38% 和 66. 71%. 这表明植物-微生物联合修 复的效果要优于单一修复技术. 近几年, 为了降低除 草剂对修复植物的氧化损伤,具有高耐受性的转基 因植物被用于除草剂的修复. Yan 等[50]研究了组合 转基因植物-微生物联合修复除草剂异丙隆,使拟南 芥叶绿体中表达细菌的 N-demethylase PdmAB 基 因,结果表明转基因后的拟南芥对异丙隆表现出了 显著的耐受性,和细菌菌株 Sphingobium sp. Strain 1017-1 联合修复异丙隆污染土壤,可在 20d 内完全 降解 15 mg·kg⁻¹和 30 mg·kg⁻¹的异丙隆,而单独植 物处理的去除率分别为75%和44.8%,单独微生物 处理的去除率分别为81.2%和51.2%.也有研究用 作物自身和微生物联合来修复除草剂. Ahmad 等[110]研究了作物小麦和双草醚降解细菌联合体 (BDAM)对双草醚的联合修复效果,结果表明在2 mg·kg⁻¹和5 mg·kg⁻¹双草醚的污染土壤中,在种植 小麦和接种降解菌 45d 后,根际土壤中双草醚的降 解率均为100%.而相同时间内,只接种降解菌的污 染土壤中双草醚的降解率分别为 96% 和 90%, 只种 植小麦的污染土壤中双草醚的降解率分别为93% 和 84%.

除此之外,还有一些联合技术也具有农田土壤除草剂修复应用的潜力.例如,材料和微生物联合修复技术.Tao等[III]用秸秆制成的生物炭与4种磷酸盐溶解菌和莠去津降解菌 Acinetobacter lwoffii DNS32联合修复莠去津污染土壤,结果表明生物炭的添加

增加了微生物的活性,在水溶液和土壤实验中,微生物-生物炭联合作用相比于单独的微生物作用,在24 h 内降解 $100 \text{ mg} \cdot \text{L}^{-1}$ 莠去津的能力提高了 49%,在3 d 内降解 $20 \text{ mg} \cdot \text{kg}^{-1}$ 莠去津的能力提高了 27%. Fang 等 [112] 利用纳米 Fe_3O_4 和土壤原生微生物联合降解污染土壤中的 2,4-D,并研究了纳米 Fe_3O_4 对土壤微生物种群和酶活性的影响,结果表明纳米 Fe_3O_4 和土壤原生微生物联合处理 2,4-D 的降解效率要高于单独纳米 Fe_3O_4 处理或土壤原生微生物处理,而纳米 Fe_3O_4 的添加不仅降解了土壤中的 2,4-D,还提高了土壤微生物数量和酶活性.

上述研究结果证实了联合修复技术对除草剂的 降解效率要高于相对应的单一修复技术. 但目前植物-微生物和材料-微生物等联合修复技术仍处于研 发阶段,仅在实验室条件下进行了模拟实验. 不过有 研究表明联合修复技术具有较好发展潜力和应用 前景.

6 展望

- (1)继续筛选具有高效降解和转化能力的植物和微生物种类;利用基因技术将除草剂降解的关键基因导入到微生物或植物体内,增强其耐受性和降解能力.
- (2)研究生物炭修复除草剂的机制和老化效应,和其对土壤肥力和作物生产力的长期影响,了解生物炭对土壤动植物的负面影响,并筛选合适的原材料.
- (3)加强联合修复技术的研究,尤其是植物-微生物联合修复技术,探究联合修复作用的机制.
- (4)修复技术在实际农业生产过程中大规模应 用时,要和区域农艺管理措施有机结合,在注重修复 效果的同时,也要充分考虑技术应用的经济成本.

7 结论

(1)农田土壤中除草剂的去除受土壤类型,除草剂理化性质、残留浓度、污染范围和气候条件等因素的影响,要因地制宜地选取适合的技术或集成技术.目前研究较多技术主要有生物修复技术和生物炭吸附固定技术.其中微生物修复、酶修复和植物修复等生物修复技术具有无破坏性、成本低、环境友好和易于操作等优点,是一种消减农田土壤除草剂残留的可持续修复技术.但目前研究中仍存在一些问题.比如,修复植物和微生物种类繁多,并且对除草剂的去除往往具有专一性,导致修复效果存在差异.生物炭具有出色的吸附性能,能够吸附固定土壤中的除草剂,还能增加土壤养分和保水能力,改

善土壤结构,因此被广泛研究.但在实际应用中也存在一些问题,例如附着在生物炭上的副产品是否会对土壤造成负面影响还不清楚,还有生物炭制备的条件、修复的时间和成本的控制等.

(2)植物-微生物和材料-微生物等联合修复技术将不同的修复技术进行有机耦合.其中植物-微生物联合修复技术是目前研究的热点,结合了植物和微生物修复的优点,并且可规模化应用,具有广泛的应用前景.但目前研究中仍存在一些问题.对植物-微生物联合技术在土壤除草剂残留修复应用中的相关研究还比较少,植物和微生物联合方式多种多样,对不同组合方式的作用机制仍不明确.而关于材料-微生物等联合修复技术,虽然已证明对除草剂具有良好的降解效果,但相关研究较少,对作用机制还不清楚.

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