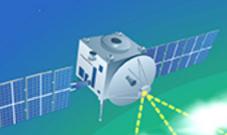


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ENVIRONMENTAL SCIENCE

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PM_{2.5}和O₃污染协同防控区的遥感精细划定与分析 李沈鑫,邹滨,张凤英,刘宁,薛琛昊,刘婧



O₃ PM_{2.5}

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PM_{2.5}



- 主办 中国科学院生态环境研究中心
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基于地理探测器的镇域尺度土壤重金属含量空间分异 及其影响因素分析

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摘要:为揭示镇域尺度规模土壤重金属含量的空间分异及其影响因素,以成都平原腹心地带某镇为研究区域,采集788份表层土壤样品,利用地累积指数法对土壤 Cd、Hg、As、Cu、Pb、Cr、Zn 和 Ni 的污染进行评价,基于地理探测器,以土壤性质、地形和距离等15种因子为自变量,各重金属含量为因变量,探析土壤重金属含量的空间分异及其影响因素.结果表明,研究区土壤 Hg、As、Pb、Cr、Cu、Ni 和 Zn 含量平均值是成都市土壤背景值的1.06~1.93倍,Cd 含量低于背景值;除 Hg 呈现轻度污染外,其余7种重金属处于清洁状态.8种重金属空间分布存在显著差异,且各重金属间存在显著相关性,并与土壤性质存在显著相关性。因子探测发现,总磷(TP)、总钾(TK)、pH、总有机碳(TOC)、高程和距铁路距离对8种重金属含量的解释力最显著.交互作用探测发现,土壤性质与其他因子交互作用是重金属空间分异的最主要影响因素,高程、距住宅区距离、距铁路距离和距工业用地距离也是重要影响因子.风险探测发现,Hg 在高程和距铁路距离的子区域的差异最显著,其余7种重金属在土壤性质影响因子子区域的差异最显著.镇域尺度规模土壤重金属的空间分布差异显著,这与研究区的土壤性质、地形和人类活动密切相关.

关键词:镇域尺度;土壤重金属;地理探测器;空间分布;地累积指数;相关性分析中图分类号: X53 文献标识码: A 文章编号: 0250-3301(2022)10-4566-12 **DOI**: 10.13227/j. hjkx. 202112077

Spatial Differentiation and Influencing Factor Analysis of Soil Heavy Metal Content at Town Level Based on Geographic Detector

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Abstract: Geographic detectors can quickly detect spatial stratified heterogeneity and quantitatively reveal the intensity of driving factors of heavy metal content, which is of great significance for the prevention, control, and remediation of soil heavy metal pollution. In order to reveal the spatial differentiation and influencing factors of soil heavy metal content on the town-scale, 788 topsoil samples were collected from a town in the hinterland of Chengdu Plain. Soil heavy metal (Cd, Hg, As, Cu, Pb, Cr, Zn, and Ni) pollution risk assessments were carried out by using the geo-accumulation index method. Additionally, based on the geographic detector model, 15 factors such as soil properties, topography, soil forming factors, and distance were taken as independent variables, and the contents of each heavy metal element were taken as dependent variables to explore the spatial differentiation and influencing factors of heavy metal content in soils. The results showed that: the average contents of Hg, As, Pb, Cr, Cu, Ni, and Zn in the study area were 1.06-1.93 times the background value of Chengdu, and the content of Cd was lower than the background value; among them, Hg reached the light pollution level, and the other seven heavy metals were at the non-pollution level. The spatial distribution of eight heavy metals was significantly different, the correlation among the elements was significant, and a significant correlation was found between most heavy meals with soil properties; however, the correlation with distance factor and topographic factor was relatively weak. The factor detection showed that TP, TK, pH, TOC, elevation, and distance from the railway had the most significant explanatory power for the heavy metal contents. Interaction detection showed that the interaction between soil properties and other factors was the dominant factors of the spatial variation in heavy metals, and elevation, distance from residential area, distance from railways, whereas the other seven heavy metals had the most signi

Key words: town level; soil heavy metals; geographic detectors; spatial distribution; index of geo-accumulation; correlation analysis

土壤是人类生存、生产和发展必不可少的重要资源,是生态系统的基本组成部分,也是各种污染的媒介[1].在过去的几十年中,土壤重金属污染因其毒性和难降解性而成为一个重要的环境问题.调查显示全国土壤 Cd、Hg、As、Cu、Pb、Cr、Zn 和 Ni 的点位超标率分别为 7.0%、1.6%、2.7%、2.1%、1.5%、1.1%、0.9% 和 4.8% [2].重金属污染会改变

土壤物理和化学性质、破坏生态环境,甚至导致灾难性的环境问题^[3,4]. 此外,重金属可以通过直接吸入、皮肤吸收或作物摄入等方式被人体吸收和积

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累,进而导致严重的公共卫生问题^[5,6].有效处理土壤重金属污染、追踪其来源和探索影响土壤重金属分布的因素都是关键.因此,近10年来,国内外对土壤重金属污染的驱动因素进行了大量的研究和分析^[7~13].

多元统计分析如主成分、因子、聚类和回归等分析,已被广泛应用于研究土壤重金属污染程度与污染源的相关性. 李伟迪等[14]利用多元统计分析与地统计分析相结合的方法,对土壤重金属空间分布及来源进行解析; Li 等[15]使用改良受体模型解析土壤重金属来源; 宋波等[16]采用相关分析和主成分分析方式解析了广西西江流域农田土壤重金属的来源; 赵靓等[17]采用主成分分析/绝对主成分分数受体模型解析了北方某市城市绿地土壤重金属来源. 然而,在多元统计分析中,重金属的空间分布与环境因素之间的关系通常被忽略,或分析了单一因素的影响,没有考虑多因素相互作用的影响强度[18].

相比之下,王劲峰等[19]开发的地理探测器可揭 示单一因素对因变量的影响,以及双因素相互作用 的影响,而不需要考虑线性,避免多变量共线性的影 响,可定量确定各因子对土壤重金属空间异质性的 影响. 因其更直观、快速和有效地衡量各因子的贡 献[20],没有较强的模型假设,解决了传统方法在分 析类别变量时的局限性[21],已被广泛用于地下 水[22,23]、土地利用[24~26]和生态脆弱性[27~30]等多个 领域. 其在土壤重金属污染领域也逐步得到应用, 如 李雨等[31]利用地理探测器模型与空间插值相结合 的方式研究了土壤 Pb、Cd、As、Cr 和 Hg 与 6 种影 响因子的相关性和交互作用. 齐杏杏等[21] 利用地理 探测器定量解释了全国范围内土壤 Cd、Pb、Zn、 As、Cu 和 Cr 与 16 种影响因子的相互作用关系. 张 军等[32]基于地统计方法及地理探测器模型揭示了 宝鸡市土壤重金属污染及其驱动因素,但研究尺度 集中在国家、省市和县级别,多以中大型尺度为主. 大中型尺度上主要揭示了土壤重金属影响因素的整 体趋势和宏观规律,其含量变化主要受自然因素控 制,大尺度的土壤重金属含量差异分布容易忽略小 尺寸或微观尺度区域污染源的影响[21,33,34]. 然而对 于镇域尺度规模土壤重金属含量的空间分布特征及 其驱动因素定量研究鲜见报道,难以准确把握微观 尺度上土壤重金属的空间变化规律. 故本研究以成 都平原腹心地带某镇为研究区域,尝试基于微观尺 度视角,应用地理探测器,定量分析: pH、总有机碳 (TOC)、总氮(TN)、总磷(TP)、总钾(TK)、高程 (X_1) 、坡度 (X_2) 、坡向 (X_3) 、土地利用类型 (X_4) 、

距工业用地距离 (X_5) 、距商服用地距离 (X_6) 、距铁路距离 (X_7) 、距住宅用地距离 (X_8) 、距公路(主干路)距离 (X_9) 和距河流距离 (X_{10}) 等影响因子对镇域规模土壤重金属含量空间分布特征、驱动因素及其交互作用,以期为我国镇域尺度规模土壤环境质量保护和管理、污染防控治理提供科学依据和指导.

1 材料与方法

1.1 研究区概况

研究区位于四川省成都市成都平原腹心地带,距离成都市区约40 km,全镇面积约80 km².属亚热带季风性湿润气候,年平均气温15.7℃,年均降雨量972 mm,年均日照时间1280.9 h. 区域内除少部分区域为浅丘台地外,大部分为平坝,90%的土质为冲击形成的黑色油沙土.全镇3/4 区域位于水源保护区范围,是成都市饮用水源保护核心区,柏条河和徐堰河等多条主要河流流经镇域.成灌高速、国道317和成灌快铁等主要交通干线贯穿全境.土地利用类型以耕地为主(约46.5%),林地(24.1%)和住宅用地(13.9%)次之,1.9%的工业用地零星分布在西南和中部等区域.

1.2 样品采集和处理

样点布设以1 km² 为单位格子,耕地和林地密度为9点·km²²,居住区、学校、工业用地和建设用地等区域密度为4点·km²².2021年4月完成样品采集.依据《土地质量地球化学评价规范》(DZ/T0295-2016),共采集表层土(0~20 cm)788个样品和16个重复样品,采样位置如图1所示.为了提高样品代表性,采用"X"型采样法从每个采样点周围20~50m区域采集5个子样品混合成一个样品,并使用便携式GPS定位采样点.样品自然风干一周,除碎屑,过10目塑料筛,外送中国地质科学院矿产综合利用研究所分析测试.pH采用离子选择电极法测定,TOC采用容量法测定,TN采用燃烧红外法测

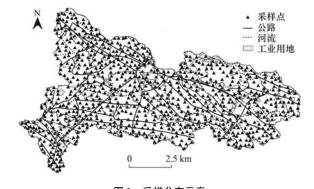


图1 采样分布示意

ig. 1 Map of sampling sites

定,As、Hg和Se采用原子荧光法测定,Cu、Pb、Zn、Ni、Cr、Cd、TP和TK采用X射线荧光法和电感耦合等离子体光/质谱法测定.采用插入国家一级土壤标准物质、重复性检验、异常点检查和空白试验等手段控制分析测试质量.

1.3 地累积指数法

地累积指数法用于比较土壤中不同重金属的浓度及其污染程度. 一般以地累积指数 I_{eee} 来表示 $[^{35}]$:

$$I_{\text{geo}} = \log_2 [C_i / (K \times B_i)]$$
 (1)

式中, I_{geo} 为重金属i的地累积指数; C_i 为土壤重金属i的测定值; B_i 为参比值,选用成都经济区背景值;K为修正系数,一般为 1.5. 污染程度划分为 7个等级: $I_{geo} < 0$ 、 $0 \le I_{geo} < 1$ 、 $1 \le I_{geo} < 2$ 、 $2 \le I_{geo} < 3$ 、 $3 \le I_{geo} < 4$ 、 $4 \le I_{geo} < 5$ 和 $I_{geo} \le 5$ 分别对应:无污染、轻污染、中污染、中-重污染、重污染、重-极重污染和极重污染.

1.4 地理探测器

地理探测器(geographical detector)通过计算分类后各自变量方差之和与因变量方差之和的比来衡量自变量对因变量的贡献,包括因子、交互、风险区和生态这4种探测器^[19].

因子探测器:用于探测因变量的空间分异性以及各自变量对因变量影响程度的解释能力,用q值度量 $^{[19]}$:

$$q = 1 - \frac{\sum_{h=1}^{\infty} N_h \sigma_h^2}{N\sigma^2} = 1 - \frac{\text{SSW}}{\text{SST}}$$
 (2)

式中, $h=1,\dots,L$ 为自变量 X 的分类数; N_h 和 N 分别为分类 h 和整个区域内单元的数量; σ_h^2 和 σ^2 分别为分类 h 和区域内因变量 Y 的方差. SSW 和 SST 分别为自变量 X 所有分类的方差之和以及区域内的总方差. q 的值域为[0,1], q 值越大,表明该自变量 X 对因变量 Y 的影响程度越大.

交互探测器:通过识别两个不同自变量交互时的 q 值,判断自变量之间的交互作用对因变量的影响程 度,判断依据: $q(X_a \cap X_b) < \min[q(X_a), q(X_b)]$ 交互 作用 为 非 线 性 减 弱; $\min[q(X_a), q(X_b)] < q(X_a \cap X_b) < \max[q(X_a), q(X_b)]$ 为单因子非线性减 弱; $q(X_a \cap X_b) > \max[q(X_a), q(X_b)]$ 为双因子增强; $q(X_a \cap X_b) = q(X_a) + q(X_b)$ 为独立交互作用; $q(X_a \cap X_b) > q(X_a) + q(X_b)$ 为非线性增强.

风险区探测:用于探测影响因子对土壤重金属 是否有具有风险性,用 t 统计量来检验:

$$t_{\bar{Y}_{h=1}-\bar{Y}_{h=2}} = (\bar{Y}_{h=1} - \bar{Y}_{h=2}) \cdot \left[\frac{\operatorname{Var}(\bar{Y}_{h=1})}{n_{h-1}} + \frac{\operatorname{Var}(\bar{Y}_{h=2})}{n_{h-2}} \right]^{-1/2}$$
(3)

式中, \bar{Y}_h 为子区域 h 内的属性均值,本研究为某元素含量; Var 为方差; n_h 为子区域 h 内样本数量; 统计量 t 近似地服从 Student's t 分布, t 值越大代表该影响因子对土壤重金属的空间分异性影响越大.

生态探测器:用于比较两个影响因子对土壤重 金属空间分布的影响是否有显著的差异,以 F 统计 量来衡量:

$$F = \frac{\text{SSW}_{X_a} N_{X_a} (N_{X_b} - 1)}{\text{SSW}_{X_b} N_{X_b} (N_{X_a} - 1)}$$
(4)

 $SSW_{X_a} = \sum_{h=1}^{L_a} N_h \sigma_h^2$, $SSW_{X_b} = \sum_{h=1}^{L_b} N_h \sigma_h^2$ (5) 式中, N_{X_a} 和 N_{X_b} 分别为两个自变量 X_a 和 X_b 的样本量; SSW_{X_a} 和 SSW_{X_b} 分别表示由 X_a 和 X_b 形成的分层的层内方差之和; L_a 和 L_b 分别为变量 X_a 和 X_b 分层数目. 其中零假设(H_0): $SSW_{X_a} = SSW_{X_b}$. 如果在 α 的显著性水平上拒绝 H_0 ,则表明两个自变量 X_a 和 X_b 对属性因变量 Y 的空间分布的影响存在着显著的差异.

1.5 因子指标选取及数据处理

参考齐杏杏等[21]、李雨等[31]和张军等[32]的因 子指标选取方法,结合土壤中重金属元素的来源,并 考虑数据获取的难易程度和研究区实际情况,选取 土壤性质(pH、TOC、TN、TP、TK)、地形因子(高 程、坡度、坡向)、成土因素(土地利用类型)和距 离因子(距公路、河流、工业用地、铁路、住宅区、 商服用地的距离)这15个因子. 高程数据 (GDEMDEM 30m)来自于地理空间数据云(http:// www.gscloud.cn).利用地理探测器对影响因素进行 分析时,因变量必须为数值量,自变量必须为类型 量,若自变量是数值量,需将其离散化处理为类型 量[19]. 参考李雨等[31]、周洋等[36]和张军等[37]的研 究,采用自然断点法将15个影响因子分别划分为6 类. 利用 SPSS 26.0 对数据进行描述性统计分析及 相关性分析,利用 ArcGIS10.8 绘制采样和空间分 布,利用 Origin 2019 完成绘图,地理探测器采用 GeoDetector 软件(http://www.geodetector.org/) 完成.

2 结果与讨论

2.1 研究区表层土壤基本性质

研究区表层土壤重金属含量及理化性质见表 1. 土壤平均 pH 值为 6. 17, 范围为 4. 16 ~ 9. 04, 采样点中呈酸性(pH \leq 6. 5)、中性(6. 5 < pH \leq 7. 5) 和碱性(pH > 7. 5) 的占比分别为 69. 0%、18. 7% 和 12. 3%. 土壤 ω (TN)、 ω (TP)、 ω (TK)和 ω (TOC)的平均值和范围分别为 1. 33、1. 16、23. 6、17. 7 和 0. 70 ~ 2. 40、

 $0.22 \sim 20.8 \times 10.3 \sim 28 \times 2.40 \sim 43.8 \text{ g} \cdot \text{kg}^{-1}$.

土壤 ω (Cd)、 ω (Hg)、 ω (As)、 ω (Pb)、 ω (Cr)、 ω (Cu)、 ω (Ni)和 ω (Zn)的平均值分别为 0. 221、0. 155、9. 76、32. 2、91. 9、35. 2、37. 1 和 108. 8 mg·kg⁻¹. 除 Cd 外, Hg、As、Pb、Cr、Cu、Ni和 Zn 的含量平均值分别是成都土壤背景值的 1. 93、1. 07、1. 06、1. 18、1. 25、1. 11 和 1. 32 倍,说明该区域土壤重金属存在一定程度富集. Cu、Hg

和 Zn 的高变异系数说明它们在不同采样点的含量存在大差异,表明它们可能受到较为明显的外部干扰因素的影响. 许多研究指出,变异系数与人类活动等外部因素的干扰程度呈正比^[38~40]. 研究区土壤中的 8 种重金属含量平均值均低于土壤污染风险筛选值(GB 15618-2018),但 Cd、Pb、Cr、Cu 和 Zn 分别在 76、1、1、6 和 2 个采样点中的含量高于土壤污染风险筛选值.

表 1 土壤组分描述性统计结果1)

Ta	bl	le	1	Descri	ptive	statistical	results	of	soil	composition	
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40 /\	亚拉佐	仁州伯子	由总法	具示估	具上店	亦巳妥粉	라	1	_壤污染	4风险筛选值	ĺ
组分	平均值	标准偏差	中位值	最小值	最大值	变异系数	成都市背景值[41]	A	В	С	D
Cd	0. 221	0.069	0. 210	0. 082	0. 83	31. 05	0. 25	0. 3	0.3	0.3	0.6
Hg	0. 155	0.090	0.130	0.022	0.88	58. 13	0. 08	1.3	1.8	2. 4	3.4
As	9. 76	2.08	9.43	4. 17	18. 0	21. 35	9. 11	40	40	30	25
Pb	32. 2	5. 00	32. 3	19.8	90. 3	15. 42	30. 3	70	90	120	170
Cr	91.9	10.0	92. 3	61.7	264	10.87	78	150	150	200	250
Cu	35. 2	21. 1	34. 4	18. 5	607	59. 93	28. 1	50	50	100	100
Ni	37. 1	3. 70	37. 2	23.6	56. 6	10.07	33. 5	60	70	100	190
Zn	108.8	62. 9	106.0	55.0	1 820	57. 80	82. 2	200	200	250	300
TN	1. 33	0. 26	1.30	0.70	2. 40	19. 6	///	N A	/	// //	3 1
TP	1. 16	0.79	1.11	0. 22	20.8	68. 14	11		/	1/5	11
TK	23.6	2. 40	24. 3	10.3	28. 0	10. 15	1º1V à	X	/	VP! (28
TOC	17.7	5. 10	17. 6	2, 40	43.8	28. 62	(PY of	4/1	/	/	D// .
pН	6. 17	1.01	5. 96	4. 16	9.04	16.40	6. 14	1 8	/	/_	1/
-		1 07		1 /11/11	15		/ 1 1			/	10

1) Cd、Hg、As、Pb、Cr、Cu、Ni 和 Zn 的单位为mg·kg⁻¹,TN、TP、TK 和 TOC 的单位为g·kg⁻¹,pH 无量纲,变异系数的单位为%;土壤污染风险筛选值参考《土壤环境质量 农用地土壤污染风险管控标准(试行)》(GB 15618-2018),A 表示 pH≤5.5,B 表示 5.5 < pH≤6.5,C 表示 6.5 < pH≤7.5,D 表示 pH >7.5; /表示文献中无相关数据

2.2 土壤重金属含量空间分布特征

图 2 展示了研究区表层土壤 8 种重金属的空间分布. 从中可知, Cu、Ni 和 Zn 的高值区总体呈带状分布于中西部区域, 低含量主要出现在西部区域; Hg 和 Pb 高含量区主要集中在研究区中部; Cd 的高值区零星分布在中部及东部区域, 东部区域总体高于中西部区域; As 的高值区主要集中在西部村落, 西部区域明显高于中东部区域; Cr 的高值区零散分布于全镇除西部少数村外的各个行政村. 可见, 镇域尺度下的土壤重金属含量空间分布存在显著性差异.

2.3 土壤重金属污染评价

8 种土壤重金属的 $I_{\rm geo}$ (表 2) 平均值依次为: Hg (0.18)、Zn(-0.22)、Cu(-0.30)、Cr(-0.36)、Ni(-0.45)、 Pb(-0.51)、 As(-0.52) 和 Cd(-0.82),除 Hg 外都小于 0,表明研究区土壤 Zn、Cu、Cr、Ni、Pb、As 和 Cd 总体处于无污染状态,而 Hg 整体处于轻度污染状态. 从采样点看,Pb、Ni 和 As 各有 1.65%、0.25% 和 4.95% 采样点的 $I_{\rm geo}$ 处于 $0\sim1$ 之间,属于轻度污染,其余采样点均处于无污染状态; Cr 和 Cd 分别有 0.63% 和 2.54% 的采

样点属轻度污染,都有 0. 13% 采样点属于中度污染 $(2 < I_{geo} \le 3)$; Cu 和 Zn 都有 0. 13% 采样点属重度 污染,分别有 0. 25% 和 0. 13% 为中度污染,属轻度 污染的分别占 3. 81% 和 7. 61%; Hg 处于无污染、轻度、中度和中度-重度污染分别占 47. 34%、40. 61%、10. 66% 和 1. 40%. 说明研究区土壤 Zn、Cu、Cr、Ni、Pb、As 和 Cd 在少部分采样点可能存在点污染源,Hg 存在面源污染源. 李冰等[42]的研究 指出,成都平原土壤 Hg 含量的总体分布趋势为北部相对较高,主要是受地质构造、城镇居民生活污染和工业企业污染物排放等的影响.

2.4 相关性分析

相关性分析结果显示多数重金属间存在显著 (P < 0.05) 相关性(表 3), 尤其是 Hg-Cd、As-Cd、Cr-Pb、Cu-Cd、Ni-Hg-Pb、Ni-Cu 和 Zn-Cr-Ni 之间存在极为显著 (P < 0.01) 的相关性. 土壤性质中,除 Zn 外的其他 7 种元素与 TN、TP、TK、pH 和 TOC中的两个及以上因子存在极为显著或显著相关性,表明土壤性质对 Cu、Pb、Cr、Ni、Cd、As 和 Hg 元素有显著影响. 地形因子中,高程与 Zn、Cr、Ni 和 Cd 有极为显著的负相关性,与 As 有极为显著正相

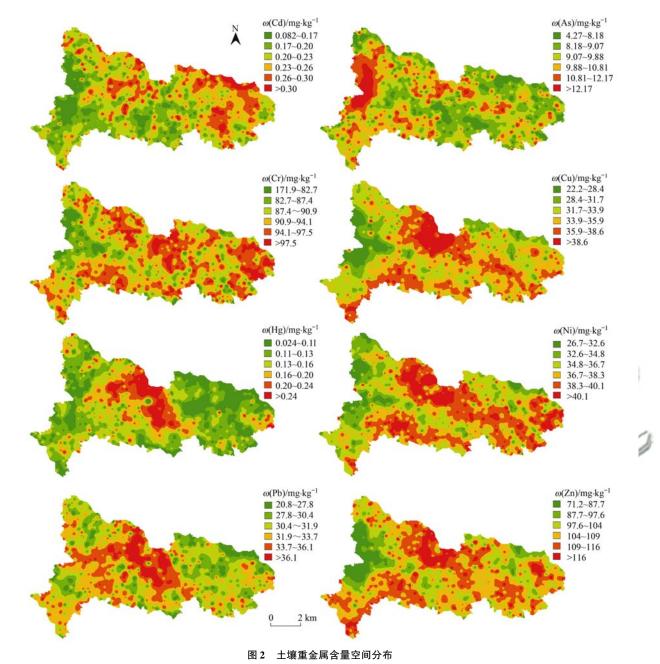


Fig. 2 Spatial distribution of the heavy metals in the topsoil

表 2 研究区土壤重金属地累积指数

Table 2 Index of geo-accumulation for soil heavy metals in study aera

元素		$I_{ m geo}$				各分级中	样品数占总样品	数比例/%		
儿系	平均值	最大值	最小值	$I_{\text{geo}} \leq 0$	$0 < I_{\text{geo}} \leq 1$	$1 < I_{\text{geo}} \leq 2$	$2 < I_{\mathrm{geo}} \leq 3$	$3 < I_{\mathrm{geo}} \leq 4$	$4 < I_{\text{geo}} \leq 5$	$I_{\rm geo} > 5$
Cu	-0.30	3. 85	- 1. 19	95. 81	3. 81	0. 25	0.0	0. 13	0	0
Pb	-0.51	0. 99	-1.20	98. 35	1. 65	0.0	0.0	0.0	0	0
Zn	-0.22	3.88	-1.16	92. 13	7. 61	0. 13	0.0	0. 13	0	0
Cr	-0.36	1.17	-0.92	99. 24	0. 63	0. 13	0.0	0.0	0	0
Ni	-0.45	0.17	-1.09	99. 75	0. 25	0.0	0.0	0.0	0	0
Cd	-0.82	1. 15	-2.19	97. 34	2. 54	0. 13	0.0	0.0	0	0
As	-0.52	0.40	-1.71	95.05	4. 95	0.0	0.0	0.0	0	0
Hg	0. 18	2. 87	- 2. 45	47. 34	40. 61	10. 66	1.40	0.0	0	0

关性,表明小尺度研究区域的高程对土壤重金属元素也有一定影响;坡度和坡向除与 Cd 和 Pb 存在显著相关性外,与其他元素无明显的相关性.土地利用

类型仅与 Ni 有显著负相关性,表明土壤利用类型对研究区土壤重金属的影响较弱. 距离因子中, As 与距工业用地距离有极为显著负相关性; 距商服用地

表 3 土壤重金属和影响因子相关性系数^[]

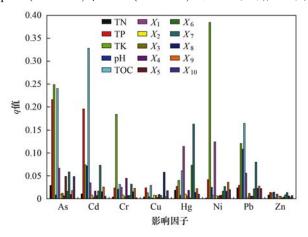
							Table 3		Pearson correlation coefficient of soil heavy metal elements and impact factors	lation co	efficient	of soil h	eavy met	al elemen	nts and in	npact fac	tors							
	Cu	Pb Zn	Cr	r. Ni	i Cd		As I	Hg	Se	NI.	TP	TK	Hd	TOC	X_1	X_2	X_3	X_4	X_5	X_{6}	X_7	X_8	X_9	X_{10}
Cī	-									4	0	6												
Pb	0.051	1								J	9	L	91	/										
Zn	0.033 0	0.076 * 1									SF		l'	ì										
Ċ	(0.004) 0	0.219 ** 0.213 **	*								3													
N.	0.250 ** 0	0.290 ** 0.314 **	** 0.454 **	4** 1																				
g	0.386 ** 0	0.148 ** 0.069	0.019	9 0.039	39 1							- 5												
As	(0.053) (0	(0.072) * (0.072) * (0.014)) * (0.01		0.086* (0.312) ***) *** 1				×	(1) !	1										
Hg	0.091 * 0	0.602 ** 0.087	* 0.087 *		0.191 ** 0.161 **	** (0.129) ***	9) ** 1				1	//		1										
æ	0.033 0	0.296 ** 0.060	0. 193 **		0.153 ** 0.346 **	(0.307) ***		0.182 ** 1			6	1												
IN	0.023 0	0.146 ** 0.010	0.046		(0.016) 0.103 **		(0.137) ** 0.12	0.123 ** 0.132	132 ** 1	1	18	0/		3	1	_								
TL	0.882 ** (0.010)	0.010) 0.007	(0.069)	9) 0.051	51 0.462 **	(0.199)	9) ** 0.052	52 0.079	*	0.038	_				ı									
TK	(0.043) 0	0.197 ** 0.038	0.370 **		0.585 ** (0.076) *	(0.140) **	0) *** (0.1;	0.156 ** 0.2	0.208 ** 0.	0.035 (0)	(0.064)	4		2										
$^{\mathrm{hd}}$	0.032 (0	(0.338) *** 0.004	(0.129) ***		(0.103) ** 0.272 **	,** 0.045	(0.028)		(0.125) *** (0.	(0.104) ** (0.025)		(0.182) **	_											
TOC	0. 143 *** 0	0.364 ** 0.021	0.114 ***		(0.033) 0.599 **		(0.482) *** 0.236 ***		0.482 *** 0.	0. 245 ** 0.	0. 286 ** 0. 002		(0.138) ****	-										
X_1	(0.001) 0	0.049 (0.098) **) ** (0.145) **		(0.274) **(0.162) **		0.218 ** (0.039)		(0.110) *** 0.	0.150 ** (0.025) (0.209) ** (0.125) **	025) (0), 209) ** (0.125) **	(0.030)	_									
X_2	0.035	0.018 (0.011)	(0.011)	1) 0.022	22 0.086*	* (0.024)	4) 0.037		(0.068) (0.	(0.034) 0	0.012 (0	(0.009)	0.147 *** (0.016) (0.121) ***	(0.016)	(0.121) ***	-								
X_3	0.030 (0	(0.074) * (0.005)	(0.032)	2) 0.011	11 0.034	0.056	(0.008)		(0.052) (0.	(0.024) 0	0.019 (0	(0.011)	0.049	(0.001) (0.066)	(0.066)	0.012	_							
X_4	(0.022)	(0.013) (0.027)	0.009		(0.078) * 0.020	0.055	5 0.047		(0.083) * (0.	(0.070) * (0.00)	(0.037) (0	(0.050)	0.113 ** (0.030)	-	0.033	0.010	(0.096) **	_						
X_5	0.028 (0	(0.051) (0.038)	(0.084)*	4) * (0.057)	57) (0.047)		(0.208) ** (0.025)	25) 0.033		(0.019) 0) * 140 !	0.077 * 0.059 (0.034)	0.034)	0.008	(0.08)	(0.056)	0.009	(0.106) **	_					
χ_{6}	(0.072) * (0.116) **	0.116) ** 0.018	(0.056)		(0.098) ** 0.020	(0.092) *		(0.224) ** 0.036		(0.043) (0	010) (0	(0.010) (0.143) ** (0.100) **	- 4	0.090 * (0.091) *	(0.091)*	(0.134) ** 0.021	* 0.021	(0.085)*	0. 282 **	_				
X_7	(0.013) (0	(0.167) ** 0.085 *	* 0.086 *		0.137 *** 0.255 ***		(0.209) ** (0.07	(0.078) * 0.054		(0.055) 0	0) 800 0	(0.014)	0. 226 ***	0.103 ***	0.103 ** (0.695) **	0.118 ** 0.062	0.062	(0.037)	0.017	0.144 **	_			
X_8	0. 202 ** (0	(0.124) ** (0.047)) (0.101) **		(0.074) * 0.023	0.046		(0.091) * (0.1	(0.133) *** (0.	(0.066) 0.). 178 ** (0	0.178 ** (0.199) ** 0.075 *	0.075 *	(0.051)	0.053	(0.008)	0.060	0.081 *	0.062	(0.011)	0.043	1		
X_9	(0.021) (0	(0.148) *** (0.034)) (0.090) *		(0.156) *** 0.147 ***		(0.109) *** (0.14	(0.141) *** (0.019)		(0.056) 0	0.040 (0	(0.219) ** 0.072 *	0.072 *	0.127	0.127 *** (0.169) ***	0.039	0.014	(0.043)	0.057	0.238 ***	0.301 ***	0. 140 ***	_	
X_{10}	0.066	0.134 ** (0.005)	0.005		0.077 * 0.000	0.205 ***	5 *** 0.031		(0.049) 0.	0.042 0.	0.005 (0	(0.081)* (0.061)	0.061)	(0.047)	0.070	0.018	(0.008)	0.009	$(0.109)^{**} (0.015)$	(0.015)	(0.124) ** 0.175 **	* 0.175 **	0.027	-
** (1	本示 D	(= = ∓ a ∗	0 0 / # 1	-0 05/ 双侧): X 表示真程	·羊 A	中田	世幹芒華 A		A 表示体向		十二十二十	米田田米田		逐盟雇用小工组名等 A 服米田医希子芒等 A	出海田州		1 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2	文里州田湖 短出产業 X		v 丰三肝姓 股	上吸肝密	V 丰云昭住	三胎在

 $1)^{**}$ 表示 P < 0.01(双侧)), * P 表示 0.05(双侧); X_1 表示语程, X_2 表示坡度, X_3 表示生地利用类型, X_5 表示距工业用地距离, X_6 表示距离服用地距离, X_7 表示距铁路距离, X_8 表示距往 今日站时旁,发生与时代的时候,发生与时间的影响,在日本与女器 宅用地距离, X_9 表示距公路距离, X_{10} 表示距离河流距离;括号表示负数

距离与 Pb、Ni 和 Hg 有极为显著负相关性; 距铁路 距离与 Pb 和 As 有极为显著负相关性, 与 Ni 和 Cd 有极为显著正相关性; 距住宅区距离与 Cu 有极为 显著正相关性,与 Pb 和 Cr 有极为显著负相关性; 距公路距离 Pb、Ni、As 和 Hg 有极为显著正相关 性,与 Cd 则有极为显著正相关性; 距河流距离与 Pb 和 As 呈极为显著正相关性. 总的来看,镇域尺度 规模的距离因素对土壤重金属存在一定影响.

2.5 土壤重金属污染影响因素的地理探测器分析 2.5.1 因子探测器

因子探测器探测 15 个因子对 8 种重金属的解 释力 q 值见图 3. 不同因子对 8 种重金属的解释力 存在一定差异,但总体上 TP、TK、pH、TOC、高程 和距铁路距离对各元素含量空间分布均具有较强解 释力. As 的首要影响因素是 TK(0.249),其次是 TOC(0.240),再次是TP(0.216).Cd的首要影响因 素是 TOC(0.328),其次是 TP(0.196),TK(0.074) 和距铁路距离(0.073)也有明显的影响. Cr 的第一 影响因子是 TK (0.184), 其次是土地利用类型 (0.045), 距铁路距离、TOC、高程、TP和距公路距 离也存在较大影响. Cu 的第一影响因子是距住宅区 距离(0.058),其次是 TOC(0.029). Hg 的首要影响 因子是距铁路距离(0.163),其次是高程(0.114), 距商服用地距离、TOC 和 TK 也有较大影响. Ni 的 首要影响因子是 TK(0.385),其次是高程(0.124). Pb 的第一、第二和第三影响因素分别是 TOC (0.165)、pH(0.121)和TK(0.108),距铁路距离和 高程也存在明显的影响. Zn 的前三影响因素分别是 pH (0.0145)、TP (0.0139) 和 距 铁 路 距 离



 X_1 . 高程, X_2 . 坡度, X_3 . 坡向, X_4 . 土地利用类型, X_5 . 距工业用地距离, X_6 . 距商服用地距离, X_7 . 距铁路距离, X_8 . 距住宅用地距离, X_9 . 距公路距离, X_{10} . 距离河流距离, 下同

图 3 影响因子对 8 种土壤重金属的解释力 q 值

Fig. 3 Effects of different factors on the explanatory power of eight heavy metals in soils with q value

(0.0135), 15 种影响因子对 Zn 的空间分布影响与对其他 7 种元素的影响相比明显偏弱,说明影响研究区 Zn 空间分布的影响因子还需进一步探究.

各影响因子对于不同重金属影响程度排序不 同,揭示了不同重金属变化机制的异质性. 总的来 看,影响土壤重金属空间分布的主要因素是农业活 动,研究区种植大面积果蔬和园艺苗圃,使用营养 土、施肥、灌溉和施药等农业活动直接影响土壤理 化性质,并伴随重金属的引入[41,42]. 土壤理化性质 的变化直接影响重金属的活性和迁移转化[43]. 相关 性分析也表明 TOC、TP、TN、TK 和 pH 与一个或多 个重金属存在极为显著相关性,因此土壤理化性质 是影响重金属空间分布的重要指标. 地形因素反映 了土壤重金属空间分布受自然因素影响的大小,影 响过程缓慢[44]. 研究区高程对 As、Cr、Hg、Ni 和 Pb 的空间分布存在明显影响,但对比 Pearson 相关性分 析发现两者结果存在一致性和差异性,一致性如高 程与 Cr、Ni 和 As 存在显著相关性,对它们的空间 分布也有显著影响;差异性如相关性分析显示 Hg 与高程无显著相关性,但因子探测分析发现高程对 Hg 空间分布影响排第二,这是由于地理探测器分析 的是重金属与影响因子间的关联性,包括线性和非 线性关系,而 Pearson 相关系数不显著,说明重金属 与影响因子之间无显著线性关系,但不代表没有非 线性关系[31]. 坡度和坡向对重金属的影响不明显, 可能是由于研究区为小尺寸区域,坡度和坡向变化 范围小的缘故. 土地利用类型仅对 Cr 的空间分布产 生较为明显的影响,对其他7种元素的空间分布无 明显影响,这与其他人的研究存在明显差异[36,45], 可能是由于研究区土地利用类型变化快,界线不明 显所致.

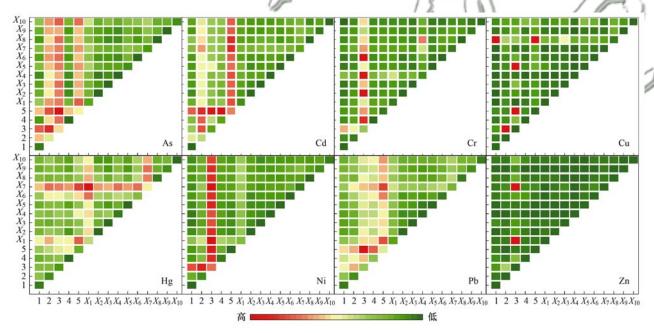
距铁路、工业用地、河流、住宅区、商服用地和公路的距离等因子都对研究区土壤中的一种或多种重金属的空间分布具有重要影响,体现为人类活动对土壤重金属空间分布变化的综合影响,人类活动改变了土壤重金属自然状态下的分布特征,形成新的空间特征[44].河流水系是农业灌溉的重要水源,由工业排放、交通运输造成水源重金属富集进而通过农业灌溉引起农田土壤富集[44,46].Cu的第一影响因子是距住宅区距离,相关研究指出住宅区是人类活动最频繁区域,居民日常生活会产生大量含有重金属的生活垃圾,引起住宅区周围土壤发生改变,随着距离住宅区距离拉大,影响也会减小[47],同时城镇密集的交通网络和频繁的人类活动,使得大量污染物通过大气沉降作用等扩散方式富集到土壤中进而产生污染[48].另外,研究区分布着建材、

塑料和印刷等工厂企业,工业活动产生的"三废"携带的重金属,通过大气沉降、雨水冲刷和渗透等方式富集到土壤[14,40,46]. 许多的研究证明了由于受汽车尾气排放和轮胎磨损等形成含有重金属的颗粒通过大气迁移与沉降进入土壤,使得公路或铁路周围的土壤重金属明显存在富集现象[44,46,49].

2.5.2 交互作用探测

土壤的成分和结构复杂,重金属空间分布和污染通常是由多种因素共同作用形成的,不可能存在单一性质的因素或单一因子影响重金属的分布和变化^[31].因此,利用交互作用探测器分析多种因子对重金属空间分布的交互影响程度,有利于精准判断影响重金属空间分布的深层驱动机制^[37].从图 4 可知,任意两个因子的交互作用对 8 种重金属空间分异性影响的解释程度均大于单个因子的解释程度,多数为非线性增强作用,少数为双因子增强,不存在减弱或独立作用类型.就 As 而言,TOC、TK 和 TP

两两的交互作用值最大,且 TOC、TK、TP 分别与其 他因子的相互作用值都不低于 0.216; 对于 Cd, TOC 与其他因子的交互作用最强,TP 与其余因子的 交互作用值都在 0.200 以上;对于 Cr,TK 与其余因 子以及 $X_8 \cap X_4$ 的交互作用明显强于其他因子的交 互作用;对于Cu,TK∩TP、TK∩TOC、TK∩X5、TN $\cap X_8$ 和 TOC $\cap X_8$ 的交互作用最强, q 值分别是 0.944、0.944、0.939、0.922 和 0.926; 就 Hg 而言, X_1 和 X_7 与其余因子的交互作用占主导地位;对于 Ni,TK 与其余因子的交互作用最强;TOC 与其余因 子以及 pH \cap TK、TP \cap TK、TK \cap X₇ 和 pH \cap X₇ 的交 互作用对 Pb 的解释力最强;对 Zn 而言, $TK \cap X$,和 $TK \cap X_7$ 的交互作用值最大(0.945). 总体而言, 土 壤性质(TN、TP、TK、TOC和pH)与其他因子交互 作用对于研究区重金属的空间分异具有重要影响, 同时高程、距住宅区距离、距铁路距离和距工业区 距离也是重金属累积分布的重要影响因子



1. TN, 2. TP, 3. TK, 4. pH, 5. TOC, 下同 不同影响因子对土壤重金属影响的交互作用

Fig. 4 Interaction of different influence factors on soil heavy metals

2.5.3 风险区探测

利用风险区探测器探测了 15 个因子各自子区域两两间重金属是否存在显著性差异及重金属在各因子子区域的高值区和低值区(图 5). 就 As 而言,其在 TOC 子区域的显著性差异最强烈,其次是 TK和 TP 的多数子区域中呈显著性差异.对于 Cd,在TOC 的 6 个子区域中均存在显著性差异,在 TP 的子区域中除第五类 〇第六类无显著性差异外,其余都存在显著性差异,另外,在 TK、pH 和距铁路距离的多数子区域有显著性差异;对于 Cr,仅在 TK 的

多数子区域具有显著性差异;对于Cu在各因子子区域的显著性差异,最强烈的是TOC,半数以上子区域显示显著性差异;就Hg来看,显著性差异最为强烈是在距铁路距离和高程的多数子区域;对于Ni,其在TK大多数子区域呈现显著性差异,其次是高程和TP的子区域;对于Pb,显著性差异最强烈的是TOC的子区域,其次是TK、pH、距铁路距离和高程的子区域;对于Zn,显著性差异最强烈的是TK的半数子区域.总的来看,Zn、Cu、Cr、Ni、Pb、As和Cd在土壤性质影响因子子区域的差异最显著.

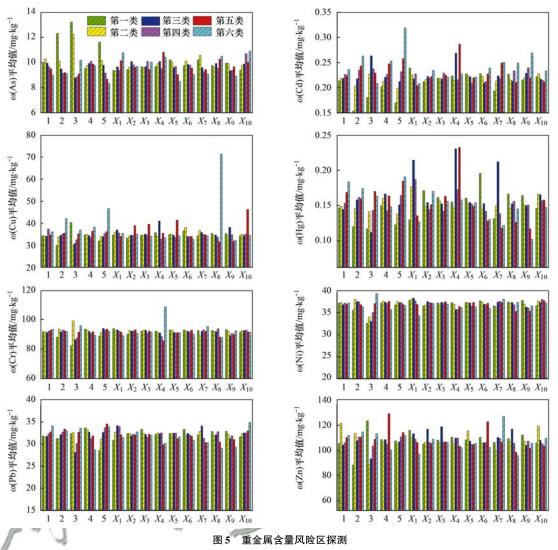


Fig. 5 Risk detection of heavy metals contents

而 Hg 则在距铁路距离和高程的子区域的差异最显著. 另外,分析显示,风险区探测的结果与因子探测结果相一致,即因子探测对重金属解释力大的影响因子,其子区域间的重金属含量也存在显著性差异.

2.5.4 生态探测

生态探测着重比较一个影响因子和另一个影响因子对土壤重金属的空间分布的影响是否有显著的差异^[31,45],若显著,则记为 *Y*,否则记为 *N*.

研究区土壤重金属的生态探测结果显示:TN与TP、TK和TOC以及TOC与pH对As的影响存在显著差异,而其他因子之间的差异不显著;TOC与TN、TP、TK和pH以及TP与TN对Cd的影响呈显著差异,其余影响因子间的差异不显著;TK与TN和TP对Cr的影响存在显著差异;对于Hg的影响,高程与TN、TP和pH以及距铁路距离与TN、TP、TK、pH、TOC、坡度、坡向、土地利用类型、距工业用地距离和距商业用地距离存在显著差异;对于Ni的影响,TK与TN和TP以及高程与TN、pH和TOC

显示显著差异;对于 Pb 的影响, TY 与 TN 和 TP 以及 TOC 与 TN 和 TP 存在显著差异;对于 Cu 和 Zn 的影响, 15 种影响因子间均不存在显著性差异.

3 结论

- (1)研究区土壤重金属 Cd 含量平均值低于成都市土壤背景值,但 Hg、As、Pb、Cr、Cu、Ni 和 Zn含量平均值是背景值的 1.06~1.93 倍,且 Cu、Hg和 Zn 的变异系数均大于 50%. 地累积指数显示 Zn、Cu、Cr、Ni、Pb、As 和 Cd 总体处于清洁状态,而 Hg整体属轻度污染,但 8 种重金属的含量平均值均低于土壤污染风险筛选值(GB 15618-2018).
- (2)研究区土壤重金属空间分布存在显著差异性,Cu、Ni和Zn高值区主要集中在研究区中西部区域,Hg和Pb高值区主要集中在中部村落,As高值区集中在西部区域,Cd和Cr的高值区呈零散分布.
 - (3)相关性分析显示 8 种重金属间多数都存在

显著相关性,土壤重金属与土壤性质的相关性最为强烈,与距离因子和地形因子的相关性次之,与成土因素的相关性最弱.

(4)镇域尺度下 8 种重金属与 15 种影响因子的因子探测结果显示, TP、TK、pH、TOC、高程和距铁路距离对 8 种土壤重金属的解释力尤为显著. 交互探测发现,交互作用解释力均显示增强效应, TN、TP、TK、TOC 和 pH 与其他因子交互作用是重金属空间分异的主导因素, 高程、距住宅区距离、距铁路距离和距工业区距离也是重要影响因子. 风险区探测显示 Hg 在高程和距铁路距离的子区域的差异最显著, 其他重金属在土壤性质影响因子子区域的差异最显著, 其对重金属解释力大的影响因子, 其子区域间的重金属含量的差异性尤为显著. 总体上, 研究区土壤重金属空间分布是多种因素共同作用的结果, 且不同影响因素对不同重金属元素的作用强弱不尽相同, 但人类活动因素对重金属空间分布的扰动最为强烈.

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