

ENVIRONMENTAL SCIENCE

ISSN 0250-3301 CODEN HCKHDV HUANJING KEXUE

2000~2018年我国大气重金属沉降通量时空变化特征 陈其永,郜允兵,倪润祥,潘瑜春,阎跃观,杨晶,刘孝阳,顾晓鹤



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第43卷 第9期 2022年9月15日

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废水排放对近海环境中抗生素抗性基因和微生物群落 的影响

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摘要: 污水处理厂是水环境中抗生素抗性基因(ARGs) 的重要来源. 可移动遗传元件(MGEs) 和微生物群落是影响 ARGs 增殖 扩散的关键因素. 为探究污水处理厂废水排放对近海环境中 ARGs 和微生物群落的影响,采用高通量荧光定量 PCR (HTqPCR) 和高通量 16S rRNA 扩增子测序技术,对杭州湾上虞(SY) 和嘉兴(JX) 两个近岸纳污区(ERAs) 及远岸湾区表层沉积物 中 ARGs、MGEs 和微生物群落的组成和分布进行调查: 结果表明,多重耐药类 ARGs 是所有样点中丰度最高的 ARGs 类型. 纳 污区沉积物中 ARGs 和 MGEs 多样性和丰度远远高于远岸湾区沉积物. JX 纳污区沉积物中微生物群落丰富度和多样性高于 SY 纳污区及远岸湾区沉积物. PCoA 结果显示 .纳污区与远岸湾区沉积物中 ARGs 、MGEs 和微生物群落分布存在显著的差异 . 说明长期的废水排放对近海环境中 ARGs、MGEs 和微生物群落影响较大. ARGs、MGEs 和细菌属的共现网络显示, 鳍袴杆菌 属、假单胞菌属、亚硫酸杆菌属、假交替单胞菌属和芽孢杆菌属等 12 种菌属与 ARGs 和 MGEs 存在显著正相关. 多重耐药类 和β-内酰胺类 ARGs 的潜在宿主最多.

关键词:近海环境;废水排放;抗生素抗性基因(ARGs);可移动遗传元件(MGEs);微生物群落 中图分类号: X171.5 文献标识码: A 文章编号: 0250-3301(2022)09-4616-09 DOI: 10.13227/j. hjkx. 202112272

Effects of Wastewater Discharge on Antibiotic Resistance Genes and Microbial Community in a Coastal Area

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Abstract: Wastewater treatment plants (WWTPs) are important sources of antibiotic resistance genes (ARGs) in aquatic environments. Mobile genetic elements (MGEs) and microbial communities are key factors that affect the proliferation of ARGs. To reveal the effects of WWTPs effluent discharge on the ARGs and microbial community in a coastal area, the structure and distribution of ARGs, MGEs, and microbial community in Shangyu (SY) and Jiaxing (JX) effluent receiving areas (ERAs) and the offshore area of Hangzhou Bay (HB) were investigated via high-throughput quantitative PCR and 16S rRNA high-throughput sequencing. The results showed that multidrug resistance genes were the most abundant ARGs across all the sampling sites. The diversity and abundance of ARGs and MGEs in the ERAs were much higher than those in the HB. Additionally, the diversities of the microbial community in the JX-ERA were higher than those in the SY-ERA and HB. PCoA showed that the distribution of ARGs, MGEs, and microbial communities in the ERAs and HB were significantly different, indicating that the long-term wastewater discharge could alter the distribution of ARGs, MGEs, and microbial communities in the coastal area. The co-occurrence pattern among ARGs, MGEs, and microbial communities revealed that 12 bacterial genera, such as Psychrobacter, Pseudomonas, Sulfitobacter, Pseudoalteromonas, and Bacillus, showed strong positive correlations with ARGs and MGEs. Most potential hosts carried multidrug and β -lactamase resistance genes.

Key words: coastal area; wastewater discharge; antibiotic resistance genes (ARGs); mobile genetic elements (MGEs); microbial community

近年来, 抗生素在医疗和养殖等领域的大量使 用甚至滥用,导致环境中的抗生素污染问题日益严 重[1]. 尤其是抗生素会诱导产生抗生素抗性基因 (antibiotic resistance genes, ARGs), 使其在各种环境 介质中广泛分布,大大增加了环境细菌获得耐药性 的可能性,从而使耐药细菌感染成为威胁人类健康 的全球性问题[2,3].

作为连接人类活动和自然环境的纽带,污水处 理厂接收了各种来源的污水,其丰富的营养物质和 高密度的细菌为 ARGs 的繁殖提供了适宜的环境, 因此污水处理厂被视作 ARGs 的储存库[4,5]. 然而现 有的污水处理技术对 ARGs 的去除作用有限,污水 处理厂出水中仍含有较高丰度的 ARGs[6,7], 是环境 中 ARGs 的重要来源[8]. 目前污水处理厂尾水通常 直接排放到河流和近海等水环境中,这一长期排放 行为对受纳水体中 ARGs 组成和分布的影响还不甚 清楚. ARGs 在环境中的传播扩散受到多种因素的 影响. 转座子、整合子和质粒等可移动遗传元件 (mobile genetic elements, MGEs)介导的水平基因转 移(horizontal gene transfer, HGT) 是 ARGs 在不同细

收稿日期: 2021-12-27; 修订日期: 2022-01-17 基金项目: 国家自然科学基金项目(51938001, 52170185)

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菌间广泛传播的主要途径^[9]. 此外,环境中 ARGs 的传播扩散还易受到抗生素、重金属、营养盐和微生物群落等因素的影响^[10,11]. 尤其是微生物群落之间复杂的相互关系,可能对 ARGs 的水平转移产生促进或抑制的作用,因此对微生物群落与 ARGs 共存关系的研究近年来备受关注^[12,13]. 有研究表明,废水排放能够引起受纳水体中微生物群落结构的显著变化^[14]. 为深入探究废水排放对受纳水体中 ARGs 的影响,进一步解析环境样本中 MGEs 和微生物群落的组成至关重要.

沿海地区人口密度高、工业发达,沿岸市政污水处理厂和工业废水处理厂众多,持续的废水排放活动使得近岸海域污染已成为全球范围内的环境问题.为探究废水排放对近海环境中 ARGs 和微生物群落的影响,本文选择我国受陆源废水排放负荷最重和污染形势最严峻的杭州湾地区为研究区域,采用高通量荧光定量 PCR (HT-qPCR)和高通量16S rRNA扩增子测序技术对杭州湾中典型近岸纳污区和远岸湾区中表层沉积物的 ARGs、MCEs 和微生物群落进行分析,并识别近海沉积物中 ARGs 的潜在宿主.

1 材料与方法

1.1 样品采集

选取我国东海杭州湾中两个典型的污水处理厂排放区及远离排放区的海湾作为研究区域. 采样点 SY1、SY2、JX1 和 JX2 位于上虞(SY)和嘉兴(JX)纳污区(effluent receiving areas, ERAs,图 1). 其中, SY1 和 JX1 位于污水处理厂排放口,长期接收污水厂尾水的直接干扰. 杭州湾的样点 HB1~HB6 与SY、JX 污水处理厂的距离分别为 32. 1~216. 9 km和 18. 0~209. 0 km,远离污水处理厂排放的干扰. 用抓泥斗(Van Veen, Germany)在各样点重复采集3份表层(0~5 cm)沉积物样品,存放于无菌塑料袋中,并立刻运回实验室,置于 −80℃冰箱中保存.

1.2 沉积物 DNA 提取

称取 0. 25 g 冻干后的沉积物,根据标准方法使用 Power Soil DNA 分离试剂盒(MoBio Laboratories, USA)提取沉积物样品 DNA. 并使用 NanoDrop 分光光度计(Thermo Fisher, USA)检测提取的 DNA 的浓度与纯度.

1.3 ARGs 和 MGEs 的高通量检测

使用 WaferGen SmartChip Real-time PCR (Warfergen Inc, USA)技术平台,对样品中 ARGs 和 MGEs 进行高通量 qPCR 分析^[15]. 使用 296 对引物检测目的基因,包括 285 对 ARGs、10 对 MGEs 和 1



Fig. 1 Location of sampling sites

对 16S rRNA 基因引物 [16]. HT-qPCR 体系包括 $1 \times \text{LightCycler}$ 480 SYBR Gree I Master, 500 nmol·L ⁻¹ 引物以及 2 ng· μ L ⁻¹ DNA 模板(100 nL). 反应程序为预变性 95°C, 10 min; 95°C, 30 s 和 60°C, 30 s (40 个循环); 最终采用熔融曲线估算扩增反应产物,扩增效率在 $1.8 \sim 2.2$ 之间 [17]. 检测阈值 (CT)设为 31. 如果目的基因在大于或等于两个重复样本中均被检出,且具有较小的偏差(<20%)和较好的拟合度,将被视为阳性检出. 根据 CT 值计算目的基因的相对丰度 [式(1)] [18]. 使用实时荧光定量 PCR (qPCR)测定 16S rRNA 的绝对丰度 [11]. 根据 16S rRNA 的绝对丰度与 Δ CT 值[式(2)] 计算 ARGs 和MGEs 的绝对丰度.

拷贝数 =
$$10^{(31-CT)/(10/3)}$$

基因相对丰度 = $\frac{$ 基因拷贝数(ARG)}{16S rRNA 拷贝数} (1)

 Δ CT = CT (目的基因) - CT (16S rRNA) 目的基因绝对丰度 = 16S rRNA 绝对丰度 $\times 2^{-\Delta CT}$ (2)

1.4 16S rRNA 基因的高通量测序

使用上游引物 515F (GTGCCAGCMGCCGCGGTAA) 和下游引物 806R (GGACTACHVGGGTWTCTAAT)对 16S rRNA 基因高变区(V4)进行扩增. 纯化后的扩增产物在 Illumina MiSeq 2500 平台(MAGIGENE, China)进行测序. 使用 Trimmomatic(v0.33, http://www.usadellab.org/cms/?page = trimmomatic)对原始序列进行质控,获得双端干净序列. 采用 FLASH 合并后(v1.2.11, https://ccb.jhu.edu/software/FLASH/),使用 UCHIME 对嵌合序列进行检查和过滤^[19]. 根据 97%的相似度,通过UPARSE 将获得的序列聚类为相同的操作分类单元(OTUs)^[20]. 最后,通过核糖体数据库(ribosomal

database poject, RDP)分类器(https://sourceforge.net/projects/rdp-classifier/)以 80% 阈值对 OTUs 进行分类^[21].

1.5 数据分析

ARGs、MGEs 和微生物群落分布柱状图以及ARGs 和 MGEs 相关性由 Excel 绘制. ARGs 和 MGEs 丰度热图由 R 语言(v1.4, https://www.r-project.org/)中的 phatmap 软件包完成. 使用 R 语言中的vegan 软件包计算微生物群落的 α 多样性指数(Chao1、ACE、Richness、Shannon、Simpson 和 PD指数). ARGs、MGEs 和微生物群落分布的主坐标分析(PCoA)由 R 语言中的vegan 软件包和OmicStudio 工具(https://www.omicstudio.cn/tool)共同完成. 使用 MENA(https://ieg4.rccc.ou.edu/mena/)和 Gephi (v0.9.2, https://gephi.org/)实现ARGs、MGEs 和细菌属的共现网络分析(cutoff = 0.8, P<0.05)^[22,23].

2 结果与讨论

2.1 近海环境中 ARGs 的多样性和组成

2.1.1 ARGs 的多样性

如图 2 所示, 所有沉积物样品共检出氨基糖苷类、β-内酰胺类、氯霉素类、磺胺类、四环素类、万古霉素类、MLSB 类(大环内酯类、林可胺类、链阳菌素类)、多重耐药类和其它类这 9 种 ARGs 类型.其中, SY 纳污区中 SY1 样点(检出 83 种 ARGs 亚型)的 ARGs 多样性最高, 说明长期的尾水排放可能导致环境中 ARGs 多样性增加.除了 HB4 (1 种ARGs 亚型)和 HB5 (9 种 ARGs 亚型)外,其他样点中 ARGs 多样性水平相似, 检出 ARGs 介于 34 (HB6)~42 (HB2)之间.

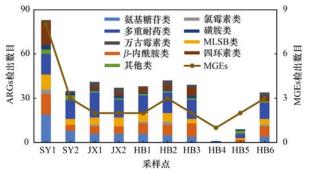


图 2 近海环境中 ARGs 和 MGEs 多样性分布

Fig. 2 Diversity of ARGs and MGEs in the coastal environment

2.1.2 ARGs 的丰度分布

沉积物中 ARGs 的相对丰度为 8.07 × 10^{-4} (HB5) ~ 5.96 × 10^{-2} (SY2) copies·cell $^{-1}$,绝对丰度为 4.52 × 10^{6} (HB5) ~ 1.23 × 10^{10} (JX1) copies· g^{-1} (图 3). SY 和 JX 纳污区沉积物中的

ARGs 丰度明显高于杭州湾沉积物,尤其是位于污水处理厂排放口的样点(SY1和 JX1). 说明长期的污水处理厂排放行为不仅增加了近岸环境中ARGs 的多样性,而且提高了ARGs 的丰度水平. 实际上,污水处理厂的处理工艺、污水来源和污泥絮凝及沉淀等特性是尾水释放 ARGs 的关键^[24,25]. 因此,以上特定因素可能是导致 SY 和 JX 纳污区ARGs 组成差异的原因.

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所有样点中,多重耐药类 ARGs 的多样性和丰度最高[平均值分别为 2. 24×10^{-2} copies·cell⁻¹ (相对丰度,下同)和 2. 03×10^9 copies·g⁻¹ (绝对丰度,下同)],其次是氨基糖苷类 (1. 93×10^{-3} copies·cell⁻¹和 2. 80×10^8 copies·g⁻¹)、MLSB类 (1. 27×10^{-3} copies·cell⁻¹和 1. 34×10^8 copies·g⁻¹)和 β -内酰胺类 ARGs (1. 13×10^{-3} copies·cell⁻¹和 1. 24×10^8 copies·g⁻¹).多重耐药类 ARGs 能够编码对多种抗生素的耐药性,因此对人类健康构成的威胁可能更大^[26,27].有研究表明,多重耐药细菌更多是由人为污染造成的^[28].环杭州湾地区频繁的人类活动可能是其环境中多重耐药类 ARGs 丰度水平高的原因,尤其是在近岸纳污区环境.

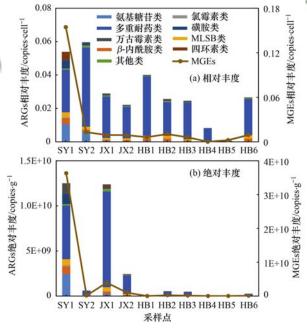


图 3 近海环境中 ARGs 和 MGEs 的丰度

Fig. 3 Abundances of ARGs and MGEs in the coastal environment

所有样点中共检出 122 种 ARGs 亚型,图 4 为杭州湾沉积物中占优势地位的 ARGs. 其中, mexF [平均值分别为 1. 66×10^{-2} copies·cell⁻¹ (相对丰度,下同)和 1. 52×10^{9} copies·g⁻¹ (绝对丰度,下同)]、pncA (3. 16×10^{-4} copies·cell⁻¹和 4. 08×10^{7} copies·g⁻¹)、qacEdeltal- 01 (3. 92×10^{-4} copies·cell⁻¹和 1. 55×10^{8} copies·g⁻¹)、sul2 (1. 18

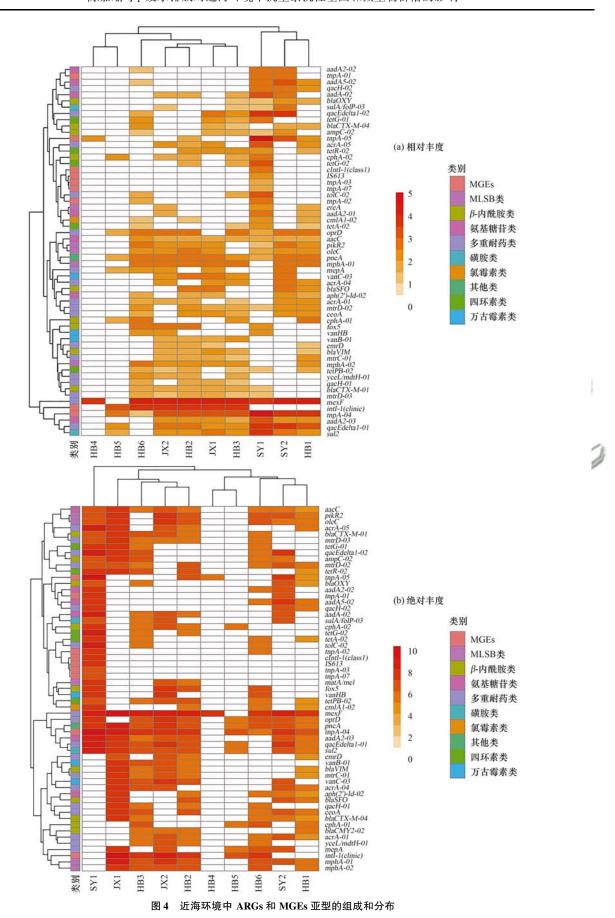


Fig. 4 Compositions and distributions of ARGs and MGEs in the coastal environment

 $\times 10^{-4}$ copies • cell $^{-1}$ 和 1. 28 $\times 10^{8}$ copies • g $^{-1}$) 、 oprD (3. 17 $\times 10^{-4}$ copies • cell $^{-1}$ 和 9. 73 $\times 10^{6}$ copies • g $^{-1}$)

和 mphA-01 (4. 52×10^{-4} copies·cell⁻¹和 3. 59×10^{7} copies·g⁻¹)的丰度最高,比其他 ARGs 亚型高 2~4

个数量级. 这些 ARGs 亚型在垃圾填埋场、养鸡场和污水处理系统等不同的人为环境被频繁检出^[29~31].

2.2 近海环境中 MGEs 的组成和分布

在所有样品中共检测到9个MGEs,包括7个转 座子(tnpA-01、tnpA-02、tnpA-03、tnpA-04、tnpA-05、 tnpA-06、tnpA-07) 和 2 个整合子(cIntI-1、intI-1) (图 2 和图 3). 其中, tnpA-05 「平均值分别为 1.02 ×10⁻²copies·cell⁻¹(相对丰度,下同)和 2.42×10⁹ copies·g⁻¹ (绝对丰度,下同)]的丰度最高, tnpA-07 的丰度最低. 与 ARGs 分布类似, MGEs 在纳污区中 的丰度高于远岸湾区(图3). 尤其是 SY1 样点, MGEs 总丰度 (1.54 × 10⁻¹ copies · cell ⁻¹ 和 3.63 × 10¹⁰copies·g⁻¹)高于其他取样点 1~4 个数量级. 这 表明污水处理厂也是 MGEs 的热点,且能够将 MGEs 释放到环境中[32,33]. 更重要的是, 纳污区和远岸湾 区沉积物中的 ARGs 与 MGEs 呈显著正相关(图 5). 高丰度的 MGEs 能够驱动 ARGs 在微生物群落 中发生水平转移,从而加快了 ARGs 在环境中的传 播扩散[34].

2.3 近海环境中的微生物群落分析

2.3.1 微生物群落的物种多样性

对所有样品的序列进行聚类,共得到19931个

OTUs. 各样点的覆盖度均为 0.99, 表明测序结果能够较好地反映沉积物样品中微生物群落的真实情况 (表 1). 各个多样性指数表明 JX 纳污区表层沉积物中微生物群落的物种丰富度和多样性均高于 SY 纳污区和远岸湾区, 尾水排放可能释放了大量人为源的微生物, 从而影响其受纳环境中微生物群落 [35]. 值得注意的是, 杭州湾沉积物中微生物群落物种丰富度与多样性沿湾内至湾外逐渐降低. 这表明环杭州湾区域废水排放等人类活动剧烈, 杭州湾海水水质一直处于富营养状态 [36]. 这为微生物生长提供了营养和能量, 从而提升杭州湾沉积物中微生物群落的多样性 [37,38].

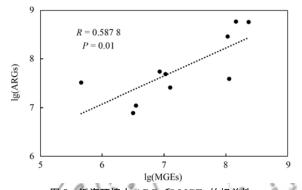


图 5 近海环境中 ARGs 和 MGEs 的相关性

Fig. 5 | Correlation between ARGs and MGEs

in the coastal environment

表 1 微生物群落 α 多样性指数

Table 1 The α-diversity of microbial community

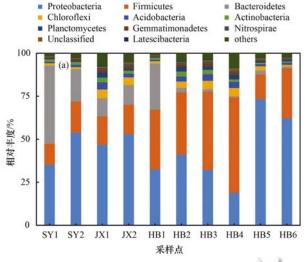
样点	覆盖度	Chao1 指数	ACE 指数	Richness 指数	Shannon 指数	Simpson 指数	PD 指数
SY1	0.99 ± 0	5 652. 20 ± 144. 3	5 900. 17 ± 147. 64	4 268.33 ± 93.70	5.45 ± 0.15	0.89 ± 0.01	213.34 ± 6.20
SY2	0.99 ± 0	$5\ 073.61 \pm 496.3$	5 221.04 ±458.95	$4\ 005.\ 67\ \pm 548.\ 22$	5.75 ± 0.25	0.93 ± 0.01	204.08 ± 21.61
JX1	0.99 ± 0	$7\ 864.65 \pm 359.3$	8 151.34 ± 395.05	6296.00 ± 303.71	7.66 ± 0.45	0.96 ± 0.01	291.41 ± 10.42
JX2	0.99 ± 0	6441.80 ± 511.5	6642.57 ± 489.50	$5\ 160.00\pm604.19$	6.32 ± 0.88	0.9 ± 0.06	249.32 ± 22.39
HB1	0.99 ± 0	3646.03 ± 372.2	3736.11 ± 410.42	2810.33 ± 220.35	4.67 ± 0.21	0.89 ± 0.03	150.47 ± 9.86
HB2	0.99 ± 0	$7\ 047.92 \pm 268.49$	$7\ 201.57\ \pm 280.17$	$5\ 562.\ 67\ \pm 135.\ 19$	5.88 ± 0.10	0.85 ± 0.02	257.91 ± 5.03
HB3	0.99 ± 0	6332.22 ± 80.04	6557.14 ± 42.28	4944.67 ± 97.41	5.43 ± 0.13	0.79 ± 0.04	233.66 ± 5.76
HB4	0.99 ± 0	$5\ 036.\ 23\pm 510.\ 56$	$5\ 032.75 \pm 536.91$	4029.33 ± 388.19	5.85 ± 0.28	0.86 ± 0.02	201.67 ± 15.18
HB5	0.99 ± 0	$4\ 315.\ 37\ \pm 429.\ 66$	$4\ 446.\ 46\ \pm 425.\ 22$	$3\ 225.\ 33\ \pm398.\ 19$	4.06 ± 0.83	0.76 ± 0.08	168.51 ± 14.65
HB6	0.99 ± 0	4 763.81 ± 474.97	4 933. 97 ± 483. 58	$3\ 547.00 \pm 434.61$	4.01 ± 0.85	0.72 ± 0.11	173.68 ± 16.71

2.3.2 微生物群落的物种组成

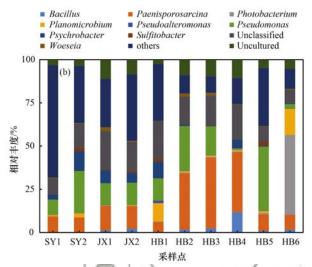
本研究共识别到细菌的 76 个门、214 个纲、471 个目、971 个科和1 102个属. 门水平上, 11 个细菌门的相对丰度均超过 0. 1% [图 6(a)]. 绝对优势菌门包括变形菌门(Proteobacteria, 44. 9%)、厚壁菌门(Firmicutes, 27. 9%)、拟杆菌门(Bacteroidetes, 12. 1%)、绿弯菌门(Chloroflexi, 3. 0%)、酸杆菌门(Acidobacteria, 2. 2%)、放线菌门(Actinobacteria, 1. 7%)和浮霉菌门(Planctomycetes, 1. 6%),相对丰度均超过 1%. 之前的研究发现,变形菌门和拟杆菌

门等细菌在污水处理厂污泥中高度富集^[39]. SY、JX 纳污区样点高丰度的拟杆菌门(21.5%)有可能是长期废水排放行为导致的. 而厚壁菌门在杭州湾远岸湾区沉积物中所占比例(35.9%)大于近岸纳污区沉积物(16%). 这与 Dai 等^[40]对杭州湾沉积物微生物群落的研究结果—致,厚壁菌门在杭州湾沉积物中占有很大的比例.

属水平上[图 6(b)],共有 13 个菌属的丰度在 所有样点中高于 0.1%,其余 201 个菌属的丰度均小 于 0.1%,被归类为其它类菌属.其中,优势菌属(相 对丰度大于 1%)为: 芽孢杆菌属(Paenisporosarcina, 17.6%)、假单胞菌属(Pseudomonas, 15.4%)和嗜冷杆菌属(Psychrobacter, 4.7%). 芽孢杆菌属和假单胞菌属在杭州湾沉积物(22.0%和15.9%)中所



占比例高于 SY 和 JX 纳污区 (11.1% 和 15.2%). 而 201 个其它类菌属(相对丰度小于 0.1%)在近岸 纳污区沉积物(41.1%)中所占比例远远高于杭州 湾远岸湾区(18.8%).



(a)门水平;(b)属水平 图 **6** 微生物群落相对丰度

Fig. 6 Relative abundance of microbial community

2.4 杭州湾近岸纳污区与远岸湾区中 ARGs MCEs 和微生物群落的分布差异

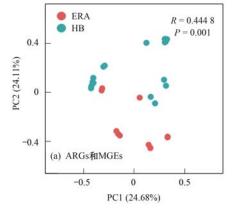
如图 7 (a) 所示, SY 和 JX 纳污区样点(SY1、SY2、JX1 和 JX2) 与远岸湾区(HB) 样点(HB1 ~ HB6)分别聚集,说明杭州湾近岸纳污区与远岸湾区沉积物中 ARGs 和 MGEs 分布差异显著(R = 0.444 8, P = 0.001).有研究发现污水处理厂能够影响环境中 ARGs 的分布. Marti 等[41] 通过对污水处理厂下游河流中 ARGs 分布的调查发现,污水处理厂排放行为可能将 ARGs 释放到环境. Su 等[42] 发现废水排放还能够驱动 ARGs 在近海环境中的传播扩散.

此外,SY和JX纳污区与远岸湾区沉积物中微生物群落分布与其ARGs和MGEs分布模式相似

[图7(b)],即纳污区沉积物与杭州湾沉积物中微生物群落具有显著差异(R=0.2984, P=0.001).有研究发现,废水排放对微生物群落有直接或间接的影响^[43].一方面,污水处理厂可以直接将人为源的微生物释放到环境中^[44].另一方面,污水处理厂通过输入有机或无机污染物改变微生物群落的生存环境^[45].因此,污水处理厂的排放行为不仅对近海环境 ARGs 和 MGEs 的分布有显著影响,而且对环境中微生物群落也能产生作用.

2.5 ARGs、MGEs 和微生物群落的相关性

在环境中, ARGs 可被某些细菌所携带, 因此ARGs 的分布在一定程度上会受微生物群落的影响^[46]. ARGs、MGEs 和细菌属的共现网络如图 8 所示. 其中, 与 ARGs 或 MGEs 呈强烈正相关的细菌被



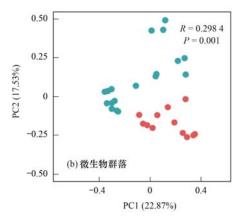


图 7 基于 Bray-Curtis 距离的主坐标分析 (PCoA)

Fig. 7 Principal coordinate analysis (PCoA) based on the Bray-Curtis distance

认为是其潜在宿主. SY 和 JX 纳污区及远岸湾区沉 积物中共识别出 12 种菌属与 34 个 ARGs、MGEs 存 在显著正相关. 例如,其它类菌属(others)与 aacC4、 cphA-02 和 tetA-02 等 14 种 ARGs 亚型呈显著正相 关; 嗜冷杆菌属(Psychrobacter)与 aacC4、acrA-01 和 sulA 等 8 种 ARGs 亚型呈显著正相关; 假单胞菌 属(Pseudomonas)与 aadA5-02、blaOXY 和 emrD 等 7 种 ARGs 亚型呈显著正相关; 亚硫酸杆菌属 (Sulfitobacter)与 aacC4、blaOXY 和 sulA 等 6 种 ARGs 亚型呈显著正相关: 假交替单胞菌属 (Pseudoalteromonas)与 blaCMY2-02 和 vanB-01 等 5 种 ARGs 亚型呈显著正相关. 这些潜在的 ARGs 宿 主,如嗜冷杆菌属和假交替单胞菌属等菌属被证明 与 sul2 等其他 ARGs 亚型也存在显著正相关[47,48]. 而假单胞菌属宿主是潜在致病菌,进一步增加了近 海环境中 ARGs 传播扩散的健康风险^[49]. 9 种 ARGs 类型中,多重耐药类和 β -内酰胺类 ARGs 的潜 在宿主最多,这可能是这两类 ARGs 在杭州湾沉积 物中广泛分布的原因之一. 而 MGEs 的潜在宿主远 远少于 ARGs, 仅有 tmpA-05 与其它类菌属存在显著 正相关.

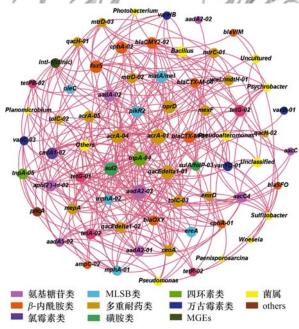


图 8 ARGs、MGEs 和细菌属的共现网络分析
Fig. 8 Co-occurrence network among ARGs, MGEs,
and microbial community

3 结论

- (1) 多重耐药类 ARGs 是杭州湾沉积物中多样性和丰度最高的 ARGs 类型,其次是氨基糖苷类、MLSB 类和 β-内酰胺类 ARGs. 芽孢杆菌属、假单胞菌属和嗜冷杆菌属是杭州湾沉积物中的优势菌属.
 - (2) 污水处理厂长期的排放行为可能影响了近

海环境中 ARGs、MGEs 和微生物群落的组成:纳污区沉积物中 ARGs 和 MGEs 的多样性和丰度远远高于远岸湾区中不受排水干扰的沉积物,尤其是位于废水排放口的样点(SYI 和 JXI);纳污区与远岸湾区沉积物中 ARGs、MGEs 和微生物群落分布存在显著差异.

- (3) MGEs 和微生物群落是 ARGs 丰度和组成的主要影响因素: MGEs 丰度和 ARGs 丰度呈显著正相关,12 种菌属与 ARGs、MGEs 呈显著正相关.多重耐药类、β-内酰胺类 ARGs 和细菌属的关联最多.参考文献:
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