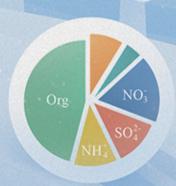


**ENVIRONMENTAL SCIENCE** 

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北京冬季奥运会历史同期大气PM<sub>2.5</sub>污染特征分析 刘玥晨,满睿琪,裘彦挺,杨佳炜,王均睿,谭瑞,汤丽姿,俞颖,宋锴,郭松,陈仕意, 曾立民,吴志军,胡敏







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## 长期秸秆还田对水稻根系碳矿化与激发效应的影响

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摘要:长期秸秆还田改变水稻土环境条件,从而会影响水稻残留根系碳在土壤中的矿化和激发效应,其影响的方向和强度尚不明确.因此,基于长期定位施肥试验,采用<sup>13</sup>C-CO<sub>2</sub> 同位素标记技术,结合室内模拟培养试验,研究不施肥(CK)、单施化肥(CF)和秸秆还田配施化肥(CFS)这3种长期处理下,水稻根系和土壤本身有机碳的矿化特征,分析根系的激发效应的强度与方向,以及各处理 CO<sub>2</sub> 释放量的来源组成.结果表明,经过120d的室内淹水培养,根系残留(R)将 CO<sub>2</sub> 累积释放量增加了617.41~726.27 mg·kg<sup>-1</sup>. CFS+R和CF+R处理根系来源 CO<sub>2</sub> 累积释放量分别为470.82 mg·kg<sup>-1</sup>和444.04 mg·kg<sup>-1</sup>,根系的矿化率分别为18.8%和17.8%,均显著高于CK+R处理(384.19 mg·kg<sup>-1</sup>,15.4%).3个处理的土壤本身有机碳产生的CO<sub>2</sub> 累积排放量无显著差异,但CFS+R处理的土壤本身有机碳矿化率(4.2%)显著低于CF+R和CK+R处理(5.4%和5.8%).CFS+R处理中根系的CO<sub>2</sub> 累积激发效应为29.6%,显著低于CK+R处理的42.5%,高于CF+R处理的14.4%.淹水水稻土CO<sub>2</sub> 累积释放量中23.47%~27.59%来源于根系,其余来源于土壤,其中,激发效应引起的CO<sub>2</sub> 释放量占比在CFS+R处理中比CK+R处理小,比CF+R处理大.综上,淹水水稻土长期秸秆还田会提高根系碳的矿化潜力,但是更有利于土壤本身有机碳的稳定.

关键词: 秸秆还田; 水稻土; 13C 同位素标记; 根系矿化; 激发效应

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# Effect of Long-term Straw Returning on the Mineralization and Priming Effect of Rice Root-carbon

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Abstract: Long-term straw returning to the field changes the environmental conditions of rice paddy soil, which affects the mineralization and priming effect of residual rice roots in the soil, but the direction and intensity of its influence is not clear. Therefore, based on a long-term fertilization field experiment,  $^{13}$ C-CO<sub>2</sub> isotopic labeling technology and laboratorial incubation were used to analyze the characteristics of mineralization of rice roots and native soil organic carbon, the intensity and direction of the priming effect, and the source partitioning of CO<sub>2</sub> emissions in three treatments, consisting of no fertilization (CK), chemical fertilizer (CF), and straw returning with chemical fertilizer (CFS). The results showed that after 120 days of flooding incubation, the root residue (R) increased the cumulative CO<sub>2</sub> emissions by 617.41-726.27 mg·kg<sup>-1</sup>. The cumulative CO<sub>2</sub> emissions from roots and root mineralized proportions in the CFS + R and CF + R treatments were 470.82 and 444.04 mg·kg<sup>-1</sup>, respectively, and 18.8% and 17.8%, respectively. These were significantly higher than those in the CK + R treatment (384.19 mg·kg<sup>-1</sup>, 15.4%). There was no significant difference in the cumulative CO<sub>2</sub> emissions from native soil organic carbon among the three treatments. However, the mineralized proportion of native soil organic carbon in the CFS + R treatment (4.2%) was significantly lower than that in the CF + R and CK + R treatments (5.4% and 5.8%). The priming effect in the CFS + R treatment was 29.6%, which was significantly lower than that in the CK + R treatment (42.5%) and higher than that in the CF + R treatment (14.4%). A total of 23.47% to 27.59% of the cumulative CO<sub>2</sub> emission of the flooded paddy soil was from the roots, and the remainder was from the soil. In addition, the proportion of CO<sub>2</sub> emission caused by the priming effect was smaller in the CFS + R treatment than that in the CK + R treatment and larger than that in the CF + R treatment. In summary, the long-term straw returni

Key words: straw returning; paddy soil; 13C isotope labeling; root mineralization; priming effect

稻田在我国粮食安全中占有举足轻重的地位, 总面积达3 400万 hm²,主要分布在亚热带地区,该 区稻田占全国总面积的 31.4%<sup>[1,2]</sup>.稻田土壤具有 明显的碳汇功能,开展稻田的固碳和温室气体排放 研究,对减缓温室气体排放和实现"双碳"计划具有 重要意义<sup>[3,4]</sup>.

长期秸秆还田可以有效地提高土壤的养分和有机碳含量,改变土壤物理、化学和生物环境,从而影响水稻土有机碳的周转和积累<sup>[5,6]</sup>,因此在我国得

到大力推广,且卓有成效. 2000~2010年间我国水稻的秸秆还田比例由不足 25%增长到近40%<sup>[7]</sup>. Wang等<sup>[8]</sup>的研究对长期(>10 a)秸秆还田试验的荟萃分析显示,水稻秸秆还田的碳固定

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效率为9.4%,长期水稻秸秆还田可将土壤有机碳(SOC)含量提高约12%. Wang等<sup>[9]</sup>的荟萃分析也显示长期水稻秸秆还田可将土壤有机碳含量提高5.4%.秸秆还田不仅对土壤有机碳库有所贡献,还会起到改良土壤物理结构,改善土壤通透性和蓄水能力的作用,其分解还会释放氮磷等养分元素<sup>[5,6]</sup>. 然而,目前秸秆还田的相关研究多集中在其对土壤有机碳库库容、养分含量以及其他物理、化学和生物条件等造成的影响,以及秸秆在土壤中的分解、矿化和转化过程,缺乏对长期秸秆还田情况下其他外源有机物进入土壤后的矿化和转化过程的研究.

水稻根系是稻田土壤有机碳的重要来源之 一[10,11],水稻根系在土壤中的分解相当缓慢,与其 他新鲜碳源如秸秆和叶片等相比,对土壤有机碳的 贡献可能更大[12,13]. 有研究显示,约8.7%~20%水 稻根系碳会转化为土壤有机碳[13,14]. 水稻根系进入 土壤中不仅可以补充土壤有机碳库,还会增加土壤 本身有机碳的矿化,即产生正激发效应[14,15],且小 于水稻地上部引起的激发效应[16]. 有机物在土壤中 的分解和矿化受土壤物理、化学和生物等环境因素 的影响,如土壤肥力、水分条件和土壤质地 等[15,17,18]. 然而,目前关于外源有机物在土壤中的 矿化研究多集中于秸秆和有机肥等有机物料或者单 一小分子有机物,针对水稻根系残留的研究极其缺 乏[10],尤其长期秸秆还田改造下的土壤中根系矿化 过程的研究很少,鉴于水稻根系碳在土壤有机碳中 的重要贡献,其分解和矿化过程及其影响因素等方 面亟待深入研究.

因此,本研究依托于湖南省宁乡县的长期定位 试验,通过室内培养试验和<sup>13</sup>C 稳定同位素标记示踪 技术,探讨和量化了长期秸秆还田对水稻根系矿化 特征及其激发效应方向和强度的影响,以期为更好 地了解稻田生态系统碳循环以及为稻田固碳减排提 供理论指导.

### 1 材料与方法

### 1.1 供试土壤

供试土壤采自湖南省宁乡县长期定位试验站 (112°18′E, 28°07′N), 种植制度为早稻-晚稻-冬闲 模式,试验始于1986年,该地区属于典型的亚热带 季风湿润气候,平均海拔 36.1 m,年均气温为 16.8℃,每年蒸发量均为1353.9 mm,每年无霜期 274 d 左右, 年均降雨量是1 553.7 mm. 采样点土壤 氮、磷和钾供应状况根据作物预期需求量. 本试验 选取田间3种长期不同施肥处理的土壤:①无施肥 处理(CK); ②化肥处理(CF), 早稻施肥量为142.5  $kg \cdot hm^{-2}$ (以N计), 54.0  $kg \cdot hm^{-2}$ (以 $P_2O_5$ 计), 63.0 kg·hm<sup>-2</sup>(以 K<sub>2</sub>O 计),晚稻施肥量为 157.5 kg·hm $^{-2}$ (以N)), 43.2 kg·hm $^{-2}$ (以 $P_2O_5$ ), 81.0 kg·hm<sup>-2</sup>(以 K<sub>2</sub>O 计); ③化肥+秸秆还田处 理(CFS),早稻秸秆的还田量为2775 kg·hm<sup>-2</sup>,晚稻 的还田量为3 600 kg·hm<sup>-2</sup>,以 CF 处理的施氮量为 标准,不足用化肥补足. 采集耕作层(0~20 cm)土 壤样品,人工挑出石块和肉眼可见动植物残体和根 系等杂物,过10目筛,充分混匀.将混匀的土壤样品 按照四分法取出一部分,经风干过筛后测定土壤样 品的基本理化性质(表1).

表 1 供试土壤基本理化性质1)

Table 1 Physiochemical properties of the soil

				FF				
处理	рН	ω[SOC(有机碳)] /g·kg <sup>-1</sup>	ω[TN(全氮)] /g·kg <sup>-1</sup>	ω[ AN(碱解氮)] /mg·kg <sup>-1</sup> )	ω[TP(全磷)] /g·kg <sup>-1</sup>	ω(黏粒) /%	ω(粉粒) /%	ω(砂粒) /%
不施肥(CK)	6. 26a	21. 87b	1.61c	161. 4b	0. 62c	8. 80b	74. 09a	17. 11c
全化肥(CF)	6. 59a	22. 70b	1. 66b	180. 95b	0.70b	10.09b	69. 02b	20. 89b
秸秆还田(CFS)	5.85a	32. 34a	2. 46a	250. 43a	1. 72a	7. 25 c	63. 38c	29. 37a

<sup>1)</sup>不同小写字母表示不同处理在 P < 0.05 水平上差异显著

### **1.2** <sup>13</sup>C 标记根系制备

在 PVC 盆中(直径 17.2 cm,高度 16.7 cm)移 栽 25 日龄的水稻幼苗,每盆 3 株,置于密闭植物生长箱(长×宽×高:150 cm×100 cm×150 cm),连续不断地输入 CO<sub>2</sub> 维持含量 400 μL·L<sup>-1</sup>. 在水稻营养生长期,包括水稻分蘖期和伸长期,对水稻连续输入<sup>13</sup>CO<sub>2</sub>(<sup>13</sup>C丰度 2.8%)进行标记,并在孕穗期开始时停止标记. 将生长箱置于稻田中,使标记植物和未标记对照物的环境条件相同. 于成熟期进行破坏性

采样,将水稻根系与土壤分离. 根系用水反复冲洗去除附着的土壤. 根系于 105 公 杀青 0.5 h 后,置于 65 化的烘箱中烘干并粉碎保存. 最终制得根系的 $^{13}$ C 丰度为 2.152%,含碳量为 37.75%.

### 1.3 土壤培养试验

对不同施肥土壤分别设置添加<sup>13</sup>C标记根系(CK+R、CF+R和CFS+R)与不加根(CK、CF和CFS)处理,共6个处理,3个重复,另外设置添加未标记根系的自然丰度对照组.将风干土淹水

至水层深约 2~3 cm,置于 25℃ 的恒温培养暗室进行 10 d 预培养. 预培养结束后,将上清液倒掉,将土样与 $^{12}$ C标记根系按 2.5 g·kg $^{-1}$ (干土计)的碳添加量均匀混合,称 20 g(干土计)加入到 500 mL 的培养瓶中,淹水至水层约为 2~3 cm,盖子松开保持通气状态,置于 25℃的恒温培养室中遮光培养120 d.

### 1.4 样品采集与分析

每隔 1~5 d 进行气体采集,用一次性针管注射器抽取培养瓶顶空气体 35 mL,注入已抽真空的 12 mL 贮藏玻璃瓶待测. 每次采气结束后,通过气泵将瓶中气体置换为新鲜空气,密封后放回 25℃恒温室继续培养. 气体 CO₂含量用气相色谱法-温室气体气相色谱仪(7890A,美国)测定; ¹³CO₂用稳定性同位素质谱仪(MAT 253,美国)测定.

### 1.5 数据处理

根系和土壤本身有机碳(SOC)的 CO<sub>2</sub> 释放量计算公式:

$$A_{\text{root}} = \frac{\delta^{13} C_{\text{SOC}} - \delta^{13} C_{\text{total}}}{\delta^{13} C_{\text{SOC}} - \delta^{13} C_{\text{root}}} \times A_{\text{total}}$$

$$A_{\text{SOC}} = A_{\text{total}} - A_{\text{root}}$$
(1)

式中, $A_{\text{total}}$ 为各处理的  $\text{CO}_2$ -C 总释放量( $\text{mg·kg}^{-1}$ );  $\delta^{13}\text{C}_{\text{total}}$ 为总释放量中 $^{13}\text{C}$  丰度;  $A_{\text{SOC}}$ 为来源于 SOC 的  $\text{CO}_2$  释放量;  $\delta^{13}\text{C}_{\text{SOC}}$ 为 SOC 的 $^{13}\text{C}$  丰度(自然丰度对照);  $\delta^{13}\text{C}_{\text{root}}$ 为根系的 $^{13}\text{C}$  丰度;  $A_{\text{root}}$ 为根系的  $\text{CO}_2$  释放量.根系的矿化率是根系  $\text{CO}_2$  累积释放量与根系添加碳量的比值(质量分数).

激发效应和相对激发效应计算公式:

$$PE = A_{SOC} - A_{Nonroot}$$
 (3)

 $RPE = (A_{SOC} - A_{Nonroot})/A_{Nonroot} \times 100\%$  (4) 式中, $A_{Nonroot}$ 为不加根处理的  $CO_2$  累积释放量.

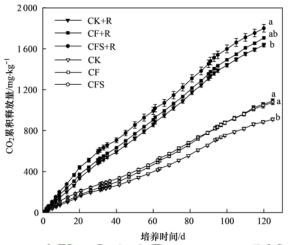
所有数据计算均为 3 次重复的平均值,运用 Microsoft Excel 2010 进行整理,用 Origin 2018 绘图 分析.用 SPSS 21.0 进行不同处理间的单因素方差 分析(One-way ANOVA),显著性水平为P < 0.05,并用 Tukey 法进行多重比较.

#### 2 结果与分析

### 2.1 水稻土 CO<sub>2</sub> 累积释放特征

在 0~120 d 培养期间,不同施肥处理土壤前期 CO<sub>2</sub> 累积释放量快速增长,之后增速逐渐减缓(图 1).添加根系土壤 CO<sub>2</sub> 的累积释放量显著高于不添加根系的土壤, 120 d 时相较于 CK、CF 和 CFS 处理,加根处理累积矿化量分别增加了 726.3、617.4 和 724.5 mg·kg<sup>-1</sup>.不添加根系的土壤中,培养前期 (0~30 d) CFS 处理的 CO<sub>2</sub> 累积释放量显著高于 CK

和 CF 处理,之后 CFS 和 CF 处理间 CO<sub>2</sub> 累积释放量的差距逐渐减小至无显著差异,在 120 d 时 CF 和 CFS 处理的 CO<sub>2</sub> 累积释放量分别比 CK 处理增加了 19.6%和 18.2% (P<0.05).添加根系的土壤中,培养期间 CFS+R和 CF+R处理的 CO<sub>2</sub> 累积释放量高于 CK+R处理,在培养结束时, CFS+R、CF+R和 CK+R处理的 CO<sub>2</sub> 累积释放量分别为1800.3、1706.8和1636.9 mg·kg<sup>-1</sup>.



CK:不施肥, CF: 化肥, CF: 秸秆还田配施化肥, CK+R: 不施肥+根系, CF+R: 化肥+根系, CFS+R: 秸秆还田配施化肥+根系; 不同小写字母表示 120 d 各不加根处理间或各加根处理间差异的显著性(P<0.05), 下同

## 图 1 不同长期施肥处理水稻土添加和不添加根系 条件下 $CO_2$ 累积释放量

Fig. 1 Cumulative  $\mathrm{CO}_2$  emission with and without root addition under different fertilization

### 2.2 水稻根系和土壤本身有机碳的矿化规律

相较于不施肥处理,单施化肥和添加秸秆处理中根系矿化生成  $CO_2$  的比例明显增加(图 2). 培养前 40 d, CFS + R 处理的根系  $CO_2$  累积释放量显著高于 CK + R 和 CF + R 处理, 之后 3 个处理的根系

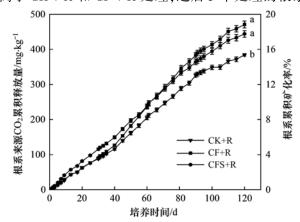


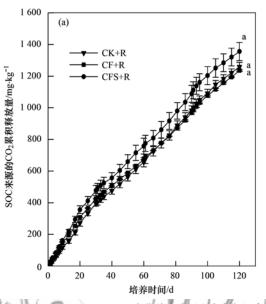
图 2 不同长期施肥水稻土根系来源的  ${
m CO}_2$  累积释放量及其矿化率

Fig. 2 Cumulative CO<sub>2</sub> emission from root and mineralized proportion of root under different long-term fertilization

 $CO_2$  释放速率同时加快,且 CF + R 处理的  $CO_2$  释放速率最快,其根系  $CO_2$  累积释放量与 CFS + R 处理的差距逐渐缩小,并在 76 d 以后逐渐超过 CFS + R 处理的  $CO_2$  累积释放量,但是差异不显著. 在 90 ~ 120 d 时各处理根系  $CO_2$  释放速率变缓. 在培养 120 d 结束时,CF + R、CFS + R 和 CK + R 处理中根系  $CO_2$  累积释放量分别为 470.8、444.0 和 384.2  $coldsymbol{mg·kg}^{-1}$ , CF + R 处理和 CFS + R 处理根系  $coldsymbol{CO}_2$  累积释放量显著高于 CK + R,三者根系的矿化率分别

为 18.8%、17.8% 和 15.4%.

SOC 矿化生成的  $CO_2$  累积释放量在前期  $0 \sim 20$  d 内快速升高,后期释放速率逐渐变缓. CFS + R 处理的 SOC 来源  $CO_2$  累积释放量略高于 CK + R 和 CF + R 处理,但差异不显著[图 3(a)]. 但因 CFS + R 土壤的 SOC 含量较高,其 SOC 的累积矿化率显著低于 CK + R 和 CF + R 处理. 120 d 时, CK + R 、CF + R 和 CFS + R 处理的 SOC 累积矿化率分别为 5.8%、5.4% 和 4.2% [图 3(b)].



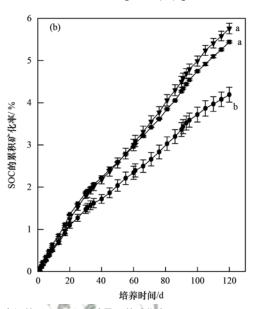


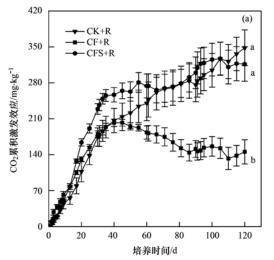
图 3 不同长期施肥水稻土本身有机碳(SOC)来源的 CO<sub>2</sub> 累积释放量及其矿化率

Fig. 3 Cumulative CO<sub>2</sub> emission and mineralized proportion of soil organic carbon (SOC) under different long-term fertilization

### 2.3 水稻根系的激发效应

水稻根系导致土壤  $CO_2$  累积释放的正激发效应. 培养前期(0~40 d), 3 个处理的正  $CO_2$  累积激发效应迅速增长,且 CFS+R 处理最快, 40 d 时 CK+R、CF+R 和 CFS+R 处理的  $CO_2$  累积激发效应分别为 234. 2、184. 9 和 256. 8  $mg \cdot kg^{-1}$ . 随后 CK+

R 和 CFS + R 处理的 CO<sub>2</sub> 累积激发效应增速减缓, CF + R 处理的正 CO<sub>2</sub> 累积激发效应逐渐降低. 在 120 d 培养结束后, CK + R、CF + R 和 CFS + R 处理的 CO<sub>2</sub> 累积激发效应分别为 386. 8、156. 2 和 316. 5 mg·kg<sup>-1</sup>[图 4(a)]. 水稻根系导致的 CO<sub>2</sub> 相对累积激发效应[图 4(a)]与 CO<sub>2</sub> 累积激发效应[图 4



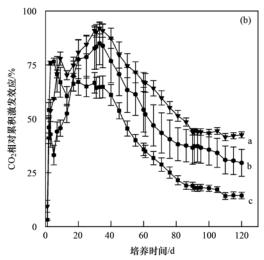


图 4 不同长期施肥水稻土中根系的  $CO_2$  累积激发效应和  $CO_2$  相对累积激发效应

Fig. 4 Cumulative and relative cumulative priming effects of roots in paddy soils with different long-term fertilization

(b)]不同, 3 个处理的  $CO_2$  相对累积激发效应 0 ~ 40 d 迅速增长,之后均逐渐下降,且 40 ~ 120 d 呈现 出: CK + R > CFS + R > CF + R,在 120 d 时, 3 种处理的  $CO_2$  相对累积激发效应分别为 42.5%、29.6% 和 14.4%.

### 2.4 添加根系水稻土 CO, 排放源分布特征

培养结束时, CO<sub>2</sub> 累积排放量 23.5%~27.6% 来源于根系, 72.4%~76.5% 来源于 SOC, 其中, 因

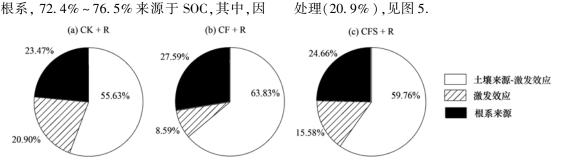


图 5 120 d 添加根系水稻土 CO<sub>2</sub> 排放来源与比例

Fig. 5 Partitioning of C sources for CO<sub>2</sub> emissions in paddy soil with added roots at 120 d

### 3 讨论

### 3.1 水稻根系的矿化和激发效应特征

农业土壤中土壤有机碳的积累对维持土壤肥力 和作物生产力至关重要,水稻根系是稻田土壤有机 碳碳库中重要的输入碳源[11]. Liu 等[11]总结了连 续标记和脉冲标记试验结果,发现水稻光合碳约 10%~13%分配在根系之中. 其中 8.7% [13] 或 20% [14] 水稻根系碳会转化为土壤有机碳,取决于土 壤水分条件. 根系中木质素和软木脂等难分解有机 碳组分的含量较高,因此根系的矿化率普遍低于秸 秆或叶片等作物其他部位的残留物[10,12,19,20]. 本文 经过120 d 淹水培养, 15.4%~18.8%的根系碳矿 化为 CO<sub>2</sub>(图 2). 类似地, Kramer 等[21] 经过 32 d 培 养,玉米根系矿化率约为12%,小于玉米叶片矿化 率(45%). Viswanath 等[22] 在保持田间持水量的情 况下,室内培养150 d后,水稻秸秆和小麦秸秆的矿 化率分别为 81% 和 73%, 均高于本文和 Kramer 等[21]的根系矿化率,虽然矿化率一定程度上与培养 时间和土壤水分状况有关,但仍能显示出根系矿化 率普遍低于秸秆矿化率. Zhu 等[16]的研究中,虽然 水稻根系矿化率为45.4%,高于秸秆31.9%,但是其 根系碳的平均滯留时间大于秸秆碳. 刘晖等[23]的室 外培养试验显示,在培养5个月时,水稻根系累积矿 化率低于秸秆碳,培养1a时,水稻根系和秸秆碳累 积矿化率相近. 根系产生 CO2 在总释放量中的占比 也小于水稻秸秆的 CO, 占比. 本文根系产生 CO, 累 积释放量不足总释放量的 30% (图 5),这与 Guo

等 $^{[10]}$ 的研究结果一致. Guo 等 $^{[10]}$ 的研究中培养前 14 d, 11%~19%的  $CO_2$  来源于根系, 28 d 以后不足 5%. 类似地, 玉米根系产生  $CO_2$  占比为 12%~17% $^{[21]}$ . 而水稻秸秆  $CO_2$  释放量占总释放量可达 30%~80% $^{[24]}$ .

根系引起的正 CO, 累积激发效应贡献了 CO, 累积

排放量的 8.6%~20.9% (图 5). CF+R 处理的根系

来源 CO, 累积排放量占比最高, CFS+R处理次之.

CF+R 处理的 SOC 来源 CO, 累积排放量占比为

72.4%,其中8.6%源于根系引起的正CO。累积激发效应.CFS+R处理的正CO。累积激发效应占比

是 CF + R 处理的近两倍(15.6%), 且低于 CK + R

水稻根系在土壤中会引起正激发效应. 因为根系碳相对土壤有机碳更易被微生物利用, 当外源有机碳的可利用性较高时, 会刺激微生物群落为了达到碳氮平衡而对土壤本身有机碳进行矿化分解, 从而挖掘更多矿质氮, 因而产生正激发效应. 本文根系CO<sub>2</sub> 相对累积激发效应为 14.4% ~ 42.5%, 且培养前 40 d CO<sub>2</sub> 相对累积激发效应更高(图 4). Zhu等<sup>[16]</sup>的研究结果与本文相似, 水稻根系的激发效应在培养 50 d 后达到最大, 为 43%, 之后下降至2.9%. 然而, 秸秆的激发效应大于根系, 可达 52% ~ 378% <sup>[16]</sup>. 这与根系和秸秆的易利用和难利用等多种组分形态有关<sup>[21]</sup>.

### 3.2 秸秆还田对水稻根系矿化和激发效应的影响

秸秆还田是提高土壤碳库和土壤肥力的重要措施,一直以来得到大力推广.不同的荟萃分析研究显示长期水稻秸秆还田可将土壤有机碳含量提高5.4%~12% [8,9].本文土壤经过约30 a 的秸秆还田,土壤有机碳含量与 CK 相比提高了 48% (表 1).有机质含量对土壤养分含量、微生物群落结构和活性和生物酶活性等具有显著影响 [14,25].本文秸秆还田土壤较 CK 全氮、碱解氮和全磷含量分别增长了53%、55%和 177%.秸秆还田还会改良土壤物理结构,改善土壤通透性,本文秸秆还田土壤较 CK 的黏

粒和粉粒含量分别降低了 18% 和 14% (表 1). 因此,秸秆还田利于土壤微生物生长和对有机物的分解矿化. 宁赵等<sup>[26]</sup>的研究显示秸秆还田会增加土壤微生物生物量和生长速率,提高微生物对羧酸类、氨基酸和碳水化合物的利用速率. 因此,本文不加根的秸秆还田土壤 CO<sub>2</sub> 累积释放量比 CK 多 18. 1% (图 1),加根的秸秆还田处理的根系产生 CO<sub>2</sub> 累积释放量比 CK 增加了 22. 5% (图 2). 另外,水稻土中有机质的转化过程以真菌为主<sup>[27,28]</sup>,长期秸秆还田土壤中真菌含量显著高于无秸秆还田土壤<sup>[27]</sup>. 真菌相比于细菌更擅长和偏好分解难降解有机物<sup>[27,28]</sup>,因此秸秆还田处理相较于 CK 处理对根系的分解能力更强,根系产生 CO<sub>2</sub> 和矿化率更高(图 2).

与单施化肥相比,本文秸秆还田处理在培养前期根系 CO<sub>2</sub> 释放量更高,之后因单施化肥处理的根系 CO<sub>2</sub> 释放速率更快,培养结束时二者根系 CO<sub>2</sub> 累积释放量无显著差异(图 2). 这可能是由于培养初期,如前面讨论所述,因秸秆还田土壤微生物生物量和活性更高,根系矿化更快,但由于矿化过程中对土壤养分如氮磷的消耗,导致养分计量比逐渐不平衡,限制了矿化速率,而化肥土壤相对秸秆还田土壤的氮磷养分占比更大<sup>[5]</sup>,因此根系的矿化能够持续高速进行,最终导致两个处理间的累积矿化率相近. 宁赵等<sup>[26]</sup>的研究也显示化肥和秸秆还田处理的外源碳代谢特征相似.

土壤本身有机碳的矿化率为4.2%~5.8%(图 3),小于根系矿化率,表明土壤中难分解有机碳组 分更多. 秸秆还田土壤本身有机碳的 CO, 累积释放 量略高于 CK 和化肥处理,但差异不显著. 但因秸秆 还田土壤有机碳含量较高,其累积矿化率显著低于 CK 和化肥处理(图3),说明秸秆还田土壤有机碳的 易矿化程度更低. 秸秆还田处理的根系 CO, 相对累 积激发效应小于 CK 处理(图 4),这可能是由于秸 秆还田土壤养分含量相对更高,根系矿化过程中微 生物通过分解土壤本身有机碳来获取速效氮磷养分 的需求强度不如 CK 大. 化肥处理的 CO, 累积激发 效应小于秸秆还田处理(图4),也是出于养分元素 计量比的原因,化肥处理的土壤中氮磷养分的相对 比例更高,因此分解土壤本身有机碳的强度不高.另 外,因真菌分解难降解有机物的能力较强且秸秆还 田土壤真菌含量更高[27,28],即在外源有机碳刺激下 对土壤本身有机碳的挖掘能力优于化肥土壤,这可 能也是导致更大激发效应的原因之一.

### 4 结论

(1)经过120 d 的室内培养,秸秆还田和化肥处

理根系的矿化率分别为 18.8% 和 17.8%, 显著高于 CK 处理的 15.4%.

- (2)3个处理的土壤本身有机碳产生的 CO<sub>2</sub> 累积排放量无显著差异,但是对于土壤有机碳矿化率,秸秆还田处理显著低于化肥和 CK 处理.秸秆还田土壤中根系引发的 CO<sub>2</sub> 相对累积激发效应为29.6%,显著低于 CK 处理的 42.5%,但高于化肥处理的 14.4%.
- (3) 淹水水稻土 CO<sub>2</sub> 释放量中 23.47% ~ 27.59%来源于根系,其余来源于土壤,其中,激发效应引起的 CO<sub>2</sub> 释放量占比在 CK 处理中最大,秸秆还田处理次之,化肥处理最小.

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