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城市污水再生处理中微量有机污染物控制的关键难题与解决思路 王文龙,吴乾元,杜烨,黄南,陆韻,魏东斌,胡洪营







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# 汞矿区周边土壤重金属空间分布特征、污染与生态风 险评价

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摘要: 为探明汞矿开采对周边土壤环境及人体健康的影响,在重庆市酉阳县某汞矿聚集区周边采集表层土壤组合样 42 件,分析了土壤中重金属(Cd、Hg、Pb、As、Cr、Cu、Zn 和 Ni)的含量及 pH,对土壤重金属含量的空间分布特征、污染程度和生态风险进行了研究. 结果显示,研究区土壤重金属表层富集明显,参照《土壤环境质量 农用地土壤污染风险管控标准(试行)》(GB15618-2018)发现,土壤 Cd、Hg、Pb、As 和 Zn 存在不同程度的超标现象. 土壤存在一定程度的污染和生态风险,土壤重金属中度-重度污染和强生态风险区均分布在矿点周围,说明矿业活动对土壤环境的影响. 土壤 Cr、Cu 和 Ni 含量可能受到母岩的风化和成土作用的影响,土壤 Hg、Pb 和 Zn 可能受到矿产开采等人为活动的影响,土壤 Cd 和 As 同时受到地质背景和人为活动的影响. 重金属对成人的健康影响较小,对儿童造成健康风险的概率较大,土壤 As 是造成人体健康风险的主要贡献因子,8种重金属均以经口摄入途径的贡献率最高. 汞矿的开采是造成研究区土壤污染及生态风险的主要原因.

关键词:汞矿区;土壤重金属;分布特征;土壤污染;生态风险

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# Spatial Distribution Characteristics, Pollution, and Ecological Risk Assessment of Soil Heavy Metals Around Mercury Mining Areas

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Abstract: To ascertain the impact of mercury mining on the surrounding soil environment and human health, 42 surface soil composite samples were collected around a mercury mining area in Youyang County, Chongqing, and the heavy metals (Cd, Hg, Pb, As, Cr, Cu, Zn, and Ni) contents and pH of the soil, the spatial distribution of heavy metals, pollution degree, and ecological risk were studied. The results show that the surface soil layer in the study area is significantly enriched in heavy metals. According to the soil environmental quality risk control standard for soil contamination of agricultural land (GB15618-2018), soil Cd, Hg, Pb, As, and Zn showed different degrees of excess. A certain degree of pollution and ecological risk was also identified in the studied soil. Moderate-to-severe pollution and strong ecological risk areas are distributed around the mining sites, indicating the impact of mining activities on the soil environment. The content of Cr, Cu, and Ni in the soil may be affected by weathering and soil formation from the parent rock; Hg, Pb, and Zn content may be affected by human activities such as mineral mining; and Cd and As content may be affected by both geological processes and human activities. Heavy metals pose less of a health risk for adults but have a greater probability of causing health risks for children. Soil As is the main contributor to human health risks, and the oral intake of the eight heavy metals has the highest contribution rate. The mining of mercury is the main cause of soil pollution and ecological risk in the study area.

Key words: mercury mining area; soil heavy metals; distribution characteristics; soil pollution; ecological risk

土壤是人类赖以生产的基础,是粮食生产的重要保障. 近些年来,随着城镇化节奏的加快以及农业生产不断规模化和机械化,土壤环境正面临着严峻的挑战<sup>[1~3]</sup>. 大量研究表明,我国中南、珠三角和长三角地区由于大量的工业及矿业活动,部分区域土壤重金属污染问题显著<sup>[4~10]</sup>. 土壤重金属具有持久性、潜伏性和易进入食物链等特点,会对农产品安全及人体健康造成严重威胁,是影响土壤环境质量、限值地区经济发展的主要因子之一<sup>[11,12]</sup>.

我国西南地区地质背景复杂,分布有多个成矿带,受到矿业活动的影响,土壤重金属超标问题显

著<sup>[13-15]</sup>. 鲍丽然等<sup>[16]</sup>对重庆秀山西北部某汞矿区周边农田土壤重金属进行了生态健康风险评价,结果表明,由于高地质背景及矿产的开采,土壤 Hg、Cd 和 As 等重金属存在不同程度的污染和生态风险,且通过不同的暴露途径,土壤重金属会对该区域儿童造成一定程度的致癌和非致癌风险,陈绍杨<sup>[17]</sup>也得到了相似的结论. 李礼等<sup>[18]</sup>对重庆某锰

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作者简介: 王锐(1994~),男,硕士,主要研究方向为生态地球化学, E-mail: 1372661182@ qq. com 矿区土壤重金属含量进行研究发现,矿区周边土壤中 Mn、Cr 和 As 等重金属污染问题显著,尾矿库下游土壤存在中等以上生态风险.可见,采矿等人为活动会对其周边土壤中释放大量重金属元素,造成严重的生态问题.前人关于渝东南矿区的研究多以数据评价为主,缺乏重金属污染程度和生态风险空间分布规律的探讨.本文选择重庆市酉阳县西北部某汞矿集中区为研究对象,通过分析土壤中重金属的含量特征及空间分布特征,评价土壤重金属污染及风险程度,并基于不同暴露途径对该地区成人及儿童进行健康风险评价,通过系统地了解当地土壤重金属污染特征及人体健康风险程度,以期为矿山污染修复和治理提供科学依据.

#### 1 材料与方法

#### 1.1 研究区概况

酉阳县位于渝鄂湘黔四省市结合部,东邻湖南省龙山县,南与秀山县、贵州省松桃、印江县接壤,西与贵州沿河县隔江(乌江)相望,西北与彭水县,正北与黔江区、湖北省咸丰、来凤县相连.酉阳县属亚热带湿润季风气候区,海拔高差大,年平均气温由海拔280 m的17℃递减到中山区的11.8℃.月平均气温1月最冷为3.8℃,7月最高为24.5℃.年降雨量一般在1000~1500 mm.本次研究区位于酉阳县西北部,该区域内存在分布较集中的汞矿点.研究区排地类型以旱地为主,大宗农作物为玉米.

### 1.2 样品采集与测试

网格化采集农用地表层土壤样,采样密度为 1点·km<sup>-2</sup>,采样深度 0~20 cm,在采样中心点周围 100 m 范围内 3~5 处多点采集组合,共采集表层土壤单点样 168 件. 待土壤样品风干、干燥后敲碎,过 0.8 mm(20目)粒级尼龙筛,将 4 km<sup>2</sup> 大格内样品等重量组合为一个分析样,组合分析样 42 件,分析组合样分布位置见图 1.

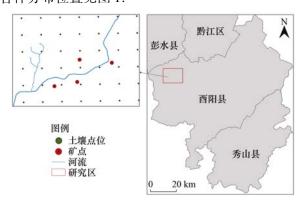


图 1 土壤样品及汞矿点示意

Fig. 1 Schematic diagram of soil samples and mercury deposits

样品在日光下自然风干,干燥过程中及时揉搓或用木棒敲打防止结块,干燥后用 20 目尼龙筛过筛,样品过筛重不低于 500 g,每件样品重新过筛后剩余部分不超过 1 g. 每加工完一个样品均对加工工具进行全面清扫,防止发生人为玷污. 样品初步加工后装于干净无污染的塑料瓶中,送至实验室进一步加工并分析.

分析检测过程按《区域地球化学样品分析方法》(DZ/T 0279-2016)<sup>[19]</sup>的检验规程进行测试,选择使用 X 射线荧光光谱法(XRF)测定土壤 Pb、Cr和 Zn 含量,使用电感耦合等离子体质谱法(ICP-MS)测定土壤 Cd、Ni和 Cu 含量,使用原子荧光法(AFS)测定土壤 As和 Hg含量,使用离子选择性电极法(ISE)测定土壤 pH. 各分析方法的最低检出限均等于或优于《多目标区域地球化学调查规范(1:250000)》(DZ/T0258-2014)<sup>[20]</sup>的要求.采用多目标分析方法配套方案,经过12次分析国家一级土壤地球化学标准物质系列样品(GBW07458~07461),分别统计各被测元素测定平均值与标准值之间的对数误差,其计算结果符合文献[20]的要求,元素报出率均达到100%.

#### 1.3 污染负荷指数

污染负荷指数法是 Tomlinson 提出的能够反映不同污染物污染等级及其对总体污染情况贡献程度的污染评价方法,被广泛应用于场地和农田土壤重金属的污染评价,计算公式如下<sup>[21]</sup>:

$$CF_i = C_i / C_n \tag{1}$$

$$P = \sqrt[n]{\text{CF}_1 \times \text{CF}_2 \times \dots \times \text{CF}_n}$$
 (2)

$$P_{\text{zone}} = \sqrt[m]{P_1 \times P \times \dots \times P_m} \tag{3}$$

式中, $CF_i$  表示重金属 i 的污染指数;  $C_i$  和  $C_n$  分别表示重金属 i 的实测含量和背景值,此处参照 Yan 等 [22] 给出的重庆市土壤背景值进行评价 (表 1),该值是通过重庆市已完成的多目标区域地球化学调查所得,统计样品共计13 976件; P 表示某采样点的污染负荷指数, $P_{zone}$  表示区域总体污染负荷指数,n 表示重金属个数,m 表示样点个数.  $P \le 1$  为无污染,  $1 < P \le 2$  为轻微污染,  $2 < P \le 3$  为中度污染, P > 3 为重度污染 [23],  $P_{zone}$  和 P 的分级标准一致.

#### 1.4 潜在生态风险评价

瑞典学者 Hakanson 于 1980 年首次提出潜在生态危害指数法,该方法体现了生物有效性、相对贡献及地理空间差异等特点,是综合反映重金属对生态环境影响潜力的指标,适合对大区域范围沉积物和土壤进行评价比较,是目前常用的重金属风险评价方法<sup>[24]</sup>.其计算公式如下<sup>[25]</sup>:

#### 表 1 重庆市土壤重金属背景值

元素	Cd	Hg	Pb	As	Cr	Cu	Zn	Ni
背景值/mg·kg <sup>-1</sup>	0. 26	0. 05	27. 03	4. 42	72. 32	23. 19	78. 61	30. 2

$$E_{\rm r}^i = T_{\rm r}^i \times \frac{C^i}{C_{\rm r}^i} \tag{4}$$

$$RI = \sum_{i}^{m} E_{r}^{i}$$
 (5)

式中,RI表示土壤中重金属的潜在生态风险指数; $E_r^i$ 是重金属i的潜在生态风险系数; $C_n^i$ 为土壤中重金属i的实测值; $C_n^i$ 表示重金属i的背景值; $T_n^i$ 是重金属i的毒性系数,重金属Cd、Hg、Pb、As、Cr、Cu、Zn 和 Ni 的毒性系数分别为 30、40、5、10、2、5、1 和  $S_n^{[26]}$ . 潜在生态风险指数分级标准见表 2.

#### 1.5 人体健康风险评价

采用 USEPA 公布的健康风险评估模型评价土

#### 表 2 重金属潜在生态风险指数分级标准

Table 2 Classification of the potential ecological

risk index of heavy metals 生态风险等级 生态风险等级 RI  $E_r^i < 40$ 轻微 RI < 150 轻微 150 ≤ RI < 300  $40 \le E_r^i < 80$ 中等  $80 \le E_r^i < 160$ 强 300 ≤ RI < 600  $160 \le E_r^i < 320$ 很强 RI≥600  $E_r^i \ge 320$ 极强

壤重金属的人体健康风险评价模型,考虑经口直接摄入、呼吸吸入和皮肤接触3种暴露途径,进行人体健康风险评价. 计算公式如下<sup>[27]</sup>:

ADD<sub>Δ□摄λ</sub> = 
$$\frac{C_i \times IR_{ing} \times EF \times ED}{BW \times AT} \times 10^{-6}$$
 (6)

$$ADD_{\text{FFB-WA}} = \frac{C_i \times IR_{\text{inh}} \times EF \times ED}{PEF \times BW \times AT} \times 10^{-6}$$
 (7)

$$\mathrm{ADD}_{\mathrm{gkk\acute{e}k\acute{e}k}} \, = \, \frac{C_i \times \mathrm{SA} \times \mathrm{AF} \times \mathrm{ABS} \times \mathrm{EF} \times \mathrm{ED}}{\mathrm{BW} \times \mathrm{AT}} \times 10^{-6}$$

 $HI = \sum HQ_i = \sum ADD_i/RfD_i$  (9)

(8)

式中,ADD表示不同途径重金属的日均暴露剂量,  $C_i$ 表示土壤中重金属 i 的含量,  $HQ_i$  表示重金属 i非致癌健康风险指数,  $RfD_i$  表示重金属 i 在不同暴 露途径下的参考剂量,对于多种重金属暴露的情况, 用指数 HI 表示产生的总危害. 若 HI < 1,表明研究 区土壤重金属暴露浓度低于参考剂量,可认为研究 区风险较小或忽略不计,若 HI > 1,则认为研究区有 发生慢性病的风险,且 HI 越高,发病的可能性就越 大[28]. 式中其余参数的含义及取值见表 3,不同暴 露途径下重金属元素的参考剂量  $RfD_i$  见表 4.

表 3 健康风险评价模型计算参数[29]

Table 3 Calculation parameters of the health risk assessment model

49	Table 5 Calcula	ation parameters of the nearth ris	sk assessment model	
参数含义	符号	成人参考值	儿童参考值	单位
每日经口摄人土壤量	$IR_{ing}$	100	200	mg•d ⁻¹
每日空气呼吸量	$IR_{inh}$	14. 5	7.5	$m^3 \cdot d^{-1}$
暴露皮肤表面积	SA	2145	1150	$\mathrm{cm}^2$
皮肤粘附系数	$\mathbf{AF}$	0. 07	0. 2	mg•cm <sup>-2</sup> •d
皮肤吸收因子	ABS	0.001	0.001	_
地表灰尘排放因子	PEF	$1.36 \times 10^9$	$1.36 \times 10^9$	$m^3 \cdot kg^{-1}$
暴露频率	EF	350	350	d•a <sup>-1</sup>
暴露年限	ED	24	6	a
平均体重	BW	56. 8	15. 9	kg
平均暴露时间	AT	$ED \times 365$	$ED \times 365$	d

#### 表 4 不同暴露途径下重金属元素的参考剂量[30]

Table 4 Reference doses of heavy metal elements

under different exposure routes

元素		RfD/mg•d•kg-	1
儿系	经口摄入	呼吸吸入	皮肤接触
Cd	$1 \times 10^{-3}$	$1 \times 10^{-3}$	$1 \times 10^{-5}$
Hg	$3 \times 10^{-4}$	$3 \times 10^{-4}$	2. $1 \times 10^{-5}$
Pb	$3.5 \times 10^{-3}$	$3.5 \times 10^{-3}$	$5.25 \times 10^{-4}$
As	$3 \times 10^{-4}$	$3 \times 10^{-4}$	$1.23 \times 10^{-4}$
$\operatorname{Cr}$	$3 \times 10^{-3}$	$3 \times 10^{-4}$	$6 \times 10^{-5}$
Cu	$4 \times 10^{-2}$	$4 \times 10^{-2}$	$1.2 \times 10^{-2}$
Zn	$3 \times 10^{-1}$	$3 \times 10^{-1}$	$6 \times 10^{-2}$
Ni	2 × 10 <sup>-2</sup>	$2 \times 10^{-2}$	$5.4 \times 10^{-3}$

对于致癌风险, USEPA 模型未给出所有重金属元素 3 种暴露途径的致癌因子,此外,对于研究区基础数据的调查缺乏,因此,本文未评价重金属的致癌风险.

#### 2 结果与讨论

#### 2.1 土壤重金属空间分布特征

研究区土壤重金属含量特征见表 5. 土壤重金属的变异系数由大到小的顺序为 Hg(1.19) > Pb(0.46) > As(0.42) > Cd(0.36) > Zn(0.34) >

Cu(0.22) > Ni(0.21) > Cr(0.07),其中 Cr 的变异系数最小,说明其含量在空间上变化不大,受到人为活动影响较小,其余重金属均表现出明显的空间变异性,尤其是 Hg,变异系数高达 1.19,说明其受到强烈的人为活动的影响<sup>[31]</sup>. 土壤 pH 的变化范围为5.08~7.84,土壤以中酸性为主.

研究区土壤 Cd、Hg、Pb、As、Cr、Cu、Zn 和 Ni 的平均值与全国土壤环境背景值<sup>[32]</sup> 的比值分别为 4.40、16.57、1.81、1.53、1.35、1.16、1.38 和 1.19,与重庆市土壤元素背景值的比值分别为 1.69、23.20、1.74、3.88、1.14、1.13、1.31 和

1.06. 对于 Cd 和 Hg 而言,最高含量分别是全国土壤环境背景值的 10.82 和 154.17 倍,说明其存在显著异常高值点.

与文献[33]中给出的重金属污染风险筛选值进行比较发现,土壤 Cd、Hg、Pb、As和 Zn超过相应筛选值的点位所占比例分别为 42.86%、28.57%、2.38%、7.14%和2.38%,说明土壤对农产品质量安全、农作物生长或土壤生态环境可能存在风险.与重金属污染风险管控值对比得出,土壤 Hg 超过管控值的点位所占比例为9.52%

表 5 土壤中重金属元素含量统计结果1)

Table 5 Statistical results of heavy metal element content in soil

元素	最小值	最大值	平均值	中位值	标准差	变异系数	$K_1$	$K_2$
Cd	0. 24	1. 05	0. 44	0.42	0. 16	0. 36	4. 40	1. 69
Hg	0.05	10.02	1. 16	0.35	2. 22	1. 91	16. 57	23. 20
Pb	30. 50	147. 01	47. 11	40. 72	21. 44	0.46	1. 81	1.74
As	4. 23	35. 91	17. 13	17. 15	7. 22	0.42	1. 53	3. 88
$\operatorname{Cr}$	69. 10	111.02	82. 18	82. 01	6. 31	0. 07	1. 35	1,147
Cu	19. 02	46. 11	26. 21	24. 55	5. 70	0. 22	1. 16	1.13
Zn	73. 40	270. 11	102. 68	93. 15	34. 91	0.34	1.38	1,31
Ni	22, 61	57. 03	32. 13	30.40	6. 81	0.21	1. 19	1.06
	7 8		1 2 1 100.7	-	1 21 3		400	

1) K<sub>1</sub> 和 K<sub>2</sub> 分别表示研究区土壤重金属含量均值与全国环境背景值和重庆土壤背景值的比值

利用 ArcGIS 10.2 中的反距离权重法对土壤重金属元素进行空间插值分析,结果见图 2.图中 Cd、Hg、Pb、As 和 Zn 含量的空间分布特征基本相似,表现为中部地区含量高,周边区域含量低,其高含量区与矿点的分布基本吻合,说明其含量可能受到矿业活动的影响.土壤 Cr、Cu 和 Ni 含量的空间分布特征不明显,且 Cr、Cu 和 Ni 的变异系数均相对较小,说明其含量受到采矿等人为活动的影响较小.

#### 2.2 重金属污染评价

#### 2.2.1 污染负荷指数

利用式(1)~(3)给出的公式对研究区土壤重金属污染状况进行评价,污染负荷指数(P)的变化范围为1.13~4.37,平均值为1.89.绘制污染负荷指数(P)的空间分布,并统计不同污染等级所占比例,结果见图3.研究区轻微污染、中度污染和重度污染所占比例分别71.43%、21.42%和7.15%,不存在无污染区,重金属污染问题显著.从空间分布上看,土壤重金属污染区以矿点为中心,中部主要为中度和重度污染,距离矿点较远的区域,重金属污染程度相对较低,进一步说明了矿业活动对土壤环境的影响.研究区总污染负荷指数(Pzone)为1.82,表现为轻度污染.总体而言,研究区土壤重金属以轻度污染为主,存在一定程度的中度和重度污染,采矿等人为活动可能是造成土壤重金属污染的主要原因.

## 2.2.2 潜在生态风险评价

利用式(4)和(5)进行土壤潜在生态风险评价,各重金属的潜在生态风险系数( $E_r^i$ )计算结果见图 4. 土壤 Pb、Cr、Cu、Ni 和 Zn 的潜在生态风险系数 均小于 40,生态风险等级较低. 土壤 As 和 Cd 的  $E_r^i$  值变化范围分别为 9. 57~81. 22 和 27. 34~121. 15,平均值分别为 38. 76 和 50. 73,以中等风险为主,存在部分中度风险点位. 土壤 Hg 的  $E_r^i$  值变化范围为 39. 36~8 016. 8,平均值为 930. 16,生态风险极强.

图 5 为土壤中重金属的潜在生态风险指数空间分布情况,土壤重金属轻微、中等、强和很强生态风险所占比例分别为 9.53%、28.57%、26.19% 和 35.71%,与矿点分布图对比发现,研究区土壤重金属强到很强生态风险均集中在矿点分布区,说明矿业活动会造成明显的生态风险.

#### 2.2.3 主成分分析

利用 SPSS 25.0 软件对表层土壤重金属含量进行 KMO 检验,得到的统计量值为0.86,Bartlett 球度检验相伴概率为0.000,说明数据适合进行因子分析<sup>[34]</sup>,对 Kaiser 标准化后的因子进行 Varimax 正交旋转,得到了2个特征值大于1的主成分,累计方差贡献率为76.576%,可解释土壤重金属元素的大部分信息<sup>[35]</sup>,分析结果见表6.

F1 载荷较高的元素为Cd、As、Cr、Cu和Ni,前

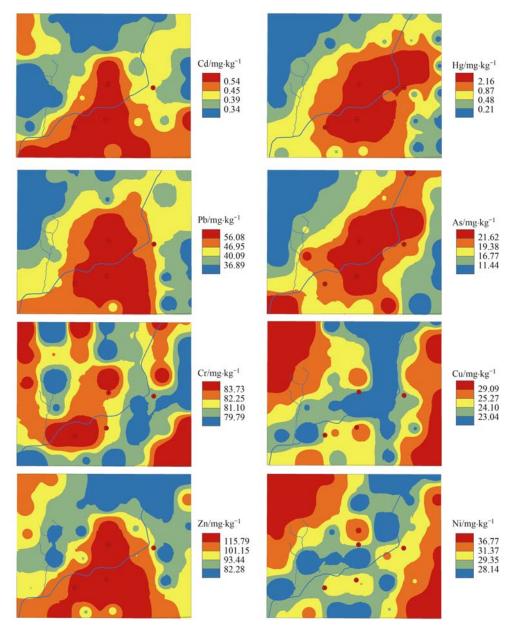


图 2 土壤重金属元素含量空间分布

Fig. 2 Spatial distribution of soil heavy metal content

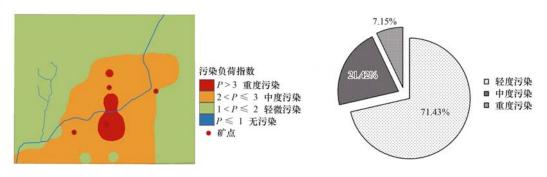


图 3 土壤重金属污染负荷指数评价结果

Fig. 3 Evaluation results of soil heavy metal pollution load index

文分析已知, Cr、Cu 和 Ni 的变异系数较小,且含量的空间分布与汞矿点分布相关性较低,受到人为活动的影响较小. 王锐等<sup>[36]</sup>对酉阳县南部土壤重金属的来源进行了研究,结果表明,土壤 Cr 及 Ni 的来源

主要受到地质背景的控制,因此,第一主成分元素可能来源于母岩的风化和成土作用.

F2 载荷较高的元素为 Cd、Hg、Pb、Zn 和 As, 从图 2 可以看出,土壤 Cd、Hg、Pb 和 Zn 的高含量

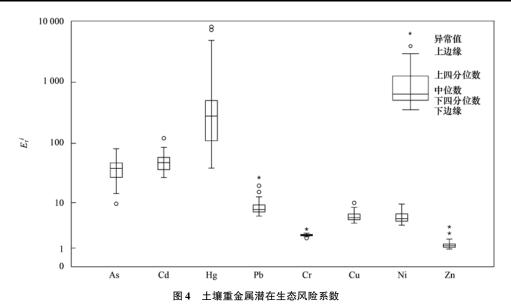


Fig. 4 Potential ecological risk coefficient of soil heavy metals

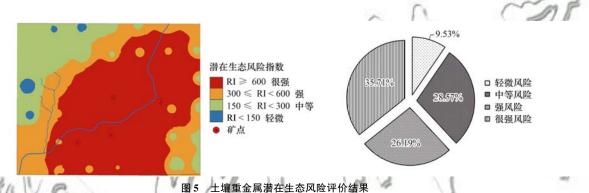


Fig. 5 Results of potential ecological risk assessment of soil heavy metals

区主要集中在汞矿点附近. 有研究表明, Hg 与 Pb、Zn 等元素的地球化学性质相似,且在成矿过程中,存在明显的伴生关系,汞矿的开采会导致土壤中 Hg 等重金属元素的输入<sup>[37,38]</sup>,因此,第二主成分可能来源于矿产开采等人为活动.

结合重金属元素含量、重金属污染和重金属生态风险分布情况可知,由于汞矿的开采,向周边土壤中输入了大量的重金属元素,导致土壤出现了以 Hg和 Cd 等元素为主的重金属污染及生态风险.

## 表 6 土壤重金属主成分分析矩阵

Table 6 Matrix analysis of soil heavy metals

元素	F1	F2
Cd	0. 626	0. 561
Hg	0. 213	0. 881
Pb	0. 344	0. 634
As	0. 641	0. 457
$\operatorname{Cr}$	0. 803	-0.171
Cu	0. 961	0. 124
Zn	0. 142	0. 940
Ni	0. 919	0. 141
初始特征值	3. 572	2. 554
方差贡献率/%	44. 649	31. 927

#### 2.3 人体健康风险评价

利用式(6)~(9)计算成人和儿童非致癌风险 指数,结果见表7及图6.对成人而言,3种暴露途 径下,8种重金属成人的健康风险指数(HQ)及总 危害指数(HI)均小于1,说明土壤重金属对成人 的风险较小, 8 种重金属对 HI 的贡献率按从大到 小的顺序依次为 As(53.33%) > Cr(27.43%) > Pb(12.56%) > Hg(3.68%) > Ni(1.50%) >Cu(0.61%) > Cd(0.47%) > Zn(0.33%),  $\pm$   $\frac{1}{3}$ As 为主要的贡献因子. 就暴露途径而言, 8 种重金 属均以经口摄入途径的贡献率最高. 对儿童而言, 土壤 As 在 3 种暴露途径下的健康风险指数(HQ) 的变化范围为 0.17~1.45,平均值为 0.69,存在 大于1的点位,存在一定的慢性病风险. 儿童的总 危害指数(HI)的变化范围为 0.67~2.55,平均值 为1.29,说明通过3种暴露途径,土壤重金属元素 对儿童造成健康风险的概率较大,8种重金属对 HI的贡献率按从大到小的顺序依次为As (53.63%) > Cr(27.13%) > Pb(12.64%) > Hg(3.69%) > Ni(1.51%) > Cu(0.62%) > Cd (0.46%) > Zn(0.32%), 土壤 As 是主要的贡献 因子. 与成人相同, 8 种重金属均以经口摄入途径的贡献率最高.

图 7 为人体健康风险评价空间分布情况,可以看出,高风险区多与矿点的分布存在一致性,说明采矿活动对人体健康存在明显的风险.

表 7 人体健康风险指数统计

rr 11 /	7 0.		C .1	human	1 1.1	. 1	. 1
Table	/ Sta	fistics.	of the	human	health	risk	index

	二丰		成人			儿童	
	元素 -	经口摄人	呼吸吸入	皮肤接触	经口摄人	呼吸吸入	皮肤接触
	最小值	4. 00E - 04	4. 27E – 08	6. 01E - 05	2. 86E - 03	7. 88E - 08	3. 29E - 04
$\operatorname{Cd}$	最大值	1.77E - 03	1.89E - 07	2. 66E – 04	1.27E - 02	3.49E - 07	1.46E - 03
	平均值	7. 42E – 04	7.91E - 08	1.11E - 04	5.30E - 03	1.46E -07	6. 10E – 04
	最小值	2. 77E – 04	2. 95E - 08	5. 94E - 06	1. 98E - 03	5. 45E - 08	3. 25E - 05
Hg	最大值	5. 64E - 02	6.01E - 06	1.21E - 03	4.03E - 01	1.11E - 05	6.62E - 03
	平均值	6. 54E – 03	6.98E - 07	1. 40E – 04	4. 67E – 02	1. 29E - 06	7. 68E – 04
	最小值	1. 47E – 02	1. 57E – 06	1. 47E - 04	1. 05E - 01	3. 76E - 07	8. 06E - 04
Pb	最大值	7.09E - 02	7. 56E – 06	7. $10E - 04$	5.07E - 01	9.48E -07	3.88E - 03
	平均值	2.27E - 02	2.42E - 06	2.27E - 04	1.62E - 01	5.34E -07	1.24E - 03
	最小值	2. 38E - 02	2. 54E - 06	8. 72E - 05	1. 70E - 01	4. 69E - 06	4. 77E – 04
As	最大值	2. 02E - 01	2. 15E – 05	7. 40E – 04	1.44E + 00	3. 98E - 05	4. 05E – 03
	平均值	9. 64E – 02	1.03E - 05	3.53E - 04	6. 89E – 01	1. 90E - 05	1. 93E <b>-</b> 03
	最小值	3.89E - 02	4. 15E – 05	2. 92E - 03	2. 78E – 01	7. 66E – 05	1. 60E – 02
Cr	最大值	6. 25E – 02	6.66E - 05	4. 69E – 03	4. 46E – 01	1. 23E - 04	2. 57E – 02
	平均值	4. 62E – 02	4.93E - 05	3.47E - 03	3. 30E - 01	9. 11E - 05	1.90E - 02
	最小值	8. 02E - 04	8. 55E – 08	4. 01E - 06	5. 73E - 03	1. 58E - 07	2. 20E - 05
Cu	最大值	1. 95E – 03	2. 07E - 07	9. 74E – 06	1. 39E - 02	3.83E - 07	5. 33E - 05
1 1	平均值	1. 11E – 03	1. 18E – 07	5. 54E – 06	7. 90E - 03	2. 18E - 07	3.03E - 05
3	最小值	4. 13E – 04	4. 40E - 08	3. 10E - 06	2. 95E - 03	8. 14E - 08	1. 70E - 05
Zn	最大值	1. 52E – 03	1.62E - 07	1. 14E – 05	1.09E -02	2. 99E - 07	6. 24E – 05
23	平均值	5. 78E – 04	6. 16E – 08	4. 34E - 06	4. 13E – 03	1. 14E – 07	2.37E - 05
(e //	最小值	1. 91E - 03	2. 03E - 07	1. 06E - 05	1. 36E - 02	3.76E - 07	5. 81E - 05
Ni	最大值	4. 81E – 03	5. 13E - 07	2. 68E - 05	3.44E - 02	9.48E - 07	1.46E - 04
108	平均值	2.71E - 03	2.89E - 07	1.51E - 05	1. 94E - 02	5. 34E - 07	8.25E - 05

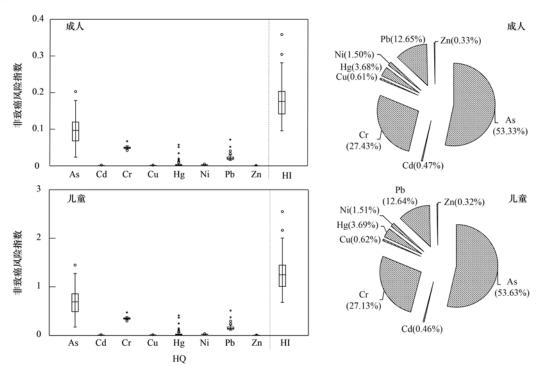


图 6 人体健康风险评价结果

Fig. 6 Results of the human health risk assessment

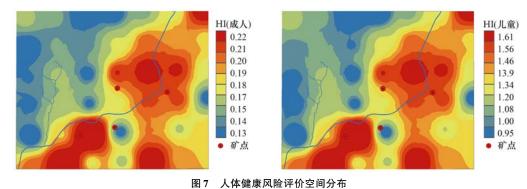


Fig. 7 Spatial distribution of the human health risk assessment

#### 结论 3

- (1)与重庆市全国土壤背景值对比发现,研究 区土壤重金属表层富集明显. 土壤 Cd、Hg、Pb、As 和 Zn 超过相应筛选值的点位所占比例分别为 42.86%、28.57%、2.38%、7.14%和2.38%,土壤 Hg 超过管控值的点位所占比例为 9.52%,研究区土 壤重金属超标问题显著. 以 Hg 为代表的土壤重金 属的变异系数较大,说明其受到强烈的人为活动的 影响. 土壤 Cd、Hg、Pb、As 和 Zn 的高含量区与矿 点的分布基本吻合,土壤 Cr、Cu 和 Ni 含量的空间 分布规律不明显.
- (2)污染负荷指数法评价结果显示,土壤以轻 微污染为主,存在部分中、重度污染区,不存在无污 染区. 潜在生态风险指数法评价结果显示, 土壤重金 属存在大面积的极强生态风险区所占比例分别为 35.71%,与矿点分布图对比发现,研究区土壤重金 属污染及生态风险区均分布在矿点周边,说明了矿 业活动对土壤安全的影响.
- (3)结合土壤重金属含量特征、分布特征及主 成分分析结果,土壤 Cr、Cu 和 Ni 含量受到人为活 动的影响较小,可能来源于母岩的风化和成土作用. 土壤 Hg、Pb 和 Zn 存在明显的伴生关系,可能受到 矿产开采等人为活动的影响. 土壤 Cd 和 As 同时受 到地质背景和人为活动的影响.
- (4)人体健康风险评价结果表明,成人的健康 风险指数(HQ)及总危害指数(HI)均小于1,说明土 壤重金属对成人的风险较小,但对儿童造成健康风 险的概率较大, 高风险区多与矿点的分布存在一致 性. 对成人和儿童而言,土壤 As 是主要的贡献因子, 8 种重金属均以经口摄入途径的贡献率最高.

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# 《环境科学》再获"百种中国杰出学术期刊"称号

2020年12月29日,中国科技论文统计结果发布会在北京举行,会议公布了"百种中国杰出学术期刊" 获奖名单.《环境科学》连续19次荣获"百种中国杰出学术期刊"称号."百种中国杰出学术期刊"是根据中国科技学术期刊综合评价指标体系进行评定.该体系利用总被引频次、影响因子、基金论文比、他引总引比等多个文献计量学指标进行统计分析,对期刊分学科进行评比,其评价结果客观公正,为我国科技界公认,并具有广泛影响.

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