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城市污水再生处理中微量有机污染物控制的关键难题与解决思路 王文龙,吴乾元,杜烨,黄南,陆韻,魏东斌,胡洪营







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西南典型区域夏季大气含氧挥发性有机化合物来源解析

陈木兰^{1,2,3}, 王赛男^{1,2,3}, 陈天舒⁴, 朱波⁵, 彭超^{1,2}, 周佳维^{1,2}, 车汉雄^{1,2}, 黄汝辉^{1,2}, 杨复沫⁶, 刘合凡⁷, 谭钦文⁷, 韩丽⁸, 陈军辉⁸, 陆克定⁹, 陈阳^{1,2,3}*

(1.中国科学院重庆绿色智能技术研究院,大气环境研究中心,重庆 400714; 2.中国科学院大学重庆学院,资源与环境学院,重庆 400714; 3.中国科学院大学资源与环境学院,北京 101400; 4.山东大学环境研究院,青岛 266237; 5.北京大学深圳研究生院环境与能源学院,城市人居环境科学与技术重点实验室,深圳 518055; 6.四川大学建筑与环境学院,成都 610065; 7.成都市环境保护科学研究院,成都 610031; 8.四川省生态环境科学研究部,成都 610041; 9.北京大学环境科学与工程学院,北京 100871)

摘要:含氧挥发性有机物(OVOCs)是大气光化学过程中的重要中间产物,是臭氧的重要来源之一.利用质子转移反应飞行时间质谱仪(PTR-TOF-MS)在成都平原对 OVOCs 进行观测,探讨其日变化特征、光化学反应活性、臭氧生成潜势和来源.结果表明,10 个 VOCs[乙醛、丙酮、异戊二烯、甲基乙基酮(methyl ethyl ketone, MEK)、甲基乙烯基甲酮(methyl vinyl ketone, MVK)、甲基丙烯醛(methacrolein, MACR)、苯、甲苯、苯乙烯、C8 芳香烃和C9 芳香烃]总浓度(体积分数)为(10.97 ± 4.69)×10⁻⁹, OVOCs 为(8.54 ± 3.44)×10⁻⁹, 芳香烃为(1.53 ± 0.93)×10⁻⁹, 生物源 VOCs 为(0.90 ± 0.32)×10⁻⁹; 光化学活性和臭氧生成潜势均排名前三的物种为:异戊二烯、乙醛和C8 芳香烃; 3 个 OVOCs 物种(乙醛、丙酮和MEK)主要来源于本地生物源和人为二次源,且丙酮有较强的区域背景值,说明该地区的污染受到较为显著的区域传输的影响.本研究可加深对西南地区臭氧的区域形成机制的认识,为科学管控臭氧污染提供依据.

关键词:含氧挥发性有机物(OVOCs);质子转移反应飞行时间质谱仪(PTR-TOF-MS); 臭氧生成潜势;光化学年龄参数法:来源分析

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Sources Apportionment of Oxygenated Volatile Organic Compounds (OVOCs) in a Typical Southwestern Region in China During Summer

CHEN Mu-lan^{1,2,3}, WANG Sai-nan^{1,2,3}, CHEN Tian-shu⁴, ZHU Bo⁵, PENG Chao^{1,2}, ZHOU Jia-wei^{1,2}, CHE Han-xiong^{1,2}, HUANG Rui-hui^{1,2}, YANG Fu-mo⁶, LIU He-fan⁷, TAN Qin-wen⁷, HAN Li⁸, CHEN Jun-hui⁸, LU Ke-ding⁹, CHEN Yang^{1,2,3*}

(1. Research Center for Atmospheric Environment, Chongqing Institute of Green and Intelligent Technology, Chinese Academy of Sciences, Chongqing 400714, China; 2. College of Resources and Environment, Chongqing School, University of Chinese Academy of Sciences (UCAS Chongqing), Chongqing 400714, China; 3. College of Resources and Environment, University of Chinese Academy of Sciences, Beijing 101400, China; 4. Environment Research Institute, Shandong University, Qingdao 266237, China; 5. Key Laboratory for Urban Habitat Environmental Science and Technology, School of Environment and Energy, Peking University Shenzhen Graduate School, Shenzhen 518055, China; 6. College of Architecture and Environment, Sichuan University, Chengdu 610065, China; 7. Chengdu Institute of Environmental Protection Science, Chengdu 610031, China; 8. Sichuan Ecological and Environmental Science Research Department, Chengdu 610041, China; 9. College of Environmental Sciences and Engineering, Peking University, Beijing 100871, China)

Abstract: Oxygenated volatile organic compounds (OVOCs) are important intermediates in the troposphere and the most important sources of ozone. Proton-transfer-reaction time-of-flight mass spectrometry (PTR-TOF-MS) was used to measure VOCs in the Chengdu Plain, Southwestern China. The diurnal variations, photochemical reactivity, O_3 formation potential, and sources were also investigated. The mixing ratios of ten kinds of VOCs (acetaldehyde, acetone, isoprene, Methyl ethyl ketone, Methyl vinyl ketone and Methacrolein, benzene, toluene, styrene, C8 aromatics, and C9 aromatics) were $(10.97 \pm 4.69) \times 10^{-9}$. The concentrations of OVOCs, aromatic hydrocarbons, and biogenic VOCs were $(8.54 \pm 3.44) \times 10^{-9}$, $(1.53 \pm 0.93) \times 10^{-9}$, and $(0.90 \pm 0.32) \times 10^{-9}$, respectively. Isoprene, acetaldehyde, and m-xylene were the top three photochemically active species with the greatest O_3 formation potentials. The dominant three OVOCs species (acetaldehyde, acetone, and MEK) were mainly derived from local biogenic sources and anthropogenic secondary sources, and acetone had a strong regional background level, indicating that pollution in this area is significantly affected by regional transmission. This study deepens the understanding of regional O_3 formation mechanisms in

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作者简介: 陈木兰(1996~),女,硕士研究生,主要研究方向为大气环境化学,E-mail; chenmulan18@ mails. ucas. ac. cn

^{*} 通信作者, E-mail: chenyang@ cigit. ac. cn

southwest China and provides a basis for the scientifically informed control of O₃ pollution.

Key words: oxygenated volatile organic compounds (OVOCs); proton-transfer-reaction time-of-flight mass spectrometry (PTR-TOF-MS); O₃ formation potential; photochemical age-based parameterization method; source apportionment

"十三五"以来,随着我国环境治理措施的不断完善和加强, $PM_{2.5}$ (空气动力学直径小于等于 2.5 μ m 的颗粒物)浓度显著下降^[1];而城市区域近地面臭氧浓度和超标天数则持续上升,污染呈持续加重趋势^[2~4].根据 2019 年中国生态环境状况公报(http://www.mee.gov.cn/),168 个主要城市中以 O_3 和 $PM_{2.5}$ 为首要污染物的超标天数分别占总超标天数的 46.4% 和 45.8%,平均超标天数比例为 27.3%;与 2018 年污染物浓度年际相比, O_3 浓度上升 13 μ g·m⁻³,占比约 8.4%; $PM_{2.5}$ 持平(44 μ g·m⁻³).近几年来,在成渝地区的夏秋季节, O_3 已代替 $PM_{2.5}$ 成为主要的空气污染物. 臭氧浓度升高影响人体健康,如形成慢性阻塞性肺病^[5],可造成过早死亡. 因此探究 O_3 形成机制,持续改善空气质量,也是污染防治攻坚战的重要组成部分.

环境挥发性有机化合物(VOCs)来源复杂,天然源主要以生物源如植被排放为主;人为源主要来自于生产生活过程,如机动车排放、工业排放以及溶剂使用等^[6].在氮氧化物(NO_x)的参与下,VOCs在大气中主要与羟基自由基(·OH)和臭氧等氧化剂发生化学反应,寿命一般在几个小时到几十天之间;生成的二次污染物包括含氧挥发性有机物(OVOCs)、O₃和过氧乙酰硝酸酯等光化学烟雾物质^[7].OVOCs同时也是生成二次有机气溶胶(SOA)的关键前体物和中间体^[8],在大气光化学中扮演重要的角色.

OVOCs 除一次排放外,也可由烷烃、烯烃和芳香烃等 VOCs 通过光化学反应二次生成. OVOCs 有着较高的光化学反应活性,自身可以发生光解,也可以与·OH和 NO₃等活性自由基发生光氧化、光解和气-固相分配反应,促进臭氧和 SOA 的生成^[9]. 大部分的 OVOCs 都对人体健康有危害,如国际癌症研究机构(International Agency for Research on Cancer)在2004 年将甲醛列为第一组人类致癌物^[10].

与非甲烷烃(NMHCs)相比,OVOCs的环境浓度变化范围更大,反应活性更高,大气寿命更短,对其分析测量方法提出了苛刻的要求:测量的时间分辨率必须足够高,并且采样及分析过程需要避免目标分析物的降解或化学反应转化^[11].目前针对OVOCs的分析以离线和在线方法为主.离线方法主要包括:①以2,4-二硝基苯肼(DNPH)^[12]或邻-(2,3,4,5,6-五氟苄基)-羟胺盐酸盐(PFBHA)^[13]等衍生剂为代

表的化学衍生法;②以离线罐采样、吸附剂采样或低温采样联合气相色谱(GC)分析方法.在线方法主要包括:在线气相色谱火焰离子化检测器(GC-FID)、气相色谱质谱联用仪(GC-MS)或 GC-FID/MS方法联用法^[8,14];质子转移反应质谱(PTR-TOF-MS)^[15]和化学诱导电离质谱(CIMS)等质谱方法^[16,17];差分吸收光谱技术(DOAS)、可调谐半导体激光吸收光谱(TDLAS)和傅里叶红外光谱仪(FTIR)等光谱方法^[18].

近年来,在我国京津冀、珠江三角洲、长三角和香港等地采用以上方法对 OVOCs 进行观测,探明其浓度水平和时空演化规律,并利用 PMF 模型 (positive matrix factorization) $^{[19]}$ 和光化学年龄参数法(photochemical age-based parameterization method)等手段进行了 OVOCs 来源解析 $^{[20]}$. 对北京、深圳和广州等城市和农村 OVOCs 观测结果来看,醇类和醛类是主要污染物,北方区域污染较南方严重 $^{[20-22]}$. Huang 等 $^{[20]}$ 利用 PTR-TOF-MS 监测对比春季北京和深圳两个城市,甲醇和甲醛均为两个地区的优势物种,且北方城市总 VOCs 浓度(体积分数,39.4×10 $^{-9}$)远高于南方城市的总 VOCs 浓度(16.7×10 $^{-9}$). 深圳污染地区丙酮、MEK 浓度分别为4.1×10 $^{-9}$ 和1.9×10 $^{-9[22]}$,广州偏远地区仅分别为0.6×10 $^{-9}$ 和0.3×10 $^{-9[23]}$.

四川盆地终年气候温热潮湿,全年风速较小,污染物不易扩散,形成了独特的大气污染物物理化学反应条件,大气复合污染较为严重,但目前对成渝地区城市臭氧污染成因方面的科学认识尚不充分,且臭氧的区域成因无明确结论.具体而言,成都平原的臭氧污染呈明显的区域性特征,只针对城市区域的臭氧成因进行研究不能全面掌握其形成规律.因此,本研究选取离成都市区不远处的农村区域进行布点,通过掌握大气中典型 OVOCs 物种的浓度水平、组成特征以及演化过程,分析其来源与大气反应途径,以期阐明区域 OVOCs 特征对四川盆地臭氧形成的影响,并为改善区域空气质量提供科学支撑.

1 材料与方法

1.1 地点和气象条件

本研究的观测点(31.01°N,104.22°E)位于成都市东北方的广汉市新平镇桂红村村委会院内(图1),是成都平原与高原的交界处,作为城市与平原、

高原相互作用的背景点. 采样口离地面约 3 m,周围 5 km 无高层建筑,无规模以上的工业企业,站点周围为稻田,是观测大气活性物质的一个较理想站点. 观测期为 2019 年 8 月 20 日 ~ 9 月 12 日,期间平均温度、湿度和矢量风速分别为(24.58 ± 3.50) $^{\circ}$ (86.45 ± 12.68)% 和(0.43 ± 0.27) m·s⁻¹, 西南-偏西风占主导.

本研究采用质子转移反应质谱仪连续监测该地区的 OVOCs 和其他 NMHCs.

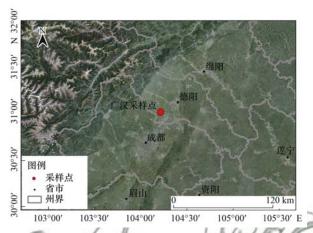


图 1 广汉采样点示意

Fig. 1 Map of the Guanghan and sampling sites

1.2 测量 OVOCs 和其他空气污染物

采用质子转移反应飞行时间质谱仪(proton-transfer-reaction time-of-flight mass spectrometry, PTR-TOF-MS, 奥地利 IONICON 公司) 进行 VOCs 的在线观测, 重点定量观测乙醛、丙酮、异戊二烯、MEK、MVK + MACR、苯、甲苯、苯乙烯、C8 芳香烃和 C9 芳香烃等 10 种 VOCs 的浓度. PTR-TOF-MS 的基本工作原理是利用 VOCs 的质子亲和性,以 H_3O^+ 作为母离子在电场作用下进入漂移管,并同 VOCs 发生离子-分子反应,即质子转移反应,产生子离子 RH^+ . 然后反应剩余的 H_3O^+ 和反应生成的 RH^+ 一起进入漂移管后端的质谱检测器进行检测. 关于 PTR-TOF-MS 的更多细节和介绍见文献[24,25].

VOCs 的采样为直上直下式,即在 PTR-TOF-MS 进气端口设置 1 个 Teflon 材质的膜托,过滤空气中的颗粒物,以免损伤仪器,干扰结果. 使用外径 1/4"的 Teflon 管作为进气管路,采样气体在流量为 2 L·min $^{-1}$ 的引流泵作用下进入采样管路,调节 PTR-TOF-MS 进样口流量保持在 $200 \sim 300 \text{ mL·min}^{-1}$,气体经过 PEEK 管 (poly-ether-ether-ketone)时加热到 80°C进入漂移管中,离子压强为 220 Pa,离子源电流为 3 mA, E/N 为 142 Td (1 Td = 10^{-17} V·cm²). E/N 值在离子迁移率研究中经常使用,其中 E 为电场, N

为漂移管中气体的数密度,值的变化会影响分子簇 离子的分布^[25,26];提高漂移管中 E/N 值,可以使离 子与空气分子间具有更高的碰撞动能,有效地减少 分子簇离子的生成. 但是,平均碰撞动能的提高也会 增加质子转移反应生成离子的断裂可能性,从而生 成较多的离子碎片,增加质谱谱图的分析难度.因此 实际操作中,需要权衡这两点来选择合适 E/N 值, 一般实际操作中 E/N 值保持在 $100 \sim 140 \text{ Td}^{[27]}$. 采 样期间选择 H,O+离子源模式,离子源取于美国艾 科浦(Aquaplore)纯水器产生的超纯水(电阻率: 18. 25 MΩ·cm, 25℃). 采用 PAMS 和 TO15 校准气 体(美国 Spectra Gases 公司)对 PTR-TOF-MS 进行 标定,每周至少进行1次.采样气路中所用的 Teflon 膜托、不锈钢接头以及 Teflon 采样管等进口于美国 Swagelok 公司,采样前分别用超纯水超声 3 次,每次 15 min.

由于 VOCs 在 PTR-TOF-MS 中和 H₃O+离子的 质子转移反应速率常数、反应时间和 H₃O+离子强 度都可以计算或测量,理论上可以根据仪器测量直接计算出所测 VOCs 各物种的浓度,更多详情见文献[25,28]. 然而,这种直接计算得到的物种浓度会存在较大的误差,主要来源于:①反应速率常数 k 的误差;②反应温度的不确定性;③离子通过漂移管传输时间和离子传输率的不确定性;④各仪器离子碎片的不确定性等. 因此,目前普遍应用标准气体标定 PTR-TOF-MS 来对物种进行定量分析,考虑到湿度的影响和不断变化的 H₃O+强度,将产物离子RH+的信号标准化为10⁶ counts·s⁻¹的标准信号,得到标准化离子信号,再根据通人标气建立的标准曲线计算得到 VOCs 浓度.

空气质量数据包括: NO、 NO_2 、 SO_2 、 O_3 和 CO,分别由热电公司 i 系列气体监测仪 42i、43i、48i 和 49i 测得,气象数据包括温度、湿度、风向和风速,由 Lufft WS501-UMB 气象仪获得.

1.3 ·OH损耗速率和臭氧生成潜势 OFP

大气中的 NMHCs 和 OVOCs 的化学反应活性差异非常大,且臭氧的生成潜势也存在差异. 本研究采用·OH损耗速率(·OH loss rate, $L_{.OH}$)来反映污染物在大气中的化学反应活性,采用 VOCs 最大增量反应活性(maximum incremental reactivity, MIR)量化 VOCs 的臭氧生成潜势(ozone formation potentials, OFP).

$$L_{\cdot \text{OH},i}$$
可以用以下公式(1)计算^[29]:

$$L_{\cdot \text{OH},i} = k_{\cdot \text{OH},i} \times [\text{VOC}]_i \tag{1}$$

式中, $L_{.OH,i}$ 表示 VOC 物种 i 的 · OH 损耗速率, s^{-1} ; $k_{.OH,i}$ 是 VOCs 物种 i 同单位浓度的 · OH的反应速率

常数, cm³·(molecule·s) $^{-1}$, 计算时 $k_{.OH,i}$ 参考文献 [30]; [VOC]_i 是 VOC 物种 i 的环境浓度, molecule·cm $^{-3}$, 实测 VOC 浓度单位为× 10^{-9} (体积分数),代入公式计算时需换算. 目前对 $L_{.OH}$ 的研究主要集中在烷烃、烯烃、芳香烃和 OVOCs.

OFP 可用以下公式计算^[29],如式(2):

$$OFP_i = MIR_i \times [VOC]_i$$
 (2)

式中,OFP_i 表示 VOC 物种 i 的臭氧生成最大潜势; (VOC]_i 是 VOC 物种 i 的环境浓度,×10⁻⁹,而实测 VOC 浓度单位为×10⁻⁹,直接代入公式计算; MIR_i 是 VOC 物种 i 的最大增量反应活性,g·g⁻¹(以 O₃/ VOC 计), MIR 值参考文献[31].

1.4 OVOCs 的源解析方法——光化学年龄参数法

本研究主要利用 De Gouw 的光化学年龄参数 法对成都平原 OVOCs 源解析^[32]. 该方法把 OVOCs 的来源主要分为人为一次源、人为二次源、生物源 和背景源,如式(3):

$$[\text{OVOC}] = \text{ER}_{\text{OVOC}} \times [\text{Tracer}] \times \exp [- (k_{\text{OVOC}} - k_{\text{Tracer}}) [\cdot \text{OH}] \Delta t] + \text{ER}_{\text{precursor}} \times [\text{Tracer}]$$

$$\times [k_{\text{precursor}} / (k_{\text{OVOC}} - k_{\text{precursor}})] \times$$

$$[\exp (- k_{\text{precursor}} [\cdot \text{OH}] \Delta t) -$$

$$\exp (- k_{\text{OVOC}} [\cdot \text{OH}] \Delta t) /$$

$$\exp (- k_{\text{Tracer}} [\cdot \text{OH}] \Delta t)] + \text{ER}_{\text{biogenic}} \times$$

$$(\text{isoprene}_{\text{source}}) + [\text{background}]$$

$$(3)$$

式中,ERovoc和 ERprecursor分别表示 OVOC 和前体物 相对于示踪物的排放速率, ER historic 表示生物排放的 OVOC 相对于异戊二烯源浓度(isoprene source)的排放 速率.「OVOC]、「Tracer]和[background]分别表示 环境中 OVOC 的浓度、示踪物的浓度和环境背景 值. kovoc、kracer和 kprecursor分别表示 OVOC、示踪物和 前体物与·OH反应的速率常数,其中 k_{ovoc} 、 k_{Tracer} 取 值于 Atkinson 等^[30]的研究. [•OH]Δt 表示•OH随时 间变化的暴露水平, isoprene_{source}和[·OH] Δt 分别由 异戊二烯的浓度和它的光化学产物 MVK 和 MACR 获得[33,34]. 需要指出的是,这里只考虑了异戊二烯 与·OH的反应,没有考虑到大气稀释作用,未考虑异 戊二烯与 NO₃ 的去除机制,只适用于白天情景.在 此方法中,一个关键性的因素是示踪物的选择.有研 究用 CO 或乙炔作为人为源的示踪物[32,35],而四川 盆地苯的来源没有乙炔的来源复杂[36],因此本研究 采用苯代替乙炔作为人为源的示踪物,此外,类似处 理方法在珠三角的研究中也有采用[20].

 ER_{OVOC} 、 $ER_{precursor}$ 、 $ER_{biogenic}$ 、 $k_{precursor}$ 和 [background]由公式(3)中利用最小二乘拟合法获得,该方法基于以下假设:①OVOCs 的人为源排放、前体物和一次排放示踪物成正比;②OVOCs 的去除

过程主要是与·OH反应; ③OVOCs 的生物源与异戊二烯排放成正比; ④可确定采样气团的光化学年龄^[32].

2 结果与讨论

2.1 VOCs 浓度水平

利用 PTR-TOF-MS 测量了成都平原地区 VOCs 环境浓度(表1和图2),常规污染物和气象参数(图 3). 结果表明 10 种主要 VOCs 总浓度(体积分数)为 (10.97 ± 4.69) × 10⁻⁹, 其中含氧挥发性有机物浓 度为(8.54±3.44)×10⁻⁹,芳香烃为(1.53±0.93) ×10⁻⁹,生物源 VOCs 为(0.90±0.32)×10⁻⁹. 丙酮 的浓度最高(4.11 ± 1.42) × 10^{-9} , 其次是乙醛 (2.26±1.01)×10⁻⁹和苯乙烯最低(0.14±0.08) ×10⁻⁹. 丙酮、乙醛、MEK 和 MVK + MACR 等 4 种 OVOCs 所占比例最大, 达77.85%. 苯、甲苯、苯乙 烯、C8 芳香烃及 C9 芳香烃等 5 种芳香烃和异戊二 烯(BVOCs)所占比例分别为 13.95% 和 8.20%.由 此可见, 3个OVOCs(丙酮、MEK和乙醛)是本次观 测中的优势物种.5个芳香烃物种同人为一次源示 踪物 CO 的时间序列趋势相似(图 4),存在明显的 一次排放特征.

如表2,典型城市和城郊中,OVOCs浓度从高到 低排列为丙酮、乙醛和 MEK, 与本研究相似. 将成 都平原与北京城郊相比,乙醛浓度基本相当(2.26 ×10⁻⁹),但成都平原丙酮(4.11×10⁻⁹)和 MEK (1.62×10⁻⁹)浓度明显较高,分别是北京郊区的近 2 倍 [2.33 × 10⁻⁹ (2.33 ppb)] 和近 3 倍 [0.52 × 10⁻⁹(0.52 ppb)][37],这可能与成都平原周边地区 分布大量的涂料、油漆工业有关,丙酮、MEK 是这 些工业上常用的溶剂. 与广州相比, 乙醛浓度, 城市 乡镇分别为 5.89×10⁻⁹(5.89 ppb)和 4.25×10⁻⁹ (4.25 ppb),是成都平原的近2倍,有可能是广州站 点受到交通排放的影响比成都平原大,导致其浓度 高,有研究指出城市区域乙醛受机动车排放的影 响[38]; 而丙酮相差不大,但 MEK 浓度,城市乡镇分 别为「0.10 × 10⁻⁹ (0.10 ppb)]和「0.26 × 10⁻⁹ (0.26 ppb)],远远低于成都平原^[23,39]. 成都平原与 成都市区 OVOCs 浓度对比而言,丙酮和乙醛浓度变 化不大,成都平原 MEK 浓度高于市区[0.75×10⁻⁹ (0.75 ppb)]^[36],也高于其它地区以往测量 值[23,37,39,40],说明成都平原区域受到溶剂源影响的 态势不容忽视,而本次研究中 OVOCs 浓度水平比其 他城市略高.

2.2 日变化

大气污染物的日变化特征是气象、排放和大气

表 1 挥发性有机物的浓度水平1)

TC 11 1	C	1 1 6 1 . 1	organic compounds
Table I	Loncentration	level of volatile	organic compounds

物种	类别	平均值 ± 标准差 × 10 ⁻⁹	最小值×10-9	最大值×10-9
丙酮	OVOC	4. 11 ± 1. 42	0. 98	14. 22
乙醛	OVOC	2.26 ± 1.01	0. 74	22. 57
MEK	OVOC	1.62 ± 0.44	0. 85	6. 26
MVK + MACR	OVOC	0.55 ± 0.30	0. 15	2. 20
异戊二烯	Biogenic	0.90 ± 0.31	0.48	5. 30
苯	Aromatic	0.18 ± 0.07	0.09	1. 19
甲苯	Aromatic	0.25 ± 0.12	0.09	4. 05
苯乙烯	Aromatic	0.14 ± 0.08	0. 07	0.81
C8 芳香烃	Aromatic	0.56 ± 0.45	0. 13	3. 09
C9 芳香烃	Aromatic	0.40 ± 0.13	0. 13	1.02

1) MVK 和 MACR 被合并为同一个物种进行测量;C8 芳香烃包括邻-二甲苯、间-二甲苯、对-二甲苯和乙苯;C9 芳香烃包括 1,3,5-三甲苯、1,2,4-三甲苯和 1,2,3-三甲苯

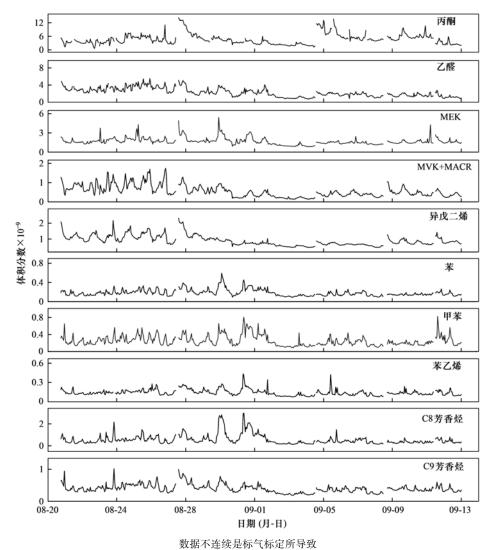


图 2 采样期成都平原夏季 10 个 VOCs 的时间变化序列

Fig. 2 Time series of 10 kinds of VOCs during the sampling periods in the Chengdu Plain in summer

化学综合反映的结果. 图 4 展示了区域内 CO、NO、 NO_2 、 O_3 和典型 VOCs 的日变化趋势. OVOCs(乙醛、丙酮和 MEK)物种一般呈现单峰或者多峰的特征,峰值一般出现在正午前后,与臭氧日变化趋势相似,并在 O_3 出现峰值之前,与深圳和北京观测到的特征一致 $[^{20,21}]$,且 OVOCs 的峰值与 CO 的谷值出现

时间一致,说明了 OVOCs 的排放不仅受到一次人为源的影响,还受到 NMHCs 的二次氧化反应生成的影响.3 个 OVOCs 物种在 20:00 之后出现上升趋势,可能由于夜间 NMHCs 与 NO₃、臭氧的清除作用和光化学反应减弱等因素导致^[43],也可能是本地人为源.

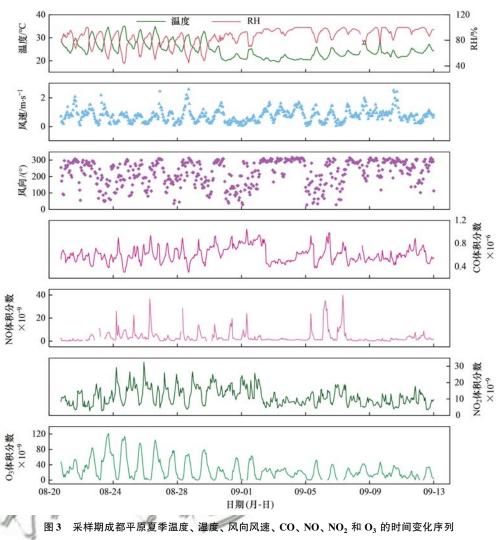


Fig. 3 Time series of temperature, humidity, wind direction and speed, CO, NO, NO₂, and O₃ during the sampling periods in the Chengdu Plain in summer

表 2 不同城市 OVOCs 浓度比较 × 10 -9

Table 2 Comparison of OVOC concentrations among the different sites $\times 10^{-9}$

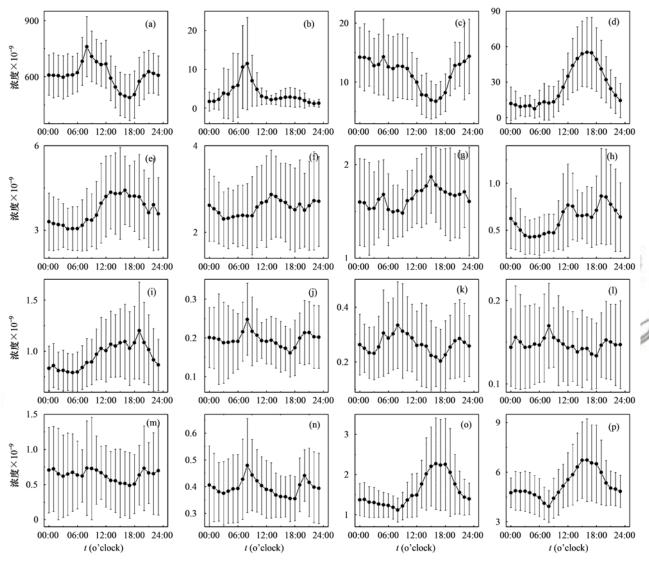
地点	类型	日期(年-月)	乙醛	丙酮	MEK	方法	文献
佩特雷(希腊)	城市	2012-06	_	2. 93	0.30	PTR-MS	[40]
雅典(希腊)	城市	2012-07	_	4. 28	0.50	PTR-MS	[40]
北京	城市	2008-07 ~ 2008-08	1. 0	3.96	1.16	PTR-MS	[41]
北京	城市	2010-06 ~ 2010-09	4.02 ± 1.44	2.90 ± 1.06	0.65 ± 0.22	DNPH/HPLC	[37]
北京	城郊	2010-06 ~ 2010-09	2.33 ± 0.92	2.34 ± 0.76	0.52 ± 0.22	DNPH/HPLC	[37]
深圳	城市	2016-08 ~ 2016-09	1.98 ± 1.53	3.78 ± 2.35	1.33 ± 1.22	PTR-MS	[21]
深圳	城市	2011-05 ~ 2011-09	1.56 ± 1.36	6.26 ± 6.86	0.95 ± 1.03	PTR-MS	[42]
广州	乡镇	2011-08	4. 25	0. 59	0. 26	GC-FID/MS	[23]
广州	城市	2005-07	5. 89	7. 21	0. 10	DNPH/HPLC	[39]
成都	城市	2016-09	2. 04	4. 70	0.75	GC-FID/MS	[36]
成都	城郊	2019-08 ~ 2019-09	2.26 ± 1.15	4. 11 ± 1. 48	1.62 ± 0.5	PTR-MS	本研究

芳香烃类各物种日变化与 CO 和 NO₂ 较为一致,共两个峰值.第一个峰值出现在 08:00 左右,和 NO 峰值相似,与早间上班高峰期相符,为典型的交通源排放特征;第二个峰值在 20:00,与人为源排放示踪物 CO 类似,但是在机动车尾气排放的晚高峰时段没有明显的高峰,因此,芳香烃的峰值可能来

自于附近的溶剂使用如喷漆等过程,以上特征表明该区域受到一次源机动车尾气和溶剂挥发使用的影响.

异戊二烯主要由植物在白天排放,释放强度与温度和太阳辐射有关;其活性较高易被·OH、NO₃和 O₃氧化.白天,随着光照增加,异戊二烯和

·OH的消耗速度逐渐增加,造成异戊二烯在13:00 左右达到峰值,此特征与深圳的观测结果类似.异 戊二烯的浓度在傍晚19:00 出现一个小高峰.图 4 (o)和图 4(p)显示异戊二烯/CO、异戊二烯/苯均 在傍晚时分呈上升趋势,说明浓度变化与边界层 高度变化无关;可能是傍晚温度下降,光化学反应的去除速率降低,异戊二烯浓度的堆积. MVK + MACR 存在 2 个峰值,它作为异戊二烯的氧化产物,两个峰值符合氧化产物的峰值特征,出现时间均在异戊二烯峰值之后.



(a) CO; (b) NO; (c) NO₂; (d) O₃; (e) 丙酮; (f) 乙醛; (g) MEK; (h) MVK + MACR; (i) 异戊二烯; (j) 苯; (k) 甲苯; (1) 苯乙烯; (m) C8 芳香烃; (n) C9 芳香烃; (o) 异戊二烯/CO 和(p) 异戊二烯/苯

图 4 夏季成都平原 CO、NO、NO₂、O₃ 和 VOCs 的日变化特征

Fig. 4 Mean diurnal variations of CO , NO , NO_2 , O_3 , and VOCs in the Chengdu Plain during summer

2.3 ·OH损耗速率和臭氧生成潜势 OFPs

臭氧生成与前体物(NO_x 和 VOCs) 呈高度非线性关系,并且同气象条件和污染源排放有关.本研究 VOCs 的 $L_{.OH}$ 范围为: $0.01 \sim 2.21 \text{ s}^{-1}$,结果见表 3. PTR-TOF-MS 检测的所有物种中,活性最强的是异戊二烯,占比为 58.83%; 其次是 3 个 OVOCs 物种,占比为 23.39%; 最低的是 5 个芳香 烃物种(17.77%). 丙酮为本次观测的优势物种,但其活性较低(0.02 s^{-1}); 乙醛由于其高·OH反应速率,加上浓度较高,故在 OVOCs 物种中活性最大(0.81

s⁻¹). 苯、甲苯和 MEK 由于低活性和低浓度导致 OFP 活性不高, 故本地生物源活性占主导.

有研究表明,北京^[20]、深圳^[20]、上海^[44]、广州^[44]和重庆^[45]属于 VOCs 控制区. 如图 3 和图 4 所示,O₃ 与 NO_x 的时间序列变化和日变化趋势呈负相关,表明该区域也可能属于 VOCs 控制区,臭氧生成对 VOCs 浓度比较敏感. 图 5 是光化学反应初期(10:00~11:00,此时是臭氧浓度上升最快的阶段)和白天(07:00~19:00,因该区域臭氧夜晚浓度较低,所以仅考虑白天 VOCs 臭氧生成潜势)的 VOCs 臭氧生

成潜势和污染物浓度. $10:00 \sim 11:00$ 和白天的总 OFP (TOFP)分别为 43.61×10^{-9} 、 39.51×10^{-9} ,表明光化 学初始阶段臭氧潜势最高. $3 \land OVOCs$ 物种臭氧生成潜势 最高,占比 47.01%,其次是 $5 \land 芳$ 香烃 (27.98%),最后是异戊二烯(25.01%). 从单个物种来看,臭氧生成潜势最高的为乙醛,其次是异戊二烯,

共占 TOFP 的 62. 89%,并且这两个物种在光化学初始 阶段 TOFP 的贡献明显大于白天,最低的是苯. 而深 圳地区四季排前 2 位 OFP 的物种为乙醛和甲苯^[21],成都夏季 OFP 主要贡献物种是烯烃和芳香烃^[46],说明成都平原区域臭氧生成以 OVOCs 和生物源为主,但芳香烃亦存在重要作用.

表 3 成都平原夏季 VOCs 的 $\cdot OH$ 损耗速率 $(L_{\cdot OH})$

Table 3 The \cdot OH loss rate($L_{\cdot,\mathrm{OH}}$) of VOCs in the Chengdu Plain in summer

物种	分类	$k(289\text{K}) \times 10^{12}$ /cm ³ ·(molecule·s) ⁻¹	MIR(O ₃ /VOC) /g·g ⁻¹	$L_{\rm \cdot OH}/{\rm s}^{-1}$
乙醛	OVOC	15	6. 54	0.81
丙酮	OVOC	0. 17	0.36	0.02
MEK	OVOC	1. 22	1.48	0.05
异戊二烯	Biogenic	100	10. 61	2. 21
苯	Aromatic	1. 22	0.72	0.01
甲苯	Aromatic	5. 63	4	0.04
苯乙烯	Aromatic	58	1.73	0. 20
C8(间-二甲苯)1)	Aromatic	23. 1	6. 57	0.31
C9(1,2,3-三甲苯)2)	Aromatic	11.8	2. 52	0.12
总计		0		3.75

1)利用间-二甲苯的 k 值代表 C_8 芳香烃类的 $L_{.0H}$ 和 OFP; 2)利用 1,2,3-三甲苯的 k 值代表 C_9 芳香烃类的 $L_{.0H}$ 和 OFP

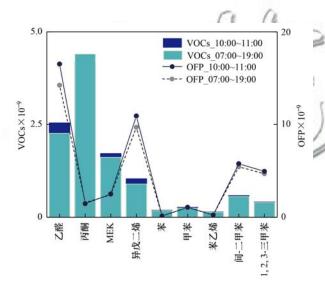


图5 光化学反应初期(10:00~11:00)和白天(07:00~19:00)的 VOCs 臭氧生成潜势和污染物浓度

Fig. 5 Mixing ratios of the measured VOCs and calculated OFPs at the initial stage of photochemical reaction (10:00-11:00) and during the whole day (07:00-19:00)

2.4 OVOCs 来源解析

本研究选取异戊二烯和其光化学氧化产物 MVK + MACR 计算光化学年龄,具体细节见文献 [34], ·OH 的 平 均 浓 度 取 2.56 × 10⁶ molecule·cm^{-3[47]},得出夏季成都平原地区 OVOCs 气团光化学年龄最大值为 1.75h,平均值(0.87 ± 0.20)h,De Gouw 等^[32]用甲苯/苯比值法计算英国整个气团的光化学年龄超过 40 h,比成都平原地区光化学年龄高的原因是选取参照物不一致,甲苯和苯在大气中化学性质较稳定,而异戊二烯较活泼,用

异戊二烯及其氧化产物计算的 OVOCs 光化学年龄 更具有参考意义. 对于 OVOCs 短寿命物种,采用 POLARPLOT 解释近局地 OVOCs 来源和臭氧生成 (图 6). NO 和 NO,来源于本地,而臭氧、乙醛、丙 酮和 MEK 均主要来自东南方,根据气团老化时间和 风速(本地风速最高为 2.59 m·s⁻¹,矢量平均风速 为 0.43 m·s⁻¹),是来源于广汉市的局地传输,采样 点臭氧的生成主要受局地污染物的影响. 乙醛主要 来自东北方:根据 OFP 结果显示,本地乙醛臭氧的 生成潜势最高,合理推测乙醛是东北方臭氧的重要 前体物. 西北-偏西方向的臭氧存在明显的长距离区 域特征,该方向无明显人为源,而生物来源较为重 要. 西南-偏南方向存在 OVOCs 和臭氧长距离污染 特征,该方向主要受成都市区影响,由于 NO,浓度由 城市至农村逐渐降低,造成臭氧在下风向累积.采样 点的光化学污染同时受局地城市和区域污染物的 影响.

利用光化学年龄参数法分析了成都平原地区 OVOCs 白天(07:00~19:00) 的污染来源,结果见表 4. 拟合结果和实测浓度的相关性较好(r 为 0. 69~0.83,P<0.05).式(3)的5个参数中,丙酮主要对 ER_{OVOC} 敏感,MEK 对 $ER_{precursor}$ 和 $k_{precursor}$ 敏感,乙醛对 ER_{OVOC} 、 $ER_{precursor}$ 和 $k_{precursor}$ 敏感,上述结果可能低估了 OVOCs 的人为源贡献,非本地 OVOCs 前体物可能在传输过程中大幅消耗.本研究表明成都平原地区乙醛、丙酮和 MEK 的主要来源是生物源和人为二次源,累积贡献达到 58.59%~79.84%;其中丙

表 4 成都平原地区 OVOCs 不同来源的占比情况

Table 4	Percentage	contributions	from	different	sources	to	OVOCs	in	the	Chengdu	Plain
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				,	
物	J种 r	人为一次源	/% 人为二次源/%	生物源/%	背景源/%
Z	.醛 0.83	9. 59	10. 96	68. 88	10. 58
丙	i 酮 0.69	12. 87	5. 87	52. 72	28. 55
M	EK 0. 72	26. 02	27. 82	35. 42	10. 74

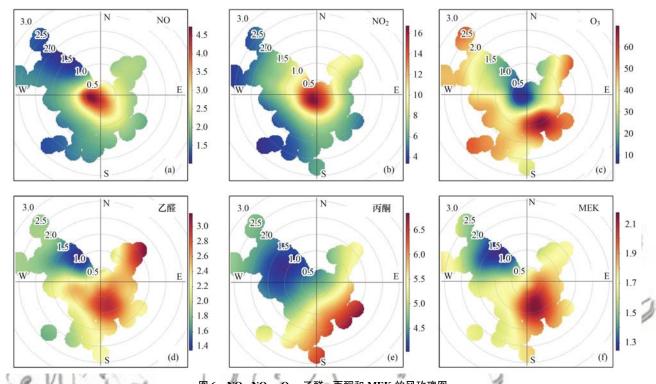
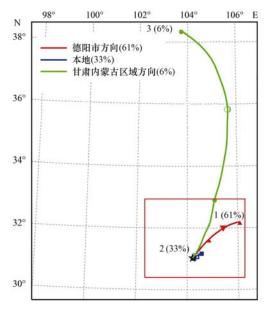


图 6 NO、NO₂、O₃、乙醛、丙酮和 MEK 的风玫瑰图

Fig. 6 Polar plots of NO, NO₂, O₃, acetaldehyde, and acetone

酮的背景源达到28.55%,且可能受区域传输影响,



括号中的值表示频率

图 7 采样期间成都平原的气团 48 h 后向轨迹图

Fig. 7 Mean 48-h back trajectories of clusters in the Chengdu Plain during the sampling periods

导致其背景浓度高于乙醛和 MEK. 而 MEK 的人为一次源(26.02%)和二次源(27.82%)较其他物种高,说明 MEK 受到人为源和生物源的共同影响.

为了进一步探讨成都平原地区 OVOCs 的高区域背景水平的影响,本文采用了 NOAA 的混合单粒子拉格朗日综合轨迹模拟系统(hybrid single particle Lagrangian integrated trajectory, HYSPLIT)来计算出采样期间污染物的后向轨迹(图7).利用软件计算每小时广汉离地 100 m 处 2 d 的反向轨迹,并根据空间分布相似性将结果分类为 3 簇,分别为德阳市方向 61%、本地方向 33%和甘肃内蒙古区域方向6%,其中来自甘肃内蒙古区域方向的气团较为洁净.综上所述,OVOCs 受到区域人为源和生物源的共同影响,在制定 0₃ 和 VOCs 控制政策时,应采取区域尺度的联合防控策略.

3 结论

(1) 本文在 2019 年夏季利用 PTR-TOF-MS 在成都平原在线监测了 4 个 OVOCs (乙醛、丙酮、MEK、MVK + MACR) 和其他 NMHCs (异戊二烯、

- 苯、甲苯、苯乙烯、C8 芳香烃、C9 芳香烃)的环境浓度. 10 个 VOCs 总浓度为 $(10.97 \pm 4.69) \times 10^{-9}$,其中含氧挥发性有机物(0VOCs)浓度为 $(8.54 \pm 3.44) \times 10^{-9}$,占比为 77. 85%,是本次观测中的优势物种;芳香烃和生物源 VOC 分别为 $(1.53 \pm 0.93) \times 10^{-9}$ 和 $(0.90 \pm 0.32) \times 10^{-9}$,分别占 13.95% 和 8.20%.
- (2) 丙酮和乙醛是该地的主要污染物;乙醛和异戊二烯是主要的 OFP 贡献物种,光化学活性也最强.乙醛、丙酮和 MEK 峰值出现时间正好是 CO 的谷值,说明 OVOCs 来源于二次源.光化学活性和臭氧生成潜势排名前三的物种均为:异戊二烯、乙醛和 C8 芳香烃,且光化学年龄参数法显示本地的生物源贡献更大,累计达到 35.42%~68.88%,说明成都平原区域受生物源影响较大.丙酮有显著的背景来源(28.55%),乙醛和 MEK 分别为 10.58%、10.74%,说明区域背景 OVOCs 的浓度对成都平原地区影响大,且敏感性分析表明低估了 OVOCs 的人为源,表明了成渝地区的区域污染水平高.
- (3) 综上结果可知,成都平原受到生物源和人为源影响较大,因生物源不好控,还是要降低人为源排放的影响,区域综合减排,臭氧和 PM_{2.5}协同控制. 参考文献:
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