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ENVIRONMENTAL SCIENCE

ISSN 0250-3301 CODEN HCKHDV HUANJING KEXUE

- 主办 中国科学院生态环境研究中心
- ■出版科学出版社





2020

Vol.41 No.10 第41卷 第10期

ENVIRONMENTAL SCIENCE

第41卷 第10期 2020年10月15日

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泉州山美水库及入库河流沉积物中多溴二苯醚的时空 分异和降解分析

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摘要:为了解城市水源水库中多溴二苯醚(PBDEs)的时空分异和同系物的降解来源及其贡献,分析了泉州山美水库及人库河流表层沉积物中 PBDEs 的含量、污染程度、空间分布、水文期变化、赋存量、同系物组成及其降解来源的贡献. 结果表明,人库河流沉积物中 \sum PBDEs 中值(1072.1 \log · g $^{-1}$)是山美水库(160.4 \log · g $^{-1}$)的 6.7 倍,山美水库单位面积沉积物中 \sum PBDEs 的赋存量(80.3 \log · kg · km $^{-2}$)是太湖的 6.3 倍,北美五大湖的 188 倍,其污染程度较国内外大多数湖库更严重,且以 BDE-209 为主 (84.5% ~ 99.2%). 水库大多数样点(r 为 0.564 ~ 0.994,P < 0.034)及河流各点(r 为 0.953 ~ 1.0,P < 0.000)间 PBDEs 组成相似度较高,入库区和入库河流样点间极显著正相关(r 为 0.779 ~ 0.964,P < 0.005)且相关性强于其他功能区,显示入库河流是水库中 PBDEs 的主污染源。库尾区与入库河流相关性较低(r 为 0.454 ~ 0.915,P < 0.128),受九都镇影响较大.各样点 \sum PBDEs 水文期变化较一致(r 为 0.617 ~ 0.714,P < 0.077),但水文期变化对 \sum PBDEs 的影响统计不显著(P = 0.178,Two-Way ANOVA),而点位变化则对 \sum PBDEs 有极显著影响(P = 0.0001),入库区和其他功能区有(近)显著差异(P 为 0.019 ~ 0.061),表明 PBDEs 在水库沉积物中的空间分布变异大于水文期变化。PBDEs 自然降解从河流到入库区再到库中区逐渐增加,且各级还原脱溴速率不同,部分 BDE 因其继续降解速率较慢而累积,丰度比值法研究表明,低溴 BDE 主要源自十溴二苯醚的逐级还原脱溴自然降解。Deca-BDE 降解产生的 Nona-BDE 约70%以上可较快降解生成 Octa-BDE,BDE-208 约85%源自 BDE-209 的降解,从 Octa-BDE 到 Penta-BDE 的降解过程中,部分 Octa-BDE 和 Hexa-BDE 同系物因降解较慢而累积,Penta-BDE 到 Tri-BDE 降解率在 70%以上.

关键词:多溴二苯醚; 沉积物; 时空分异; 降解; 山美水库; 入库河流; 水源水库

中图分类号: X52 文献标识码: A 文章编号: 0250-3301(2020)10-4525-14 DOI: 10.13227/j. hjkx. 202003096

Spatiotemporal Differentiation and Degradation Analysis of Polybrominated Diphenyl Ethers in Sediments of Shanmei Reservoir and Its Inflowing River, Quanzhou, China

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Abstract: To investigate the spatiotemporal differentiation of polybrominated diphenyl ethers (PBDEs) in urban water-source reservoirs and degradation sources of BDE homologues and their contributions, we analyzed the contents, pollution degrees, spatial distributions, hydrological period changes, inventories, profiles, and degradation source contributions of PBDEs in the surface sediments of Shanmei Reservoir and its inflowing river, Quanzhou, China. The results showed that the median \sum PBDEs (1 072. 1 ng·g⁻¹) in the inflowing river sediment was 6.7 times than that of the reservoir (160.4 ng·g⁻¹) and the total amount of \sum PBDEs in sediments per unit area (80.3 kg·km⁻²) was 6.3 times than that of Taihu Lake and 188 times than that of the Great Lakes in North America. The pollution degrees of PBDEs in Shanmei Reservoir were more severe than those of most lakes and reservoirs at home and abroad, which was dominated by BDE-209 (84.5%-99.2%). Most of the sampling sites in the reservoir (r 0.564-0.994, P<0.034) and the inflowing river (r 0.953-1.0, P<0.000) had high similarity in the composition of PBDEs. Significantly positive correlations (r 0.779-0.964, P<0.005) were observed between the reservoir entry area and river sampling sites, which were stronger than the other functional areas, indicating that the inflowing river was a major pollution source of PBDEs in the Shanmei Reservoir. The tail region of the reservoir had low correlations with the inflowing river (r 0.454-0.915, P<0.128), and was relatively much more affected by Jiudu Town. The changes in hydrological period of the \sum PBDEs were relatively consistent at each sampling site (r 0.617-0.714, P<0.077), but the impact of the changes in the hydrological period on the \sum PBDEs was not statistically significant (P=0.178, Two-Way ANOVA). However, the site changes had a significant influence on the \sum PBDEs (P=0.0001), and significant or nearly

收稿日期: 2020-03-09: 修订日期: 2020-04-13

基金项目: 国家自然科学基金项目(41203077); 福建省自然科学基金项目(2018J01065)

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differences were observed between the reservoir entry area and other functional areas (P 0.019-0.061), indicating that the spatial distribution variations of the PBDEs in reservoir sediments were greater than the changes in hydrological period. The natural degradation of the PBDEs gradually increased from the river to the reservoir entry area and then to the central reservoir area. The reductive debromination rates varied at different brominated levels, and some BDE homologues accumulated due to their slowly continued degradation velocities. Research on abundance ratios indicated that the lower brominated BDE homologues were mainly derived from the natural degradation of decabromodiphenyl ether by stepwise reductive debromination. Approximately 70% of Nona-BDE produced by Deca-BDE degradation could rapidly be degraded to form Octa-BDE. Approximately 85% of BDE-208 was derived from the degradation of BDE-209. During the degradation process from Octa-BDE to Penta-BDE, some Octa-BDE and Hexa-BDE homologues accumulated due to relatively slower degradation velocities, and the degradation rates of Penta-BDE to Tri-BDE were above 70%.

Key words: polybrominated diphenyl ethers (PBDEs); sediment; spatiotemporal differentiation; degradation; Shanmei Reservoir; inflowing river; water-source reservoir

多溴二苯醚(polybrominated diphenyl ethers, PBDEs)是全球用量最大的溴代阻燃剂(BFRs)之一^[1,2]. 近年来,PBDEs 主要商用品(五、八和十溴二苯醚)陆续被列入《斯德哥尔摩公约》的持久性有机污染物(POPs)清单^[3],在各类环境介质^[4~8]及食物^[9]中广泛存在.

PBDEs 在我国淡水水体等环境介质中的赋存量较大. Zhou 等^[10]估算太湖沉积物中 PBDEs 的总量约 30 t,且同系物以 BDE-209 为主(>80%). Wang 等^[11]报道太湖沉积物中 PBDEs 含量(3.8~347 ng·g⁻¹,平均 72.8 ng·g⁻¹)远高于六溴环十二烷(HBCD)和四溴双酚 A(TBBPA)等其他 BFRs. 其中,太湖梅梁湾和贡湖湾等靠近注入河流的沉积物中 PBDEs 含量高于其他点位,表明河流注入是太湖中 PBDEs 的重要来源.

沉积物中 BDE-209 可通过改变磷矿化细菌的群落组成^[12,13]和促进碱性磷酸酯酶的活性,从而减少沉积物中的有机磷,并增加水体和沉积物中的生物可利用磷^[14],进而改变湖库中磷的生物地球化学循环,加重湖库的富营养化. BDE-209 及其还原脱溴生成的低溴 BDE^[15~17]等二次污染物对生物体的内分泌、神经、生殖^[18,19]和遗传系统^[20,21]及智力^[22]等有潜在的毒害作用.

虽然 BDE-209 易于脱溴降解^[23],但其在环境介质中彻底降解为完全脱溴产物二苯醚 (diphenylether,DE)是一个漫长的过程^[12,19,24,25].环境介质中的低溴 BDE 究竟主要来自五溴和八溴二苯醚商用品的直接使用、或是高溴 BDE 的降解,另外其贡献率如何,这些目前尚不清楚,相关研究也十分缺乏,直接影响了人们对 PBDEs 同系物来源的正确认识,以及对十溴二苯醚商用品环境风险的准确评估.

供水安全是城市经济社会持续发展的基石,作为至2019年末GDP连续21年全省第一的地级市,泉州人均水资源量仅为福建全省人均的1/3,是典型的沿海缺水型城市.泉州地处闽南地区,该地区总面积仅占福建的1/5,GDP却占福建的1/2.笔者前

期的研究发现,泉州市环境中 PBDEs 含量在闽南地区最高,这与其制造等产业发达,溴代阻燃剂用量较多有关.另一方面,泉州人口稠密(2019 年末常住人口 874 万,居全省首位),PBDEs 潜在的环境和健康风险值得关注.

山美水库是泉州首要的饮用水源地,位于泉州市第一大江——晋江上游东溪中段,是晋江流域唯一集供水、灌溉、防洪、发电和生态等综合利用于一体的大型水利枢纽工程,流域面积1023 km²,总库容6.55亿m³,担负着晋江下游600多万人口的生活和生产用水^[26].2018年8月5日,源自山美水库的水直供金门,开启了两岸"共饮一江水"的时代^[27].近年来,库区水质随流域社会经济快速发展而明显下降,总氮等甚至为劣V类^[26].

现有城市水源水库的报道多集中于库区无机或有机污染物的组成和空间分布等,且大多为一次采样,缺乏对水库及入库河流沉积物中典型 POPs 的时空分异和降解来源及其贡献的研究. 就闽南地区而言,湖库研究多集中于氮磷等生源要素方面,PBDEs 等 POPs 的报道稀少,且集中于海洋沉积物与海洋生物方面^[28],而与城市供水安全和人体健康息息相关的水源水库中 PBDEs 时空分异和降解来源的研究尚未见报道.

基于以上讨论,笔者采集典型城市水源水库——泉州山美水库及入库河流代表性点位的表层沉积物样品,研究 PBDEs 的含量、污染程度、空间分布、水文期变化、赋存量、同系物组成及其降解来源的贡献,以期为城市水源水库中 PBDEs 的污染治理提供基础资料,并为水环境中 PBDEs 同系物的自然降解来源研究提供科学依据.

1 材料与方法

1.1 采样

使用 GPS 定位,抓斗采样器采集表层沉积物. 山美人库河流采集了 5 个样品(图 1),时间为 2013 年 9 月 13 日. 山美水库采集了 26 个样品(图 2,丰 水期 S8 样品损失),时间分别为 2012 年 7 月 1 日 (丰水期)、2012 年 11 月 4 日(枯水期)和 2013 年 3 月 9 日(平水期).其中,采样点 S1 和 S2 位于人库区,S3 和 S4 位于库尾区,S5、S6 和 S7 位于库中区,S8 和 S9 位于坝前区,入库区连接入库河流,库尾区为闭合区域,紧邻九都镇,坝前区位于水坝上游.

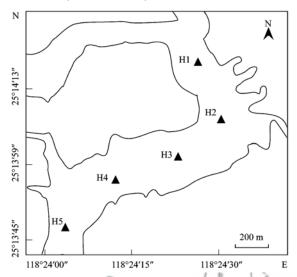


图1 山美入库河流沉积物采样点示意

Fig. 1 Map of sediment sampling sites in the inflowing

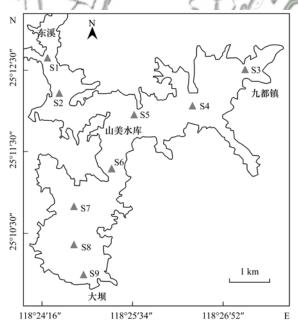


图 2 山美水库沉积物采样点示意

Fig. 2 Map of sediment sampling sites in Shanmei Reservoir

沉积物样品采集后装入洁净铝箔袋(马弗炉中450℃,4h)中密封,外套聚乙烯自封袋,带回实验室置于超低温冰箱(<-70℃)中保存.用冻干机-50℃下72h冻干后研磨过100目筛,置于洁净铝箔袋中密封,外套聚乙烯自封袋密封,于冰箱中冷冻保存.

1.2 样品预处理与分析

1.2.1 标样与试剂

PBDEs 标样购自 Accustandards (New Haven, CT, USA), ¹³C₁₂-CB-141、¹³C₁₂-CB-208 和¹³C₁₂-CB-209 购自 Cambridge Isotope Laboratories (Andover, MA, USA). 所有溶剂为分析纯或更优(上海国药集团化学试剂有限公司),经全玻璃系统重蒸后使用.

1.2.2 样品预处理和仪器分析

样品预处理和仪器分析方法基于已有方法修改 而来[5,6]. 简述如下:四分法分样称取沉积物样品10 g,用抽提过的洁净滤纸包好,烧瓶中加入预先活化 的铜片和 200 mL 丙酮/正己烷(1:1) 索氏抽提,抽 提液旋蒸浓缩后,置换溶剂为正己烷,过多层硅胶/ 氧化铝柱净化,用70 mL二氯甲烷/正己烷(1:1)淋 洗,淋洗液旋蒸浓缩后转入1.5 mL 棕色玻璃瓶中, 氮吹置换溶剂为正己烷,定容至50 μL. 样品进样前 加入适量 ¹³C₁₂- CB-208 内标, 用安捷伦 GC-MS (Agilent 7890N GC/5975 MS) 负化学电离(NCI) 选 择离子扫描(SIM)法测定 PBDEs 的组成. 扫描离子 3~7 溴 BDE 为 79 和 81, 8~9 溴 BDE 为 79、81、 408.7 ± 2、486.7 和 488.7, BDE-209 为 79、81、 486.7 和 488.7. 内标¹C₁₂-CB-208 为 475.8 ± 2. 回收 率指示物¹³C₁₂-CB-141 和¹³C₁₂-CB-209 分别为 371.9 ±2 和 509.7 ± 2. 目标化合物为 BDE-17、BDE-28、 BDE-71, BDE-47, BDE-66, BDE-77, BDE-100, BDE-99、BDE-85、BDE-118、BDE-154、BDE-153、 BDE -138 , BDE-183 , BDE-190 , BDE-197 , BDE-203、BDE-196、BDE-208、BDE-207、BDE-206 和 BDE-209,共22种.内标法五点校正曲线定量.仪器 检测限(IDL)用约5倍信噪比(S/N)的标样连续测 定 6~8 次,取其标准差的 3.36 倍作为 IDL, 3~7 溴 BDE 为 0.41~0.75 pg, BDE-209 为 3.74 pg. 空白 样品中只有少量的 BDE-47 被检出,含量小于大多 数样品的5%. 流程空白加标回收率3~7 溴 BDE 为 105.0% ± 4.7%, BDE-209 为 86.8% ± 17.1%; 基 质空白加标回收率 3~7 溴 BDE 为 103.1% ± 5.1%, BDE-209 为 70.7% ± 1.7%. 回收率指示物 ¹℃₁₂-CB-141、¹℃₁₂-CB-209 的回收率分别为 106.3% ± 16.3% 和 97.4% ± 13.4%. 数据未经回收率 校正.

2 结果与讨论

2.1 沉积物中 PBDEs 的含量和污染程度

人库河流和山美水库沉积物中 BDE-209 中值含量(1 067.2 $\operatorname{ng} \cdot \operatorname{g}^{-1}$, 155.9 $\operatorname{ng} \cdot \operatorname{g}^{-1}$)均比 $\sum_{21} \operatorname{PBDEs}$ (4.92 $\operatorname{ng} \cdot \operatorname{g}^{-1}$, 3.15 $\operatorname{ng} \cdot \operatorname{g}^{-1}$, 除 BDE-

209 外的 21 种 BDE 含量之和)高 2~3 个数量级,表明山美水库流域的 PBDEs 以 BDE-209 为主. 入库河流沉积物中 \sum PBDEs (22 种 BDE 含量之和,含BDE-209)中值含量(1072.1 ng·g⁻¹,均值1406.7 ng·g⁻¹)是山美水库中值(160.4 ng·g⁻¹,均值 267.7 ng·g⁻¹)的 6.7 倍,显示人库河流是山美水库的主污染源.

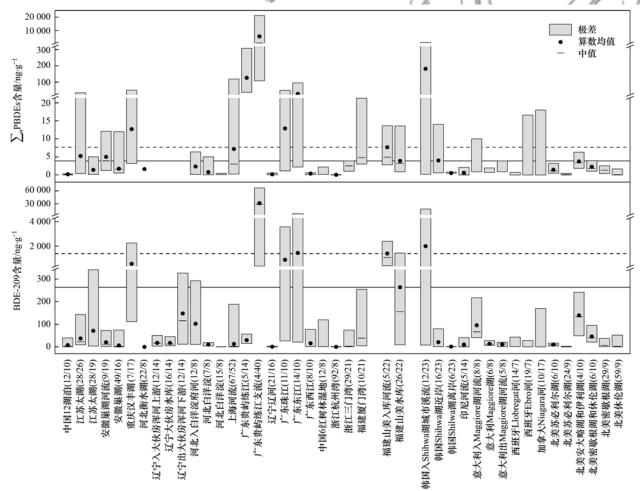
注入湖库的河流因河床侵蚀和接纳各种污染排放,通常是湖库中 PBDEs 等 POPs 的首要污染源 $[^{29]}$. 人巢湖河流沉积物中 BDE-209 中值含量 $(12.4~\mathrm{ng}\cdot\mathrm{g}^{-1[30]})$ 是巢湖沉积物($2.5~\mathrm{ng}\cdot\mathrm{g}^{-1[31]}$)的 $5.0~\mathrm{fe}(\mathbb{B}\,3)$. 流经保定市区的府河沉积物中 BDE-209 含量算数均值 $(102.6~\mathrm{ng}\cdot\mathrm{g}^{-1}$,未报道中值,下同)是其注人的白洋淀 $(10.4~\mathrm{ng}\cdot\mathrm{g}^{-1})$ 的 $9.9~\mathrm{fe}^{[32]}$. 韩国人 Shihwa 湖城市溪流中 \sum PBDEs $(2~180~\mathrm{ng}\cdot\mathrm{g}^{-1})$ 是 Shihwa 湖近岸 $(25~\mathrm{ng}\cdot\mathrm{g}^{-1})$ 的 $87.2~\mathrm{fe}$,离岸 $(1.8~\mathrm{ng}\cdot\mathrm{g}^{-1})$ 的 $1~211.1~\mathrm{fe}^{[33]}$. 意大利人湖河流沉

积物中 BDE-209 的含量(67.0 ng·g⁻¹)是 Maggiore 湖(12.1 ng·g⁻¹)的 5.5 倍^[34]. Melymuk 等^[35]报道 城市河流输入和污水处理厂排放是安大略湖中 PBDEs 的两个主要来源,其输入贡献率(>85%)远高于大气干湿沉降途径.

与国内外研究相比(图 3),山美水库及人库河流表层沉积物中 PBDEs 的污染程度较国内外大多数湖库^[7,10,11,31~34,36~46] 与河流^[30,32~34,36,47~54] 更严重,且主要体现为 BDE-209 的污染较重,表明山美水库 PBDEs 污染程度较高,流域内商用十溴二苯醚的用量较多.

2.2 PBDEs 的空间分布

山美水库及入库河流每个样点 22 种 BDE 的含量经多种检验 (Kolmogorov-Smirnov、Shapiro-Wilk、Anderson-Darling 和 D'Agostino-K squared test) 均不符合正态分布,但符合对数正态分布,因此以下采用不基于正态分布前提非参数的 Spearman 相关系数



 \sum PBDEs 表示除 BDE-209 外测定的其他 BDE 同系物总和,横坐标括号中数字(i/j)表示样本量/BDE 同系物数量,

参考线表示山美水库(实线) 及入库河流(虚线) \sum_{n} PBDEs 或 BDE-209 算数均值

图 3 山美水库及入库河流表层沉积物中 PBDEs 的含量与国内外其他地点的比较

Fig. 3 Comparisons of the contents of PBDEs in the surface sediments of Shanmei Reservoir and its inflowing river,

Ouanzhou, China with other sites at home and abroad

讨论.

河流各样点间 BDE 同系物呈极显著高度正相 关(r 为 0.953~1, P < 0.000),表明河流各样点 PBDEs 的组成高度相似,来源相同. H1 和 H2 位于上游来水口(图 1),水流在此处变缓,污染物易于沉积, \sum PBDEs 含量最高(2 422.5 ng·g⁻¹和2 329.3 ng·g⁻¹,图 4). H4 因下游河道转弯变窄而水流减缓,且 H4(水深 15 m)较 H3(水深 10 m) 地势低,污染物易沉积汇集,致 H4 (1 072.1 ng·g⁻¹)高于 H3 (409.2 ng·g⁻¹). H5 水流因河道变窄而加快,底泥可能被搅动悬浮于水体中,且 H5 水深(14 m)比 H4 略浅,其 \sum PBDEs (800.5 ng·g⁻¹) 略低于 H4. 上游样点(H1 和 H2) \sum PBDEs 是下游(H4 和 H5)的 2.5 倍,显示 PBDEs 在从河流上游到下游的过程中逐渐沉积.

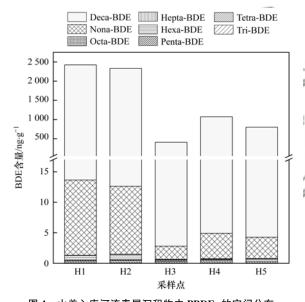


图 4 山美入库河流表层沉积物中 PBDEs 的空间分布

Fig. 4 Spatial distributions of PBDEs in the inflowing river surface sediments of Shanmei Reservoir

山美水库各水文期 \sum PBDEs 的空间分布见图 5. 除丰水期 S1 和 S3(r = 0. 493,P = 0. 062)、枯水期 S2 和 S9(r = 0. 507,P = 0. 054)及平水期 S5 和 S7(r = 0. 420,P = 0. 175) BDE 同系物含量间 Spearman 相关系数不显著外,其他点位间均显著相关(r 为 0. 564 ~ 0. 994,P < 0. 034),显示库区大多数样点的 PBDEs 组成相似度较高,有共同的主要来源.

丰水期 \sum PBDEs 算数均值含量入库区(538.2 ng·g⁻¹) >库中区(192.7 ng·g⁻¹) >库尾区(104.5 ng·g⁻¹) > 坝前区(27.5 ng·g⁻¹). 人库区与入库河流样点间相关系数在水库各功能区中最高(r 为 0.915 ~ 0.964, P < 0.000), 结合前文入库河流

 \sum PBDEs 远高于库区的情况,显示人库区 PBDEs 直接来自入库河流. 丰水期正值汛期,地处亚热带沿海区域的泉州台风暴雨高发,上游来水丰富(河流平均透明度:134 cm),携带大量颗粒物(水库平均透明度,丰:152 cm,枯:230 cm)进入库区后水域面积扩大,水流减缓,悬浮颗粒易在入库区沉积,致其 \sum PBDEs 最高;库中区位于入库区下游,受入库河流影响较小(r 为 0.690 ~ 0.903, P ≤ 0.017), \sum PBDEs 低于入库区;库尾和坝前区受上游来水影响最小,但库尾区受附近九都镇人类活动影响明显,使库尾近岸的 S3 与入库区 S1 相关性不显著;坝前区丰水期排水增加,部分底泥随水流排出, \sum PBDEs 最低.

枯水期 \sum PBDEs 入库(378.1 ng·g⁻¹) > 库尾(166.4 ng·g⁻¹) > 坝前(165.3 ng·g⁻¹) > 库中(77.6 ng·g⁻¹). 枯水期上游来水最少,但入库区 \sum PBDEs 仍最高,入库区和入库河流样点间相关系数在各功能区中也最高(r 为 0. 779 ~ 0. 915,P < 0. 005),与丰水期一致,说明枯水期上游来水仍是水库的主污染源;库尾区 \sum PBDEs 偏高,显示枯水期九都镇对库尾区影响较大;库中区低于坝前区,与枯水期大坝排水少有关;S9 与入库区样点相关系数较低或不显著(r 为 0. 507 ~ 0. 654,P 为 0. 008 ~ 0. 054),且坝前区与入库河流相关系数在各功能区中最低(r 为 0. 314 ~ 0. 903,P < 0. 346),均显示枯水期因来水少,坝前区 \sum PBDEs 受来水影响程度减弱.

平水期 \sum PBDEs 库中(607.3 ng·g⁻¹) > 入库(498.0 ng·g⁻¹) > 坝前(175.1 ng·g⁻¹) > 库尾(123.8 ng·g⁻¹). 平水期水位稳定,库中区 \sum PBDEs 最高受 S7 附近从枯水期后开始鱼类养殖的影响,鱼体内易富集大量的 PBDEs,其活动可能造成 S7 偏高^[55],S7 和库中区 S5、S6 及库尾区 S3 的相关系数不显著或较低(r 为 0. 420 ~ 0. 622,P 为 0. 018 ~ 0. 175),也显示其受到了局地污染源的影响;人库区高于库尾区,且人库区与人库河流相关系数在各功能区中仍最高(r 为 0. 779 ~ 0. 964,P < 0. 005),而库尾区与人库河流相关系数则最低(r 为 0. 454 ~ 0. 882,P < 0. 128),表明上游来水仍是水库的主污染源.

综上,除平水期 S7 外,各水文期入库区 \sum PBDEs 均最高,入库区和入库河流样点间 BDE

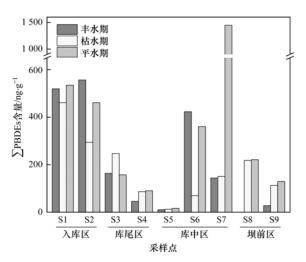


图 5 山美水库表层沉积物中 $\sum PBDEs$ 的时空分异

Fig. 5 Spatiotemporal differentiation of \sum PBDEs in surface sediments of Shanmei Reservoir

含量显著正相关(r 为 0. 779 ~ 0. 964, P < 0. 005)且在各功能区中相关性最强,显示入库河流是山美水库 PBDEs 的主污染源;库尾区近九都镇的 S3 各水文期 \sum PBDEs 均高于 S4(1.7~3.6 倍),且库尾区与入库河流相关性较低,显示库尾区受城镇影响较大;坝前区近坝的 S9 受排水及大坝拦截作用使沉积物被带走或悬浮于水体中,其 \sum PBDEs 低于 S8. 样点间 \sum PBDEs 含量不同且差别较大,显示各样点污染源及影响因素有所不同[10].

除入库区个别低溴 BDE 的丰度略小于入库河流外(水库/河流:~0.8),山美水库各功能区 Tri~Nona-BDE 的平均丰度均高于入库河流,其年均比值为3.3~147.8,表明 PBDEs 在水库中发生了逐级还原脱溴自然降解,且各级还原脱溴速率存在差异^[24,56],部分 BDE 因继续还原脱溴速率较慢而累积.山美水库各功能区 BDE-209 的丰度均小于入库河流(99.4%),且入库区 BDE-209 丰度最高(丰:99.0%,枯:98.7%,平:98.9%),库中区最低(丰:94.4%,枯:93.2%,平:96.6%),显示 PBDEs 在从入库河流进入水库再到库中区的过程中自然降解程度逐渐加深.

2.3 PBDEs 的水文期变化

各水文期 \sum PBDEs 的 Pearson 相关系数不显著,但丰水期和枯水期呈近显著正相关(r=0.699,P=0.054). 尽管各水文期 \sum PBDEs 经 Kolmogorov-Smirnov 检验符合正态分布,但多种更严格的正态检验 方 法 (Shapiro-Wilk、Anderson-Darling 和 D'Agostino-K squared test)显示,枯水期 \sum PBDEs

不符合正态分布(P<0.013),且丰水期正态检验的相伴概率也较低(P为0.083~0.251).因此,进一步通过 Spearman 相关系数分析显示,丰水与其他水文期均显著正相关(均为r=0.714,P=0.047),而枯水和平水期近显著正相关(r=0.617,P=0.077),表明各样点 \sum PBDEs 的水文期变化较一致,水文期变化对沉积物中 \sum PBDEs 含量有一定的影响.

由于部分水文期 \(\sum_{PBDEs}\) 不符合正态分布, 但均符合对数正态分布(P=1,K-S test),故经对数 转换后作双因素方差分析(Two-Way ANOVA)显示, 水文期变化对 \sum PBDEs 的影响统计不显著(P =0.178),而采样点位变化则对 ∑PBDEs 有极显著 影响(P = 0.0001),表明水库沉积物中 $\sum PBDEs$ 的空间分布变异大于水文期变化. 基于 Fisher 最小 显著差异法(LSD)的多重比较(multiple comparisons)进一步显示, S5 与其他所有样点均有 显著差异,表明 S5 沉积情况与其他点位显著不同. S5 地势最高,沉积物以砂质为主,沉积柱采样长度 仅为其他样点平均长度的57%~67%, ∑PBDEs 在 所有样点中也最低,且低溴 BDE 的比例与其他点位 有较大差异(见2.5节). 入库区和其他水库功能区 之间有(近)显著差异(P 为 0.019~0.061),结合前 文人库区 $\sum PBDEs$ 最高的情况,显示人库区因接 纳了人库河流悬浮颗粒的大量沉积, \(\sumperpreserve) PBDEs (近)显著高于其他距来水较远的功能区.

不同水文期库区 \sum PBDEs 空间分布的影响因素存在差异. 丰水期 \sum PBDEs 在各功能区的分布主要受上游来水的影响; 枯水期地下水上涌可能会使下层沉积物中污染物重新进入表层,加之九都镇的人为活动,使库尾区枯水期 \sum PBDEs 高; 平水期除 受 九都镇影响较大的 S3 外,其他点位 \sum PBDEs 高于枯水期,除上游来水因素外,平水期库区水体 pH(均值 8.3,范围 7.9 ~ 8.6)较枯水期(均值 7.5,范围 7.3 ~ 7.7)更偏碱性可能也有一定影响,因为碱性条件更有利于沉积物中有机质被矿物质吸附.

各样点 \sum PBDEs 的水文期变化也有差异. S1 和 S2 处 \sum PBDEs 算数均值丰 > 平 > 枯,与上游来水量一致. S3 枯 > 丰 > 平,枯水期最高显示其主要受九都镇的影响. S4 枯平相当,丰水期最低,说明库

2.4 PBDEs 的赋存量

沉积物既是 PBDEs 的重要储库,也是水环境中 PBDEs 的二次释放源. 虽然高溴 BDE 尤其是 BDE-209 因 K_{ow} 大,理论上应以固相赋存为主,但研究者在英国亚耳河水(17~295 $ng\cdot L^{-1}$)和北美五大湖水中监测到大量的 BDE-209(丰度 $18\% \pm 3\%$),因为 PBDEs 能以胶体态存在于水中,且其在胶体中的分配比例随溴代数增加而提高^[19].

为评估山美水库沉积物作为 PBDEs 源和汇的 潜力,用公式(1)^[7]计算其单位面积赋存量:

$$I = \sum kC_{i}d\rho \tag{1}$$

式中,I 为 PBDEs 的单位面积赋存量($kg \cdot km^{-2}$),k 为单位转换常数, C_i 为 PBDEs 的平均含量($ng \cdot g^{-1}$),d 为沉积物厚度($20 \text{ cm}^{[10]}$), ρ 为沉积物密度($1.5 \text{ g} \cdot \text{cm}^{-3[10,31]}$).

赋存总量用公式(2)[7]计算:

$$TI = \sum kIA \tag{2}$$

式中,TI 为总赋存量(t),A 为山美水库水域面积(23.75 $km^{2[27]}$).

山美水库沉积物中 BDE-209、 \sum_{21} PBDEs 和 \sum PBDEs 的单位面积赋存量分别为 79.1、1.2 和

80. 3 kg·km⁻²,相应赋存总量分别为 1. 88、0. 03 和 1. 91 t. 虽然山美水库中 \sum PBDEs 赋存总量小于国内的太湖(30 t,水域面积2 338. 1 km^{2[10]})和北美五大湖中的苏必利尔湖、密歇根湖和休伦湖(85. 3 t,水域面积总计199 500 km^{2[7]}),但其单位面积沉积物中 \sum PBDEs 的赋存量(80. 3 kg·km⁻²)却是太湖(12. 8 kg·km⁻²)的 6. 3 倍,北美五大湖(0. 427 kg·km⁻²)的 188 倍,表明山美水库 PBDEs 污染较为严重,应深入研究其作为城市水源水库潜在的环境和健康风险,也应及时对其污染源进行管控和治理.

由于山美水库 PBDEs 等有机污染和氮磷等生源要素污染均以河流输入为主,且 PBDEs 等 POPs可增加沉积物中磷等生源要素的释放量和生物可利用率^[14],从而加重湖泊的富营养化,因此,一方面应管控上游来水河流流域工业点源污染、农业面源污染和居民生活垃圾等外源输入,另一方面应疏浚污染较重的人库河流等区域的沉积物,以实现外源管控、内源削减的综合治理.

2.5 PBDEs 的同系物组成

测定的 22 种 3~10 溴 BDE 按溴代数划分:Tri-BDE 包括 BDE-17 和 BDE-28, Tetra-BDE 包括 BDE-71、BDE-47、BDE-66 和 BDE-77, Penta-BDE 包括 BDE-100、BDE-99、BDE-118 和 BDE-85, Hexa-BDE 包括 BDE-154、BDE-153 和 BDE-138, Hepta-BDE 包括 BDE-183 和 BDE-190, Octa-BDE 包括 BDE-197、BDE-203 和 BDE-196, Nona-BDE 包括 BDE-208、BDE-207 和BDE-206, Deca-BDE 为 BDE-209.

山美水库沉积物中不同溴代数 BDE 的组成表明(图 6): BDE-209 丰度最高(中值 98.3%,均值 96.9%,范围 84.5%~99.2%),这与国内主要湖库沉积物以 BDE-209 为主的 PBDEs 组成一致[10,11,31,32,36,39,40]. 各水文期 S5 的 Deca-BDE 丰度均最小(84.5%~92.0%),而 Nona-BDE(3.2%~

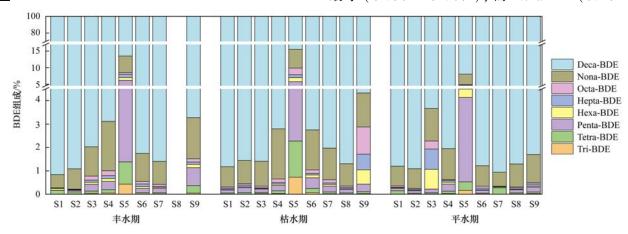


图 6 山美水库沉积物中各溴代数 BDE 的组成

Fig. 6 Compositions of BDE with different brominated levels in the sediments of Shanmei Reservoir

5.7%)、Penta-BDE (3.5% ~ 4.8%) 和 Tetra-BDE (0.4%~1.6%)则最高,可能是该处沉积物以砂质为主,且地势在水库各样点中最高,水体中溶解氧(DO)也最高(8.1 mg·L⁻¹,水库各样点范围:6.1~8.1 mg·L⁻¹),Deca-BDE 较易降解为低溴 BDE.

Nona-BDE 丰度(中值 1.2%,均值 1.6%,范围 0.6%~5.7%) 仅次于 Deca-BDE, Octa-BDE 丰度中 值为 0.1% (均值 0.2%, 范围 0~2%), 而 Tri~ Hepta-BDE 合计丰度中值仅 0.4% (均值 1.3%,范围 0.2%~8.0%). 商用十溴二苯醚 Saytex 102E 和 Bromkal 82- 0DE 主要同系物丰度为 BDE-209 (96.8% 和 91.6%)、BDE-206(2.19% 和 5.13%) 和 BDE-207(0.24% 和 4.10%)等[57],相应的 BDE-209: BDE-206: BDE-207分别为: 403:9:1和22:1:1, 山美水库沉积物中该比值(173:1:1)介于以上两者 之间,且这些同系物在所有样品中均为主要同系物, 说明山美水库流域使用了十溴二苯醚商用品. BDE-208 在商用 PBDEs 中丰度均很低(<0.02%~ 0.19%),而在山美水库沉积物中丰度(中值 0.31%,均值 0.43%, GM 0.37%,范围 0.16%~ 1.34%)则较高,表明其主要来自十溴二

原脱溴降解. 山美水库沉积物中 BDE-208 (0.43%)和 BDE-197(0.12%)均值丰度都高于商用十溴二苯醚 Saytex 102E 和 Bromkal 82-0DE (BDE-208:0.06%和 0.07%, BDE-197:0 和 0.03%),而 BDE-206(0.60%)则低于商用十溴二苯醚(2.19%和 5.13%)^[57],表明沉积物中的Deca-BDE 先脱溴还原为 Nona-BDE,随后继续还原为 Octa-BDE.

山美水库不同溴代数 BDE 及 \sum PBDEs 含量之间的 Spearman 相关系数显示(表 1),各水文期 \sum PBDEs 与 Deca-BDE 极显著完全正相关(r=1),进一步说明 BDE-209 是山美水库沉积物中首要的 PBDEs 同系物;各水文期 Deca-BDE 与 Nona-BDE 间均极显著高度正相关(r 为 0.905 ~ 1.0, $P \le$ 0.002),说明两者有共同的来源且环境行为一致[10];各水文期 Octa-BDE 和 Hepta-BDE 间(近)显著中度正相关(r 为 0.619 ~ 0.750,P < 0.102),说明两者有部分共同来源.其中,Octa-BDE 来自十溴二苯醚商用品[57] 及其逐级还原脱溴降解,而 Hepta-BDE 则可能主要来自 Octa - BDE 的继续还原脱

表 1 山美水库各水文期不同溴代数 BDE 含量的 Spearman 相关系数¹⁾

Ta	ble 1 Spearman	correlations for	or contents of	BDE with diff	ferent bromina	ated levels in var	rious hydrolo	gical periods o	f Shanmei R	eservoir
水文期	to Va	Tri-BDE	Tetra-BDE	Penta-BDE	Hexa-BDE	Hepta-BDE	Octa-BDE	Nona-BDE	Deca-BDE	$\sum PBDEs$
g.	Tri-BDE Tetra-BDE	0.024	1 8	al				9		
	Penta-BDE	0.833 *	0.143	1						
	Hexa-BDE	0.190	0.667	0.262	1					
丰水期	Hepta-BDE Octa-BDE	0.857**	0.286	0.810*	0.381	1	1			
	Octa-BDE Nona-BDE	0.310	0.357	0.381	0.690	0.619	1	1		
	Deca-BDE	0. 167 -0. 095	0.714 * 0.762 *	0.286 -0.048	0.619 0.452	0.619 0.333	0. 786 * 0. 595	1 0. 905 **	1	
									1	
	\sum PBDEs	-0.095	0.762 *	-0.048	0.452	0.333	0.595	0.905 **	1 **	1
	Tri-BDE	1								
	Tetra-BDE	0.700 *	1							
	Penta-BDE	0.483	0.783 *	1						
	Hexa-BDE	0.133	-0.067	0.350	1					
枯水期	Hepta-BDE	-0.267	0.100	0.483	0.450	1				
	Octa-BDE	0.167	0.200	0.583	0.533	0.750*	1			
	Nona-BDE	-0.200	0.350	0.417	-0.467	0.367	0.083	1		
	Deca-BDE	-0.183	0.333	0.467	-0.333	0.483	0.183	0. 983 **	1	
	$\sum PBDEs$	-0.183	0.333	0.467	-0.333	0.483	0.183	0. 983 **	1 **	1
	Tri-BDE	1								
	Tetra-BDE	0.717 *	1							
	Penta-BDE	0.450	0.333	1						
	Hexa-BDE	0.033	0.317	0.000	1					
平水期	Hepta-BDE	-0.167	0.050	-0.383	0.150	1				
	Octa-BDE	0.267	0.517	-0.183	0.467	0.633	1			
	Nona-BDE	0.683 *	0.950 **	0.333	0.300	0.167	0.650	1		
	Deca-BDE	0.683 *	0.950 **	0.333	0.300	0.167	0.650	1 **	1	
	$\sum PBDEs$	0.683*	0.950 **	0.333	0.300	0.167	0.650	1 **	1 **	1

^{1)*}表示P<0.05(双尾); **表示P<0.01(双尾)

溴降解[19,25].

此外,丰水期 Hepta-BDE 和 Penta-BDE 间 (r = 0.810, P = 0.015),枯水期 Penta-BDE 和 Tetra-BDE 间 (r = 0.783, P = 0.013),丰水期 Penta-BDE 和 Tri-BDE 间 (r = 0.833, P = 0.010),枯水和平水期 Tetra-BDE 和 Tri-BDE 间 $(r \to 0.700 \sim 0.717, P < 0.036)$ 也显著中度正相关,丰水和平水期高溴 BDE (Deca-BDE 和 Nona-BDE) 与低溴 BDE (Tetra-BDE 和/或 Tri-BDE) 显著中高度正相关 $(r \to 0.683 \sim 0.950, P < 0.047)$,进一步说明高溴 BDE 逐级还原脱溴是山美水库沉积物中低溴 BDE 的重要来源.

2.6 PBDEs 同系物的降解来源分析

商用五溴二苯醚 DE-71 和 Bromkal 70-5DE 中主要同系物丰度为 BDE-99 (48.60% 和 44.80%)、BDE-47 (38.20% 和 42.80%)、BDE-100 (13.10% 和 7.82%)、BDE-153 (5.44% 和 5.32%)和 BDE-154 (4.54% 和 2.68%)等[57],相应地 BDE-99: BDE-47: BDE-100: BDE-153: BDE-154 分别为 11:8:3:1:1和 17:16:3:2:1,山美水库沉积物中该比值 (0.2:0.7:0.1:0.2:1)较其低 1~2 个数量级,表明这些低溴 BDE 并非来自商用五溴二苯醚. 商用八溴二苯醚 DE-79 和 Bromkal 79-8DE 中主要同系物丰度为

BDE-183 (42.00% 和 12.60%)、BDE-197 (22.20% 和 10.50%)、BDE-207 (11.50% 和 11.20%)、BDE-196 (10.50% 和 3.12%)、BDE-153 (8.66% 和 0.15%)、BDE-203 (4.40% 和 8.14%)、BDE-206 (1.38% 和 7.66%)和BDE-209 (1.31% 和 49.60%)等[57],相应的 BDE-183:BDE-197:BDE-207:BDE-196:BDE-153:BDE-203:BDE-206:BDE-209 分别为:30:16:8:8:6:3:1:1和 1.6:1.4:1.5:0.4:0.02:1:1:6.5,水库沉积物中该比值(0.2:0.2:0.9:0.1:0.05:0.05:1:162)较其大多相差 1~2个数量级,表明这些 BDE 并非来自商用八溴二苯醚.综上,山美水库表层沉积物中低溴 BDE 主要来自商用十溴二苯醚的逐级还原脱溴自然降解.

山美水库表层沉积物和主要 PBDEs 商用品中BDE 同系物丰度比值对照见表 2. 其中, R_s 表示样品中不同BDE 丰度比值, R_c 表示商用品中相应BDE 丰度比值, $R_{(m/n)} = R_s/R_c$ (m/n 表示特定BDE相比). $R_{(m/n)} = 1$ 表示样品中BDE 丰度比值与商用品一致,该BDE 主要来自商用品使用; $R_{(m/n)} > 1$ 表示样品中BDE 丰度比值大于商用品,表示其可能来自样品中更高溴BDE的降解,且其继续降解为更低溴BDE的速率较小(除BDE-209外), $R_{(m/n)} < 1$ 表

表 2 山美水库沉积物和主要 PBDEs 商用品中 BDE 丰度比值对照

Table 2	Comparisons of the BDE a	abundance ratios	in sediments o	f Shanmei	Reservoir ar	nd the major	commercial	products of PBD	Es
Table 4	Companisons of the DDL a	inuliuance rangs	in scuments o	1 phanner	reservoir ar	nu me major	Commercial	products of 1 D	ப

V9 11/01	F		19	(a	۲.	<i>₽</i> ≥				$R_{(m)}$	'n)		
同系物	$R_{ m s}$	五溴	二苯醚	八溴	二苯醚	十溴二	二苯醚	五溴	二苯醚	八溴	二苯醚	十溴二	二苯醚
间示视	平均值	DE-71	Bromkal 70-5DE	DE-79	Bromkal 79-8DE	Saytex 102E	Bromkal 82-0DE	DE-71	Bromkal 70-5DE	DE-79	Bromkal 79-8DE	Saytex 102E	Bromkal 82-0DE
BDE-209/Nona-BDE	61.06	_	_	_	_	38. 9	9.8	_	_	_	_	1.57	6. 20
BDE-209/BDE-208	226. 64	_	_	_	_	1 613. 3	1 308. 6	_	_	_	_	0.14	0. 17
BDE-209/BDE-207	172. 75	_	_	_	_	403. 3	22. 3	_	_	_	_	0.43	7. 73
BDE-209/BDE-206	161. 93	_	_	_	_	44. 2	17. 9	_	_	_	_	3.66	9. 07
BDE-207/BDE-197	4. 84	_	_	0. 52	1. 07	_	136. 7	_	_	9.3	4. 5	_	0.04
BDE-207/BDE-196	6. 99	_	_	1. 10	3. 59	_	8.9	_	_	6. 4	1.9	_	0. 78
BDE-206/(BDE-203 + 196)	5. 33	_	_	0.09	0.68	_	9. 7	_	_	57. 6	7. 8	_	0. 55
BDE-206/BDE-203	18.71	_	_	0.31	0. 94	_	73. 3	_	_	59. 7	19. 9	_	0. 26
BDE-206/BDE-196	7. 45	_	_	0. 13	2. 46	_	11. 2	_	_	56. 7	3. 0	_	0. 67
(BDE-197 + 196)/BDE-183	1.45	_	_	0.78	1.08	_	_	_	_	1.9	1. 3	_	_
BDE-197/BDE-183	0.85	_	_	0.53	0.83	_	_	_	_	1.6	1.0	_	_
BDE-196/BDE-183	0. 59	_	_	0. 25	0. 25	_	_	_	_	2. 4	2. 4	_	_
BDE-183/BDE-154	0.89	0.02	0. 12	39. 3	315.0	_	_	40. 5	7. 3	0.02	0.003	_	_
BDE-183/BDE-153	5. 02	0.02	0.06	4. 8	84.0	_	_	273. 1	80. 9	1.0	0.06	_	_
BDE-183/BDE-138	8. 98	0. 14	0. 62	67. 7	_	_	_	65. 6	14. 4	0. 13	_	_	_
BDE-138/BDE-85	0.04	0. 25	0. 25	_	_	_	_	0.14	0. 15	_	_	_	_
BDE-154/BDE-99	4. 61	0.09	0.06	_	_	_	_	49. 3	77.0	_	_	_	_
BDE-153/BDE-99	0. 82	0. 11	0. 12	_	_	_	_	7. 3	6. 9	_	_	_	_
BDE-99/BDE-47	0.30	1. 27	1.05	_	_	_	_	0. 24	0. 29	_	_	_	_
BDE-47/(BDE-28 + 17)	1.62	119. 4	285. 3	_	_	_	_	0.014	0.006	_	_	_	_
BDE-47/BDE-28	5. 23	152. 8	428. 0	_	_	_	_	0.034	0.012	_	_	_	_
BDE-47/BDE-17	2. 36	545. 7	856. 0		_			0.004	0.003				

示样品中 BDE 丰度比值小于商用品,表示其可能已发生还原脱溴,部分降解为更低溴 BDE.

 $R_{ ext{(BDE-209/Nona-BDE)}}$ 为 1.57 和 6.20,显示 Deca-BDE 降解产生的 Nona-BDE 迅速降解生成更低溴 BDE, 这与沉积物中微生物降解 BDE-209 时, Nona-BDE 先增加后又减少波动的报道一致[56]. $R_{\text{(BDE-209/BDE-208)}}$ 为 0.14 和 0.17, 表明 BDE-209 降解产生了 BDE-208; $R_{\text{(BDE-209/BDE-207)}}$ 为 0.43 和 7.73, $R_{\text{(BDE-209/BDE-206)}}$ 为 3.66 和 9.07,显示 BDE-209 降解为 BDE-207 和 BDE-206 的速率高于其继续降解为更低溴 BDE 的 速率. R_(BDE-207/BDE-197) < R_(BDE-207/BDE-196), 显示 BDE-207 降解为 BDE-197 (间位脱溴)的速率高于其降解为 BDE-196(邻位脱溴). 据报道[58], 商用八溴二苯醚 降解时,邻、间、对位溴原子的取代率分别为24%、 73%和54%,其中,间位溴原子取代率最高,因为其 空间位置比邻位溴原子离醚键远,空间位阻较小,故 该处重叠的电子云斥力较小, 溴原子更易被氢原子 取代. 另外, PBDEs 间位溴原子数量约是对位的 2 倍,故降解过程中,间位最易被取代,对位次之,邻位 最难被取代. R_(BDE-197/BDE-183) < R_(BDE-196/BDE-183), 因为 BDE-196 可通过间位脱溴降解为 BDE-183, 而 BDE-197 还可通过间位脱溴还原为 BDE-184^[58,59] R_(BDE-196/BDE-183) > 1,显示 BDE-183 没有出现累积,因 为其易于间位脱溴继续降解产生 BDE-154[58]. 以商 用五溴二苯醚 R_c 为分母, R_(BDE-183/BDE-154) 为 40.5 和 7.3, 而以商用八溴二苯醚 R。为分母, R(BDE-183/BDE-154) 为 0.02 和 0.003,两者差异大的原因可能是五溴二 苯醚中 BDE-183 和 BDE-154 均非主要同系物,且八 溴二苯醚中 BDE-154 非主要同系物[57],这需要继续 研究. R_(BDE-138/BDE-85) 为 0.14 和 0.15,显示 BDE-138 较快降解为 BDE-85. R_(BDE-154/BDE-99) 为 49.3 和 77.0, R_(BDE-153/BDE-99)为 7.3 和 6.9,结合 R_(BDE-196/BDE-183)和 R(BDE-99/BDE-47),说明 BDE-154 和 BDE-153 出现累积, 较少降解为 BDE-99(邻位脱溴^[58]). R_(BDE-99/BDE-47) 为 0.24 和 0.29,说明 BDE-99 较多降解为 BDE-47(间 位脱溴[60]),可能系微生物还原脱溴所致[56]. R_(BDE-47/BDE-28) 为 0.034 和 0.012, R_(BDE-47/BDE-17) 为 0.004 和 0.003,显示 BDE-47 降解产生了 BDE-28 和 BDE-17,且 BDE-47 降解为 BDE-28(邻位脱溴) 的速率低于其降解为 BDE-17(对位脱溴)的速率, 与理论研究一致[58,61].

根据降解途径中样品和商用品 BDE 丰度比值 (表 2),选择主要同系物,分析山美水库沉积物中不同溴代数 BDE 的降解率. Deca-BDE \rightarrow Nona-BDE: BDE-209 降解为 BDE-207 和 BDE-206 是主要途径^[59],同商用十溴二苯醚相比, $R_{\text{(BDE-209/BDE-207)}}$ 为

0.43 和 7.73, R_(BDE-209/BDE-206) 为 3.66 和 9.07, 显示 Nona-BDE 丰度较低,说明 Nona-BDE 可较快继续降 解为低溴 BDE. R_(BDE-209/BDE-208) 为 0.14 和 0.17,显示 BDE-208 约83%~86%源自BDE-209的降解. Nona-BDE \rightarrow Octa-BDE: $R_{\text{(BDE-207/BDE-197)}}$ $\not\supset$ 0.04, R_(BDE-207/BDE-196)为 0.78, BDE-207 更易降解为 BDE-197^[58]. $R_{\text{(BDE-206/BDE-196)}}$ 为 0.67, $R_{\text{(BDE-206/BDE-203)}}$ 为 0.26, BDE-206 更易降解为 BDE-203, 所以 BDE-207 →BDE-197(降解率约96%)和BDE-206→BDE-203 (降解率约74%)为 Nona-BDE 降解至 Octa-BDE 的 主要涂径. Octa-BDE → Hepta-BDE: $R_{[(BDE-197+BDE-196)/BDE-183]}$ 为 1.9 和 1.3,说明 Octa-BDE 降解至 Hepta-BDE 过程较慢,使 Octa-BDE 累 积. Hepta-BDE →Hexa-BDE: BDE-183 降解为 BDE-154、BDE-153 和 BDE-138 为主要途径^[59,62]. 以商 用八溴二苯醚 R_c 为分母, R_(Hepta-BDE/Hexa-BDE) 为 0.003 ~1.0, 而以商用五溴二苯醚 R。为分母, $R_{\text{(Hepta-BDE/Hexa-BDE)}}$ 为 7.3 ~ 273.1,两者矛盾的原因 可能是商用五溴二苯醚中 BDE-183、BDE-154和 BDE-138 丰度很小,且商用八溴二苯醚中 BDE-154 和 BDE-138 丰度很小[57],以及样品中 BDE-153 和 BDE-138丰度小. Hexa-BDE →Penta-BDE: BDE-138 →BDE-85 和 BDE-154 \BDE-153 →BDE-99 为主要 途径. R(BDE-138/BDE-85) 为 0.14 和 0.15,显示其降解率 约 85%, 而 BDE-154 和 BDE-153 → BDE-99 的 $R_{(m/n)}$ 为 6.9 ~ 77.0, 显示其降解速度较慢, 使 BDE-154 和 BDE-153 累积. Penta-BDE → Tetra-BDE: BDE-99 → BDE-47 为主要途径. R_(BDE-99/BDE-47) 为 0. 24 和 0. 29,显示 BDE-99 →BDE-47 降解率大 致在 71% ~ 76% 之间. Tetra-BDE → Tri-BDE: BDE-47 → BDE-28 和 BDE-17 为 主 要 途 径, $R_{\text{[BDE-47/(BDE-28+BDE-17)]}}$ 为 0.006 和 0.014,说明 Tetra-BDE 至 Tri-BDE 降解率约99%.

山美人库河流表层沉积物和主要 PBDEs 商用品中 BDE 丰度比值对照见表 3. 相比山美水库, Deca ~ Penta-BDE 的 $R_{(m/n)}$ 同水库类似, 大部分 > 1 且大于水库, 而 Penta ~ Tri-BDE 值 < 1, 且小于水库, 可能是入库河流水流湍急, 水环境不稳定, Deca-BDE 降解较库区少所致.

淡水湖库是我国众多城市的主水源,PBDEs 尤其是 BDE-209 在我国多区域历史用量和环境赋存量大,其自然降解过程复杂漫长,一系列毒性更大的低溴 BDE 产物给水环境和人体健康带来风险,并可能对磷等生源要素的生物地球化学循环产生影响,进而对湖库环境构成更大的威胁.本研究表明,我国部分城市水源水库 PBDEs 污染程度可能较高,河流注人

是首要污染源,PBDEs 在水库沉积物中的空间分布变异大于水文期变化,PBDEs 自然降解从河流到入库区再到库中区逐渐增加,且各级还原脱溴速率不同,部

分 BDE 因继续降解速率较慢而累积,应继续研究和评估其环境和健康风险,并通过外源管控和内源削减等综合措施治理改善,保障城市供水安全.

表 3 山美入库河流沉积物和主要 PBDEs 商用品中 BDE 丰度比值对照

Table 3 Comparisons of the BDE abundance ratios in the inflowing river sediments of Shanmei Reservoir and the major commercial products of PBDEs

				Ì	$R_{\rm e}$					$R_{(m)}$	/n)		
同系物	$R_{ m s}$	五溴	二苯醚	八溴	二苯醚	十溴二	二苯醚	五溴	五溴二苯醚		二苯醚	十溴二	二苯醚
111/1/10	平均值	DE-71	Bromkal 70-5DE	DE-79	Bromkal 79-8DE	Saytex 102E	Bromkal 82-0DE	DE-71	Bromkal 70-5DE	DE-79	Bromkal 79-8DE	Saytex 102E	Bromkal 82-0DE
BDE-209/Nona-BDE	212. 31	_	_	_	_	38. 9	9.8	_	_	_	_	5.46	21. 56
BDE-209/BDE-208	1 250. 81	_	_	_	_	1 613. 3	1 308. 6	_	_	_	_	0.78	0.96
BDE-209/BDE-207	711.50	_	_	_	_	403. 3	22. 3	_	_	_	_	1.76	31.85
BDE-209/BDE-206	399. 18	_	_	_	_	44. 2	17. 9	_	_	_	_	9.03	22. 36
BDE-207/BDE-197	32. 88	_	_	0. 52	1. 07	_	136. 7	_	_	63.5	30.8	_	0. 24
BDE-207/BDE-196	28.94	_	_	1. 10	3. 59	_	8. 9	_	_	26. 4	8. 1	_	3. 25
${\rm BDE\text{-}206/(BDE\text{-}203+196)}$	51. 59	_	_	0.09	0.68	_	9. 7	_	_	557.0	75.8	_	5. 33
BDE-206/BDE-203	_	_	_	0.31	0.94	_	73. 3	_	_	_	_	_	_
BDE-206/BDE-196	51. 59	_	_	0. 13	2.46	_	11. 2	_	_	392. 5	21.0	/	4. 63
(BDE-197 + 196)/BDE-183	3.41	_	_	0.78	1.08	_	_	_	A	4. 4	3. 2	1	8
BDE-197/BDE-183	1.60	_	_	0.53	0.83	_	_	1	, 	3.0	1. 9	#-/	£-
BDE-196/BDE-183	1. 82	_	_	0. 25	0. 25	_	_	H	+ 1/	7.3	7. 3	/-/	7
BDE-183/BDE-154	_	0.02	0. 12	39. 3	315.0	_	_	14	+ //	\ -	- (#)—]]
BDE-183/BDE-153	1. 93	0.02	0.06	4. 8	84. 0	_	- 1	105. 0	31.1	0.4	0.02	2	/#
BDE-183/BDE-138	0. 13	0. 14	0. 62	67. 7			- /	0. 9	0.2	0.00	- '	1000 M	B)
BDE-138/BDE-85	1. 21	0. 25	0. 25	<i>λ/-£</i>		_	- 9	4. 91	4. 94	//-	_	_	
BDE-154/BDE-99	11-	0.09	0.06	10		_	_	/-1) =	(-	- (_	-
BDE-153/BDE-99	ľ –	0. 11	0. 12	P.	"F	_	-/	(A)	-24	1-	- 1	1	9 1
BDE-99/BDE-47	0.00	1. 27	1. 05	1/30	//- (\-	-\	0.00	0.00	}-	_	_	S/-
BDE-47/(BDE-28 + 17)	0.86	119. 4	285. 3	1 /	(- /	7	_ y	0.007	0.003	V)	_	_	_
BDE-47/BDE-28	1.48	152. 8	428. 0	4			_	0.010	0.003	_	_	_	_
BDE-47/BDE-17	2. 05	545.7	856. 0	-)	-	_	_	0.004	0.002	_	_	_	_

3 结论

- (1) 入库河流沉积物中 \sum PBDEs 中值 (1072.1 ng·g⁻¹)是山美水库(160.4 ng·g⁻¹)的6.7倍,山美水库单位面积沉积物中 \sum PBDEs 赋存量 (80.3 kg·km⁻²)是太湖的6.3倍,北美五大湖的188倍,其污染程度较国内外大多数湖库更严重,且以BDE-209为主(中值98.3%,均值96.9%,范围84.5%~99.2%).
- (2)水库大多数样点(r 为 0.564~0.994,P < 0.034)及河流各样点(r 为 0.953~1,P < 0.000)间 BDE 同系物呈(极)显著中至高度正相关,显示各样点 PBDEs 组成相似度较高,有共同主要来源. 入库区和入库河流样点间极显著正相关(r 为 0.779~0.964,P < 0.005)且相关性强于其他功能区,显示人库河流是水库 PBDEs 的主污染源.
- (3)各样点 \sum PBDEs 水文期变化较一致(r 为 0.617~0.714, $P \le 0.077$),显示水文期变化对沉积

- 物 \sum PBDEs 有一定影响, 但统计不显著 (P = 0.178, Two-Way ANOVA), 而点位变化对 \sum PBDEs 则有极显著影响 (P = 0.0001), 表明 PBDEs 在水库沉积物中的空间分布变异大于水文期变化.
- (4) PBDEs 自然降解从河流到入库区再到库中区逐渐增加,且各级还原脱溴速率不同,部分 BDE 因脱溴速率较慢而累积.入库区和其他功能区有(近)显著差异(P 为 0.019 ~ 0.061, Two-Way ANOVA, LSD).库尾区近九都镇S3 各水期 \sum PBDEs 均比S4高(1.7~3.6倍),且库尾区与入库河流相关性较低(r 为 0.454~0.915,P \leq 0.128),显示其受九都镇影响较大.
- (5)丰度比值法研究表明,低溴 BDE 主要源自 十溴二苯醚的逐级还原脱溴自然降解. Deca-BDE 降 解产生的 Nona-BDE 约 70%以上可较快继续降解生 成 Octa-BDE, BDE-208 约 85% 源自 BDE-209 的降 解,从 Octa-BDE 到 Penta-BDE 的降解过程中,部分 Octa-BDE 和 Hexa-BDE 同系物因降解相对较慢而累

积,Penta-BDE 到 Tri-BDE 的降解率在 70% 以上. 参考文献:

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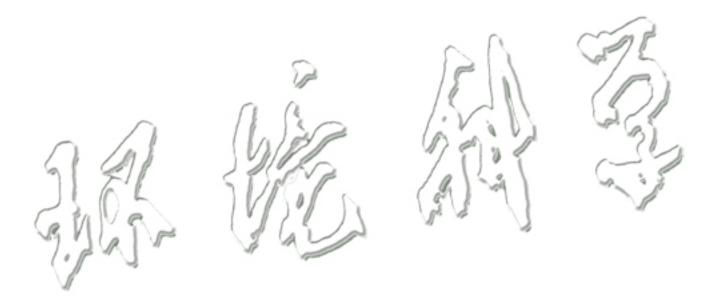
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