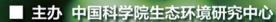
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贵州某规模化养猪场废水中抗生素的污染特征及去除 效果

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摘要:为了解贵州某规模化养猪场废水处理过程中抗生素的污染特征及废水处理工艺对抗生素的去除效果,使用固相萃取-液相色谱-串联质谱仪(SPE-LC-MS)技术,对 2 家规模化养猪场(猪场 A 和猪场 B)废水处理工艺各处理单元进出水中 10 种兽用抗生素的去除进行调查研究. 结果表明, 2 家规模化养猪场废水处理工艺对常规污染物(COD、NH $_4^+$ -N、TN 和 TP)的去除率在 88. 10%以上. 2 家养猪场废水处理工艺各处理单元进出水中均有多种抗生素检出,检出浓度范围在 ND ~ 120 842. 74 ng·L $^-$ 1之间. 其中,主要的污染单体为磺胺间甲氧嘧啶(SMM)、磺胺对甲氧嘧啶(SMD)、土霉素(OTC)和氧氟沙星(OFL),最高单体污染浓度达120 842. 74 ng·L $^-$ 1(SMM). 调查的 10 种抗生素在处理工艺中的去除效果较好,总去除率为 99. 23% ~ 100. 00%,在猪场 A 的废水处理工艺中,"USR + 2 级 A/O + 消毒池 + 氧化塘"组合工段能有效去除废水中残留的抗生素,其中对 SMM、SMD 和 OTC 的总去除率达 100. 00%;在猪场 B 的废水处理工艺中,"超滤(UF) + 纳滤(NF)"组合工段能有效去除废水中的抗生素,最终废水中 99. 23%以上的抗生素被去除. 但最终出水中大部分抗生素的浓度高于欧盟水环境抗生素阈值(10 ng·L $^-$ 1). 通过冗余分析发现废水中常规指标(COD、NH $_4^+$ -N、TN、TP 和 pH)与部分抗生素的降解具有相关性.

关键词:养猪场废水;兽用抗生素;污染特征;去除效果;冗余分析

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Pollution Characteristics and Removal Effects of Antibiotics in Wastewater from Large-Scale Pig Farms in Guizhou

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Abstract: To understand the pollution characteristics and removal effect of antibiotics in the wastewater treatment process of large-scale pig farms in Guizhou, solid-phase extraction-liquid chromatography-tandem mass spectrometry (SPE-LC-MS) was used to investigate the removal of ten veterinary antibiotics from the influent and effluent of each treatment unit during the wastewater treatment process in two large-scale pig farms (named Farm A and Farm B). The results showed that the removal rates of conventional pollutants [including chemical oxygen demand (COD), NH_4^+ -N, total nitrogen (TN), and total phosphorus (TP)] in Farm A and Farm B were above 88. 10%. The antibiotics concentrations detected in the influent and effluent ranged from ND-120 842. 74 $ng \cdot L^{-1}$. The main antibiotics were sulfamonomethoxine (SMM), sulfamethoxazole (SMD), oxytetracycline (OTC), and ofloxacin (OFL), and the SMM concentration was highest at 120 842. 74 $ng \cdot L^{-1}$. The removal rate of the ten antibiotics was 99. 23% -100. 00% in Farm A and Farm B. In the wastewater treatment process of Farm A, the treatment section "USR +2A/O + disinfection pond + oxidation pond" removed antibiotics in wastewater effectively, with the total removal rate of SMM, SMD, and OTC reaching 100.00%. In the wastewater treatment process of Farm B, the treatment section "ultrafiltration (UF) + nanofiltration (NF)" removed antibiotics effectively by more than 99. 23%. However, the concentrations of antibiotics investigated in the effluent were higher than the EU water environment antibiotic threshold (10 $ng \cdot L^{-1}$). Finally, through redundancy analysis, it was found that conventional indicators (COD, NH_4^+ -N, TN, TP, and pH) in wastewater were related to the degradation of some antibiotics.

Key words: swine wastewater; veterinary antibiotics; pollution characteristics; removal effect; redundancy analysis

当前,随着畜禽养殖业的集约化、规模化发展, 兽用抗生素被广泛用于预防和治疗动物疾病[1]. 据统计,全球兽用抗生素用量是人用抗生素用量的 2 倍^[2]. 据报道,2012 年美国兽用抗生素的使用量为 1.46×10⁴ t^[3];在2013 年中国抗生素使用量达 1.62×10⁵ t,其中52%被用于畜禽养殖业^[4]. 虽然兽用抗生素能有效预防和治疗动物疾病,但不能被动物完全吸收和代谢. Massé 等^[5]的研究发现,约70%~90%的兽用抗生素以原始形态或代谢产物的形式通

过粪便和尿液排出. Zhou 等[6] 和 Jiang 等[7] 已在养猪场废水、粪便和废水处理过程中产生的污泥中检测到残留抗生素. 然而, 当前养殖场废水处理设施主

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要针对废水中常规污染物的削减而设计,并未考虑抗生素的去除,这使得残留抗生素不断从养殖场排出,最终进入自然环境^[8]. Tasho 等^[9]的研究发现,进入土壤中的抗生素不仅可以通过杀死植物根际微生物来改变土壤微生物的结构和功能,而且还能增加土壤中抗性基因(ARGs)发生的频率和丰度.同时,残留在土壤和水体中的抗生素可通过食物链或饮用水途径进入人体,诱导体内病原体产生抗性,从而降低抗生素治疗疾病的能力^[10]. 因此,最大限度地削减残留在养猪场废水中的抗生素成为当前研究者关注的焦点^[11,12].

当前,国内大多数养猪场只配备简单的处理设 施,如氧化塘和厌氧消化池[13].而有研究发现养殖 场中的氧化塘和厌氧消化池不能有效去除废水中的 抗生素[6]. 近年来, 一些养猪场开始使用更先进的 废水处理单元,如上流式厌氧污泥床(UASB)和厌 氧 + 好氧氧化^[14,15]、膜生物反应器(MBR)^[16]、超滤 (UF)和纳滤(NF)[17],然而当前关于这些处理单元 组合的处理工艺对废水中抗生素的去除报道较少. 同时目前关于贵州养猪场废水中抗生素的研究主要 集中在养猪场污灌区域土壤中重金属和抗生素的复 合污染状况[18],而关于探讨养猪场废水处理工艺中 各处理单元对废水中抗生素的去除鲜见报道. 故本 研究选取贵州 2 家规模化养猪场废水中 10 种兽用 抗生素(6种磺胺类抗生素、3种四环素类抗生素和 1种喹诺酮类抗生素)进行调查,分析养猪场废水中 抗生素的污染特征及其在处理单元中的去除效果, 以期为规模化养殖业中兽用抗生素的污染控制及环 境治理提供理论参考.

1 材料与方法

1.1 仪器与试剂

LC-MS(LC: Agilent Technologies 1290 Infinity; MS: AB SCIEX QTRAP 6470,美国 Agilent 公司);色谱柱(ZORBAX Eclipse Plus C18 1.8 μm 3.0 × 50 mm Column,美国 Agilent 公司);固相萃取装置(24 孔,美国 Waters 公司);HLB 固相萃取小柱(6 mL/500 mg,美国 Waters 公司);电子天平(中国上海仪天科学仪器有限公司);旋转蒸发仪(德国 Heidolph公司).

抗生素标准品包括 3 种四环素类(TCs)抗生素:盐酸四环素(tetracycline hydrochloride,TC)、土霉素(oxytetracycline dihydrate,OTC)和金霉素(chlorotetracycline hydrochloride,CTC),6种磺胺类(SAs)抗生素:磺胺对甲氧嘧啶(sulfameter,SMD)、磺胺甲基噻唑(sulfamethoxazole,SMX)、磺胺间甲氧

嘧啶 (sulfamonomethoxine, SMM)、磺胺二甲嘧啶 (sulfamethazine, SMZ)、磺胺吡啶 (sulfapyidine, SPD) 和磺胺嘧啶 (sulfadiazine, SD), 1 种喹诺酮类 (FQs) 抗生素:氧氟沙星 (ofloxacin, OFL), 购置于百灵威科技有限公司, 纯度均大于 95.0%. 甲醇和氨水 (色谱纯) 均购置于德国的 MERCK 公司, 乙二胺四乙酸二钠、磷酸和磷酸二氢钠 (分析纯) 购置于百灵威科技有限公司. 实验用水为超纯水 (Milli-Q 超纯水系统, 美国 Waters 公司), 0.45 μ m 玻璃纤维滤膜购自 Millipore 公司, 0.22 μ m 有机针孔滤膜购置于南京荣华科学器材有限公司.

10 种抗生素标准品先分别以甲醇为溶剂配制为1 000 mg·L⁻¹的标准储备液,然后再以甲醇作为溶剂,将标准储备液配制为 10 mg·L⁻¹的混合标准物质的储备液,均存储在 - 20℃冰箱中,备用.

1.2 样品的采集

选取 2 种不同处理工艺的规模化养猪场(猪场 A 和猪场 B)为研究对象. 2 家养猪场的存栏量均为 5 000头(母猪),场内配备有废水处理设施,处理规模均为 200 m³·d⁻¹. 2019 年 4 月对这 2 家养猪场按废水处理工艺流程采集样品,猪场 A 的处理工艺为"固液分离+物化沉淀+USR(升流式固体厌氧反应器)+2 级 A/O+消毒池+氧化塘",出水最终排向周边农耕地;猪场 B 的处理工艺为"厌氧发酵+SBR 生化+超滤(UF)+纳滤(NF)",出水最终排向周边水体. 猪场 A 共采集 12 个样品,猪场 B 共采集 6 个样品,具体采样点如图 1 所示. 每个采样点分别采集 2 次并混合均匀,每次采集 1.0 L,共采集 2.0 L. 所有采集水样均为瞬时水样,使用有机玻璃采水器采集表层水样,储存在预先洗净的采样瓶中,低温避光保存,并尽快运回实验室,保存在 4℃的冰箱内待测.

1.3 分析方法

1.3.1 常规指标的分析

废水中常规指标包括:pH、化学需氧量(COD)、 氨氮(NH₄⁺-N)、总氮(TN)和总磷(TP).pH采用 PHSJ-3F计(上海精科雷磁化学仪器公司)进行测 定;NH₄⁺-N采用纳氏试剂光度法测定;TN采用碱性 过硫酸钾消解-紫外分光光度法测定;TP采用钼酸 铵分光光度法测定;COD采用快速消解分光光度法 测定.

1.3.2 抗生素的分析

样品前处理:水样经 $0.45~\mu m$ 的玻璃滤膜过滤去除小颗粒物,然后用磷酸水溶液调节水样 pH 约为 3,每 200~m L 水样加入 1~m L 的 Na_2EDTA ($100~g \cdot L^{-1}$),用 HLB 小柱进行富集,流速控制在 $3~5~m L \cdot m in^{-1}$. 富集前,依次用 6~m L 甲醇、3~m L 超纯

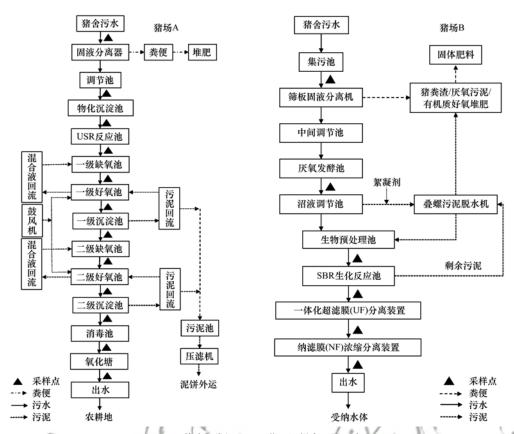


图 1 养猪场粪污处理工艺及取样点设置示意

Fig. 1 Swine treatment process and sampling point setting in a pig farm

水、6 mL pH 约为 3 的磷酸-磷酸二氢钠水溶液(100 g·L⁻¹)活化小柱. 富集完成后用 6 mL 超纯水淋洗,负压真空干燥 40 min 后依次用 6 mL 甲醇、6 mL 2% 氨水甲醇(体积比)洗脱 HLB 小柱,洗脱液用鸡心瓶接收. 洗脱液在 45℃下旋蒸至干, 然后用甲醇溶液定溶至 1 mL, 涡旋振荡 1 min 后, 待 LC-MS 分析.

LC 检测条件: 柱温 40%,流速 0.2 mL·min^{-1} ,流动相 A 为 0.1% 甲酸水 $+2 \text{ mmol·L}^{-1}$ 乙酸铵,流动相 B 为乙腈,进样体积 2μ L,梯度洗脱(洗脱分离步骤见表 1). LC-MS 质谱测定条件:采用电喷雾离子源(ESI)正离子模式、多反应离子监测扫描定量分析目标物.

表 1 梯度洗脱分离步骤

Table 1 Gradient elution separation step

| _ | | | 1 | 1 |
|---|------------|------------------|------------------|-----------------------------|
| | 时间 /min | 流动相 A (体积比)/% | 流动相 B (体积比)/% | 流速 /mL·min ⁻¹ |
| | 5 | 95 | 5 | 0.6 |
| | 17 | 80 | 20 | 0.6 |
| | 18 | 10 | 90 | 0.6 |
| | 20 | 10 | 90 | 0.6 |
| _ | 22 | 95 | 5 | 0.6 |
| | | | | |

1.4 质量控制

采用外标法对样品进行定量分析,以8个不同

梯度的标准物质溶液作定量曲线,线性方程浓度范围为 0~200 μ g·L⁻¹,相关系数 (R^2)值均大于 0.995.同时为了保证实验数据的可靠性,本次实验设置回收率实验.在不含目标化合物的超纯水水样 (100 mL)中添加抗生素标准溶液,按照前述样品处理方法对样品进行前处理.经检测空白水样中目标物回收率为 80.06%~138.43%,相对标准偏差为 1%~10.65%,目标抗生素的检出限为 0.034~1.40 ng·L⁻¹.

1.5 抗生素去除率的计算

抗生素的去除率计算公式如下式所示[11]:

$$AR = [(c_{inf} - c_{enf})/c_{inf}] \times 100\%$$

式中,AR 是每种抗生素的水相去除率,%; c_{inf} 为处理单元进水中抗生素浓度, $ng\cdot L^{-1}$; c_{enf} 为处理单元出水中抗生素浓度, $ng\cdot L^{-1}$.

1.6 数据处理

用 Excel 2013 对实验数据进行处理和分析,采用 Origin 9.1 和 Canoco 5 软件进行作图.

2 结果与讨论

- **2.1** 养猪场废水处理工艺各处理单元对常规水质指标的削减情况
 - 2 家规模化养猪场废水处理工艺各处理单元对

常规水质指标的削减情况如表 2 所示. 猪场 A 和猪场 B 的废水处理工艺对废水中常规污染物(COD、NH₄⁺-N、TN 和 TP)的去除率分别为 97. 10%~99. 75%和 88. 10%~98. 10%. 整体而言,对废水中COD的去除,猪场 A 优于猪场 B;在猪场 A 的 USR反应池中对 COD的去除贡献最大,猪场 B 废水处理中生物预处理池、SBR 生化反应池及 UF 处理单元对 COD的去除效果较好. 在猪场 B 中,NH₄⁺-N和 TN仅在 NF 处理单元被有效去除,这可能是由于纳滤膜孔径和膜表面电荷的共同作用^[19]. 在猪场 A 的

USR 反应池 TP 的浓度高于上一级处理单元,这可能是因为 USR 反应池处于厌氧环境,聚磷菌会在此环境下打开体内聚合磷酸盐的高能磷酸键^[20],从而释放出磷,使得 USR 反应池中磷酸盐浓度增加. 在猪场 A 的出水中 COD、NH₄⁺-N、TN 和 TP 的浓度均满足《畜禽养殖业污染物排放标准》(GB 18596-2001)的要求;在猪场 B 的出水中仅 TP 指标满足GB 18596-2001 的要求,其余 3 项指标均超标,故仍需进行后续处理. 综上,猪场 A 废水处理工艺对废水中常规污染物的去除效果优于猪场 B.

表 2 养猪场废水处理工艺各处理单元对常规水质指标的削减情况

| Table 2 | Reduction of | conventional | water quality | indicators in | various treatmen | t units durin | or the swine | wastewater treatment p | rocece |
|---------|---------------|--------------|---------------|---------------|------------------|----------------|--------------|------------------------|--------|
| rabic 2 | recuuction of | Conventional | water quarity | muicators in | various ireaunen | t units uuring | g me swine | wastewater treatment p | 10005 |

| 猪场 | 采样点 | COD/mg·L ⁻¹ | NH ₄ ⁺ -N/mg•L ⁻¹ | TP/mg·L ⁻¹ | TN/mg·L ⁻¹ | рН |
|-----|-----------|----------------------------|--|-----------------------|--------------------------|-----------------|
| | 原水池 | 7550.00 ± 50.00 | 922.00 ± 18.00 | 288. 10 ± 12. 20 | 1 159.50 ± 35.50 | 6.25 ± 0.02 |
| | 调节池 | 5840.00 ± 45.00 | 505.50 ± 16.00 | 118.16 ± 4.02 | 540.75 ± 16.25 | 6.72 ± 0.00 |
| | 物化沉淀池 | 3750.00 ± 50.00 | 551.75 ± 16.75 | 89.73 ± 0.01 | 562.50 ± 2.00 | 7.34 ± 0.00 |
| | USR 反应池 | $3\ 300.\ 00\ \pm 25.\ 00$ | 220.50 ± 0.70 | 112.96 ± 3.18 | 283.99 ± 14.22 | 7.30 ± 0.01 |
| | 一级缺氧池 | 925.00 ± 75.00 | 181.30 ± 4.30 | 78.66 ± 0.01 | 249.38 ± 3.63 | 7.61 ± 0.01 |
| A | 一级好氧池 | 605.00 ± 35.00 | 38.25 ± 1.65 | 76.50 ± 2.65 | 96.70 ± 0.01 | 7.34 ± 0.01 |
| А | 一级生化沉淀池 | 420.00 ± 30.00 | 30.50 ± 3.90 | 79.44 ± 0.26 | 60.05 ± 3.85 | 7.38 ± 0.01 |
| | 二级缺氧池 | 250.00 ± 10.00 | 32.98 ± 0.76 | 72.97 ± 2.55 | 43.45 ± 1.45 | 7.52 ± 0.01 |
| | 二级好氧池 | 360.00 ± 10.00 | 24.40 ± 0.00 | 70.74 ± 1.48 | 56.70 ± 0.90 | 7.56 ± 0.01 |
| | 二级生化沉淀池 | 272.50 ± 7.50 | 8.49 ± 0.21 | 65.45 ± 0.85 | 54.05 ± 0.01 | 7.50 ± 0.01 |
| | 消毒池 | 238.00 ± 15.50 | 2.89 ± 0.06 | 10.46 ± 0.01 | 34.06 ± 0.01 | 7.63 ± 0.01 |
| - | 氧化塘 | 146.50 ± 3.50 | 2.29 ± 0.42 | 4.64 ± 0.01 | 33.65 ± 0.01 | 8.06 ± 0.03 |
| / | 去除率/% | 98. 06 | 99.75 | 98. 39 | 97. 10 | ~ |
| 7 | 原水池 | 8900.00 ± 65.00 | 1973.00 ± 19.00 | 65.19 ± 0.01 | 2019.50 ± 49.50 | 7.32 ± 0.01 |
| | 厌氧发酵池 | 7625.00 ± 50.00 | 1925.00 ± 25.00 | 70.53 ± 0.75 | $2\ 197.00 \pm 50.50$ | 7.32 ± 0.01 |
| | 生物预处理池 | 5725.00 ± 50.00 | 1771.50 ± 3.00 | 31.67 ± 0.18 | $2\ 087.50 \pm 66.50$ | 8.01 ± 0.01 |
| B | SBR 生化反应池 | $4\ 275.\ 00\ \pm 75.\ 00$ | 1670.31 ± 33.55 | 40.03 ± 0.25 | $2\ 017.50 \pm 22.50$ | 8.24 ± 0.01 |
| 100 | UF / | $1\ 387.50 \pm 22.50$ | 1786.50 ± 5.50 | 13.08 ± 0.01 | $2\ 289.\ 00\pm 69.\ 00$ | 8.21 ± 0.01 |
| - (| NF | 975.00 ± 25.00 | 198.70 ± 7.50 | 1.24 ± 0.01 | 240.25 ± 11.75 | 8.66 ± 0.01 |
| _ \ | 去除率/% | 88. 04 | 89. 93 | 98. 10 | 88. 10 | _ |

2.2 目标抗生素在各处理单元水相中的残留情况

养猪场废水处理单元水相中抗生素的残留情况 如图 2 所示. 在猪场 A 废水处理工艺的废水中,共 检出6种抗生素,包括3种SAs、2种TCs和1种 FQs 类抗生素,浓度范围在 ND~47 029.25 ng·L⁻¹ 之间. SAs 类抗生素中 SMM 和 SMD 的检出率较高, 均为58.33%,其中,SMM 在原水池、调节池、物化沉 淀池、USR反应池、二级生化沉淀池和消毒池中的 残留浓度较高,浓度分别为3338.82、4955.09、 11 561.13、 6 450.26、 47 029.25 和 1 770.93 ng·L⁻¹,这与周婧等^[21]的研究结果相似. TCs 类抗 生素中 OTC 在废水中的检出率为 91.67%,浓度范 围在 ND~37 976.04 ng·L⁻¹之间,其中在进水处的 检出浓度最高;同时也发现 OTC 是猪场 A 废水中被 检出浓度最高的抗生素,这与 Ben 等[22]的研究结果 相似,表明在猪场 A 中 OTC 的使用量较大. TC 在废 水进水中未检出,而在出水中被检出,这可能是因为 进水中存在目标抗生素的代谢产物,随后在生物处理过程中转化为其母体化合物^[15]. FQs 类抗生素中OFL 在各处理单元中的检出率为83.33%,浓度在ND~22714.43 ng·L⁻¹之间,这可能是由于OFL 在废水中的降解速率较慢和在废水中的累积效应所导致^[23]. 整体而言,猪场 A 废水处理工艺中抗生素残留规律为OFL(136157.92 ng·L⁻¹)>OTC(117472.08 ng·L⁻¹)>SMM(66837.02 ng·L⁻¹)>SMD(60109.32 ng·L⁻¹)>TC(9611.92 ng·L⁻¹)>SD(91.87 ng·L⁻¹),在二级生化沉淀池处的浓度最高,总浓度达124151.44 ng·L⁻¹.

在猪场 B 废水处理工艺中, 共检出 7 种抗生素,包括 4 种 SAs、2 种 TCs 和 1 种 FQs 类抗生素,浓度范围为 ND ~120 842. 74 ng·L $^{-1}$. SAs 类抗生素中 SD、SMD 和 SMM 的检出率分别为 66.67%、100.00%和 100.00%,在厌氧发酵池出水处 3 个物质均被检出,浓度分别为 81.09、108 730.47 和

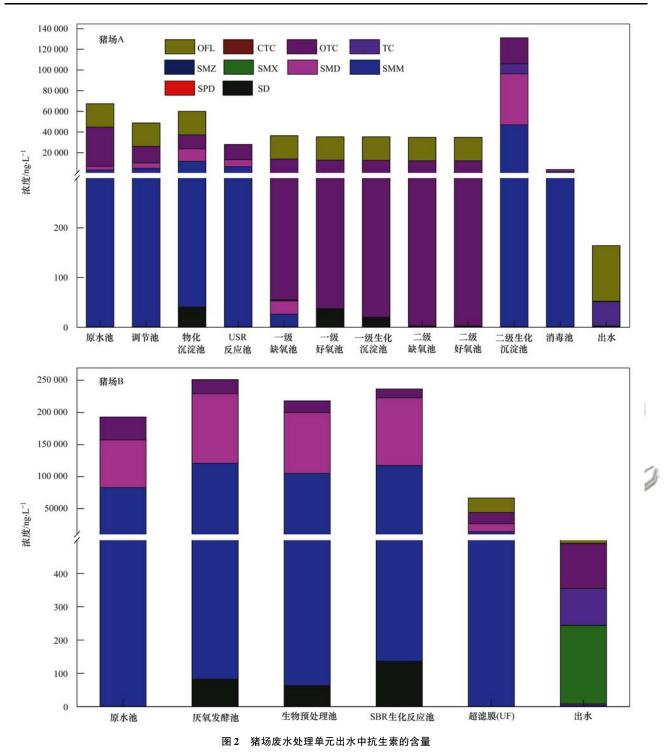


Fig. 2 Content of antibiotics in the effluent of the pig farm wastewater treatment unit

120 842. 74 $ng \cdot L^{-1}$, 高于陈永山等^[24]的研究. TCs 类抗生素中 OTC 的检出率为 100%, 这与 Wang 等^[12]的研究结果相似; TC 仅在出水处被检出、OFL 仅在 UF 和出水处检出,这可能是由于废水中 TC 和 OFL 的共轭形态向游离形态转化的结果^[25]. 整体而言,猪场 B 废水处理工艺中抗生素的残留规律为 SMM(357 427. 75 $ng \cdot L^{-1}$) > SMD(321 614. 68 $ng \cdot L^{-1}$) > OTC(71 756. 52 $ng \cdot L^{-1}$) > OFL (23 094. 08 $ng \cdot L^{-1}$) > SD (278. 55 $ng \cdot L^{-1}$) > TC

(110.76 $\operatorname{ng} \cdot L^{-1}$),在厌氧发酵池出水处的总浓度最高,达251 123.56 $\operatorname{ng} \cdot L^{-1}$.

在2家猪场废水处理工段中,SPD、SMX、SMZ和CTC均未被检出,而SMZ和CTC在陈永山等^[24]的研究中均被检出,这可能是由于不同地区养殖场对兽用抗生素的利用模式差异所致^[26].同时发现在2家养猪场废水中OTC的检出浓度均较高,这与其它地区养猪场^[24,27]和养牛场^[28]废水中OTC的残留情况相似,表明OTC为大部分养殖场所热衷使用的抗生素.

2.3 不同处理单元对养猪场废水中抗生素的去除效果

养猪场废水处理单元中抗生素的去除情况如表 3 所示. 在猪场废水处理工艺的一级处理过程中,猪场 A 对 \sum SAs、 \sum TCs 和 \sum FQs 的去除率分别为 -244.81%、64.29% 和 -0.11%. 其中对 SMM、SMD 和 OFL 的去除均出现负迁移现象(除在物化沉淀池对 OFL 的去除为正外),这可能是由于废水中

细小的粪便或颗粒包裹进水中部分抗生素(SMM、SMD和OFL),在处理单元中被微生物破坏了抗生素与粪便或颗粒之间的屏障,从而增加该处理单元中 SMM、SMD和OFL的浓度^[29].而对OTC的去除均为正向去除,其中在调节池中的去除率较高,为57.69%,这可能是因为在调节池处理单元时,OTC的去除除通过污泥吸附外,还可能通过光降解、水解等途径进行降解^[30].

表 3 猪场各处理单元中抗生素的水相去除率/%

Table 3 Water phase removal rate of antibiotics in each treatment unit of the farm/%

| | | Table 3 | water phase | removal rate | or antibiotic | | | | | | |
|-----|-------|-----------|-------------|--------------|---------------|-----------------|--------|----------|-----------------|--------|---------------|
| 猪场 | 友 | 上理单元 | SD | SMM | SMD | $\sum { m SAs}$ | TC | OTC | $\sum { m TCs}$ | OFL | $\sum FQs$ |
| | | 调节池 | _ | -48.41 | -45.7 | -47.12 | _ | 57.69 | 57.69 | -0.31 | -0.31 |
| | 一级处理 | 物化沉淀池 | _ | -133.32 | -134.66 | -134.38 | _ | 15.60 | 15.60 | 0.20 | 0.20 |
| | | 总去除率 | _ | -246.26 | -241.9 | -244.81 | _ | 64.29 | 64.29 | -0.11 | -0.11 |
| | | USR 反应池 | 100.00 | 44.21 | 44. 21 | 44.31 | _ | -9.28 | -9.28 | 100.00 | 100.00 |
| | | 一级缺氧池 | _ | 99.61 | 99.64 | 99.62 | _ | 6.46 | 6.46 | _ | _ |
| | | 一级好氧池 | _ | 100.00 | 100.00 | 24.63 | _ | 7.76 | 7.76 | 0.22 | 0.22 |
| | 二级处理 | 一级生化沉淀池 | 50.84 | _ | _ | 50.84 | _ | 1.31 | 1.31 | -0.63 | -0.63 |
| A | 二次是生 | 二级缺氧池 | 100.00 | _ | _ | 100.00 | - | 2.70 | 2.70 | 0.00 | 0.00 |
| | | 二级好氧池 | _ | -0 | _ | _ | + 0 | 0.22 | 0.22 | 0.19 | 0.19 |
| | | 二级生化沉淀池 | _ | | _ | _ | + % | - 106.09 | - 184. 15 | 100.00 | 100.00 |
| | | 总去除率 | | -306.79 | - 306.95 | -306.13 | H | -86.18 | -86.18 | 100.00 | 100.00 |
| | | 消毒池 | <i></i> | 96.23 | 96. 26 | 96.25 | 100.00 | 99.81 | 99.86 | لحا | 4 |
| _ | 三级处理 | 氧化塘 | + 1 | 100.00 | 100.00 | 100.00 | (m | 100.00 | -0.33 | 2.42 | 2.42 |
| | 1 / | 总去除率 | +0 | 100.00 | 100.00 | 100.00 | 99.49 | 100.00 | 99.86 | _ | (- |
| (0) | Tag | 七总去除率 | 20 | 100.00 | 100.00 | 100.00 | 1-0 | 100.00 | 99.87 | 99.51 | 99.51 |
| - 1 | 16 | 厌氧发酵池 | 71 | -45.86 | -45.84 | -45.91 | / #\ | 39.78 | 39.78 | -00 | ` <u>"</u> ji |
| _ | 二级处理 | 生物预处理池 | 23.33 | 12.86 | 12.85 | 12.86 | 1 2 / | 14.46 | 14.46 | _ | SF |
| (- | | SBR 生化反应池 | -117.6 | - 11.43 | - 11. 44 | -11.47 | No 10 | 23.86 | 23.86 | _ | _ |
| 19 | 1110 | 总去除率 | - +2 | -41.63 | -41.65 | -41.73 | _ | 60.78 | 60.78 | _ | _ |
| В/ | Po II | UF | 100.00 | 88.13 | 88. 14 | 88.14 | _ | -25.4 | -25.4 | _ | _ |
| 1 | 三级处理 | NF | _ | 99.97 | 99.97 | 99.08 | _ | 99.24 | 98.61 | 98.23 | 98.23 |
| A | | 总去除率 | 99.99 | 99.99 | 99.99 | 99.89 | _ | 99.04 | 98.25 | _ | _ |
| | 工艺 | 总去除率 | _ | 99.99 | 99.99 | 99.85 | _ | 99.23 | 99.31 | _ | _ |

在二级处理中,猪场 A 废水处理工段对 \sum SAs、 \sum TCs 和 \sum FQs 的去除率分别为 -306.13%、-86.18% 和 100.00%. 在 USR 反应 池、一级缺氧池和二级缺氧池中,对 SMM、SMD 和 SD 的去除效果较好(去除率为 44.21%~100.00%),原因可能是 SMM、SMD 和 SD 分子结构中存在氮原子和硫原子,在厌氧条件下易于生物降解 $[^{31,32}]$. 据报道 SAs 类抗生素的 pK_a 比较低,在酸性条件下带正电,而在碱性条件下带负电 $[^{33}]$. 猪场 A 的二级处理过程中废水的 pH 均大于 7,然而 SAs 类抗生素又属于高亲水化合物 $[^{34}]$,很难从水相转移到污泥相,因此在此阶段中猪场废水中 SAs 类抗生素的 $1gK_d$ 值在 $3.7 \sim 4.1$ 1.0000 L·kg 1.0000 L·kg

抗生素的去除^[32],这可能是厌氧发酵池中 OTC 去除效果较好的原因之一.

在猪场 A 和猪场 B 的三级处理中对 ∑ SAs 和 ∑ TCs 的去除率分别为 100.00%、99.86% 和 99.89%、98.25%. 其中在猪场 A 的氧化塘处理单元中,对 SMM、SMD 和 OTC 的去除率均为 100.00%,这可能是由于氧化塘处理单元中废水的水力停留时间长^[38],较长的水力停留时间能延长抗生素和微生物的接触时间,同时也能延长植物对抗生素吸收的时间,有利于对抗生素的去除^[39].在猪场 B 的 UF 处理单元中对 SD、SMM 和 SMD 的去除率高于 88.13%. NF 对目标抗生素的去除率均高于 99.23%,表明 NF 能有效去除猪场废水中的抗生素;这可能是因为 NF 的膜孔径一般小于 2 nm,能够有效截留水中的抗生素^[40].同时已有研究表明,多种膜处理技术组合能更好地去除废水中的抗生素^[41].

综上,从对目标抗生素的去除效果而言,猪场 A的"固液分离+物化沉淀+USR+二级 A/O+消毒池+氧化塘"处理工艺优于猪场 B的"厌氧发酵+SBR生化反应+UF+NF"处理工艺.猪场 A废水处理工艺去除效果较好的原因归因于"USR+2级 A/O"处理工段的生物降解和氧化塘处理单元中的水力停留时间较长;猪场 B废水中抗生素的去除大部分归因于 UF和 NF的组合工艺的去除.同时发现2家猪场废水处理工艺对兽用抗生素的去除效果都较好(总去除率>99.23%).但最终出水中兽用抗生素的浓度大部分都高于欧盟水环境抗生素阈值(10 ng·L⁻¹)[42],因此,最终出水中残留的兽用抗生素排放可能会引起一系列的环境风险需要引起高度重视.

2.4 常规指标与抗生素之间的相关关系

采用冗余分析(RDA)分析了 2 家规模化养猪 场废水的常规指标(COD、NH₄⁺-N、TN、TP 和 pH)与 废水中抗生素之间的关系,结果如图 3 所示. 猪场 A 和猪场 B 的废水中常规参数与抗生素在第 1 轴、第 2 轴的解释量分别为 79. 10%、19. 17% 和 97. 04%、2. 11%,累计解释量分别为 98. 27% 和 99. 15%,由此 可见前两轴能够很好地反映废水常规指标与抗生素 之间的关系,且主要由第 1 轴决定. 在猪场 A 中,pH 与 TC、SMD 和 SMM 呈正相关,与 OFL 和 OTC 呈负相关;COD、NH₄⁺-N、TN 和 TP 与 OTC 和 OFL 呈正相关,与 TC 和 SMX 呈正相关,与 SMD、SMM、SD 和 OTC 呈负相关;COD、NH₄⁺-N、TN 和 TP 与 SMD、SMM 和 OTC 呈正相关,与 OFL、TC 和 SMX 呈负相关. 综上,

表明废水中常规指标(COD、NH₄⁺-N、TN、TP和pH)与抗生素的降解具有相关性,其中NH₄⁺-N与OTC、pH与TC具有较大的正相关性,COD、NH₄⁺-N、TN和TP与TC呈负相关性,pH与OTC呈负相关性,这可能意味着部分抗生素会随着微生物对有机物的生物降解而降解^[12].但是,在各处理单元中抗生素的去除机制还有待进一步研究.

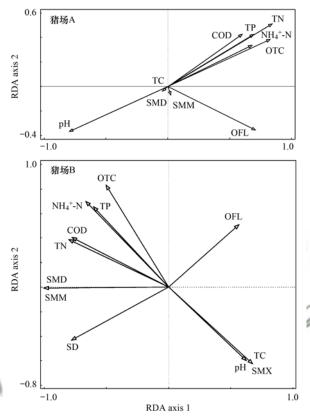


图 3 养猪场废水中常规指标与抗生素之间的关系

Fig. 3 Relationship between conventional indicators and antibiotics in pig farm wastewater

3 结论

- (1)2 家规模化养猪场(猪场 A 和猪场 B)废水处理工艺对常规污染物(COD、 NH_4^+ -N、TN 和 TP)的 去除率在 88.10% 以上,猪场 A 出水中 COD、 NH_4^+ -N、TN 和 TP 均满足《畜禽养殖业污染物排放标准》(GB 18596-2001)的要求,而猪场 B 出水中仅 TP 满足 GB 18596-2001 的要求.
- (2)目标抗生素在不同的养猪场和不同的废水处理单元中的残留情况不尽相同,猪场 A 和猪场 B 的废水处理工艺中抗生素的残留规律分别为:OFL > OTC > SMM > SMD > TC > SD 和 SMM > SMD > OTC > OFL > SD > TC. 主要的污染单体为 SMM、SMD、OTC 和 OFL,最高单体污染浓度达120 842.74 $\operatorname{ng} \cdot \operatorname{L}^{-1}(\operatorname{SMM})$.
 - (3)2家养猪场废水处理工艺对兽用抗生素的

去除效果都较好(总去除率 > 99.23%). 猪场 A 废水处理工艺中"USR + 2 级 A/O + 消毒池 + 氧化塘"组合工段能有效去除废水中的抗生素,其中对SMD、SMM和OTC的去除率均达100%;在猪场 B 废水处理工艺中,"UF + NF"组合工段对目标抗生素的去除效果较好,最终废水中99.23%以上的抗生素被去除. 总体上,去除效果猪场 A 废水处理工艺优于猪场 B.

(4)通过冗余分析发现,在 2 家养猪场废水中, pH 与 TC 均呈正相关关系、与 OTC 呈负相关关系; COD、TN、 NH_4^+ -N和 TP 与 OTC 呈正相关关系、与 TC 呈负相关关系.

参考文献:

- [1] Sarmah A K, Meyer M T, Boxall A B A. A global perspective on the use, sales, exposure pathways, occurrence, fate and effects of veterinary antibiotics (VAs) in the environment [J]. Chemosphere, 2006, 65(5): 725-759.
- [2] Aarestrup F. Sustainable farming: get pigs off antibiotics [J]. Nature, 2012, 486 (7404): 465-466.
- [3] Teillant A, Laxminarayan R. Economics of antibiotic use in U. S. swine and poultry production [J]. Choices, 2015, 30(1): 1-11.
- [4] Zhang Q Q, Ying G G, Pan C G, et al. Comprehensive evaluation of antibiotics emission and fate in the river basins of China: source analysis, multimedia modeling, and linkage to bacterial resistance [J]. Environmental Science & Technology, 2015, 49(11): 6772-6782.
- [5] Massé D I, Saady N M C, Gilbert Y. Potential of biological processes to eliminate antibiotics in livestock manure: an overview [J]. Animals, 2014, 4(2): 146-163.
- [6] Zhou L J, Ying G G, Zhang R Q, et al. Use patterns, exerction masses and contamination profiles of antibiotics in a typical swine farm, South China [J]. Environmental Science: Processes & Impacts, 2013, 15(4): 802-813.
- [7] Jiang H Y, Zhang D D, Xiao S C, et al. Occurrence and sources of antibiotics and their metabolites in river water, WWTPs, and swine wastewater in Jiulongjiang River basin, South China [J]. Environmental Science and Pollution Research, 2013, 20 (12): 9075-9083.
- [8] Ebele A J, Abou-Elwafa Abdallah M, Harrad S. Pharmaceuticals and personal care products (PPCPs) in the freshwater aquatic environment [J]. Emerging Contaminants, 2017, 3(1): 1-16.
- [9] Tasho R P, Cho J Y. Veterinary antibiotics in animal waste, its distribution in soil and uptake by plants; a review[J]. Science of the Total Environment, 2016, 563-564; 366-376.
- [10] Li W H, Shi Y L, Gao L H, et al. Occurrence, distribution and potential affecting factors of antibiotics in sewage sludge of wastewater treatment plants in China [J]. Science of the Total Environment, 2013, 445-446; 306-313.
- [11] Zhang M, Liu Y S, Zhao J L, et al. Occurrence, fate and mass loadings of antibiotics in two swine wastewater treatment systems
 [J]. Science of the Total Environment, 2018, 639: 1421-1431.
- [12] Wang R, Feng F, Chai Y F, et al. Screening and quantitation of residual antibiotics in two different swine wastewater treatment systems during warm and cold seasons [J]. Science of the Total Environment, 2019, 660: 1542-1554.
- [13] He L Y, Ying G G, Liu Y S, et al. Discharge of swine wastes

- risks water quality and food safety: antibiotics and antibiotic resistance genes from swine sources to the receiving environments [J]. Environment International, 2016, **92-93**: 210-219.
- [14] Nuengjamnong C, Rachdawong P. Performance analysis of the combined plug-flow anaerobic digester (PFAD) and upflow anaerobic sludge blanket (UASB) for treating swine wastewater in Thailand[J]. Thai Veterinary Medicine, 2016, 46(3): 435-442.
- [15] Göbel A, Thomsen A, McArdell C S, et al. Occurrence and sorption behavior of sulfonamides, macrolides, and trimethoprim in activated sludge treatment [J]. Environmental Science & Technology, 2005, 39(11): 3981-3989.
- [16] García Galán M J, Díaz-Cruz S, Barceló D. Removal of sulfonamide antibiotics upon conventional activated sludge and advanced membrane bioreactor treatment [J]. Analytical and Bioanalytical Chemistry, 2012, 404(5): 1505-1515.
- [17] 张鹏飞,刘晓文,李杰,等. 养殖废水中抗生素去除处理工艺的研究现状[J]. 净水技术, 2018, 37(4): 60-65, 95.

 Zhang P F, Liu X W, Li J, et al. Current research in treatment processes for antibiotics removal from livestock wastewater [J].

 Water Purification Technology, 2018, 37(4): 60-65, 95.
- [18] 刘艳萍, 刘鸿雁, 吴龙华, 等. 贵阳市某蔬菜地养殖废水污灌土壤重金属,抗生素复合污染研究[J]. 环境科学学报, 2017, 37(3): 1074-1082.

 Liu Y P, Liu H Y, Wu L H, et al. Co-contamination of heavy metals and antibiotics in soils under husbandry wastewater irrigation in Guiyang City [J]. Acta Scientiae Circumstantiae, 2017, 37(3): 1074-1082.
- [19] 徐兵,杨程. 纳滤膜处理微污染原水的中试研究[J]. 给水排水,2016,42(11),11-15.
- [20] 任健, 李军, 苏雷、等. 酸化液对厌氧释磷好氧吸磷速率的影响研究[J]. 环境工程, 2011, **29**(S1): 103-107, 317. Ren J, Li J, Su L, *et al.* Study on the effect of dosing hydrolysate on anaerobic phosphorus release and aerobic phosphorus uptake rate[J]. Environmental Engineering, 2011, **29**(S1): 103-107, 317.
- [21] 周婧, 支苏丽, 宫祥静, 等. 三类抗生素在两种典型猪场废水处理工艺中的去除效果[J]. 农业环境科学学报, 2019, **38**(2): 430-438.

 Zhou J, Zhi S L, Gong X J, *et al.* The removal effect of three classes of antibiotics in two typical swine wastewater treatment

systems [J]. Journal of Agro-Environment Science, 2019, 38

[22] Ben W W, Pan X, Qiang Z M. Occurrence and partition of antibiotics in the liquid and solid phases of swine wastewater from concentrated animal feeding operations in Shandong Province, China[J]. Environmental Science: Processes & Impacts, 2013, 15(4): 870-875.

(2): 430-438.

- [23] 姜凌霄. 鄱阳湖区典型养猪场废水抗生素污染特征及催化降解研究[D]. 南昌:南昌航空大学,2012.
- [24] 陈永山,章海波,骆永明,等. 典型规模化养猪场废水中兽 用抗生素污染特征与去除效率研究[J]. 环境科学学报, 2010, **30**(11): 2205-2212.
 - Chen Y S, Zhang H B, Luo Y M, et al. A preliminary study on the occurrence and dissipation of antibiotics in swine wastewater [J]. Acta Scientiae Circumstantiae, 2010, 30 (11): 2205-2212.
- [25] Hu J, Zhou J, Zhou S Q, et al. Occurrence and fate of antibiotics in a wastewater treatment plant and their biological effects on receiving waters in Guizhou [J]. Process Safety and Environmental Protection, 2018, 113: 483-490.

- [26] Managaki S, Murata A, Takada H, et al. Distribution of macrolides, sulfonamides, and trimethoprim in tropical waters: ubiquitous occurrence of veterinary antibiotics in the Mekong Delta [J]. Environmental Science & Technology, 2007, 41 (23): 8004-8010.
- [27] 魏瑞成, 葛峰, 陈明, 等. 江苏省畜禽养殖场水环境中四环 类抗生素污染研究[J]. 农业环境科学学报, 2010, **29**(6): 1205-1210.
 - Wei R C, Ge F, Chen M, et al. Pollution of tetracyclines from livestock and poultry farms in aquatic environment in Jiangsu Province, China [J]. Journal of Agro-Environment Science, 2010, 29(6): 1205-1210.
- [28] 陈乾,赵润,牟美睿,等.天津市规模化奶牛养殖场废水中典型抗生素处理效果及生态风险评估[J].环境科学,2019, 40(11):5015-5023.
 - Chen Q, Zhao R, Mou M R, et al. Treatment effect and ecological risk assessment of typical antibiotics in wastewater from large-scale dairy farms in Tianjin [J]. Environmental Science, 2019, 40(11): 5015-5023.
- [29] Blair B, Nikolaus A, Hedman C, et al. Evaluating the degradation, sorption, and negative mass balances of pharmaceuticals and personal care products during wastewater treatment[J]. Chemosphere, 2015, 134: 395-401.
- [30] 李慧. 四环素类抗生素(TCs)在活性污泥处理系统中的去除 行为研究[D]. 泰安: 山东农业大学, 2013.
- [31] Wijekoon K C, McDonald J A, Khan S J, et al. Development of a predictive framework to assess the removal of trace organic chemicals by anaerobic membrane bioreactor [J]. Bioresource Technology, 2015, 189: 391-398.
- [32] Mohring S A I, Strzysch I, Fernandes M R, et al. Degradation and elimination of various sulfonamides during anaerobic fermentation; a promising step on the way to sustainable pharmacy? [J]. Environmental Science & Technology, 2009, 43(7): 2569-2574.
- [33] Haller M Y, Müller S R, McArdell C S, et al. Quantification of veterinary antibiotics (sulfonamides and trimethoprim) in animal manure by liquid chromatography-mass spectrometry[J]. Journal of Chromatography A, 2002, 952(1-2): 111-120.
- [34] Rosal R, Rodríguez A, Perdigón-Melón J A, et al. Occurrence

- of emerging pollutants in urban wastewater and their removal through biological treatment followed by ozonation $[\ J\]$. Water Research, 2010, 44(2): 578-588.
- [35] 李士俊, 谢文明. 污水处理厂中抗生素去除规律研究进展[J]. 环境科学与技术, 2019, **42**(3): 17-29.

 Li S J, Xie W M. Research advances in antibiotics removal in wastewater treatment plants: a review[J]. Environmental Science & Technology, 2019, **42**(3): 17-29.
- [36] 靳红梅, 黄红英, 管永祥, 等. 规模化猪场废水处理过程中 四环素类和磺胺类抗生素的降解特征[J]. 生态与农村环境 学报, 2016, **32**(6): 978-985. Jin H M, Huang H Y, Guan Y X et al. Characteristics of degradation tetracyclines and sulfonamides during wastewater treating processes in an intensive swine farm [J]. Journal of Ecology and Rural Environment, 2016, **32**(6): 978-985.
- [37] Cheng D L, Ngo H H, Guo W S, et al. Bioprocessing for elimination antibiotics and hormones from swine wastewater [J]. Science of the Total Environment, 2018, 621: 1664-1682.
- [38] Knight R L, Payne Jr V W E, Borer R E, et al. Constructed wetlands for livestock wastewater management [J]. Ecological Engineering, 2000, 15(1-2): 41-55.
- [39] Liao P, Zhan Z Y, Dai J, et al. Adsorption of tetracycline and chloramphenicol in aqueous solutions by bamboo charcoal: a batch and fixed-bed column study [J]. Chemical Engineering Journal, 2013, 228: 496-505.
- [40] 赵涛, 丘锦荣, 蒋成爱, 等. 水环境中磺胺类抗生素的污染现状与处理技术研究进展[J]. 环境污染与防治, 2017, 39 (10): 1147-1152.

 Zhao T, Qiu J R, Jiang G A, et al. Research progress in pollution status and treatment technologies of sulfonamides in aquatic environment [J]. Environmental Pollution & Control, 2017, 39(10): 1147-1152.
- [41] Košutić K, Dolar D, Ašperger D, et al. Removal of antibiotics from a model wastewater by RO/NF membranes [J]. Separation and Purification Technology, 2007, 53(3): 244-249.
- [42] Calamari D, Zuccato E, Castiglioni S, et al. Strategic survey of therapeutic drugs in the Rivers Po and Lambro in Northern Italy [J]. Environmental Science & Technology, 2003, 37 (7): 1241-1248.

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| Sunflower-Straw-Derived Biochar-Enhanced Fe(\mathbb{II})/S $_2$ 0 $_8$ $^-$ System for Degradation of Benzoic Acid | G Yu, et al. (23 | 301) |
| Non-activated Peroxymonosulfate-Induced Degradation of Sulfasalazine; Kinetics and Mechanism Investigations DING Xi, ZHANG Xue-wei, ZHOU Run- | sheng, et al. (23 | 310) |
| Pollution Characteristics and Removal Effects of Antibiotics in Wastewater from Large-Scale Pig Farms in Guizhou | -liang, et al. (23 | 320) |
| Nitrogen Removal in Low-C/N Rural Sewage Treatment by Anoxic/Oxic Biofilter Packed with New Types of Fillers ZHAO Yuan-zhe, YANG Yong-zhe, WANG H. | ai-yan, et al. (23 | 329) |
| Nitrogen Removal Performance and Nitrogen/Carbon Balance of Oligotrophic Aerobic Denitrifiers | ng-lin, et al. (23 | 339) |
| Bacterial Community Shifts and Nitrogen Removal Characteristics for a SNAD Process Treating Anaerobic Digester Liquor of Swine Wastewater (ADLSW) in a Continuous-Flow B | | |
| Reactor (CFBR) | Hang, et al. (23 | 349) |
| Characteristics of ANAMMOX Granular Sludge and Differences in Microbial Community Structure Under Different Culture Conditions JIANG Ying, GUO Meng-lei, XIE Jun- | xiang, et al. (23 | 358) |
| Start-up of an Integrated Process of Denitrifying Phosphorus Removal Coupled with Partial Nitritation and Anaerobic Ammonium Oxidation | | |
| CHEN Ya, YIN Wen, ZHANG Xin | g-xing, et al. (23 | 367) |
| Construction and Application of an Evaluation System for Soil Environmental Carrying Capacity | | |
| Synthesis of Magnetic Biochar and Its Application in the Remediation of Heavy-Metal-Contaminated Soils | | |
| Remediation of Heavy-Metal-Contaminated Soil by EGTA Washing Enhanced with Reduction Solubilization ZHANG Jin-yong, ZHU Yu-ting, WANG Min | | |
| Toxicity of Chromium to Root Growth of Barley as Affected by Chromium Speciation and Soil Properties | | |
| Effect of Fertilizer Reduction and Biochar Application on Soil Nitrogen Loss in Purple Upland WANG Shu, WANG Zi-fang, LO! | | |
| Relationship Between the Composition of Soil Aggregates and the Distribution of Organic Carbon Under Long-Term Abandoned Restoration | | .00) |
| WANG Xing, ZHONG Ze-kun, ZHANG | Xin-vi et al (24 | 416) |
| Effects of Biochar Application on Soil Microbial Nutrient Limitations and Carbon Use Efficiency in Lou Soil WANG Qiang, GENG Zeng-chao, XU Cher | | |
| N ₂ O Emissions from Tea Plantations with Sorghum Intercropping and Application of Big Urea Pills | | |
| | | |
| Effects of Different Forest Vegetation Types on Soil Nitrogen-Related Microbial Communities and Functions in Jinyun Mountain WANG Ying-yan, WANG Fu-hua, LUO Doi Nitrogen Order Enjoying and Depitrifying Restorial Communities as Affected by Drip Injection with Soline Water in Cotton Fields | | |
| Nitrous Oxide Emission and Denitrifying Bacterial Communities as Affected by Drip Irrigation with Saline Water in Cotton Fields GUO Hui-nan, MA Li-juan, HUANG Z | | |
| | | |
| Effect of Organic Matter Promotion on Nitrogen-Cycling Genes and Functional Microorganisms in Acidic Red Soils | Imam 1 / 04 | |
| Effects of Simulated Acid Rain on Soil Fungi Diversity in the Transition Zone of Moso Bamboo and Broadleaf Forest WANG Nan, PAN Xiao-cheng, WANG Chuan | | |
| | ie-zhu, et al. (24 | 485) |