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目 次

V 为 Z 这种 max 八 大板 D 没 在 D 中 取 的 D 主	(1012)
长江经济带 $PM_{2.5}$ 分布格局演变及其影响因素	(1013)
阳泉市秋冬季 PM。化学组分及来源分析 ····································	(1025)
泰安市夏季 PM _{2.5} 中正构烷烃和糖类化合物的化学组成及其来源······	()
	(1045)
更出秋李天气颗粒物埋化符性 下逸韵,银燕,土红霜,陎魁 北碚区与滚晾光学厚度蛙征及甘与颗粒物浓度的相关性	(1056)
淄博市重点工业行业 VOCs 排放特征 ····································	(1007)
鄂州市大气 VOCs 污染特征及来源解析 ····································	(1085)
泰安市夏季 PM _{2.5} 中正构烷烃和糖类化合物的化学组成及其来源	(1093)
基丁二乙过程的金属包表业 VOCs 打架行证	(1099)
················ 牛真真, 孔少飞, 严沁, 郑淑睿, 郑煌, 曾昕, 姚立全, 吴剑, 张颖, 吴方琪, 程溢, 覃思, 刘玺, 燕莹莹, 祁士华	(1107)
精细化工园区工艺过程 VOCs 产生量核算方法····································	(1116)
2017 年春李常州 HONO 观测及对大气氧化能力影响的评估 ····································	(1123)
中国工任机械使用特征及共尾气採成趋势	(1132)
在用汽油和柴油车排放颗粒物的粒径分布特征实测	(11.0)
王瑞宁, 胡磬遥, 任洪娟, 马冬, 徐冲, 赵玺乾, 王孟昊, 徐为标, 安静宇, 黄成	(1151)
参数选取对畜禽养殖业大气氨排放的影响:以长三角地区为例 ************************************	(1150)
□ 张琪, 黄凌, 殷司佳, 王倩, 李红丽, 王杨君, 王军, 陈勇航, 李莉乌海市煤矿区及周边春季降尘污染特征及来源分析 □ 吴红璇, 史常青, 张艳, 赵廷宁, 胡平, 郑肃, 陈章	(1158)
长江中下游地区丰水期刊、湖水图、英门位系组成特值 学静,天华武,周水独,赵中华,土晓龙,祭水久,负斌,陈芠,孙伟	(11/6)
伊洛河流域河水来源及水化学组成控制因素 刘松韬 张东 李玉红 杨锦媚 邹霜 干永涛 黄兴宇 张忠义 杨伟 贾保军	(1184)
城镇化进程中新疆塔城盆地浅层地下水化学演变特征及成因 ····································	(1197)
基丁 Sentinel-2 MSI 彩啄的何例系统小体态仔初至间分开遥恐监测:以安徽自开壶例与连按长江投为例 工作有,工杰, 崔玉小苗十高原高浑沌水体 CDOM 光学特性及影响因素 梁晓文 邵田田 王涛	(1207)
黄土高原高浑浊水体 CDOM 光学特性及影响因素 梁晓文,邵田田,王涛人工强制混合充氧及诱导自然混合对水源水库水质改善效果分析	(1217)
从上盘间记台尤氧及防守日然记台对水体水库水质设备双来力划。	(1227)
滤速与水质对低温含铁锰氨地下水中氨去除的影响····································	(1236)
汤肝河着牛硅藻群落及其与环境因子的关系	(1256)
盐龙湖水源生态净化系统 FG 和 MBFG 演替特征及水质响应性评价 ········ 王莲,李璇,马卫星,邹立航,赵强强,丁成,吴向阳	(1265)
二、映作区有温丘小流,项级解制, 出版。	(12/6)
化肥减重配施生物灰对紫色土坡耕地解流矢的影响 ····································	(1286)
不用玩以住力解看[F为]自住復二4种任间不停内砾嶙的样放 ************************************	(1308)
3种典型多孔高温改性固废材料对磺胺二甲嘧啶的吸附特性 王静,朱晓丽,韩自玉,胡健,秦之瑞,焦文涛	(1319)
新制和老化微塑料对多溴联苯醚的吸附	(1329)
化肥碱量配施生物灰对紧色土玻料地解流失的影响	(1338)
·····································	(1346)
一个一个一个一个一个一个一个一个一个一个一个一个一个一个一个一个一个一个一个	(1357)
北京某污水处理厂及受纳水体中典型有机磷酸酯的污染特征和风险评估 ············ 张振飞, 吕佳佩, 裴莹莹, 王春英, 郭昌胜, 徐建基于短程反硝化厌氧氨氧化的低碳源城市污水深度脱氮特性 ····································	(1368)
基于短程反硝化炭氧氨氧化的低碳源项币污水深度脱氮特性····································	(1377)
王秋颖,于德爽,赵骥,王晓霞,袁梦飞,巩秀珍,楚光玉.何彤晖	(1384)
三秋颖,于德爽,赵骥,王晓霞,袁梦飞,巩秀珍,楚光玉,何彤晖中试 MBBR 反应器启动 CANON 工艺及其短程硝化	(1393)
抗生素对耐药型反硝化菌反硝化过程及微生物群落结构的影响 ······ 代莎,李彭,彭五庆,刘玉学,王拯,何义亮,沈根祥,胡双庆	(1401)
多价近水-嗪气的对氧枫恒行化系统关键 剩余污泥碱性发酵产物对硝化过程及性能的影响 ·································· 邱圣杰 刘瑾瑾 李夕耀 彭永臻	(1409)
加生素对耐约型及硝化菌及硝化过程及减生物群洛结构的影响 "代沙,孝彭,彭五庆,刘玉字,土拯,何义是,沉艰祥,胡双庆 多次进水-曝气的好氧颗粒污泥系统实验 张杰,王玉颖,李冬,刘志诚,曹思雨 剩余污泥碱性发酵产物对硝化过程及性能的影响 邱圣杰,刘瑾瑾,李夕耀,彭永臻 硫酸盐对污泥高级厌氧消化过程中甲基汞迁移转化的影响 邱圣杰	(1110)
工业城市农田土壤重金属时空变异及来源解析 何湘琳,刘吉宝,阴永光,谭颖锋,朱爱玲,左壮,高山,解立平,魏源送 耕地土壤重金属健康风险空间分布特征 姬超,侯大伟,李发志,包广静,邓爱萍,沈红军,孙华	(1425)
上业城市农田土壤重金属时至变异及米源解析····································	(1432)
黄河三角洲盐碱土根际微环境的微生物多样性及理化性质分析	(1449)
不同土地利用方式对土壤细菌分子生态网络的影响 李冰,李玉双,魏建兵,宋雪英,史荣久,侯永侠,刘厶瑶	(1456)
红壤丘陵区土壤有机碳组分对土地利用方式的响应特征 ············ 章晓芳,郑生猛,夏银行,胡亚军,苏以荣,陈香碧	(1466)
有70加一	(14/4)
等镁磷肥对石灰、海泡石组配修复镉污染稻田土壤的影响····································	(1491)
桉树遗态磷灰石材料对铅污染土壤的钝化修复效应 方雅莉,朱宗强,赵宁宁,朱义年,李超,张立浩	(1498)
耕地土壤重金属健康风险空间分布特征	(1505)
	(1303)
# 1 Not 1 4 / word 14 4 / (100 1) # 11 14 / world at 14 / 17 11 11 11 11 11 11 11 11 11 11 11 11	



3 种典型多孔高温改性固废材料对磺胺二甲嘧啶的吸附特性

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摘要:以畜禽粪便、农作物秸秆和采煤废弃物这 3 种典型多孔固体废料为原料,用低氧控温炭化法制成牛粪炭和秸秆炭以及用煅烧后的煤矸石炭对磺胺二甲嘧啶(SMZ)进行批处理吸附实验.通过吸附动力学和等温吸附平衡研究牛粪炭、秸秆炭和煤矸石炭对 SMZ 的吸附特性,并结合 FE-SEM、FT-IR、Boehm 滴定、BET 及 Zeta 电位滴定分析表征手段探讨了其吸附机制.结果表明,3 种炭材料对 SMZ 的吸附在 24 h 时基本达到平衡.3 种炭材料对 SMZ 的吸附动力学均符合准二级动力学方程, R^2 在0.996 8 ~ 0.999 9之间,吸附速率随着炭材料表面有效吸附位点的减少而减小.吸附过程主要由膜扩散、颗粒内扩散和平衡阶段这 3 个步骤组成,颗粒内扩散和膜扩散共同控制吸附速率.等温吸附数据更符合 Freundlich 模型, R^2 在0.987 4 ~ 0.999 7 之间,主要为物理吸附,是自发的放热反应.3 种炭材料的最大吸附量依次为牛粪炭(19.64 mg·g $^{-1}$) >煤矸石炭(12.06 mg·g $^{-1}$) > 秸秆炭(9.16 mg·g $^{-1}$). SMZ 在 3 种炭材料上的吸附机制主要有:分子间的氢键作用、多分子层的表面静电吸附作用和孔隙填充等.其中,静电吸附为主要吸附机制.牛粪炭吸附性能最佳可能是由于其具有较为丰富的含氧官能团、较多的负电荷和较大的比表面积和孔容.

关键词:高温改性; 固废; 吸附; 磺胺二甲嘧啶; 特性

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Adsorption Characteristics of Sulfamethazine on Three Typical Porous Hightemperature Modified Solid Waste Materials

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Abstract: Three typical porous solid wastes, including livestock manure, crop straw, and coal mining waste, were used as raw materials to prepare cattle manure charcoal, straw charcoal, and coal gangue charcoal by low-oxygen controlling temperature carbonization and calcination. Batch adsorption experiments of sulfamethazine (SMZ) in water were carried out. Adsorption kinetics and isothermal adsorption equilibrium were used to investigate the adsorption characteristics of SMZ on cattle dung charcoal, straw charcoal, and coal gangue charcoal, and the adsorption mechanism was discussed by means of field-electron scanning electron microscope, Fourier transform infrared spectroscopy, Boehm titration, Brunauer-Emmett-Teller measurement, and zeta potentiometric titration. The results showed that the adsorption of SMZ on the three carbon materials reached equilibrium at 24 h. The adsorption kinetics of SMZ on three kinds of carbon materials agreed with the quasi-second-order kinetics equation. R² ranged from 0.996 8 to 0.9999, and the adsorption rate decreased with the decrease in effective adsorption sites on the surface of carbon materials. The adsorption process mainly consists of three steps; membrane diffusion, intraparticle diffusion, and the equilibrium stage. Both intraparticle diffusion and membrane diffusion control the adsorption rate. Isothermal adsorption is more consistent with the Freundlich model. R2 is between 0.9874 and 0.9997. It is mainly physical adsorption and spontaneous exothermic reaction. The maximum adsorption capacity of the three kinds of carbon materials was cattle dung carbon (19.64 mg·g⁻¹) > coal gangue carbon (12.06 mg·g⁻¹) > straw carbon (9.16 mg·g⁻¹). The adsorption mechanism of SMZ on the three kinds of carbon materials mainly includes hydrogen bonding between molecules, surface electrostatic adsorption of multi-molecular layers, and pore filling. Of these, electrostatic adsorption is the main adsorption mechanism. The best adsorption performance of cattle manure charcoal may be due to its rich oxygencontaining functional groups, more negative charges, and larger specific surface area and pore volume.

Key words: high temperature modification; solid waste; adsorption; sulfamethazine; characteristics

近些年,抗生素在医疗卫生、畜禽养殖等领域的使用越来越普遍.由于其利用率低^[1,2]以及污水处理厂对其处理能力有限^[3]等原因,导致其最终流入环境,并随之引发一系列环境问题.在常用抗生素中,磺胺类抗生素吸附作用最弱、迁移性最强^[4].磺胺二甲嘧啶(SMZ)在地表水、土壤和地下水中均有

检出^[5]. 我国部分流域地表水中检测到磺胺类抗生素的平均浓度为 93. 93 ng·L^{-1[6]}. 成玉婷等^[7]对广

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州市典型有机蔬菜基地土壤中 8 种磺胺类抗生素 (SAs)的污染特征和风险水平进行了调查,分析表明 SAs 在土壤中检出率大于等于 94%,总含量为 0.73~973 μg·kg⁻¹,生态风险评价显示磺胺二甲嘧啶风险最高,其次为磺胺嘧啶.对一线城市地下水抗生素进行检测,发现磺胺类抗生素的检出率达 78.9% [8].因此,磺胺类抗生素污染物在环境中的高检出率问题已引起了广泛关注.

吸附法作为一种成本较低,无副产物且可回收 再利用的方法,在阻控抗生素污染等方面有很大的 应用潜力^[9]. Zhang 等^[10]研究了离子形态对磺胺甲 恶唑在改性碳纳米管上吸附造成的影响,结果表明 此吸附过程主要受静电作用、疏水性作用和氢键作 用控制. Chen 等[11] 研究了氧化石墨烯对水中磺胺 甲恶唑(sulfamethoxazole,SMX)的吸附,结果表明该 材料对 SMX 的吸附主要受 π-π 电子供-受体作用的 控制. 鲍晓磊等[12]利用一种新型磁性纳米复合材料 对水中常见的磺胺类抗生素进行吸附,结果表明 15℃ 时的平衡吸附量在 68.9 μg·g⁻¹ (磺胺二甲嘧 啶)和99.6 μg·g⁻¹ (磺胺甲二唑)之间,氢键是 CoFeM48 表面官能团和磺胺之间的一个主要作用 力,分子筛外壳有序的二氧化硅结构和磺胺之间的 π-π 电子共轭作用也可能促进两者之间的吸附. Wang 等[13] 用磁性离子交换树脂对磺胺甲恶唑进行 了去除研究. 结果表明,该树脂对磺胺甲恶唑的最大 吸附容量可达到 789. 32 μg·mL⁻¹,主要通过阴离子 交换进行吸附. 以上研究均表明吸附法对抗生素有 良好的去除效果,吸附剂的选择是成功阻控抗生素 的关键. 我国是一个农业大国,如果抗生素吸附剂的 制备建立在以废弃资源循环利用的基础之上,那么 找到一种来源广泛、价格低廉且吸附效果相对较好 的废弃资源作为吸附剂就会显得尤为重要. 如此,不 仅能从根本上解决资源浪费问题,还能实现有效避 免环境污染的目标.

畜禽粪便、农作物秸秆及煤矸石废料来源广泛,因其处理不当和无组织堆放占用了大量土地资源,对生态环境造成了不良影响[14,15].但对其进行高温改性后不仅可以减少土地资源的无效利用问题还可以对环境产生一系列积极效应.例如:经畜禽粪便热解后得到的畜禽粪便基炭由于其灰分含量高,对酸性土壤具有石灰效应[16]以及原料中营养物质丰富[17,18]等特点,可以直接作为生物肥料从而有效减小环境风险.秸秆炭电荷密度高、孔隙度发达,对无机离子和极性、非极性化合物有较强的吸附性[19],并且含有较高的固定碳,稳定性较强[20,21].将秸秆炭添加到土壤中既利于改善土壤结构和持水

性,又可降低肥料养分流失^[22]. 煤矸石经煅烧后其表面和内部形成大量微孔,比表面积增大且结构呈疏松状态,有较好的吸附和离子交换性能^[23],可作为资源丰富且价格低廉的吸附剂. 目前,有很多关于畜禽粪便基炭、秸秆炭对水中氨氮^[24]、土壤中石油烃^[25]及重金属的吸附^[26]及煤矸石炭对生活污水中磷等污染物的吸附研究^[27],但关于这 3 种炭材料对于一些新型污染物如磺胺类抗生素的吸附特性研究十分有限,且吸附机制有待进一步明确.

综上,本文以"以废治污"为指导,利用畜禽粪便、农作物秸秆及采煤废弃物制备了3种典型多孔高温改性固废炭材料,并通过批处理吸附实验研究了它们对磺胺二甲嘧啶的吸附特性,同时结合场发射扫描电子显微镜(FE-SEM)、傅里叶红外光谱(FT-IR)、Boehm滴定法、比表面积分析(BET)和Zeta 电位滴定分析探讨了吸附机制,以期为畜禽粪便基、木质基及煤基炭对磺胺类抗生素的吸附提供理论依据,对促进废弃资源循环利用和环境的可持续发展具有十分重要的意义.

1 材料与方法

1.1 实验材料

磺胺二甲嘧啶(SMZ,纯度99%)、甲醇(色谱纯)及甲酸(色谱纯)购自北京金博瑞祺科技发展有限公司.其余盐酸、氢氧化钠等化学品均为分析纯,所有溶液均用超纯水制备.磺胺二甲嘧啶基本性质见表1.

表 1 磺胺二甲嘧啶基本性质

1.2 实验方法

1.2.1 炭材料的制备

根据前期探索实验得到牛粪炭、秸秆炭和煤矸石炭的 pH 值分别为 10.8(±0.1)、9.9(±0.1)和7.5(±0.1),均呈碱性,经检测其本身均不含抗生素.有研究表明,在600℃的热解条件下有利于牛粪炭孔隙的形成、微孔数量增多、比表面积和孔容变大,其碘吸附值和亚甲基蓝吸附值最大^[30].在500℃热解温度下制备的秸秆生物炭表面含有一定数量的羧基等含氧官能团,内部含有丰富的中孔及

大孔等孔隙结构,具有很好地吸附性能^[31]. 前期探索实验表明,煤矸石煅烧温度在300~500℃之间对抗生素的吸附效果差异较小,故选择300℃作为煤矸石的煅烧温度. 因此,分别选取600、500和300℃为3种炭材料的热解温度.

牛粪炭和秸秆炭分别由牛粪(购于北京某奶牛养殖基地)和茄子秸秆(购于北京市某农村)经清洗干燥后利用低氧控温炭化法在马弗炉中(炭化温度分别为600℃和500℃)炭化2h得到.煤矸石炭由煤矸石废料在300℃的温度下煅烧2h得到.所有炭材料均研磨过100目筛,储存于干燥器中备用.

1.2.2 炭材料的表征

用场发射扫描电子显微镜(SU-8020,日立,日本)在10.0 kW 加速功率和10 mm 工作距离以及高真空模式下,观测用金膜覆盖后炭材料的性状和表面特征.使用傅立叶变换红外光谱仪(Nicolet 8700, Thermo Fisher Scientific,America)记录样品的傅立叶变换红外光谱,使用 KBr 颗粒在4 000~500 cm⁻¹范围内.炭材料的 Boehm 滴定分析:根据酸或碱的消耗量来计算不同种类官能团数量. NaOC₂H₅ 可以中和所有的酸性基团;由 HCl 与炭材料反应的量来计算表面碱性基团的量;Na₂CO₃ 中和羧基和内酯基;NaHCO₃ 仅中和羧基;NaOH 可以中和羧基、内酯基和酚羟基.用全自动比表面积及微孔物理吸附仪(ASAP2460,麦克默瑞提克,美国)进行比表面积、孔径及孔容的估算.用 Zeta 电位分析仪(ZS90,马尔文,英国)对炭材料表面电位进行分析.

1.2.3 吸附实验

由表1可知,磺胺二甲嘧啶(SMZ)的解离常数为

 pK_{al} = 2.6, pK_{a2} = 7.7.有研究表明,SMZ 的 3 种离子形态在吸附剂中吸附能力排序为:阳离子形态 > 中性分子形态 > 阴离子形态 | 32~34]. 当 $pH \le 7.7$ 时,SMZ 的部分氨基会结合 H^+ 以阳离子形态存在. 但当 pH 值过低时,过多的 H^+ 会在炭材料表面竞争吸附位点,反而会降低 SMZ 的吸附效果. 为提高吸附效果和避免 H^+ 竞争吸附位点,因此设置反应体系 pH 为 4 (±0.1),使 SMZ 主要以阳离子形态存在.

吸附动力学:在50 mL 聚四氟乙烯离心管中,加入10 mL 浓度为10 mg·L⁻¹的 SMZ 溶液,再分别加入0.05 g 炭材料. 将离心管放入摇床(25℃,180 r·min⁻¹)恒温振摇,在预设时间点分别取样并离心15 min(3 500 r·min⁻¹),取上清液过0.22 μ m 针头滤膜于2 mL 进样瓶,待测. 等温吸附:设置 SMZ 溶液的浓度为5、10、15、20 和 25 mg·L⁻¹,分别与炭材料在15、25 和 35℃的条件下振摇 24 h 至吸附平衡,其余条件同吸附动力学实验. 以上处理均做3个重复,未含目标抗生素的处理作为空白,未含生物质炭的处理作为对照.

1.2.4 抗生素分析检测方法

抗生素浓度采用高效液相色谱检测(仪器型号:ACQUITY UPLC, Waters, 美国), 仪器参数为:流动相为甲醇(A)和 0.1% 甲酸(B)(体积比 = 25/75),流速 0.8 mL·min⁻¹; 色谱柱: Eclipse plus C18色谱柱(4.6 mm×150 mm); 柱温: 30℃; UV 检测波长:270 nm; 进样量:20 μL(针清洗后进样,用甲醇洗针).

1.2.5 数据处理

实验数据处理所用公式见表 2.

表 2 数据处理涉及的公式

Table 2 Formulas for data processing

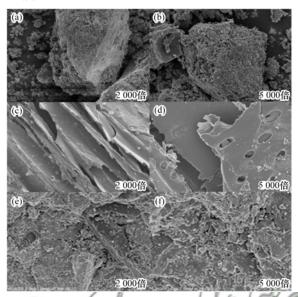
类型	公式
吸附动力学	准一级动力学方程: $\ln(Q_e - Q_t) = \ln Q_e - k_1 t$ 准二级动力学方程: $t/Q_t = 1/(k_2 \times Q_e^2) + t/Q_e$ 颗粒内扩散: $q_t = K_3 t^{1/2} + C$ 液膜扩散: $\ln\left(1 - \frac{q_t}{q_e}\right) = -K_4 t + A$
等温吸附	标准偏差: $\Delta q = 100 \sqrt{\frac{\sum \left[\left(Q_e - Q_{e, cal} \right) / Q_e \right]^2}{N - 1}}$ Freundlich 模型: $\ln Q_e = \ln K_{\rm F} + 1/n \ln c_e$ Langmuir 模型: $c_e / Q_e = c_e / Q_{\rm m} + 1/(K_{\rm L} \times Q_{\rm m})$
热力学	$K_{\rm d} = \frac{c_0 - c_e}{c_e} \times \frac{V}{m} \times 1000; \ \Delta G^\theta = -RT\lnK^\theta; \lnK_{\rm d} = \frac{\Delta S^\theta}{R} - \frac{\Delta H^\theta}{RT};$ 式中, $K_{\rm d}$ 为分配系数, $m{\rm L} \cdot {\rm g}^{-1}$; c_0 为初始浓度, $m{\rm g} \cdot {\rm L}^{-1}$; c_e 为平衡浓度, $m{\rm g} \cdot {\rm L}^{-1}$; V 为溶液体积, ${\rm L}$; m 为吸附剂质量, ${\rm g}$; R 为理想气体常数, ${\rm 8.314}{\rm J} \cdot ({\rm mol} \cdot {\rm K})^{-1}$; T 为热力学温度, ${\rm K}$; 以 ${\rm ln}K_{\rm d}$ 对 $1/T$ 作图,根据线性回归分析得到的截距和斜率的数值可分别算出 ΔS^θ 和 ΔH^θ 的值

2 结果与分析

2.1 3种炭材料的形貌结构分析

2.1.1 3种炭材料表面物理结构分析

牛粪炭、秸秆炭和煤矸石炭的扫描电镜图像如图 1 所示.



(a)、(b)牛粪炭;(c)、(d) 秸秆炭;(e)、(f)煤矸石炭 图 1 3 种炭材料的扫描电镜图像

Fig. 1 Scanning electron micrographs of the three carbon materials

图 1(a) 和 1(b) 表面粗糙不平, 结构疏松, 表面含有孔状结构. 图 1(c) 和 1(d) 表面有排列紧密的管状结构, 且在5 000倍下能看到许多不规则孔. 生物炭在高温环境下均显现出多孔或管状结构, 并且有明显分层现象, 这是由于生物质热解过程中有机质分解, 在纤维素、木质素框架结构中留下很多孔, 造成生物质炭具有多孔结构^[35]. 图 1(e) 和 1(f) 表面有较多褶皱, 呈多层结构, 这是由于煅烧后的煤矸石结构疏松, 内部形成微孔, 有较好的吸附和离子交换性能^[23]. 3 种炭材料表面的多孔及层状结构可有效促进其对抗生素的吸附.

2.1.2 3种炭材料表面官能团分析

由 3 种炭材料对 SMZ 吸附前后的红外谱图(图 2)可知, 3 种炭材料在1 000 cm -1 附近都有吸收峰,此处为羟基、酯和醚中 C—O 伸缩振动^[36]. 牛粪炭和秸秆炭分别在1 596 cm -1 和1 749 cm -1 处有吸收峰,这是由于芳香环结构上 C —C 双键和羰基中 C —O 双键的伸缩振动^[36]. 牛粪炭在3 224 cm -1 处存在宽吸收峰,为羧基和酚羟基中 O—H 的伸缩振动. 煤矸石炭在3 650 cm -1 处出现吸收峰,为自由羟基 O—H 的伸缩振动. 与牛粪炭相比,秸秆炭和煤矸石炭在1 465 ~1 340 cm -1 和3 000 ~2 850 cm -1 均未出现吸收峰,表明有活性的 C—H 键可能是影响吸

附的关键因素之一. 综上,影响 3 种炭材料对 SMZ 吸附效果的官能团主要有: C—O、C —C、C —O、—OH 和 C—H 等. 牛粪炭表面有较多的官能团,这可能是牛粪炭吸附效果较好的原因之一.

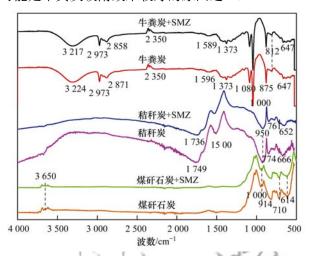
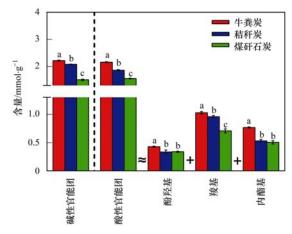


图 2 3 种炭材料对 SMZ 吸附前后的傅里叶红外光谱图

Fig. 2 Fourier transform infrared spectra of the three carbon materials before and after SMZ adsorption

2.1.3 3 种炭材料 Boehm 滴定分析

对3种炭材料表面官能团进行 Boehm 滴定分析,计算得到相应官能团数量,如图3所示.从中可知,炭材料表面的碱性官能团和酸性官能团数量相近.酸性官能团含量约为酚羟基、羧基和内酯基含量总和.牛粪炭的酸性官能团含量最高,易与极性相对较强的 SMZ 通过化学键力而发生表面化学吸附^[37],表明酸性官能团含量可能对 SMZ 在炭材料上的吸附有较大影响.由显著性差异分析结果可知,牛粪炭的3种基团(酚羟基、羧基和内酯基)与秸秆炭和煤矸石炭均呈显著性差异.秸秆炭和煤矸石炭只有羧基呈显著性差异,且秸秆炭的羧基含量大于



不同小写字母表示不同炭材料上官能团的差异显著性(P<0.05)

图 3 和炭材料表面官能团数量

Fig. 3 Number of surface functional groups on the three carbon materials

煤矸石炭,与吸附结果相反,表明在此吸附过程中除了羧基含量,可能还有其他更主要的因素(比表面积、孔容和表面电位等)影响吸附.

2.1.4 3种炭材料比表面积、孔径和孔容分析

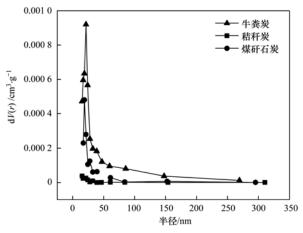
比表面积、孔容和孔径是影响生物炭吸附性能的关键因素^[38]. 由表 3 可知, 3 种炭材料的比表面积和总孔容从大到小排序为牛粪炭(26. 29 $\text{m}^2 \cdot \text{g}^{-1}$; 33. $17 \times 10^{-3} \text{ cm}^3 \cdot \text{g}^{-1}$) > 煤矸石炭(8. 27 $\text{m}^2 \cdot \text{g}^{-1}$; 4. $95 \times 10^{-3} \text{ cm}^3 \cdot \text{g}^{-1}$) > 秸秆炭(3. 34 $\text{m}^2 \cdot \text{g}^{-1}$; 2. 64 $\times 10^{-3} \text{ cm}^3 \cdot \text{g}^{-1}$),与其平衡吸附量的排序相同,表明炭材料的吸附量与其比表面积和总孔容成正相关关系; 三者的平均孔径 r 分别为 34. 13、45. 27 和 50. 60 nm.

表 3 炭材料的比表面积、总孔容及平均孔径

Table 3 Specific surface area, total pore volume and average

	pore size of the	e carbon materials	
吸附剂	比表面积	总孔容	平均孔径r
	$/\mathrm{m}^2 \cdot \mathrm{g}^{-1}$	$/\text{cm}^3 \cdot \text{g}^{-1}$	/nm
牛粪炭	26. 29	33. 17 \times 10 $^{-3}$	34. 13
秸秆炭	3. 34	2.64×10^{-3}	45. 27
煤矸石炭	8. 27	4.95×10^{-3}	50. 60

由图 4 的孔径分布曲线可知,牛粪炭的中孔集中在 20~27 nm 范围内;秸秆炭的中孔峰型不明显;煤矸石炭的中孔分布段较宽,分别在 17~20nm和 25~33 nm 处有一定数量的孔.由于 3 种炭材料在 20~30 nm 范围内的孔数量排序与吸附量排序相同,均为牛粪炭>煤矸石炭>秸秆炭(由峰高决定),因此推测 20~30 nm 范围内的孔可能会对SMZ 有较好的吸附效果.此外,观察到 3 种炭材料的比表面积相对较小,可能除了原材料本身性质差异外,还有炭化时间较长导致热解过于充分等原因.通过添加化学活化剂[39]和负载改性[40,41]等方法增大其比表面积可能有利于提高其吸附能力,活化后的炭材料对抗生素的吸附效果有待进一步研究探讨.



2.1.5 3 种炭材料 Zeta 电位滴定分析

对 3 种炭材料进行 Zeta 电位测量,分析其表面电荷信息. 牛粪炭、秸秆炭和煤矸石炭表面电荷都呈负值,且依次为(-12.6±0.29)、(-7.8±1.13)和(-8.64±0.41)mV. 3 种炭材料的表面负电荷量较大,为其以静电吸附形式吸附 SMZ 提供了较大的可能. 牛粪炭表面所带负电荷最多,其次为煤矸石炭和秸秆炭. 煤矸石炭表面所带负电荷多于秸秆炭,是由于其主要成分为 SiO₂ 和 Al₂O₃等氧化物^[42],所含矿物质常以高岭石的形式存在,而高岭土表面主要是负电荷. 由此表明,炭材料表面所带负电荷可影响 SMZ 在炭材料上吸附效果. 牛粪炭表面有较多的负电荷可能是其吸附效果最好的原因之一.

2.2 吸附动力学研究

准一级、准二级动力学模型对吸附数据的拟合 以及牛粪炭、秸秆炭和煤矸石炭对 SMZ 的吸附量 随时间的变化过程见图 5. 从中可知, SMZ 在 3 种炭 材料上吸附量在1h内呈现快速增大,而后缓慢增 加最后趋于稳定. 3 种炭材料对 SMZ 的吸附在 24 h 时达到平衡. 采用准一级动力学、准二级动力学模 型对吸附动力学的数据结果进行拟合,结果见表 4. 从中可知,准一级动力学方程的 R^2 在0.2339~ 0.8489之间,准二级动力学方程的 R² 在0.9968~ 0.999 9之间. 在此基础上采用归一化标准偏差 Δq 进一步验证. Δq 的值越小,表明理论数据与实际数 据偏差越小. 准二级动力学方程的标准偏差 Aq (%)远小于准一级动力学,且远低于1%,表明准二 级动力学方程能更好地对该吸附过程进行拟合,3 种炭材料对 SMZ 的吸附速率与炭材料上吸附活性 点位的平方成线性关系,吸附速率随着炭材料表面 有效吸附位点的减少而减小[43]. 3 种炭材料对 SMZ 的平衡吸附量的排序为:牛粪炭($2.26 \text{ mg} \cdot \text{g}^{-1}$) >

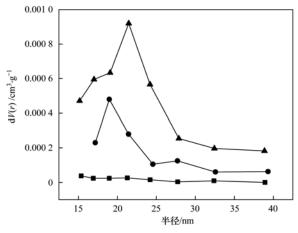


图 4 3 种炭材料孔径分布

Fig. 4 Aperture distribution of the three carbon materials

表 4 炭材料对 SMZ 的拟一级、拟二级动力学模型参数

Table 4 Or	uasi-first-order an	d quasi-second-	order kir	netic model	parameters o	of SMZ for	carbon i	materials
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		准-	一级动力学力	7程			准二级	动力学方程	
吸附剂	$q_{ m e}$ /mg·g ⁻¹	$q_{ m e, cal}$ /mg \cdot g $^{-1}$	K_1	$R_1^{\ 2}$	$\Delta q_1/\%$	$q_{\rm e}$ /mg·g ⁻¹	K_2	$R_2^{\ 2}$	$\Delta q_2/\%$
牛粪炭	2. 26	0. 522	0. 41	0. 652 7	53. 69	2. 169	5. 38	0. 999 9	0. 017
秸秆炭	1.66	0. 648	0. 11	0. 848 9	43. 27	1.669	1. 11	0. 998 8	0.008
煤矸石炭	2. 07	0. 363	0.07	0. 233 9	57. 99	2.003	0.89	0. 996 8	0.509

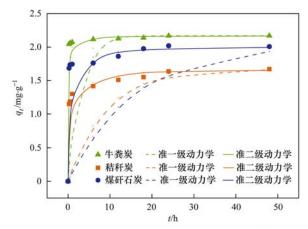


图 5 准一级动力学和准二级动力学模型对 SMZ 在 3 种炭材料上吸附过程的拟合

Fig. 5 Pseudo-first-order and pseudo-second-order kinetics for adsorption of SMZ on carbon materials at 25°C

煤矸石炭(2.07 mg·g⁻¹) > 秸秆炭(1.76 mg·g⁻¹).

为研究3种炭材料对SMZ吸附速率、吸附机制 和传质现象等方面的信息,采用颗粒内扩散及液膜 扩散模型对吸附动力学的数据结果进行拟合,颗粒 内扩散模型拟合见图 6. 从中可知,图形中的曲线未 过原点且呈多线性,表明在吸附过程中有两个或多 个步骤发生. 吸附过程根据斜率由大到小变化分为 3个阶段. 第一阶段斜率较大是由于吸附剂外表面 可以利用的吸附位点与水溶液中抗生素分子易于接 触,属于外表面吸附或瞬时吸附,主要代表膜扩 散[4].表5为颗粒内扩散和膜扩散的相关参数,其 中由液膜扩散模型的 K_4 值可知,第一阶段中吸附速 率排序为牛粪炭>煤矸石炭>秸秆炭,与吸附量的 排序相同,说明吸附初始阶段牛粪炭表面具有更多 可利用的吸附位点. 第二阶段斜率变小,吸附在吸附 剂孔内发生,代表颗粒内扩散[4].根据表5颗粒内 扩散模型中的 K, 值可知, 此阶段吸附速率排序为秸 秆炭>煤矸石炭>牛粪炭,与第一阶段恰好相反,由 此表明,吸附位点的数量可能是引起3种炭材料吸 附性能差异的主要原因之一. 第三阶段是平衡阶 段[44].

2.3 等温吸附分析

采用 Freundlich 和 Langmuir 等温吸附模型对 3 种炭材料在 15、25 和 35℃下的等温吸附数据进行

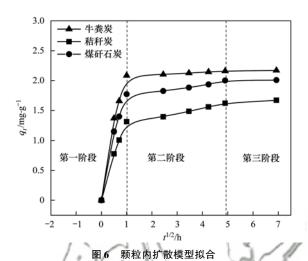


图 6 秋恒门 0 秋至10日

Fig. 6 Fitting chart of intraparticle diffusion model

表 5 炭材料对 SMZ 的颗粒内扩散和液膜扩散模型参数

Table 5 Model parameters of intragranular diffusion and liquid film diffusion of SMZ by carbon materials

l	吸附剂	30	颗粒内扩散			液膜扩散	31
J	P)X [P]1 7[1]	K_3	R_3^2	\boldsymbol{C}	K_4	R_4^2	A
	牛粪炭	0.02	0. 914 9	2. 05	0. 021	0. 803 5	-1.45
	秸秆炭	0.08	0. 920 5	1. 17	0.002	0.6250	-1.60
	煤矸石炭	0.05	0.7008	1.67	0.012	0.6366	-2.16

拟合,见图 7. 从中可知,Freundlich 和 Langmuir 模型均能较好地对 SMZ 在 3 种炭材料上的吸附进行拟合,Freundlich 模型相对更符合此吸附数据. 平衡吸附量跟反应温度呈负相关关系,说明升高温度不利于反应进行. 因此,最大吸附量根据反应温度为15℃时的吸附数据确定. 表 6 为 15℃下 SMZ 的Freundlich 和 Langmuir 模型拟合特征参数. 从中可知,Freundlich 模型的 R^2 值在0.993 3 ~ 0.999 2之间,均大于 Langmuir 模型的 R^2 值,表明 Freundlich 模型更符合本研究的吸附过程,该吸附是多分子层吸附,吸附量会随着 SMZ 浓度的增加而不断增加 $[^{34}]$. Freundlich 模型的 1/n 值均小于 1,表明 SMZ 在吸附剂上的吸附容易进行 $[^{45}]$. 牛粪炭、秸秆炭和煤矸石炭对 SMZ 的最大吸附量分别为:19.64、9.16和 12.06 mg·g $^{-1}$.

2.4 吸附热力学分析

通过对吸附后的数据进行吉布斯自由能变化 (ΔG^{θ}) 、焓变 (ΔH^{θ}) 和熵变 (ΔS^{θ}) 计算,相关公式见

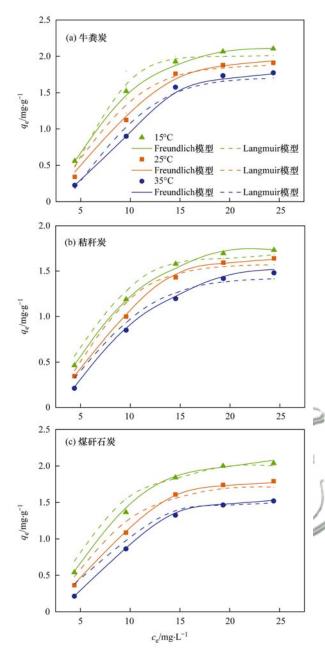


图 7 Freundlich 和 Langmuir 等温吸附模型在 15、25 和 35℃下对 SMZ 在 3 种炭材料上吸附过程的拟合

Fig. 7 Freundlich and Langmuir isotherms for adsorption of SMZ on carbon materials at 15 , 25 , and 35 $^\circ\! C$

0.849

吸附剂

牛粪炭 秸秆炭

煤矸石炭

表 2, 并以此来判断吸附过程中的方向和驱动力^[30,46]. 相关热力学参数见表 7.

由表 7 可知, 3 种炭材料对 SMZ 的吸附反应的 ΔG^{θ} 均为负值,且均在 $-20 \sim 0$ kJ·mol⁻¹范围内,表明吸附反应自发进行,静电引力作用与孔隙填充是该反应的主要吸附机制,该吸附主要为非专性吸附^[47,48]. 3 种炭材料对 SMZ 吸附的 ΔH^{θ} 为负值,表明该吸附过程是放热反应. 3 种炭材料对 SMZ 吸附的 ΔS^{θ} 值均为负值,该吸附过程为熵减的自发反应过程.

3 3 种炭材料对磺胺二甲嘧啶的吸附机制讨论

红外光谱图显示,在吸附 SMZ 后吸收峰均有不 同程度的偏移,表明含氧官能团在吸附过程中发挥 着重要作用[49]. 在吸附 SMZ 后分别偏移至1589 cm⁻¹和1736 cm⁻¹,表明炭材料表面存在羧基可能 与 SMZ 上的氨基发生酸碱反应从而形成离子 键[50]. 吸附 SMZ 后,偏移至3 217 cm⁻¹,表明牛粪炭 表面含有羟基能与 SMZ 中的苯环形成氢键[36,51,52]. 此外, Boehm 滴定分析显示, 3 种炭材料中的酚羟基 含量相对其他两种官能团较小,且秸秆炭和煤矸石 炭的羟基含量无显著差异,与其吸附能力不同步,表 明由羟基形成的氢键作用不是该吸附的主要机制. 由比表面积、孔径和孔容分析可知,3种炭材料的 比表面积和总孔容的排序与其吸附量的排序相同, 表明该吸附过程中存在孔隙填充作用. 然而, 三者吸 附量的差异明显小于其比表面积和总孔容的差异, 表明孔隙填充作用不是影响该吸附过程的最主要因 素. Zeta 电位滴定分析中显示 3 种炭材料表面均带 负电荷,且电荷量为牛粪炭>煤矸石炭>秸秆炭,与 吸附能力同步. SMZ 中含有氨基,其解离常数为 2.6 和 7. 7^[28, 29], 见表 1. 当 pH 值接近 2. 6 时, SMZ 的氨 基基团会与H+结合,主要呈现阳离子状态,有利于 被表面带负电荷的炭材料由于静电作用而被吸

0.001

0.6503

表 6 炭材料对 SMZ 的 Freundlich 和 Langmuir 模型参数
Table 6 Freundlich and Langmuir model parameters of SMZ for carbon materials

						_
	Freundlich 模型			Langmuir 模型		_
1/n	$K_{ m F}$	R_1^2	$q_{\rm m}/{ m mg}\cdot{ m g}^{-1}$	$K_{ m L}$	R_2^2	_
0. 689	0. 420	0. 999 2	19. 64	0. 095	0. 694 6	_
0.899	0. 141	0.9982	9. 16	0.010	0. 893 4	

表 7 3 种炭材料吸附 SMZ 的相关热力学参数

0.9933

Table 7 Relevant thermodynamic parameters of SMZ adsorption by the three carbon materials

	rubie / recevant in	ermoaynamie parameters	or come description by the	timee carbon materials	
吸附剂		$\Delta G^{\theta}/\mathrm{kJ \cdot mol^{-1}}$		$\Delta H^{ heta}$	$\Delta S^{ heta}$
火門加	15℃	25℃	35℃	/kJ⋅mol -1	$/J \cdot (\bmod \cdot K)^{-1}$
牛粪炭	- 14. 25	- 13. 69	- 13. 01	-30.71	- 6. 97
秸秆炭	- 12. 64	-11.91	-11.32	- 29. 86	−7.33
煤矸石炭	- 13. 42	- 12. 61	- 12. 04	- 30. 99	-7.48

附^[53]. 由此可知, 静电作用是该吸附过程的主要机制.

4 结论

- (1)吸附动力学实验结果表明:拟二级动力学模型可以对吸附过程进行很好地拟合,吸附速度由液膜扩散和颗粒内扩散共同控制.等温吸附实验结果表明:3种炭材料对抗生素的吸附均符合Freundlich和 Langmuir等温模型,相比之下更符合Freundlich模型.最大吸附量顺序为:牛粪炭>煤矸石炭>秸秆炭.热力学分析表明该吸附反应是自发放热的过程.
- (2) SMZ 在 3 种炭材料上的吸附机制主要有:① 分子间的氢键作用;② 多分子层的表面静电吸附作用;③ 孔隙填充作用.其中,静电作用为主要的吸附机制,其次为孔隙填充和氢键作用.牛粪炭对SMZ 的吸附量最高主要原因是其拥有较丰富的含氧官能团、较多的负电荷以及较大的比表面积和孔容.

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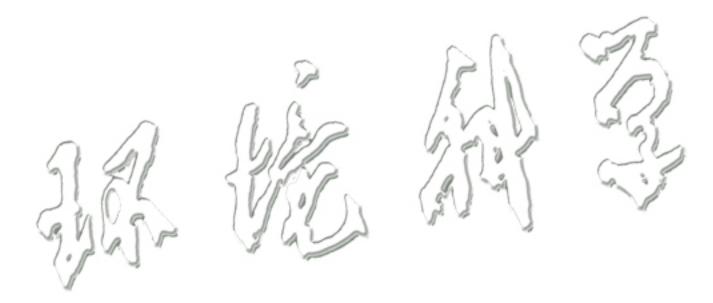
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HUANJING KEXUE

Environmental Science (monthly)

Vol. 41 No. 3 Mar. 15, 2020

CONTENTS

Evolution of the Distribution of PM _{2, 5} Concentration in the Yangtze River Economic Belt and Its Influencing Factors	
Chemical Characteristics and Source Apportionment of Water-Soluble Ions in Atmosphere Aerosols over the East China Sea Island D	uring Winter and Summer
	FANG Yan, CAO Fang, FAN Mei-yi, et al. (1025)
Analysis of Chemical Components and Sources of PM _{2.5} During Autumn and Winter in Yangquan City ·····	WANG Cheng, YAN Yu-long, XIE Kai, et al. (1036)
Chemical Compositions and Sources of n-Alkanes and Saccharides in PM _{2.5} from Taian City During the Summer	······ YI Ya-nan, HOU Zhan-fang, YANG Qian-cai, et al. (1045)
Physical and Chemical Characteristics of Atmospheric Particles in Autumn in Mt. Huangshan	
Characteristics of Aerosol Optical Depth in the Urban Area of Beibei and Its Correlation with Particle Concentration	
Emission Characteristics of Volatile Organic Compounds from Typical Industries in Zibo	······ WANG Yu-yan, WANG Xiu-yan, DU Miao, et al. (1078)
Analysis of Pollution Characteristics and Sources of Atmospheric VOCs in Ezhou City	····· FU Yu-meng, YANG Hong-gang, LU Min-yu, et al. (1085)
Producing Coefficients and Emission Coefficients of Volatile Organic Compounds from the Automobile Manufacturing Industry in Zhe	jiang Province ·····
	·· TENG Fu-hua, YANG Zhong-ping, DONG Shi-bi, et al. (1093)
Pollution Characteristics of Volatile Organic Compounds Emission from the Metal Packaging Industry Based on Analysis of Process	
Profile Characteristics of VOCs from Wood and Economic Crop Burning	NIU Zhen-zhen, KONG Shao-fei, YAN Qin, et al. (1107)
Accounting Methods of VOCs Emission Associated with Production Processes in a Fine Chemical Industrial Park	
HONO Observation and Assessment of the Effects of Atmospheric Oxidation Capacity in Changzhou During the Springtime of 2017	
Analysis of Activity and Its Emissions Trend for Construction Equipment in China	
Air Pollutant Emission Inventory from LTO Cycles of Aircraft in the Beijing Tianjin-Hebei Airport Group, China	
Particle Size Distribution of PM Emission from In-use Gasoline and Diesel Vehicles	
Impact of Parameterization on the Estimation of Ammonia Emissions; A Case Study over the Yangtze River Delta	
Characteristics and Source Apportionment of Dustfall Pollution in the Coal Mine Area and Surrounding Areas of Wuhai City in Sprin	
Characteristics and Source Apportionment of Dustian Foliution in the Coal mine Area and Surfounding Areas of Wuhai City in Sprin	WILLHORS man SHI Chang sing 7HANC Van et al. (1167)
Variations of Stable Oxygen and Deuterium Isotopes in River and Lake Waters During Flooding Season Along the Middle and Lower	
variations of stable Oxygen and Deutertum Isotopes in Aiver and Lake waters During Frooding Season Atong the Middle and Lower	Reaches of the Tangize River Regions
Water Sources and Factors Controlling Hydro-chemical Compositions in the Yiluo River Basin	LI Jing, WU Hua-wu, ZHOU 10ng-qiang, et al. (1170)
Chemical Evolution of Groundwater in the Tacheng Basin of Xinjiang in the Process of Urbanization	
Remote Sensing Monitoring on Spatial Differentiation of Suspended Sediment Concentration in a River-Lake System Based on Senting	
Yangtze River Section in Anhui Province	
CDOM Optical Characteristics and Related Environmental Factors of High-turbidity Waters on the Loess Plateau	
Effects of Artificial Destratification and Induced-natural Mixing on Water Quality Improvement in a Drinking Water Reservoir	
Effect of Filter Speed and Water Quality on Ammonia Removal in Groundwater Containing Iron, Manganese, and Ammonia at Low	l'emperature
Long-term Variation Characteristics of Zooplankton Community Structure in Meiliang Bay, Lake Taihu	
Community of Benthic Diatoms and Their Relationship with Aquatic Environmental Factors in the Tangwang River, China	····· XUE Hao, WANG Ye-yao, MENG Fan-sheng, et al. (1256)
Succession Characteristics and Water Quality Responsiveness Evaluation of FG and MBFG in Yanlong Lake Water Source Ecologica	Purification System
	l Purification System
Characteristics of Nitrogen and Phosphorus Output and Loss Flux in the Shipanqiu Watershed, Three Gorges Reservoir Area	l Purification System
	l Purification System
Characteristics of Nitrogen and Phosphorus Output and Loss Flux in the Shipanqiu Watershed, Three Gorges Reservoir Area Effect of Optimized Fertilization and Biochar Application on Phosphorus Loss in Purple Soil Sloping Farmland Use of Iron-modified Calcite as an Active Capping Material to Control Phosphorus Release from Sediments in Surface Water Bodies	l Purification System
Characteristics of Nitrogen and Phosphorus Output and Loss Flux in the Shipanqiu Watershed, Three Gorges Reservoir Area Effect of Optimized Fertilization and Biochar Application on Phosphorus Loss in Purple Soil Sloping Farmland Use of Iron-modified Calcite as an Active Capping Material to Control Phosphorus Release from Sediments in Surface Water Bodies Preparation of Tea Waste Biochar and Its Application in Tetracycline Removal from Aqueous Solution	l Purification System WANG Lian, LI Xuan, MA Wei-xing, et al. (1265) CHEN Shi-qi, LONG Yi, YAN Dong-chun, et al. (1276) LUO Dong-hai, WANG Zi-fang, LONG Yi, et al. (1286) BAI Xiao-yun, LIN Jian-wei, ZHAN Yan-hui, et al. (1296) FAN Shi-suo, LIU Wen-pu, WANG Jing-tao, et al. (1308)
Characteristics of Nitrogen and Phosphorus Output and Loss Flux in the Shipanqiu Watershed, Three Gorges Reservoir Area Effect of Optimized Fertilization and Biochar Application on Phosphorus Loss in Purple Soil Sloping Farmland Use of Iron-modified Calcite as an Active Capping Material to Control Phosphorus Release from Sediments in Surface Water Bodies Preparation of Tea Waste Biochar and Its Application in Tetracycline Removal from Aqueous Solution Adsorption Characteristics of Sulfamethazine on Three Typical Porous High-temperature Modified Solid Waste Materials	l Purification System WANG Lian, LI Xuan, MA Wei-xing, et al. (1265) CHEN Shi-qi, LONG Yi, YAN Dong-chun, et al. (1276) LUO Dong-hai, WANG Zi-fang, LONG Yi, et al. (1286) BAI Xiao-yun, LIN Jian-wei, ZHAN Yan-hui, et al. (1296) FAN Shi-suo, LIU Wen-pu, WANG Jing-tao, et al. (1308) WANG Jing, ZHU Xiao-li, HAN Zi-yu, et al. (1319)
Characteristics of Nitrogen and Phosphorus Output and Loss Flux in the Shipanqiu Watershed, Three Gorges Reservoir Area Effect of Optimized Fertilization and Biochar Application on Phosphorus Loss in Purple Soil Sloping Farmland Use of Iron-modified Calcite as an Active Capping Material to Control Phosphorus Release from Sediments in Surface Water Bodies Preparation of Tea Waste Biochar and Its Application in Tetracycline Removal from Aqueous Solution	l Purification System WANG Lian, LI Xuan, MA Wei-xing, et al. (1265) CHEN Shi-qi, LONG Yi, YAN Dong-chun, et al. (1276) LUO Dong-hai, WANG Zi-fang, LONG Yi, et al. (1286) BAI Xiao-yun, LIN Jian-wei, ZHAN Yan-hui, et al. (1296) FAN Shi-suo, LIU Wen-pu, WANG Jing-tao, et al. (1308) WANG Jing, ZHU Xiao-li, HAN Zi-yu, et al. (1319)
Characteristics of Nitrogen and Phosphorus Output and Loss Flux in the Shipanqiu Watershed, Three Gorges Reservoir Area Effect of Optimized Fertilization and Biochar Application on Phosphorus Loss in Purple Soil Sloping Farmland Use of Iron-modified Calcite as an Active Capping Material to Control Phosphorus Release from Sediments in Surface Water Bodies Preparation of Tea Waste Biochar and Its Application in Tetracycline Removal from Aqueous Solution Adsorption Characteristics of Sulfamethazine on Three Typical Porous High-temperature Modified Solid Waste Materials	l Purification System WANG Lian, LI Xuan, MA Wei-xing, et al. (1265) CHEN Shi-qi, LONG Yi, YAN Dong-chun, et al. (1276) LUO Dong-hai, WANG Zi-fang, LONG Yi, et al. (1286) BAI Xiao-yun, LIN Jian-wei, ZHAN Yan-hui, et al. (1296) FAN Shi-suo, LIU Wen-pu, WANG Jing-tao, et al. (1308) WANG Jing, ZHU Xiao-li, HAN Zi-yu, et al. (1319) XU Peng-cheng, GUO Jian, MA Dong, et al. (1329)
Characteristics of Nitrogen and Phosphorus Output and Loss Flux in the Shipanqiu Watershed, Three Gorges Reservoir Area Effect of Optimized Fertilization and Biochar Application on Phosphorus Loss in Purple Soil Sloping Farmland Use of Iron-modified Calcite as an Active Capping Material to Control Phosphorus Release from Sediments in Surface Water Bodies Preparation of Tea Waste Biochar and Its Application in Tetracycline Removal from Aqueous Solution	l Purification System WANG Lian, LI Xuan, MA Wei-xing, et al. (1265) CHEN Shi-qi, LONG Yi, YAN Dong-chun, et al. (1276) LUO Dong-hai, WANG Zi-fang, LONG Yi, et al. (1286) BAI Xiao-yun, LIN Jian-wei, ZHAN Yan-hui, et al. (1296) FAN Shi-suo, LIU Wen-pu, WANG Jing-tao, et al. (1308) WANG Jing, ZHU Xiao-li, HAN Zi-yu, et al. (1319) XU Peng-cheng, GUO Jian, MA Dong, et al. (1329) FANG Zhi-qing, WANG Yong-min, WANG Xun, et al. (1338)
Characteristics of Nitrogen and Phosphorus Output and Loss Flux in the Shipanqiu Watershed, Three Gorges Reservoir Area Effect of Optimized Fertilization and Biochar Application on Phosphorus Loss in Purple Soil Sloping Farmland Use of Iron-modified Calcite as an Active Capping Material to Control Phosphorus Release from Sediments in Surface Water Bodies Preparation of Tea Waste Biochar and Its Application in Tetracycline Removal from Aqueous Solution Adsorption Characteristics of Sulfamethazine on Three Typical Porous High-temperature Modified Solid Waste Materials Sorption of Polybrominated Diphenyl Ethers by Virgin and Aged Microplastics	l Purification System WANG Lian, LI Xuan, MA Wei-xing, et al. (1265) CHEN Shi-qi, LONG Yi, YAN Dong-chun, et al. (1276) LUO Dong-hai, WANG Zi-fang, LONG Yi, et al. (1286) HAI Xiao-yun, LIN Jian-wei, ZHAN Yan-hui, et al. (1296) FAN Shi-suo, LIU Wen-pu, WANG Jing-tao, et al. (1308) WANG Jing, ZHU Xiao-li, HAN Zi-yu, et al. (1319) XU Peng-cheng, GUO Jian, MA Dong, et al. (1329) FANG Zhi-qing, WANG Yong-min, WANG Xun, et al. (1338) KAN Ke-cong, GU Xiao-hong, LI Hong-min, et al. (1346)
Characteristics of Nitrogen and Phosphorus Output and Loss Flux in the Shipanqiu Watershed, Three Gorges Reservoir Area Effect of Optimized Fertilization and Biochar Application on Phosphorus Loss in Purple Soil Sloping Farmland Use of Iron-modified Calcite as an Active Capping Material to Control Phosphorus Release from Sediments in Surface Water Bodies Preparation of Tea Waste Biochar and Its Application in Tetracycline Removal from Aqueous Solution Adsorption Characteristics of Sulfamethazine on Three Typical Porous High-temperature Modified Solid Waste Materials Sorption of Polybrominated Diphenyl Ethers by Virgin and Aged Microplastics Spatial Distribution and Risk Assessment of Heavy Metals in Sediments of the Ruxi Tributary of the Three Gorges Reservoir Distribution and Risk Assessment of OCPs in Surface Water, Sediments, and Fish from Lake Gucheng and Inflow and Outflow Rive Occurrence and Ecological Risk Assessment of Typical Persistent Organic Pollutants in Hengshui Lake	l Purification System WANG Lian, LI Xuan, MA Wei-xing, et al. (1265) CHEN Shi-qi, LONG Yi, YAN Dong-chun, et al. (1276) LUO Dong-hai, WANG Zi-fang, LONG Yi, et al. (1286) BAI Xiao-yun, LIN Jian-wei, ZHAN Yan-hui, et al. (1296) FAN Shi-suo, LIU Wen-pu, WANG Jing-tao, et al. (1308) WANG Jing, ZHU Xiao-li, HAN Zi-yu, et al. (1319) XU Peng-cheng, GUO Jian, MA Dong, et al. (1329) FANG Zhi-qing, WANG Yong-min, WANG Xun, et al. (1338) TS WAN Ke-cong, GU Xiao-hong, LI Hong-min, et al. (1346) ZHANG Jia-wen, WEI Jian, LÜ Yi-fan, et al. (1357)
Characteristics of Nitrogen and Phosphorus Output and Loss Flux in the Shipanqiu Watershed, Three Gorges Reservoir Area Effect of Optimized Fertilization and Biochar Application on Phosphorus Loss in Purple Soil Sloping Farmland Use of Iron-modified Calcite as an Active Capping Material to Control Phosphorus Release from Sediments in Surface Water Bodies Preparation of Tea Waste Biochar and Its Application in Tetracycline Removal from Aqueous Solution Adsorption Characteristics of Sulfamethazine on Three Typical Porous High-temperature Modified Solid Waste Materials Sorption of Polybrominated Diphenyl Ethers by Virgin and Aged Microplastics Spatial Distribution and Risk Assessment of Heavy Metals in Sediments of the Ruxi Tributary of the Three Gorges Reservoir Distribution and Risk Assessment of OCPs in Surface Water, Sediments, and Fish from Lake Gucheng and Inflow and Outflow Rive Occurrence and Ecological Risk Assessment of Typical Persistent Organic Pollutants in Hengshui Lake	l Purification System WANG Lian, LI Xuan, MA Wei-xing, et al. (1265) CHEN Shi-qi, LONG Yi, YAN Dong-chun, et al. (1276) LUO Dong-hai, WANG Zi-fang, LONG Yi, et al. (1286) BAI Xiao-yun, LIN Jian-wei, ZHAN Yan-hui, et al. (1296) FAN Shi-suo, LIU Wen-pu, WANG Jing-tao, et al. (1308) WANG Jing, ZHU Xiao-li, HAN Zi-yu, et al. (1319) XU Peng-cheng, GUO Jian, MA Dong, et al. (1329) FANG Zhi-qing, WANG Yong-min, WANG Xun, et al. (1338) KAN Ke-cong, GU Xiao-hong, LI Hong-min, et al. (1346) ZHANG Jia-wen, WEI Jian, LÜ Yi-fan, et al. (1357)
Characteristics of Nitrogen and Phosphorus Output and Loss Flux in the Shipanqiu Watershed, Three Gorges Reservoir Area Effect of Optimized Fertilization and Biochar Application on Phosphorus Loss in Purple Soil Sloping Farmland Use of Iron-modified Calcite as an Active Capping Material to Control Phosphorus Release from Sediments in Surface Water Bodies Preparation of Tea Waste Biochar and Its Application in Tetracycline Removal from Aqueous Solution Adsorption Characteristics of Sulfamethazine on Three Typical Porous High-temperature Modified Solid Waste Materials Sorption of Polybrominated Diphenyl Ethers by Virgin and Aged Microplastics Spatial Distribution and Risk Assessment of Heavy Metals in Sediments of the Ruxi Tributary of the Three Gorges Reservoir Distribution and Risk Assessment of OCPs in Surface Water, Sediments, and Fish from Lake Gucheng and Inflow and Outflow Rive Occurrence and Ecological Risk Assessment of Typical Persistent Organic Pollutants in Hengshui Lake Pollution Characteristics and Risk Assessment of Typical Organophosphate Esters in Beijing Municipal Wastewater Treatment Plant a	l Purification System WANG Lian, LI Xuan, MA Wei-xing, et al. (1265) CHEN Shi-qi, LONG Yi, YAN Dong-chun, et al. (1276) LUO Dong-hai, WANG Zi-fang, LONG Yi, et al. (1286) FAN Shi-suo, LIU Wen-pu, WANG Jing-tao, et al. (1308) XU Peng-cheng, GUO Jian, MA Dong, et al. (1319) FANG Zhi-qing, WANG Yong-min, WANG Xun, et al. (1338) KAN Ke-cong, GU Xiao-hong, LI Hong-min, et al. (1346) ZHANG Jia-wen, WEI Jian, LÜ Yi-fan, et al. (1357) and the Receiving Water ZHANG Zhen-fei, LÜ Jia-pei, PEI Ying-ying, et al. (1368)
Characteristics of Nitrogen and Phosphorus Output and Loss Flux in the Shipanqiu Watershed, Three Gorges Reservoir Area Effect of Optimized Fertilization and Biochar Application on Phosphorus Loss in Purple Soil Sloping Farmland Use of Iron-modified Calcite as an Active Capping Material to Control Phosphorus Release from Sediments in Surface Water Bodies Preparation of Tea Waste Biochar and Its Application in Tetracycline Removal from Aqueous Solution Adsorption Characteristics of Sulfamethazine on Three Typical Porous High-temperature Modified Solid Waste Materials Sorption of Polybrominated Diphenyl Ethers by Virgin and Aged Microplastics Spatial Distribution and Risk Assessment of Heavy Metals in Sediments of the Ruxi Tributary of the Three Gorges Reservoir Distribution and Risk Assessment of OCPs in Surface Water, Sediments, and Fish from Lake Gucheng and Inflow and Outflow Rive Occurrence and Ecological Risk Assessment of Typical Persistent Organic Pollutants in Hengshui Lake	l Purification System WANG Lian, LI Xuan, MA Wei-xing, et al. (1265) CHEN Shi-qi, LONG Yi, YAN Dong-chun, et al. (1276) LUO Dong-hai, WANG Zi-fang, LONG Yi, et al. (1286) BAI Xiao-yun, LIN Jian-wei, ZHAN Yan-hui, et al. (1296) FAN Shi-suo, LIU Wen-pu, WANG Jing-tao, et al. (1308) WANG Jing, ZHU Xiao-li, HAN Zi-yu, et al. (1319) XU Peng-cheng, GUO Jian, MA Dong, et al. (1329) FANG Zhi-qing, WANG Yong-min, WANG Xun, et al. (1338) KAN Ke-cong, GU Xiao-hong, LI Hong-min, et al. (1346) ZHANG Jia-wen, WEI Jian, LÜ Yi-fan, et al. (1357) and the Receiving Water ZHANG Zhen-fei, LÜ Jia-pei, PEI Ying-ying, et al. (1368)
Characteristics of Nitrogen and Phosphorus Output and Loss Flux in the Shipanqiu Watershed, Three Gorges Reservoir Area Effect of Optimized Fertilization and Biochar Application on Phosphorus Loss in Purple Soil Sloping Farmland Use of Iron-modified Calcite as an Active Capping Material to Control Phosphorus Release from Sediments in Surface Water Bodies Preparation of Tea Waste Biochar and Its Application in Tetracycline Removal from Aqueous Solution Adsorption Characteristics of Sulfamethazine on Three Typical Porous High-temperature Modified Solid Waste Materials Sorption of Polybrominated Diphenyl Ethers by Virgin and Aged Microplastics Spatial Distribution and Risk Assessment of Heavy Metals in Sediments of the Ruxi Tributary of the Three Gorges Reservoir Distribution and Risk Assessment of OCPs in Surface Water, Sediments, and Fish from Lake Gucheng and Inflow and Outflow Rive Occurrence and Ecological Risk Assessment of Typical Persistent Organic Pollutants in Hengshui Lake Pollution Characteristics and Risk Assessment of Typical Organophosphate Esters in Beijing Municipal Wastewater Treatment Plant and Advanced Nitrogen Removal Characteristics of Low Carbon Source Municipal Wastewater Treatment via Partial-denitrification Coupled	l Purification System WANG Lian, LI Xuan, MA Wei-xing, et al. (1265) CHEN Shi-qi, LONG Yi, YAN Dong-chun, et al. (1276) LUO Dong-hai, WANG Zi-fang, LONG Yi, et al. (1286) FAN Shi-suo, LIU Wen-pu, WANG Jing-tao, et al. (1308) XU Peng-cheng, GUO Jian, MA Dong, et al. (1319) FANG Zhi-qing, WANG Yong-min, WANG Xun, et al. (1338) KAN Ke-cong, GU Xiao-hong, LI Hong-min, et al. (1346) ZHANG Jia-wen, WEI Jian, LÜ Yi-fan, et al. (1357) and the Receiving Water ZHANG Zhen-fei, LÜ Jia-pei, PEI Ying-ying, et al. (1368) ed with ANAMMOX MA Bin, XU Xin-xin, GAO Mao-hong, et al. (1377)
Characteristics of Nitrogen and Phosphorus Output and Loss Flux in the Shipanqiu Watershed, Three Gorges Reservoir Area Effect of Optimized Fertilization and Biochar Application on Phosphorus Loss in Purple Soil Sloping Farmland Use of Iron-modified Calcite as an Active Capping Material to Control Phosphorus Release from Sediments in Surface Water Bodies Preparation of Tea Waste Biochar and Its Application in Tetracycline Removal from Aqueous Solution Adsorption Characteristics of Sulfamethazine on Three Typical Porous High-temperature Modified Solid Waste Materials	l Purification System
Characteristics of Nitrogen and Phosphorus Output and Loss Flux in the Shipanqiu Watershed, Three Gorges Reservoir Area Effect of Optimized Fertilization and Biochar Application on Phosphorus Loss in Purple Soil Sloping Farmland Use of Iron-modified Calcite as an Active Capping Material to Control Phosphorus Release from Sediments in Surface Water Bodies Preparation of Tea Waste Biochar and Its Application in Tetracycline Removal from Aqueous Solution Adsorption Characteristics of Sulfamethazine on Three Typical Porous High-temperature Modified Solid Waste Materials Sorption of Polybrominated Diphenyl Ethers by Virgin and Aged Microplastics Spatial Distribution and Risk Assessment of Heavy Metals in Sediments of the Ruxi Tributary of the Three Gorges Reservoir Distribution and Risk Assessment of OCPs in Surface Water, Sediments, and Fish from Lake Gucheng and Inflow and Outflow Rive Occurrence and Ecological Risk Assessment of Typical Persistent Organic Pollutants in Hengshui Lake Pollution Characteristics and Risk Assessment of Typical Organophosphate Esters in Beijing Municipal Wastewater Treatment Plant a	l Purification System
Characteristics of Nitrogen and Phosphorus Output and Loss Flux in the Shipanqiu Watershed, Three Gorges Reservoir Area Effect of Optimized Fertilization and Biochar Application on Phosphorus Loss in Purple Soil Sloping Farmland Use of Iron-modified Calcite as an Active Capping Material to Control Phosphorus Release from Sediments in Surface Water Bodies Preparation of Tea Waste Biochar and Its Application in Tetracycline Removal from Aqueous Solution Adsorption Characteristics of Sulfamethazine on Three Typical Porous High-temperature Modified Solid Waste Materials Sorption of Polybrominated Diphenyl Ethers by Virgin and Aged Microplastics Spatial Distribution and Risk Assessment of Heavy Metals in Sediments of the Ruxi Tributary of the Three Gorges Reservoir Distribution and Risk Assessment of OCPs in Surface Water, Sediments, and Fish from Lake Gucheng and Inflow and Outflow Rive Occurrence and Ecological Risk Assessment of Typical Persistent Organic Pollutants in Hengshui Lake Pollution Characteristics and Risk Assessment of Typical Organophosphate Esters in Beijing Municipal Wastewater Treatment Plant a	l Purification System WANG Lian, LI Xuan, MA Wei-xing, et al. (1265) CHEN Shi-qi, LONG Yi, YAN Dong-chun, et al. (1276) LUO Dong-hai, WANG Zi-fang, LONG Yi, et al. (1286) FAN Shi-suo, LIU Wen-pu, WANG Jing-tao, et al. (1296) WANG Jing, ZHU Xiao-li, HAN Zi-yu, et al. (1308) WANG Jing, ZHU Xiao-li, HAN Zi-yu, et al. (1319) FANG Zhi-qing, WANG Yong-min, WANG Xun, et al. (1329) FANG Zhi-qing, WANG Yong-min, WANG Xun, et al. (1338) WANG Jia-wen, WEI Jian, LÜ Yi-fan, et al. (1357) and the Receiving Water ZHANG Zhen-fei, LÜ Jia-pei, PEI Ying-ying, et al. (1368) ed with ANAMMOX MA Bin, XU Xin-xin, GAO Mao-hong, et al. (1377) I Denitrification and Denitrifying Phosphorus Removal WANG Qiu-ying, YU De-shuang, ZHAO Ji, et al. (1384) FU Kun-ming, YANG Zong-yue, LIAO Min-hui, et al. (1393)
Characteristics of Nitrogen and Phosphorus Output and Loss Flux in the Shipanqiu Watershed, Three Gorges Reservoir Area Effect of Optimized Fertilization and Biochar Application on Phosphorus Loss in Purple Soil Sloping Farmland Use of Iron-modified Calcite as an Active Capping Material to Control Phosphorus Release from Sediments in Surface Water Bodies Preparation of Tea Waste Biochar and Its Application in Tetracycline Removal from Aqueous Solution	l Purification System
Characteristics of Nitrogen and Phosphorus Output and Loss Flux in the Shipanqiu Watershed, Three Gorges Reservoir Area Effect of Optimized Fertilization and Biochar Application on Phosphorus Loss in Purple Soil Sloping Farmland Use of Iron-modified Calcite as an Active Capping Material to Control Phosphorus Release from Sediments in Surface Water Bodies Preparation of Tea Waste Biochar and Its Application in Tetracycline Removal from Aqueous Solution Adsorption Characteristics of Sulfamethazine on Three Typical Porous High-temperature Modified Solid Waste Materials	l Purification System
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