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参数选取对畜禽养殖业大气氨排放的影响:以长三角 地区为例

张琪¹,黄凌²,殷司佳²,王倩²,李红丽²,王杨君²,王军¹,陈勇航¹,李莉^{2*}

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摘要: 氨对于大气细颗粒物中二次无机盐的生成具有十分重要的作用,近年来引起了国内外学者的高度关注. 相对准确地定 量氨排放对于深入研究二次气溶胶的理化过程、实现较好的数值模拟性能,以及开展氨排放精细化管控具有极为重要的科学 意义和现实意义. 已有研究表明,农业源是大气氨的主要排放源,其中,畜禽养殖业的氨排放占比最大. 已有针对畜禽养殖业 氨排放的研究大多采用排放系数法建立氨排放清单,然而,不同参数的选取会对研究结果造成较大差异.本文从活动水平和 排放系数选取上做出了多种假设,构建了8种情景,以2017年为基准年,分别计算了长三角地区畜禽养殖业大气氨排放.结果 表明,选取不同的活动水平对清单估算结果的影响最大,选取出栏量计算的结果较选取存栏量高出27.6%~34.1%.选取更细 致的月均温的计算结果高出以年均温结果0.3~0.4万 t.此外,清单的时空分布特征也与该两项参数密切相关,以存栏量进行 估算的结果中,舟山地区排放强度最低,淮南市最高;以出栏量进行估算的结果中,丽水排放强度最低,南京最高.以月均温度 估算情景中将获得更准确的月排放廓线,全年中5~9月排放量最高,冬季(12、1 和2月)排放量最小.

关键词: 氨排放; 畜禽养殖; 情景分析; 排放因子; 长三角

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Impact of Parameterization on the Estimation of Ammonia Emissions: A Case Study over the Yangtze River Delta

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Abstract: Atmospheric ammonia plays an important role in the formation of secondary inorganic composition of PM25, which has attracted a high level of attention from researchers both in China and abroad. Quantifying ammonia emissions is of great scientific significance regarding research on the formation of secondary aerosol, realizing better model performance, and control of ammonia emissions. Previous studies have shown that agricultural activities are the dominant source of atmospheric ammonia, of which livestock and poultry farming contribute the most. Existing studies on estimating ammonia emissions from livestock and poultry farming activities are mostly based on emission factors and activities. However, the choice of different emission activities could lead to large differences in estimated ammonia emissions. This study makes a variety of assumptions from the selection of activity levels (volume vs. inventory) and emission coefficients (monthly vs. annual average temperature), and establishes eight scenarios from which to calculate atmospheric ammonia emissions from livestock and poultry farming in the Yangtze River Delta region in 2017. The results show that selection of different activity levels has the greatest impact on estimated ammonia emissions; estimation based on volume is higher than that based on inventory by 27,6%-34,1%. Calculation based on a more detailed monthly average temperature is higher than using average annual temperature by 3 000 to 4 000 tons per year. In addition, the spatial and temporal distributions of the ammonia emissions are also closely related to the choice of volume vs. inventory and the choice of monthly average temperature vs. annual average temperature. When using inventory as the emission activity, Zhoushan (Zhejiang Province) has the lowest ammonia emissions, while Huainan (Anhui Province) has the highest. In contrast, when volume is used, Lishui (Zhejiang Province) has the lowest ammonia emissions and Nanjing (Jiangsu Province) has the highest. Emissions calculations based on monthly average temperature are supposed to be more representative than those based on annual average temperature, with the highest emissions from May to September and the lowest in the winter (December, January, and February).

Key words; ammonia; livestock feeding; scenario analysis; emission factor; Yangtze River Delta

在地球的岩石圈和生物圈中,氨(NH、)是大气 成分中最丰富的碱性痕量气体[1],也是大气成分中 含量仅次于 N_2 和 N_2 O 的第三大含氮气体^[2], 当 NH, 浓度超过 1 μg·m⁻³时会对植物造成损害^[3]. NH、是大气酸性成分的主要中和剂, NH、可以和酸 性气体反应形成铵盐进而增加大气中细颗粒物

(PM25)的浓度[4],影响全球辐射平衡,降低大气能

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见度,危害人体健康^[5-7]. NH₃ 还对新粒子暴发具有重要的作用^[8]. 过去几十年间, NH₃ 的减排相比硫氧化合物和氮氧化合物的削减并不明显^[9]. NH₃ 对大气细颗粒物(PM_{2.5})的贡献以及对生态系统的影响日益受到重视^[10].

近年来,国内外学者对我国的氨排放开展了大 量研究,董文煊等[11]利用排放因子法,基于畜禽养 殖、氮肥施用、工业生产及人体排放这4类源估算 了 1994~2006 年我国人为氨排放的时空分布情况, 表明 NH、排放呈明显增加趋势; 李新艳等[12]的研 究表明,我国 NH, 排放在 1997~1999 年呈下降趋 势,随后呈平稳上升趋势; Huang 等[4] 将统计数据 与 MODIS 数据相结合,构建了我国全面精细的大气 氨排放清单,表明氮肥施用和畜禽养殖是清单中排 放最高的两类源; Zhang 等[5] 基于 GEOS-Chem 伴 随模型优化了中国目前采用自下而上的研究方式估 算的大气氨清单;长三角地区人为源氨排放的研究 表明[13],畜禽养殖和氮肥施用对大气氨排放的贡献 分别为48%和40%;这些研究均表明畜禽养殖是氨 清单较为重要的排放源. 大气氨排放清单中畜禽养 殖源大多采用自上而下的方式并基于排放系数法进 行估算,但是不同的研究所采取的细致估算方法不 一致,在长三角区域,董艳强等[14]研究的氨排放清 单中畜禽养殖源的活动水平采用畜禽年末存栏数, 而在同一研究区域刘波等[15] 对养殖时间小于 1a 的 畜禽采用年末出栏数,这使得后者的研究结果在畜 禽养殖量较小的情况下却高于前者.此外,在珠三 角^[16]、四川^[17]、西安^[18]、南通^[19] 和山东省^[20] 等 的氨清单研究中,畜禽养殖源在基于活动水平与排 放系数的选取原则上存在较大差异,导致不同研究 和不同地区之间的可比性较差,然而已有研究中并 未详细报道不同参数选取方式对研究结果的影响, 以及何种参数方案更加符合实际情况. 因此,本文以 长三角为例,着重探讨了不同估算方法和参数方案 选取对畜禽养殖业氨排放估算造成的差异,在此基 础上,提出更符合实际情况的畜禽养殖业的氨排放 估算方法.

1 材料与方法

本文以长三角区域为例,研究范围为长三角三省一市(江苏省、浙江省、安徽省和上海市,共41个城市).已有研究多采用排放系数法构建畜禽养殖业 NH₃ 排放清单,其中关键影响因素为活动水平和排放因子.本文首先构建了8种基于不同参数选取方式的情景,再基于排放系数法进行氨排放估算,比较不同情景方案的差异,研究不同参数选取方式

对氨排放的影响,获得适用于长三角区域的畜禽养殖氨排放合理计算方式.

1.1 氨排放清单计算方法

畜禽养殖业氨排放的测算采用排放系数法. 氨排放总量即为活动水平和排放系数的乘积. 计算公式为:

$$E_{i,j,y} = A_{i,j,y} \times \mathrm{EF}_{i,j,y} \times \gamma$$
 (1)

式中,i 为地区(省、直辖市、自治区或县),j 为排放源,y 为年份, $E_{i,j,y}$ 为y 年 i 地区j 排放源的排放量,A 为活动水平,EF 为排放系数, γ 为氮-大气氨转换系数,针对畜禽养殖业,取 $1.214^{[21]}$.

畜禽养殖业产生的大气氨排放主要是由动物排 泄物进行释放. 对于粪便量所产生大气氨排放的计 量方式随着实地测算的完善,逐渐以文献[21]中所 描述的分阶段计算为主. 其中,粪便包括室内和户外 两部分,室内粪便在圈舍中停留一段时间后,会汇集 进行存储腐熟处理,最后进行施肥. 因此粪便管理阶 段包括户外、圈舍内、粪便存储处理和后续施肥. 后3种方式属于室内粪便管理,具有尿液和粪便两 种形态,动物户外排泄的尿液和粪便通常混合在一 起.由文献[21]中给出的排放系数可知,畜禽的粪 便在各阶段的排放系数各不相同,并且圈舍阶段的 排放系数与温度密切相关. 畜禽排泄物释放大气氨 的阶段包括户外、圈舍-液态、圈舍-固态、存储-液 态、存储-固态、施肥-液态和施肥-固态共7个.公式 (2)表示畜禽氨排放估算方式按文献[21]中分阶段 计算.

$$E_{i} = \sum_{j} L_{j,l} \sum_{l=1}^{4} (ef_{i,j,l})$$
 (2)

式中,i、j 和 l 分别为市、畜禽种类和 NH₃ 排放的 4 个类别(户外、圈舍、存储和施肥),L 为畜禽的数量,ef 为排放因子.

1.2 活动水平数据

活动水平的选取对于 NH₃ 排放的影响十分突出. 苏航等^[18]研究建立的西安市 2013 年人为源大气氨排放清单中,畜禽养殖业是最主要的排放源,其中畜禽(大牲畜)的活动水平采用年末存栏量; 栗世学等^[22]在研究兰州市 2016 年农业源氨排放清单时,肉猪、肉牛等大牲畜的活动水平选取其年内出栏量. 随着生活水平的不断提高,畜禽的需求量不断扩大,这使得各市每年的各畜禽年末存量与年内出栏量比例不断变化. 这种变化在清单排放估算工作中的影响较大,甚至可能对畜禽养殖氨排放估算的结果产生数量级上的差异. 因此,在畜禽养殖业的活动水平选取上需要认真考究.

在长三角区域,产生大气氨排放的畜禽主要包

括肉牛、奶牛、母猪、生猪、山羊、绵羊及家禽等,各畜禽的统计数据见表 1. 长三角各地级市畜禽养殖活动水平数据和对排放系数有较大影响的温度数据均来自于各城市统计年鉴. 本文统计各畜禽的年末存栏量及年内出栏量. 长三角的畜禽养殖主要以

集约化为主,并结合少量的散养. 刘波等^[15]认为长三角的畜禽中,猪、牛、家禽和羊的散养占总养殖数的比例分别为 10%、5%、5% 和 50%,本文针对2017 年长三角区域的畜禽养殖大气氨排放,根据文献[23]确认各畜禽养殖方式的比例.

表 1 2017 年长三角各地级市畜禽存栏量和出栏量1)

Table 1 Stock capacity and stocks of livestock in different cities in the Yangtze River Delta in 2017

最			存栏	量				出	栏量	
城市	母猪	肉猪	奶牛	肉牛	山羊	绵羊	家禽	肉猪	肉牛	山羊
上海市	4. 5	56	3. 8	/	19. 6	/	1 252	112	/	/
南京市	6. 6	81	0. 3	0.3	5.3	/	2 007	343	226. 3	/
无锡市	1. 1	14	0. 4	0.0	1.7	/	1 220	56	/	2. 6
徐州市	18. 3	225	/	11.8	119.6	9.8	7 401	463	/	/
常州市	2. 3	28	0. 1	/	5.4	/	3 197	64	/	9. 1
苏州市	2. 2	27	1. 5	/	1.6	4. 0	1 197	69	/	8.8
南通市	17. 5	216	0. 7	0.3	179. 5	/	10 264	343	/	226. 3
连云港市	16. 3	134	7. 0	/	23. 4	/	1 235	268	10.8	39.6
淮安市	14. 7	123	0. 9	4. 6	23. 1	/	7 430	253	2. 0	36.0
盐城市	36. 3	318	3. 0	1.9	96.0	7. 2	14 345	653	1.70	159. 9
扬州市	4. 1	57	0. 4	100	6. 7	/	4 217	128	0.3	13.3
镇江市	2. 8	38	/	1	4. 2	/	1 664	53	//	6. 2
泰州市	13.5	129	1.3	0.5	15.9	/	2 631	271	0.3	22.9
宿迁市	8.9	118	7. 5	7.	28.9	//	7 320	249	5. 5	45. 5
杭州市	8.7	95	0.6	2.1	21.6	//	1 204	237	3. 1	27. 4
宁波市	4,9	54	0.7	0.6	5.8	/ 4	715	108	0.5	7.3
嘉兴市	(1.6	17	45. 2	185	0.0	/	3 027	36	/ (57. 2
湖州市	1.9	17	5/11	18/21	32. 1	//	2 691	48	/	42. 0
绍兴市	6. 5	57	0. 3	0.6	11.3	/ (1 522	137	0.7	12. 1
舟山市	0.3	4	[/V]/	1111	1.3	- / \u	39	1	/	1. 2
温州市	4. 1	41	0.6	2. 8	14.5	/	2 399	70	3.0	14. 4
金华市	7. 2	66	1. 3	0.7	6. 2	/	1 442	172	1.4	8. 3
衢州市	8.6	66	/	1. 7	7. 0	/	3 653	212	0. 2	11. 2
台州市	4. 2	47	/	2. 2	6. 5	/	629	87	/	/
丽水市	3.4	40	/	2.4	8.4	0.0	1 032	74	/	/
合肥市	13.9	122	3. 2	2.4	9. 2	1.0	15 841	278	/	/
淮北市	3.0	39	0. 5	1. 1	20. 1	0. 1	1 967	70	/	/
亳州市	11. 2	150	/	13.4	137. 6	/	2 717	310	/	/
宿州市	31.9	268	1. 0	18.4	262. 1	4. 5	5 973	450	/	/
蚌埠市	7. 1	95	/	23. 2	81.1	/	8 654	208	/	/
阜阳市	33. 9	267	0. 3	36. 0	191.0	/	6 886	569	/	/
准南市	8.6	78	1. 1	10. 1	40. 2	/	5 215	160	/	/
滁州市	10. 1	134	/	1. 3	41.0	/	5 109	341	/	/
六安市	16. 3	171	/	12. 9	29. 3	/	8 048	306	/	/
马鞍山市	1.3	18	/	1.4	6. 3	/	772	38	/	/
芜湖市	3. 0	40	/	2. 0	3. 3	/	1 577	84	/	/
宣城市	3. 7	49	/	4. 4	4. 9	/	2 699	105	/	/
铜陵市	3. 7	21	/	1. 3	0.6	/	1 817	36	/	/
池州市	2. 5	33	0. 5	0. 6	1. 1	/	1 874	74	/	/
安庆市	11. 8	131	/	11. 2	9. 0	2. 1	6 955	260	/	,
黄山市	3. 9	62	0. 2	1. 0	2. 1	0. 1	455	91	/	,

1)对应不同的禽类,单位分别为万只、万头、万羽;数据来自于各省统计年鉴[24~27]

1.3 排放系数的选取

国外对氨排放的研究工作开展较早,对不同的 排放源和排放特征都有深入的研究. 随着国内学者 对霾污染机制研究的深入,大气氨排放清单的研究 方式也有了变化. 随着对畜禽养殖氨排放认识的不断深入,学者对畜禽养殖业氨排放的估算也从畜禽在户外、室内等多阶段的排放进行分工段计算. 文献[21]中的排放系数对应于畜禽养殖存在的各个

阶段.

表 2 列出了同类研究所采用的排放系数,由表可知,不同的研究人员所采用的排放系数相差较大.表 3 为文献[21]中不同畜禽所产生粪便铵态氮的参数,表 4 和表 5 为文献[21]中不同畜禽种类、不同养殖方式、不同排泄物管理阶段和不同粪便形态的排放

系数,气温条件为重要的影响因子. 由于计算施肥过程的活动水平时需要考虑粪便存储过程中氮的损失(以 N_2O 、NO 和 N_2 形式释放),因此粪便存储过程中的 N_2O 、NO 和 N_2 的排放系数也一并列出. 此外,散养时畜禽排泄物在室内户外各占 50%,集约化养殖条件下畜禽排泄物在室内户外分别占 100% 和 0.

表2 各畜禽的氨排放系数/kg·(头·a)-1

Table 2 Emission factors of different animals/kg·(unit·a) -1

						U	,			
畜禽种类					文献					本研究
宙為性失	[20]	[11]	[14]	[19]	[28]	[29]	[30]	[17]	[18]	(均值)
母猪	5. 66	4. 8	6. 29	5. 357	5. 357	11.54	5. 2	2. 82	11. 55	6.5 ± 2.84
肉猪	5.66	4. 8	6. 29	5. 357	5. 357	2. 82	4.8	2.82	2. 82	4.52 ± 1.28
奶牛	21.76	16. 4 ~ 26. 1	20.62	30. 971	26. 52	37. 6	26. 52	37.61	37. 61	28.8 ± 7.0
肉牛	21.76	9. 7	20.62	30. 971	6.8	37. 6	6.8	22.58	22. 58	20.3 ± 10.1
山羊	2. 59	1. 2	1.88	1.697	0.73	4. 93 ~ 5. 09	0.73	4. 93	4. 93	2.54 ± 1.62
绵羊	2. 59	1. 2	1.88	1.697	0.73	4. 93 ~ 5. 09	0.73	4. 93	4. 93	2.55 ± 1.64
家禽	0. 23	0. 22	0. 23	0. 245	_	0. 24 ~ 0. 49	_	0. 24	0. 24	0.24 ± 0.04

表 3 畜禽粪便排泄物铵态氮量的估算相关参数1

Table 3 Estimation parameters of ammonium nitrogen

in livestock manure excrement

	111 1110	Stock man	iic cacioni	,,,,,	
畜禽种类 -	排泄量	∄ 2)	含氮	量/%	铵态氮
亩岗件矢	尿液	粪便	尿液	粪便	比例/%
肉牛	10	20	0.9	0.38	60
奶牛	19/1	40	0.9	0.38	60
山羊	0.75	2. 6	1. 35	0.75	50
绵羊	0.75	2.6	1. 35	0. 75	50
母猪	5. 7	2. 1	0.4	0.34	70
肉猪	3. 2	1.5	0.4	0. 34	70
7/4	6. 5	15	1.4	0.2	/ 60
驴。	6. 5	15	1.4	0. 2	60
骡	6. 5	15	1.4	0. 2	60
蛋鸡		0. 12		1.63	70

1)各畜禽尿液和粪便中铵态氮分别在尿液和粪便中总氮的占比相同;2)单位畜禽(每头、每只或每羽)每天产生的尿液或粪便的量,单位为 kg

1.4 情景设置

综合已有研究^[31],在畜禽养殖业 NH,排放估算所需基础活动水平数据选取以及计算方法方面的不同,本文综合考虑活动水平、排放系数和气象参数选取上的差异,设置了8种计算情景,对比分析了不同估算方式下畜禽养殖业氨排放清单结果的差异.本文计算的畜禽活动水平种类涉及母猪、肉猪、奶牛、肉牛、山羊、绵羊和家禽这7大类.由于已有的研究在活动水平的选取差异主要体现在母猪、肉猪、奶牛、肉牛、山羊和绵羊等6种,家禽的活动水平数量均取其年内出栏量.因此本文讨论的活动水平主要是肉猪、肉牛和山羊数量的选取不同.情景设置条件见表6.其中,情景1(S1)和情景2(S2)分别表示活动水平采用畜禽年末存栏、年内出栏数据,排放系数采用表2国内外文献排放系数的均值.

表 4 散养畜禽养殖业氨排放系数及参数1)/%

Table 4 Ammonia emission coefficients and parameters of free-range livestock and poultry breeding industries/%

孙米	EF 户外		EF _{圖舍-液态}			EF _{圏舍-固态}			EF存	储-液态			EF存	者-固态		EE	EE
一件关	EF户外	T < 10	10 < T < 20	T > 20	T < 10	10 < T < 20	T > 20						N_2O		N_2	- EF _{施肥-液态}	LF 施肥-固态
肉牛	53	9.3	14	18. 7	9.3	14	18.7	20	1	0.01	0.3	27	8	1	30	55	79
奶牛	30	9.3	14	18. 7	9.3	14	18.7	20	1	0.01	0.3	27	8	1	30	55	79
山羊	75	9. 3	14	18. 7	9.3	14	18.7	28	7	0.01	0.3	28	7	1	30	90	81
绵羊	75	9.3	14	18. 7	9.3	14	18.7	28	7	0.01	0.3	28	7	1	30	90	81
母猪	0	9. 2	14. 7	20. 2	9. 2	14. 7	20. 2	14	0	0.01	0.3	45	5	1	30	40	81
肉猪	0	6. 2	10. 2	14. 2	6. 2	10. 2	14. 2	14	0	0.01	0.3	45	5	1	30	40	81
马	0	9. 3	14	18.7	9.3	14	18.7	35	0	0.01	0.3	35	8	1	30	90	81
驴	0	9.3	14	18. 7	9.3	14	18. 7	35	0	0.01	0.3	35	8	1	30	90	81
骡	0	9. 3	14	18.7	9.3	14	18.7	35	0	0.01	0.3	35	8	1	30	90	81
蛋鸡	69	24. 9	45. 2	56. 5	24. 9	45. 2	56.5	0	0	0	0	14	4	1	30	0	63

1) EF 为排放系数,单位为%(总铵态氮挥发率); EF $_{\text{\tiny | Be-ida}}$ EF $_{\text{\tiny | Be-ida}}$: 粪便排出阶段,室内环境下液态和固态粪便的氨挥发率; EF $_{\text{\tiny | Ph}}$: 粪便排出阶段,室外环境中的氨挥发率; EF $_{\text{\tiny fill-ida}}$: 存储阶段,液、固态粪便氨挥发率; EF $_{\text{\tiny inll-ida}}$: 施肥阶段,液、固态粪便氨挥发率; $_{\text{\tiny inll-ida}}$: 施肥阶段,液、固态粪便氨挥发率; $_{\text{\tiny inll-ida}}$: 施肥阶段,液、固态粪便氨挥发率; $_{\text{\tiny inll-ida}}$: $_{\text{\tiny inll-i$

表 5 集约化养殖畜禽氨排放系数及参数/%

Table 5 Ammonia emission coefficients and parameters of intensive culture of livestock and poultry/%

							1										
和米	EF户外		EF _{圖舍-液态}			EF _{圈舍-固态}			EF存	储-液态			EF存	诸-固态		FF	FF
作矢	上上户外	T < 10	10 < T < 20	T > 20			T > 20				N_2	$\overline{\mathrm{NH_{3}}}$	N_2O	NO	N_2	- EF _{施肥-液态}	EF _{施肥-固态}
肉牛	53	9. 3	14	18. 7	9. 3	14	18.7	15.8	1	0.01	0.3	4. 2	8	1	30	55	79
奶牛	30	9.3	14	18. 7	9.3	14	18.7	15.8	1	0.01	0.3	4. 2	8	1	30	55	79
山羊	75	9.3	14	18. 7	9.3	14	18.7	15.8	7	0.01	0.3	4. 2	7	1	30	90	81
绵羊	75	9. 3	14	18. 7	9.3	14	18.7	15.8	7	0.01	0.3	4. 2	7	1	30	90	81
母猪	0	8. 9	14. 3	19.7	8. 9	14. 3	19.7	3.8	0	0.01	0.3	4.6	5	1	30	40	81
肉猪	0	11.3	18. 5	25. 7	11.3	18. 5	25.7	3.8	0	0.01	0.3	4.6	5	1	30	40	81
马	0	9. 3	14	18.7	9.3	14	18.7	15.8	0	0.01	0.3	4. 2	8	1	30	90	81
驴	0	9. 3	14	18.7	9.3	14	18.7	15.8	0	0.01	0.3	4. 2	8	1	30	90	81
骡	0	9. 3	14	18.7	9.3	14	18.7	15.8	0	0.01	0.3	4. 2	8	1	30	90	81
蛋鸡	69	0	0	0	19. 7	35.9	44.9	0	0	0	0	3.7	4	1	30	0	63

这与杨新明等^[20]的研究方式相同;情景 3(S3)、情景 4(S4)、情景 5(S5)和情景 6(S6)是采取文献 [21]中畜禽粪便的铵态氮在不同管理阶段(户外、圈舍内、粪便存储和后续施肥)中不同的挥发效率,最终求和各阶段氨排放量的估算结果. 其中存、出栏量的粪便铵量表示各畜禽存、出栏活动水平以表 3 的估算参数计算而得. 年均温、月均温排放系数 为各地区在年均温度和月均温度下对应表 4 和表 5 的排放系数. Huang 等^[4]研究的 2006 年中国大气氨排放清单中畜禽养殖采用了 S6 这一计算方式. 此外刘波等^[15]以 S3(年均温与存栏量的活动水平)计算了长三角畜禽养殖氨排放水平;情景 7(S7)和情景 8(S8)的活动水平以生猪、肉牛、山羊等饲养周期小于 1 a 的畜禽的出栏量计算,而饲养周期大于 1 a 的畜禽以存栏量进行计算.

表 6 氨排放计算情景设置1)

Table 6 Scenarios of the estimation of ammonia emissions

情景	活动水平选取	排放系数选取
S1	畜禽的年末存栏量	文献排放系数均值
S2	畜禽的年内出栏量	文献排放系数均值
S3	畜禽存栏量的粪便铵量	年均温排放系数
S4	畜禽出栏量的粪便铵量	年均温排放系数
S5	畜禽存栏量的粪便铵量	月均温排放系数
S6	畜禽出栏量的粪便铵量	月均温排放系数
S7	畜禽存、出栏量的粪便铵量	年均温排放系数
S8	畜禽存、出栏量的粪便铵量	月均温排放系数

¹⁾ S2、S4 和 S6 中采用出栏量的畜禽为肉猪、肉牛和山羊,其他畜禽仍采用存栏量

2 结果与讨论

2.1 活动水平与排放系数对清单结果的影响

各情景下长三角地区畜禽养殖业大气 NH₃ 排放量结果见图 1. 从中可见,2017 年长三角区域畜禽养殖业大气氨排放量在各个情景(估算方式)下的排放量分别为67.5、86.1、21.3、28.7、21.6、29.1、28.6 和28.9 万 t. 情景 2(采用出栏量的活动水平数据与表 2 中对应排放系数)的测算结果最

高,为86.1万t.最小的估算结果是情景3(采用存 栏量的活动水平数据与文献[21]中对应的各活动 阶段的排放系数),仅有21.3万t.S4计算的排放量 为 28. 7 万 t, 高出 S3 情景 34. 7%, 这比 S1 和 S2 的 改变幅度(S2 高出 S1 的结果 27.6%)还要高. 而 S5 和 S6 的排放量略高于 S3 和 S4. 结果表明,在取用 同一排放系数时,活动水平(大牲畜)选用出栏量比 选用存栏量高约 27.6%~34.7%; 比较 S3 与 S5、S4 与 S6 的估算结果可以发现,在选用同样的活动水平 情况下,由于畜禽圈舍状态下的粪便氨排放系数受 温度影响较大,会使得地区整年的估算结果出现 0.3~0.4万t的偏差;比较S7和S8与前6种计算 结果,可以发现,综合考虑存、出栏粪便总量的计算 结果甚至低于单一以出栏量计算的结果(S4、S6). 这是因为 S7 和 S8 统计的存、出栏量中,出栏量所 产生的粪便量考虑了畜禽的饲养周期,由于饲料的 使用及饲养技术的提升,饲养周期可能是3、4、5或 6个月.调查长三角一些养殖基地对于肉猪、肉牛、 肉羊的饲养情况,本文中这3种畜禽在S7和S8情 景中分别选定饲养周期为3、6和3个月.由于S3~ S8 的估算方式中活动水平为畜禽当年的粪便排放 量,因此在计算出栏畜禽的粪便量时,计算结果将比 S4 和 S6(以个各畜禽 1 a 内均在排放)的粪便铵量 小4、2和4倍. 因此S7和S8情景的计算结果小于 S4 和 S6.

S1 与 S2、S3 与 S4、S5 与 S6 的结果均显示出活动水平选取的不同对清单结果的影响. 情景 1、情景 2 的结果相差 18.6 万 t,表明直接以活动水平与排放系数的乘积估算的结果受活动水平影响最为明显. 以文献 [21] 中的计算方法进行估算的各种情景,排放量远低于情景 1 与情景 2 的排放量. 本文中8 种估算情景的结果均表明,活动水平及排放系数选取的不同标准会对地区畜禽养殖大气氨排放产生一定程度的影响. 其中,以出栏量数据估算的结果会

远高于以存栏量数据进行估算的结果. 此外,受环境 因素(温度)影响的排放系数也会使得清单估算结 果出现一定程度的偏差.

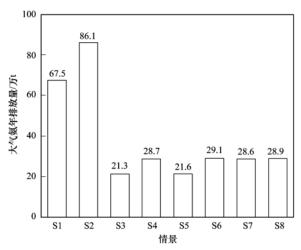


图 1 2017 年长三角畜禽养殖大气氨排放情景计算结果

Fig. 1 NH₃ emissions from livestock under different scenarios in the Yangtze River Delta in 2017

2.2 温度对清单结果的影响

由于文献[21]中排放系数在温度小于10℃ 10~20℃之间和大于20℃各不相同,因此计算了长 三角地区 2017 年畜禽养殖的各月均温下的氨排放. 由表 4 和表 5 的排放系数可知,温度对圈舍状态下 的排放系数影响较大. 比较 S3 与 S5 及 S4 与 S6 的 结果可知,选取同样的存、出栏量畜禽活动水平情 况下,选取各月不同温度的条件系数(月均温)计算 的结果高于以年均温度进行计算的排放量 0.3~ 0.4万t.图2为S8的条件下,3种月均温的长三角 畜禽养殖业大气氨的月排放量. 结果表明,温度在小 于10℃、10~20℃之间和大于20℃条件下的月排 放量分别为 2.24、2.39 和 2.54 万 t. 长三角地区的 年均温均在10~20℃之间,而大多数城市的月均温 有 3 个月 10℃以下, 4 个月处于 10~20℃之间, 高 于20℃有5个月左右. 故此造成以月均温的排放系 数来进行计算的结果高于年均温的结果,这也是在 5~10 月间畜禽养殖大气氨排放较高的主要原因.

2.3 不同情景下区域排放源清单的空间分布

本研究基于 ArcGIS 空间分析技术对氨排放进行空间分配,利用各情景计算的排放量,依据排放强度,采用 ArcGIS 中的插值技术对各地区排放强度进行比较分析,图 3 为各市在不同计算方式下的排放强度. 从中可见,各个排放情景的空间分布主要与畜禽选取的活动水平有关. 存、出栏分别计算的排放量有较大的差异. 活动水平的不同选取原则对长三角地区的排放强度有着明显的影响. 2017 年,以存栏量估算的清单结果显示,最高排放强度的 3 个地

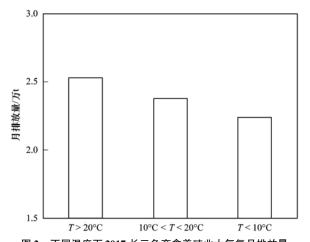


图 2 不同温度下 2017 长三角畜禽养殖业大气氨月排放量

Fig. 2 Monthly ammonia emissions from livestock in the Yangtze River Delta at different temperatures

区是铜陵市、蚌埠市、淮南市,舟山市、丽水市、黄山市是排放强度最小的3个市.而以年内出栏量估算的清单中丽水市、舟山市和黄山市是较小的3个市,嘉兴市、淮南市和南京市分别以6.0、8.2和10.2 t·km⁻²占据排放强度的前3位.

2.4 不同清单结果比较

综合8种情景的选取方式及结果分析,第8个 情景(S8)在活动水平选取上,考虑到肉类畜禽的饲 养周期,因此计算的粪便量更符合实际情况;且 S8 使用的各阶段排放系数考虑到温度的影响,使用了 更为详细的月均温度. 两者结合会使清单的估算结 果更为科学准确. 本文将 S8 的估算结果与其他同类 清单结果进行比较,从表7可知,与同一研究区域的 董艳强等[14]及刘波等[15]的清单结果相比,本文在 研究区域远多于前者研究情况下,清单结果却没有 远高于他们两人的研究. 对比发现, 董艳强等[14]的 研究主要结合国内外的排放因子,取以平均值进行 计算,对比本研究采用的计算方法,两者差异较大. 本文在分阶段考虑畜禽大气氨排放后,由于管理的 规范性及排放因子选取更符合地区性,本研究采取 文献[21]中的排放因子经过综合计算后,远低于董 艳强等[14]的平均值,而这也可能是表2中出现国内 外的同类研究在排放系数出现较大差异的原因. 而 对比刘波等[15]的研究结果发现,其在考虑畜禽活动 水平统计时,提供肉类产品的畜禽(饲养周期小于 1a) 统计了出栏量, 本研究 S8 中将生猪、肉牛、山 羊的活动水平(出栏量)分别考虑饲养周期,使得在 计算结果时出现倍数下降.

2.5 不确定性分析

氨排放量估算具有一定的不确定性,主要原因 有两个方面:一是精准畜禽活动水平数据获取的困 难性;二是排放因子数据匹配的问题.本文研究中

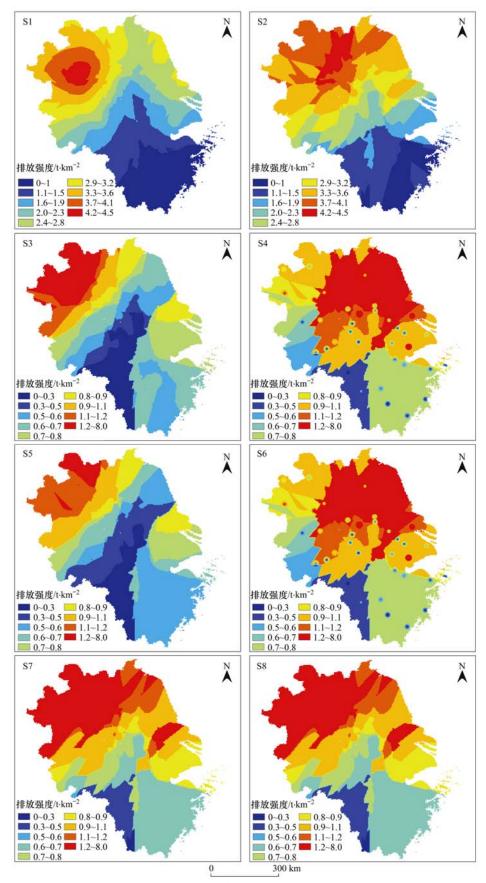


图 3 各估算情景下畜禽养殖业 NH₃ 排放的空间分布

Fig. 3 Spatial distribution of NH₃ emissions from livestock under different scenarios

畜禽养殖业排放因子一是设置了表 2 中文献的实测排放因子的均值,二是根据文献[21]推荐的系数进

行估算. 与国外的方法相比, 文献[21]中的方法所确定的排放因子更符合我国畜禽养殖实际情况,减

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	Comparison						

核算年份	地区	排放量/万 t	排放强度/t·km ⁻²	文献
2009	南通市	3. 2	3.46	[19]
2016	江苏省	46. 1	4. 48	[29]
2004	长三角(16 个城市)	20. 3	1. 85	[14]
2015	长三角(16 个城市)	21.4	1.94	[15]
2006	全国	531. 2	_	[4]
2017	长三角地区	28. 9	1. 14	本研究

小了对长三角地区氨排放估算的不确定性影响.本文研究中采用的畜禽活动水平数据来自长三角地区各市统计年鉴,非规模化与规模化畜禽养殖比例数据不全,对各市区非规模化与规模化畜禽活动水平的确定产生一定影响.

本文虽尽量多考虑活动水平和排放系数对清单结果的影响,设置了8种计算方式情景.但是由于数据来源的局限性,仍有不可避免的误差存在.年鉴中的猪、牛和羊的统计数目较为准确,但是却没有较好地与文献[21]中对应活动水平相匹配.如很多市并没有将山羊、绵羊,奶牛和肉牛等做详细的统计分类说明,这对估算工作有影响.此外,本研究通过S3~S8的情景结果已经得知温度对估算结果的影响,长三角区域,以月均温的计算方式会高于以年均温进行计算的结果.而更具体的外界条件必然能对估算结果有更进一步的提高.

3 结论

- (1)8 种估算方案的结果表明,以出栏量与排放系数计算获得的 NH₃ 排放结果最高,为 86.1 万 t·a⁻¹;以年末存栏量、年均温下及对应排放系数计算的 NH₃ 排放结果最小.各个情景的结果差异均表明清单估算中,不同活动水平选取的标准对结果的影响较为严重.一般以出栏量统计大牲畜会相对比以统计存栏量的结果高出 27.6%~34.1%.
- (2)在长三角区域,基于月均温度估算的结果 将比以年均温估算的结果高出 0.3~0.4万 t.
- (3)2017年,以存栏量估算的清单结果显示最高排放强度的三个地区是铜陵市、蚌埠市、淮南市; 舟山市、丽水市、黄山市是排放强度最小的三个市. 而以年内出栏量估算的清单中丽水市、舟山市、黄山市是较小的三个市,嘉兴市、淮南市、南京市分别以 6.0、8.2 和 10.2 t·km⁻²居排放强度的前三位.
- (4)本文经对比各个估算方式得出 2017 年长 三角畜禽养殖大气氨排放量为 28.9 万 t,其中南京 市以 2.6 万 t 排放最高,最低排放量的市是舟山市. 排放强度最高的是嘉兴市,最低的是丽水市.从整个

长三角的空间分布来看,安徽省北部及江苏省的西北部是主要的高排放区域,浙江省普遍排放较低.

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