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高爽、白霉、白岩、雷团团、土刚、李时海、陆朝阳、七娜、郝明亮、黄同峰(1575) 2015~2017年北京及近周边平房燃煤散烃及其污染排放遥感测算 起文意、李令军、鹿海峰、姜磊、张立坤、王新辉、邱昀(1594) 基于地基遥感的杭州地区气溶胶光学特性 齐冰、车慧正、徐婷婷、杜荣光、胡德云、梁卓然、马千里、姚杰(1604) 四川省人为源挥发性有机物组分清单及其臭氧生成潜势 周子航、邓也、谭钦文、吴柯颖、宋丹林、黄凤霞、周小玲(1613) 餐饮源挥发性有机物组成及排放特征 高雅琴、王红丽、许睿哲、景盛翱、刘跃辉、彭亚荣(1627) 广州番禺大气成分站一次典型光化学污染过程 PAN 和 O3 分析 邹宇、邓雪娇、李菲、殷长秦(1634) 北京市典型道路扬尘化学组分特征及年际变化 胡月琪、李萌、颜起、张超(1645) 南昌市扬尘 PM、中多环芳烃的来源解析及健康风险评价 于瑞莲、郑权、刘贤荣、王珊珊、敖旭、张超(1646) 现实工况下挖掘机尾气排放特征分析 马帅、张凯山、王帆、庞凯莉、朱怡静、李臻、毛红梅、胡宝梅、杨锦锦、王斌(1670) 雾。罐天人体平均呼吸高度处不同粒径气溶胶的微生物特性 杨唐、韩云平、李珠、《敬(1688) 支持向量机回归在臭氧预报中的应用 苏筱倩、安俊琳、张玉欣、梁静舒、刘静达、王鑫(1697) 基于中国电网结构及一线典型城市车辆出行特征的 PHEV 二氧化碳排放分析 郝旭、王贺武、李伟峰、欧阳明高(1705) 岩溶槽谷区地下河硝酸盐来源及其环境效应:以重庆龙风槽谷地下河系统为例 标准,生工工建、吴韦、彭学义、刘九维(1715) 胶州湾表层水体中邻苯二甲酸酯的污染特征和生态风险 刘成、孙翠竹、张哿、唐缭、邹亚丹、徐擎擎、李锋民(1726) 湛江湾沉积物中六六六(HCHs)、滴滴涕(DDTs)有机氯农药的分布特征与风险评估 张哿、唐缭、邹亚丹、徐擎擎、李锋民(1726) 湛江湾沉积物中六六六(HCHs)、滴滴涕(DDTs)有机氯农药的分布特征与风险评估 张哿、唐缭、邹亚丹、徐擎擎、李锋民(1726) 港上发化系统中,DOM 米层性特别,聚如用是点任意、以为成准系系为,阅读,陈法锦、于赤灵、李嘉诚、梁字钊、宋建中(1734)
内蒙古河套濯区不同盐碱程度土壤 CH。收収现律 物义柱,焦燕,物铭德,温息片(1950)水稻光合碳在植株-土壤系统中分配与稳定对施磷的响应 王莹莹,肖谋良,张昀,袁红朝,祝贞科,葛体达,吴金水,张广才,高晓丹(1957)土壤水分和温度对西南喀斯特棕色石灰土无机碳释放的影响 徐学池,黄媛,何寻阳,王桂红,苏以荣(1965)黄土丘陵区侵蚀坡面土壤微生物量碳时空动态及影响因素 覃乾,朱世硕,夏彬,赵允格,许明祥(1973)农用地土壤抗生素组成特征与积累规律 孔泉 是明明 经基础 人,是是,张世文,是超甲,胡青贵(1981)
  生物发酵制药 VOCs 与嗅味治理技术研究与发展 ··· 王东升,朱新梦,杨晓芳,焦茹媛,赵珊,宋荣娜,吕明晗,杨敏(1990)《环境科学》征订启事(1612) 《环境科学》征稿简则(1787) 信息(1663,1796,1833)
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## 尾水受纳河流中 PhACs 在传统水相中的分布及环境风险

王硕, 刘建超\*, 郑超亚, 张吉琛, 许嘉诚, 徐莹莹, 陆光华

(河海大学环境学院,浅水湖泊综合治理与资源开发教育部重点实验室,南京 210098)

摘要: 胶体介质不仅是水环境中污染物一个重要的"汇",还是污染物生物地球化学循环中至关重要的调控单元. 本研究利用错流超滤、固相萃取、液相色谱质谱联用仪等前处理和分析检测方法调查了 10 种典型的药物活性化合物(PhACs)在污水厂尾水受纳河流传统水相中的分布和环境风险水平. 结果表明, 10 种 PhACs 在水体溶解相和胶体中的含量分别达到 27.2~168.1  $\operatorname{ng\cdot L^{-1}}$ 和 164.5~751.1  $\operatorname{ng\cdot g^{-1}}$ . 布洛芬(IPF)、罗红霉素(ROX)和红霉素(ETM)是两种介质中最为主要的污染物,三者污染浓度占到总浓度的 80% 以上. 胶体对 ROX、酮康唑、ETM 和舍曲林都表现出较强的吸附性能,胶体/水分配系数( $\operatorname{lg}K_{\operatorname{col}}$ )在 3.2~4.0 之间,吸附率达到 21.1%~34.5%. 10 种 PhACs 对绿藻、溞和鱼的急、慢性毒性风险评估结果中,仅IPF 对鱼类产生高等慢性风险,其余为中等风险及以下. 值得注意的是,相对于急性风险来说,更多的 PhACs 对高等水生生物产生慢性不利影响.

关键词:药物活性化合物(PhACs);溶解相;胶体;吸附;环境风险

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## Distribution and Environmental Risk of Pharmaceutically Active Compounds in the Traditionally Aqueous Phase of Effluent-receiving Rivers

WANG Shuo, LIU Jian-chao<sup>\*</sup>, ZHENG Chao-ya, ZHANG Ji-chen, XU Jia-cheng, XU Ying-ying, LU Guang-hua (Key Laboratory for Integrated Regulation and Resources Development on Shallow Lakes of Ministry of Education, College of Environment, Hohai University, Nanjing 210098, China)

Abstract: Colloid media are not only an important "sink" for pollutants in the aquatic environment, but also a crucial regulating unit for the biogeochemical cycle of pollutants. In this study, the distribution and environmental risk levels of ten typical pharmaceutically active compounds (PhACs) in the water phase of effluent-receiving rivers were investigated using cross-flow ultrafiltration, solid-phase extraction, and liquid chromatography-tandem mass spectrometry as the pretreatment and analysis methods. The results showed that the total concentrations of the ten PhACs in the dissolved phase and colloidal phase ranged from 27. 2 to 168. 1  $\text{ng} \cdot \text{L}^{-1}$  and 164. 5 to 751. 1  $\text{ng} \cdot \text{g}^{-1}$ , respectively. Ibuprofen (IPF), roxithromycin (ROX), and erythromycin (ETM) are the dominating pollutants in the dissolved phase and colloidal phase, accounting for more than 80% of the total concentration. Strong adsorption properties for ROX, ketoconazole, ETM, and sertraline were found in the colloid phase, their colloid/water distribution coefficients ( $\text{lg}K_{col}$ ) ranged from 3. 2 to 4. 0, and the percentage of PhACs absorbed to the colloidal phase reached 21. 1% -34. 5%. The risk assessment of acute and chronic toxicity to algae, daphnia, and fish showed that only IPF presented a high chronic risk to fish, while the risk levels of the other PhACs were at or below medium risk. It is worth noting that, in comparison with their acute risk, most PhACs have chronic negative effects on higher aquatic organisms.

Key words: pharmaceutically active compounds (PhACs); dissolved phase; colloid; adsorption; environmental risk

药物活性化合物(pharmaceutically active compounds, PhACs)作为一类新型有机污染物,主要包括抗生素、抗真菌剂、抗抑郁类、镇癫痫类、降血脂类药物等<sup>[1]</sup>.由于 PhACs 在我国大量生产,并广泛应用于农业、水产养殖业、畜牧业和人类健康维护等,造成相当数量的 PhACs 及其代谢产物排入自然环境<sup>[2,3]</sup>.近年, PhACs 在污水、地表水、地下水、甚至饮用水中都有检出,直接威胁非靶向水生生物及人类健康安全,因此 PhACs 污染问题已经成为环境科学界的热点课题<sup>[4-7]</sup>.

目前, PhACs 在水环境中的分布主要集中在传统水相、悬浮颗粒物相、沉积相等介质界面赋存研究<sup>[8-11]</sup>, 忽略了胶体等典型微界面的存在, 这对 PhACs 环境风险准确评估具有一定影响<sup>[12,13]</sup>. 天然

水体中胶体介质是粒径介于 1 nm ~ 1 μm 的无机和有机非均相颗粒物,具有体积小、比表面积大、吸附位点多等特点<sup>[14]</sup>.现有研究发现水体中胶体是PhACs 类污染物的重要"汇",能够改变或控制污染物的降解、转运、生物利用等环境行为<sup>[15,16]</sup>.在长江三角洲地区 40 余种 PhACs 的水体介质分布研究中,发现胶体能够吸附传统水相中 1/3 左右的PhACs,其吸附能力是悬浮颗粒物的 2 ~ 4 倍,并指

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作者简介: 王硕(1998~), 男, 主要研究方向为污染生态化学, E-mail:1065746843@ qq. com

\* 通信作者,E-mail:jianchao-liu@hhu.edu.cn

出污水受纳水体应重点关注[17,18].

本研究以城市污水厂尾水受纳水体为研究区域,选择公众关注度极高的典型 PhACs,调查其在胶体、溶解相等水环境介质中的赋存水平,评估其在该区域中的环境风险水平,以期为新型污染物的环境标准和环境容量制定提供基础数据.

#### 1 材料与方法

#### 1.1 试剂与仪器

红霉素(ETM)、罗红霉素(ROX)、酮康唑(KTC)、布洛芬(IPF)、苯扎贝特(BZB)、立定痛(CBZ)、吉非罗齐(GFB)、吲哚美辛(IMC)、普萘洛尔(PRP)、舍曲林(SER)10种 PhACs购买于梯希爱(上海)化成工业发展有限公司,纯度>98%;丙酮、乙腈、甲醇(HPLC级)等试剂购买于德国MercK公司;固相萃取小柱(HLB,200 mg,6 mL)购买于美国Waters公司.

#### 1.2 实验方法

采样区域及方法:选取南京市江宁污水厂(J1、J2)、城东污水厂(G1、C2)、科学园污水厂(K1、K2)上下游 1.5 km 等 6 处作为采样区域. 于 2017年7月份在采样区域进行河流断面采样,每个断面设置 3 个采样点(采样断面中心点和离两岸各 1.5 m 远处点),在每个采样点水面以下 0.5 m 处利用 1 L 棕色玻璃瓶进行地表水样品采集. 具体采样方法参照《地表水和污水监测技术规范》(HJ/T 91-2002)及相关监测分析方法规定的采样要求进行. 采集后的水样品放入含有干冰的容器中储存,并迅速送回实验室做进一步处理.

水样分离及预处理:运回实验室的水样用 1  $\mu$ m 玻璃纤维滤膜进行砂滤过滤,去除悬浮颗粒物,获得传统水相. 传统水相再通过切向超滤装置和 1 ×  $10^3$  的聚醚砜超滤膜进行胶体和溶解相分离,体积比为 1: 9(胶体浓缩液/溶解相)<sup>[19]</sup>,样品分离完毕后,两种溶液分离用固相萃取装置进行 PhACs 的提取、浓缩处理. 处理前,依次用 10 mL 甲醇和 10 mL 超纯水活化 Oasis HLB 固相萃取柱. 以 3 ~ 5 mL·min <sup>-1</sup>的流速对胶体浓缩液和溶解相进行富集. 富集后的萃取柱用 10 mL 超纯水淋洗,用  $N_2$  干燥去除萃取柱中水分. 然后用  $2 \times 5$  mL 甲醇进行洗脱,洗脱液收集于 15 mL 玻璃管中并在 15 mL 棕色色谱瓶中,存放于 15 mL 收集在 1.5 mL 棕色色谱瓶中,存放于 15 mL 收集在 1.5 mL 棕色色谱瓶中,存放于 15 mL 收集在 1.5 mL 棕色色谱瓶中,存放于 15 mL 吹集在 1.5 mL 棕色色谱瓶中,存放于 15 mL 棕色色

胶体质量的测定:取1 L 经1 μm 玻璃纤维滤膜 过滤后的水样,用切向超滤装置进行胶体浓缩,胶 体浓缩液与溶解相的比达到1:9以上时,将胶体浓 缩液倒入预烘干的 1×10<sup>3</sup> 的聚醚砜超滤膜折叠杯中,将此折叠杯置于60℃的烘箱中烘至质量不变为止,通过质量差值法计算水体中胶体的质量浓度.

#### 1.3 仪器检测与质量控制

利用超高效液相色谱质谱联用仪对 10 种PhACs 进行定性、定量分析,分别以含有 0.05% 甲酸的甲醇水溶液和含有 5 mol·L<sup>-1</sup>乙酸铵的乙腈水溶液为正负离子流动相.采用多重反应监测模式(MRM)对目标化合物进行定量分析.ROX、ETM、IPF、IMC、GFB、PRP、CBZ、KTC 和 SER 以正离子模式检测,而 BZB 用负离子模式检测.样品前处理过程中采用严格的质量保证与控制,每批次处理 12 个样品,其中包含 1 个溶剂空白、1个野外空白、1个基质加标和 9个野外样品.胶体相回收率实验采用的是经过高温灭菌后的天然胶体物质.10 种 PhACs 的定量限为 0.05~1.2 ng·L<sup>-1</sup>,回收率为 75%~110%,相对标准偏差低于 20%,溶剂空白和野外空白中未检出目标化合物.

#### 1.4 数据分析方法

为研究水体中 PhACs 对水生生物的影响,采用风险熵值法(RQ)进行生态风险评估,其计算公式如下:

$$RQ = MEC/PNEC$$
 (1)

PNEC = (LC<sub>50</sub>/AF 或 EC<sub>50</sub>/AF 或 NOEC/AF)

(2)

式中, MEC 为环境检出浓度; PNEC 为预测无效应浓度.  $LC_{50}$ 、 $EC_{50}$ 和 NOEC 是半致死浓度、半效应浓度和无可见效应浓度; AF 为评价因子.  $LC_{50}$ 、 $EC_{50}$ 和 NOEC 根据前期研究获得, AF 选取欧盟 Water Framework Directive 的推荐值 $^{[12,20]}$ .

胶体-溶解相分配系数  $(K_{col})$  和胶体吸附率  $(A_{col})$  是衡量 PhACs 在胶体、溶解相间分配关系和胶体对 PhACs 吸附能力的参数,其计算公式如下:

$$K_{\rm col} = (C_{\rm col} \times 1000)/C_{\rm water} \tag{3}$$

$$A_{\rm col} = \frac{(C_{\rm col} \times m_{\rm col})}{C_{\rm col} \times m_{\rm col} + 1000 \times C_{\rm water}}$$
(4)

式中, $C_{col}$ 为胶体中目标污染物的含量 $(ng \cdot g^{-1})$ ; $m_{col}$ 为水体中胶体的质量浓度 $(ng \cdot L^{-1})$ ; $C_{water}$ 为溶解相中目标污染物的质量浓度 $(ng \cdot L^{-1})$ . SigmaPlot 12.5 数据处理包对数据进行处理和图形绘制.

#### 2 结果与讨论

#### 2.1 PhACs 溶解相中的分布特征

10 种 PhACs 在 3 个污水厂上下游水体中均被 检出,除 KTC 和 CFB 外,其余 8 种 PhACs 检出率 为 100%. PhACs 在污水厂上下游总质量浓度范围分别为 27.2~104.3 ng·L<sup>-1</sup>和 48.0~168.1 ng·L<sup>-1</sup> [图 1(a)],其中 IPF 浓度最高,在城东污水厂下游(C2)达到 101.0 ng·L<sup>-1</sup>,其次是 ROX 和 ETM,最高浓度分别达到 34.3 ng·L<sup>-1</sup>和 26.1 ng·L<sup>-1</sup>. 从空间分布来看,城东污水厂排放口附近 PhACs 污染浓度最高,这是因为 3 个污水厂受纳河流中,城东污水厂尾水受纳河流流量最小,稀释效应最低. 此外,3 个排放口下游水体中 PhACs 浓度都明显高于下游,表明污水厂尾水仍是 PhACs 进入自然水环境的重要途径. 从单一药物贡献率来看[图 1(b)], IPF、ROX 和

ETM 是检出浓度较高的 3 种 PhACs, 三者浓度平均 贡献率达到 80%, 这与我国对 IPF、ROX 和 ETM 的 生产和使用量直接相关. 2013 年在南京秦淮河水体 中同样广泛检出了 IPF、ROX 和 ETM 的存在, 最高 浓度分别达到 86.0、66.5 和 85.3 ng·L<sup>-1[12]</sup>, 这与本 研究结果相似, 但都明显高于长江水体(南京段)中 3 种污染物的检出浓度(IPF、ROX 和 ETM 平均浓度分别为 36.0、22.1 和 9.7 ng·L<sup>-1</sup>)<sup>[21]</sup>. 相似的浓度分布 在我国珠江、长江、黄河、海河、辽河等江河流域均 被发现, 而且 IPF 被视为检验城市尾水排放的典型 特征性物质<sup>[5]</sup>.

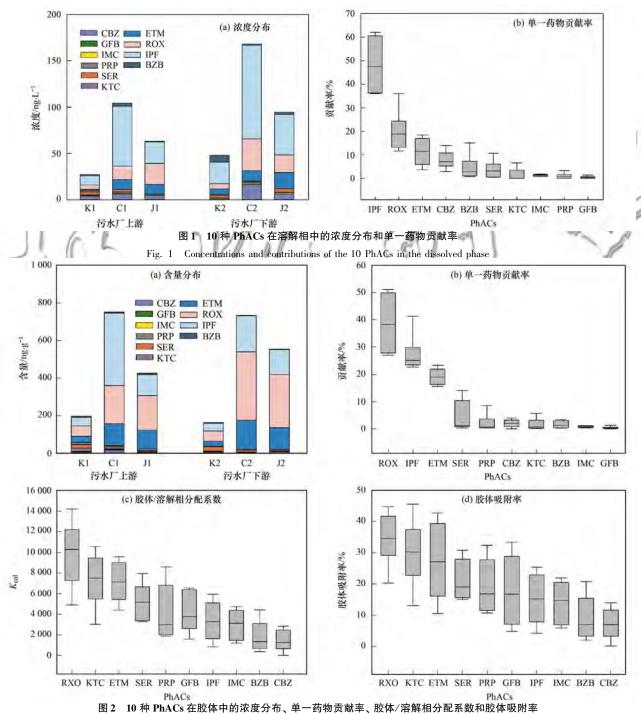


Fig. 2 Concentrations, contributions, colloid/water distribution coefficients ( $K_{\rm col}$ ), and colloidal adsorption of the 10 PhACs in the colloids

#### 2.2 PhACs 在胶体中的分布特征

在胶体相中,10种 PhACs 均被100%检出,在尾水排放口上下游的总含量分别为196.9~751.1 ng·g<sup>-1</sup>和164.5~733.1 ng·g<sup>-1</sup>[图2(a)].与溶解相中 PhACs 的空间分布相似,城东污水处理厂排放口附近是 PhACs 污染最为严重区域,其次是江宁污水处理厂、科学园污水处理厂,但是排放口上下游水体胶体中 PhACs 含量并没有明显差异.从单一药物贡献率来看[图2(b)],ROX、IPF和ETM是胶体介质吸附含量最高的3种 PhACs,平均贡献率分别达到38.7%(平均含量为190.6 ng·g<sup>-1</sup>)、29.0%(平均含量为150.7 ng·g<sup>-1</sup>)和19.2%(平均含量为91.6 ng·g<sup>-1</sup>);三者总贡献率为87%,这与溶解相质量浓度分布相似。但从胶体/水分配系数(K<sub>col</sub>)和胶体吸附率来看[图2(c)和2(d)],ROX、KTC、ETM和SER都表现出较强的吸附性能,平均lgK<sub>col</sub>

在  $3.2 \sim 4.0$  之间,胶体对这 4 种 PhACs 的吸附率分别达到了 34.5%、29.9%、27.2% 和 21.1%. 由于 IPF 平均  $\lg K_{col}$  仅为 3.5,因此其在胶体中吸附含量有所下降.先前研究发现,在污水厂尾水受纳水体中 IPF 的  $\lg K_{col}$  在  $2.35 \sim 3.06$  之间  $^{[17]}$ ,略低于本研究结果.但本研究结果又低于长江三角洲地区水体中 PhACs 的  $\lg K_{col}$  ( $5.75 \sim 7.58$ ),这与河流水质特征(盐度、温度等)和胶体理化性质(颗粒组成、有机碳含量等)直接相关  $^{[22]}$ .

#### 2.3 PhACs 环境风险评价

根据水体中 PhACs 溶解相浓度和目标污染物的急慢性毒理数据,基于风险熵值计算模型对目标污染物进行生态风险评估.评估结果一般分为 4 个等级: RQ < 0.01 为无风险;  $0.01 \le RQ < 0.1$  为低风险;  $0.1 \le RQ < 1$  中等风险;  $RQ \ge 1$  高风险 [23.24].评估结果如图 3 所示,从急性风险评估结果来看

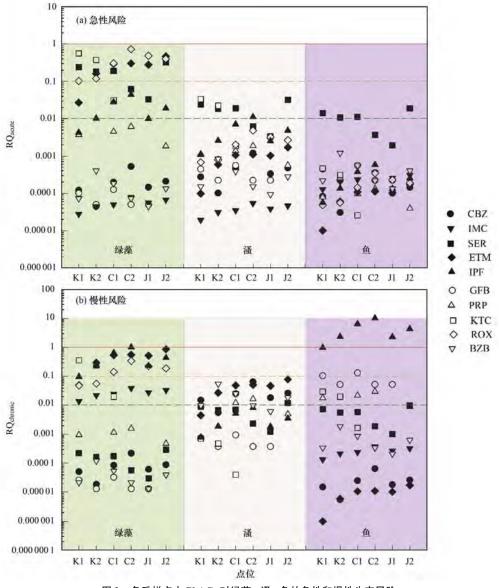


图 3 各采样点中 PhACs 对绿藻、溞、鱼的急性和慢性生态风险

 $Fig.~3~~RQ_{acute}~and~RQ_{chronic}~of~each~PhAC~based~on~acute~and~chronic~data~for~algae,~daphnids, and~fish$ 

[图3(a)], ROX、ETM、KTC 和 SER 在大部分采 样点对绿藻 RQ 均大于 0.1, 表明监测区域水体中 这4种 PhACs 的赋存水平可能对低等水生生物绿 藻产生中等急性风险,其它 PhACs 对绿藻为低等急 性风险或无风险. 所有 PhACs 对水溞和鱼类的急性 风险值都低于0.1,属于低等急性风险或无风险. 整体来看, PhACs 对水生生物的急性毒性风险值随 着水生生物营养级的增大而逐步降低. 从慢性风险 评估结果来看[ 图 3( b) ],ROX、ETM、KTC 和 IPF 对绿藻表现出中等慢性毒理风险, IMC 属于低等慢 性风险, 其他 PhACs 对绿藻为无风险. 值得注意的 是相对于急性风险水平, 更多的 PhACs 对鱼类产生 了中等或低等慢性风险,如 GFB、PRP、KTC 和 IPF. 尤其是 IPF, 其在所有采样点均对鱼类产生高 等慢性毒理风险,可能在鱼类的生殖系统产生不利 影响, 如雌鱼孵化延迟、怀卵次数降低、雄鱼卵黄 蛋白升高等[25]. 根据目前 IPF、ETM 和 ROX 的使 用量、污水厂去除效能、水体污染水平和潜在生态 风险, 多项研究建议对这3种 PhACs 应加强排放管 理和优先控制[5,12,26]

#### 3 结论

本文调查了10种典型 PhACs 在尾水受纳河流 溶解相和胶体中的污染水平及环境风险现状. 结果 显示:10 种 PhACs 在研究区域被广泛检出, 其中城 东污水处理厂尾水受纳河流, 因其流量最小, 成为 PhACs 污染最为严重区域, 在溶解相和胶体中的最 高浓度分别达到 168.1 ng·L-1和 751.1 ng·g-1; 且 在溶解相中, 尾水排放口下游 PhACs 质量浓度明显 高于上游, 表明尾水仍是城市河流中 PhACs 的重要 来源. ROX、ETM 和 IPF 是溶解相和胶体介质中污 染丰度最高的3种 PhACs, 三者总贡献率达到80% 以上. 10 种 PhACs 的  $\lg K_{col}$ 和吸附率分别为 3.2 ~ 4.0 和 7.2% ~ 34.5%, 其中 ROX、KTC、ETM 和 SER 表现出较高的吸附特性, 胶体吸附量占传统水 相的 20% 以上. 相对于急性风险, PhACs 对水生生 物产生的慢性毒理风险更应引起注意,尤其是 IPF 对高等水生生物鱼类产生的慢性不利影响.

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# **HUANJING KEXUE**

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