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## 生物炭对污泥堆肥及其利用过程重金属有效态的影响

周楫,余亚伟,蒋越,杨雨浛,张成\*

(西南大学资源环境学院,重庆 400715)

摘要:以城市脱水污泥为研究对象,设置 2 种处理(A 组:添加水稻生物炭; B 组:未添加生物炭)进行污泥堆肥,并将污泥堆肥产品进行土地利用,研究污泥堆肥及其利用过程重金属(Cd、Pb、Cu、Zn、Ni)的变化特征及其钝化效果,同时考察添加生物炭的影响作用.结果表明:在污泥堆肥及其短期利用过程中,除 Ni 外,重金属总量没有显著变化,水稻生物炭对 5 种重金属总量的影响也不显著.污泥堆肥过程对 5 种重金属具有一定钝化作用,添加生物炭能显著降低重金属有效态含量,并具有显著的钝化效果(P < 0.05),钝化率达到 16.39%~43.10%,其中 Zn、Ni 的钝化效果更为显著;而未添加生物炭的污泥堆肥过程对重金属有效态的钝化效果不显著(P > 0.05)。施用污泥堆肥会增加土壤重金属含量,短期内,生物炭对污泥堆肥土壤利用后的重金属有效态具有一定影响,但效果不显著.

关键词:污泥堆肥;土地利用;重金属有效态;生物炭;钝化

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# Effect of Biochar on Available Heavy Metals During Sewage Sludge Composting and Land Application of Compost

ZHOU Ji, YU Ya-wei, JIANG Yue, YANG Yu-han, ZHANG Cheng\*

(College of Resources and Environment, Southwest University, Chongqing 400715, China)

Abstract: The effects of biochar addition to compost on change characteristics and passivation effect of heavy metals (Cd, Pb, Cu, Zn, Ni) were investigated during the process of sludge composting with two different composts (group A: with biochar; group B: without biochar) and land application of compost. The results indicated that the total amount of heavy metals (except Ni) did not change significantly during the process of sludge composting and land application of compost. Additionally, biochar addition had little effect on the total amount of heavy metals. During the sludge composting process, five heavy metals (Cd, Pb, Cu, Zn, Ni) were passivated. Sludge composting with the addition of biochar can decrease the available contents of heavy metals, and the passivation effect of heavy metals was significant (P < 0.05). The passivation rate of the five examined heavy metals (Cd, Pb, Cu, Zn, Ni) ranged from 16.39%-43.10%, and the passivation effect for Zn and Ni was more significant. However, the passivation effect was not significant in the sludge composting process without the addition of biochar (P > 0.05). The concentrations of heavy metals in soil increased with the application of sewage sludge compost products. In the short term, biochar had a certain passivation effect on the available heavy metals in soils with sludge compost application, but the effect was not significant.

Key words: sludge compost; land use; available heavy metals; biochar; passivation

随着我国工业化与城市化速度加快, 我国污水 处理量日益增加,相应地污泥产量也逐年上升, 2016 年污泥产量已突破4 500万 t. 国家新环境保护 法及《水污染防治行动计划》强调了我国污泥处理 处置的迫切性, 堆肥及其土地利用被认为是一种具 有成本效益的有机废物处理和再利用的方式[1],但 其含有重金属可能造成环境的二次污染, 因而限 制了污泥堆肥土地利用的推广和应用. 研究表明, 在堆肥过程中添加一些外源物质可降低污泥中重 金属有效态含量. 如杨坤等[2]通过在堆肥处理中 施用不同钝化剂,发现重金属钝化效果不同,如 膨润土和磷矿粉对交换态铅的钝化效果较好, 硅 藻土对交换态镉的钝化效果较好. 在污泥堆肥过 程添加粉煤灰和生石灰, 能降低重金属的生物有 效性,增加 Zn、Ni 的残渣态含量[3]. 近年来,生 物炭作为一种新型高效安全的功能型材料,被广 泛应用于土壤修复、环境安全等领域[4,5]. 生物炭 直接施入土壤后,短期能显著降低土壤重金属有效态含量<sup>[6]</sup>.不同用量生物炭均能降低土壤 Pb 和 Cd 的含量<sup>[7]</sup>,秸秆生物炭能显著降低土壤中 Cu 的酸溶态<sup>[8]</sup>.生物炭在堆肥过程中也具有一定的重金属钝化效果,侯月卿等<sup>[9]</sup>在猪粪堆肥过程中加入不同生物质炭,发现花生壳炭、玉米秸秆炭和木屑炭分别对 Cu、Pb、Cd 具有较好的钝化效果. Zhou等<sup>[10]</sup>发现木屑炭和麦草炭也能钝化猪粪堆肥过程中的 Pb、Cu 等重金属. Chen 等<sup>[11]</sup>在猪粪堆肥中加入竹炭后,发现对于 Cu 和 Zn 的钝化效果较为显著. 这些研究表明生物炭对重金属的形态变化具有一定的影响作用,但关于生物炭在

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作者简介: 周楫(1998~), 女, 主要研究方向为环境污染化学, E-mail:747594755@qq. com

\* 通信作者,E-mail:zhcheng@126.com

城市污泥堆肥过程中对重金属的影响作用,以及将污泥堆肥施入土壤后,生物炭是否会持续产生钝化作用还缺乏研究.本试验将城市污泥与秸秆进行混合堆肥,并添加水稻生物炭作为调理剂,堆肥后再进行土地利用,研究污泥堆肥及其短期利用过程重金属的变化特征,同时分析添加水稻生物炭对重金属的影响及钝化作用,以期为进一步开展城市污泥土地利用的重金属污染控制和环境风险研究提供基础数据.

#### 1 材料与方法

#### 1.1 污泥堆肥试验设计

污泥堆肥试验于 2015 年 4 月进行,以城市脱水污泥和玉米秸秆作为原料,以水稻生物炭作为调理剂,按照质量比分别将污泥: 秸秆: 生物炭 = 20:5 :1(A组处理)、污泥: 秸秆 = 4:1(B组处理)充分混合,调节含水率在 60% ~ 70% 后放入自制堆肥箱(图1),堆肥原料总质量为 50 kg,体积约为 0.279 m³. 简易堆肥箱为 PVC 材料外壳以及泡沫箱制成内层制作而成,有效尺寸为 0.90 m× 0.45 m× 0.71 m(长×宽×高),有效体积约为 0.288 m³. 堆肥期间采用强制通风 + 人工翻堆的方式进行处理[12],

前 2 周每周翻堆 2 次,随后每周翻堆 1 次;以 50 L·min<sup>-1</sup>的小型鼓风机供气,采用通气管按照设计量进行分流,设置循环通风/关闭为 20 min/20 min,通风量约为 0.1 m³·(min·m³)<sup>-1</sup>.

堆肥所用污泥来自重庆市北碚区某污水处理厂脱水污泥,玉米秸秆来源于重庆市西南大学试验农场(剪碎至2~3 cm). 水稻生物炭购于商丘市三利新能源有限公司,为水稻秸秆在500℃条件下热解炭化3h,破碎后过0.35 mm 筛所得. 堆肥物料基本理化性质见表1.

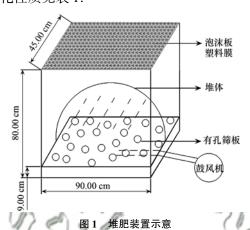


Fig. 1 Sketch map of composting process

表 1 堆肥物料基本理化性质

Table 1 Basic characteristics of composting materials

7 20 11 / 6			T C		
名称	pH	含水率/%	TC/g⋅kg <sup>-1</sup>	$TN/g \cdot kg^{-1}$	TOC/g·kg <sup>-1</sup>
原污泥	7. 23	81. 02	598. 49	27. 79	301. 67
玉米秸秆	_	15.03	492. 07	8. 92	419. 63
水稻生物炭	10. 24	2. 22	578. 35	7. 68	249. 60
A 组处理	8. 01	66. 08	536. 55	25. 30	353. 61
B组处理	7. 82	67.71	512. 23	24. 12	312.76

#### 1.2 堆肥土地利用试验设计

经检测,上述污泥堆肥所得的堆肥产品重金属含量(表 2)均能达到《农用污泥中污染物控制标准》(GB 4284-84),因此进一步进行土地利用试验.试验于2015年7月10日~8月14日于西南大学试验大棚中进行,试验周期为36  $d^{[13]}$ ,共设置7个处理: CK(不施肥)、A1、A2、A3(施肥量分别为0.40、0.80、1.19  $kg\cdot m^{-2}$ )、B1、B2、B3(施肥量分别为0.41、0.83、1.24  $kg\cdot m^{-2}$ ). 供试植物为耐热且早熟的不结球白菜(*Brassica campestris* L. ssp.

chinensis Makino),供试地点位于西南大学试验大棚. 每块试验田为1 m², 试验田之间用 PVC 板隔开深入土壤表层 30 cm,隔绝地表径流,处理之间间隔约1 m,避免不同处理组之间相互干扰. 将磨细的堆肥产品均匀撒在各土壤样方表面后混合均匀,并浇水,后期不追肥. 种植密度为40 穴·m², 每穴3 株,每天上午 08:00 前浇水灌溉,作物成熟后收割<sup>[13]</sup>. 按照传统管理模式: 施肥、间苗、翻耕、灌溉等管理菜地. 供试土壤和污泥堆肥基本理化性质见表 2.

表 2 供试土壤和污泥堆肥基本理化性质

Table 2 Basic characteristics of tested soil and sludge compost

类别	рН	TC/g·kg <sup>-1</sup>	TN/g•kg <sup>-1</sup>	Cd/mg·kg <sup>-1</sup>	Pb/mg·kg <sup>-1</sup>	Cu/mg•kg <sup>-1</sup>	Zn/mg·kg <sup>-1</sup>	Ni/mg•kg <sup>-1</sup>
原污泥	7. 23	598. 49	27. 79	1. 87	83. 60	96. 46	989. 76	46. 32
供试土壤	7. 93	64.40	1. 39	0. 64	19. 38	38. 58	58. 50	24. 90
A 组堆肥	7.58	414. 98	24. 34	1. 29	46.73	74. 92	469. 67	45. 41
B组堆肥	7. 33	338. 73	24. 92	1.31	44. 19	76. 81	483. 67	43. 01

#### 1.3 样品采集与分析

#### 1.3.1 堆肥固体样品

分别于堆肥初期和堆肥结束后采集堆肥固体样品. 采用剖面采样法, 采集不同深度的固体样品, 混合均匀后取 300 g 左右带回实验室, 测定堆体基本性质及重金属含量. 堆体温度分别在堆体四周 4 个点、中心部位选择 3 个固定点, 深入堆料 30 cm 内测定, 取平均值.

#### 1.3.2 土壤样品

在土地利用初期和利用结束时分别采集土壤样品.土壤样品的采集选取作物种植附近的表层土壤(0~10 cm).土壤每次采集样品约 250 g 左右,各处理组每次采集3个样品;采集的样品放入样品袋中,放入冻干机中冻干,研磨并过 100 目尼龙筛测定.堆肥固体样品基本性质、土壤基本性质及重金属含量测定方法参见文献[14,15].其中,重金属有效态含量采用二乙基三胺五乙酸(DTPA)浸提原子吸收分光光度计测定.

#### 1.4 重金属的钝化效果

污泥堆肥及其土地利用过程中,重金属的钝化效果主要通过对重金属有效态的钝化来体现.重金属有效态的钝化效果计算公式如下:

$$P_{i} = \frac{C_{i}}{C} \times 100\%$$

$$N_{i} = \frac{P_{1} - P_{2}}{P_{1}} \times 100\%$$
(1)

式中, $P_i$  为各重金属有效态含量占其总量的质量分数(%); $C_i$  为重金属的有效态含量( $mg \cdot kg^{-1}$ );C 为重金属总量( $mg \cdot kg^{-1}$ ); $N_i$  为重金属有效态钝化率(%); $P_1$  为钝化前重金属有效态含量占总量的质量分数(%); $P_2$  为钝化后重金属有效态含量占总量的总量的质量分数(%).

本研究采用空白试验、平行试验和标准样品进行质量控制,所测浓度值与标准值相差均在6%以内,Cd、Pb、Cu、Zn、Ni的加标回收率为88.6%~104.8%.采用SPSS 20.0和 Origin 9.0进行统计分析和作图.

#### 2 结果与讨论

#### 2.1 污泥堆肥基本性质变化特征

两种处理堆体温度及环境温度变化如图 2. 两种处理的堆体温度在堆肥开始时急速上升,快速进入高温期(50~70℃),高温持续时间均超过 1 周,且添加生物炭的堆体高温持续时间更长. 10 d 后堆体温度逐渐下降,并在 1 个月后逐渐达到环境温度,堆肥时间持续 50 d. 这可能是生物炭的多孔结

构为氧气在堆体内的运输和分布提供有利条件,更利于微生物的活动释放热量,提高堆体温度,加快堆肥的腐熟进程<sup>[16]</sup>. 两种处理 pH 值均表现为先升高后下降的趋势,这主要是由于前期堆体中含氮有机物不断降解,产生的氨逐渐积累导致 pH 值增大<sup>[17]</sup>. 随着堆体温度下降,有机物分解产生有机酸,pH 值开始下降.

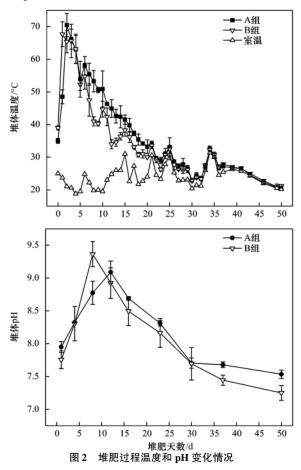


Fig. 2 Changes in temperature and pH during composting

#### 2.2 生物炭对污泥堆肥过程重金属总量的影响

污泥堆肥前后,两种处理重金属总量变化如表 3 所示.污泥堆肥产品重金属(Cd、Pb、Cu、Zn、Ni)含量均低于《农用污泥中污染物控制标准》(GB 4284-84)最低限值.在堆肥前后,两种处理堆体中 Ni 含量分别降低了 19% 和 16%.其他 4 种重金属(Cd、Pb、Cu、Zn)含量略有变化,但总体变化不显著(P>0.05).在污泥堆肥过程中,重金属的形态会发生变化<sup>[18]</sup>,但其总量难以被降解,重金属依然存在于堆肥产品中.如冯春等<sup>[19]</sup>研究发现,污泥堆肥后,堆体中交换态 Cu、Pb、Zn、Cd 和 As 的含量显著降低或升高.

## **2.3** 生物炭对污泥堆肥过程重金属有效态的影响和钝化

A、B两种处理堆肥前后重金属有效态含量有

#### 表 3 堆肥前后重金属含量变化/mg·kg-1

m 11 a	7.7	1	1 6 1	c 1		,	?	1
Table 3	Heavy metal	changes	before and	atter the	composting	process/	mg•kg i	

		, 0	1 01	0 0	
处理组	Cd	Pb	Cu	Zn	Ni
A 组堆肥前	1. 290 ± 0. 007	46. 73 ± 4. 53	$74.92 \pm 4.00$	469. 67 ± 7. 73	45. 41 ± 2. 40
A 组堆肥后	$1.270 \pm 0.021$	$48.07 \pm 1.85$	73. $58 \pm 6.41$	474. 42 ± 11. 31	$36.79 \pm 1.56$
B组堆肥前	$1.314 \pm 0.039$	44. $19 \pm 3.62$	76. $81 \pm 2.96$	$483.64 \pm 8.79$	43. 01 $\pm$ 2. 07
B组堆肥后	$1.322 \pm 0.021$	44. 11 ± 5. 11	$78.66 \pm 2.96$	$479.59 \pm 12.93$	$35.95 \pm 1.13$

一定变化(见表 4), 其中 A 组处理重金属有效态含量下降显著,而 B 组处理变化不显著.根据堆肥前后不同重金属有效态含量占总量的质量分数来计算2 种堆肥处理过程重金属的钝化效果(见表 4),添加生物炭(A 组处理)对污泥堆肥过程 5 种重金属有效态均有较好的钝化效果,Cd、Pb、Cu、Zn、Ni的钝化率分别为:23.54%、16.39%、27.16%、43.10%、37.82%,其中,Zn、Ni钝化效果尤为显著.与 A 组处理相比,未添加生物炭的污泥堆肥(B组处理)对重金属钝化效果不显著,其中 Cd、Pb、Zn 钝化率仅为 1.65% ~8.12%,Cu、Ni 有效态所

占质量分数反而增加了 3.76%、1.17%.

这主要是 A 组处理堆肥中添加了水稻生物炭,其具有较大的比表面积、较强的阳离子交换能力等性质,与污泥混合后能够增强透气性,为堆肥中微生物提供适宜生存繁殖的环境<sup>[20]</sup>,也可通过吸附作用降低重金属的溶解,有效地降低重金属在环境中的迁移<sup>[21]</sup>.同时生物炭呈碱性,含有碱性物质,其有机质含量高,大量添加使用后有利于提高 pH 值<sup>[22,23]</sup>,使重金属的残渣态含量显著升高,生物可利用态下降,因而 A 组堆肥重金属钝化效果较好.

表 4 堆肥前后重金属有效态含量及钝化率

Table 4 Available contents and passivating effect of heavy metals in sludge composting

	Tub	ic + Available contents and passive	ating effect of fically if	ictais in staage comp	osting	to the second
- 重金属	取样时期	A组	(0	1 00	B组	431
里並周	以作的别	含量/mg·kg-1 分配率/%	钝化率/%	含量/mg·kg-1	分配率/%	钝化率/%
S Cd	堆肥前	$0.182 \pm 0.007$ 14.11	- /	0. 167 ± 0. 004	12. 71	
1 u Z	堆肥后	$0.137 \pm 0.012$ $10.79$	23. 54	$0.158 \pm 0.008$	11. 95	5. 96
Pb	堆肥前	5. 29 ± 0. 54 11. 32	_( @	$4.93 \pm 0.36$	11. 16	A
C 1/1 V 0	堆肥后	4. 55 ± 0. 48 9. 47	16. 39	$4.84 \pm 0.21$	10. 97	1. 65
VA VIII	堆肥前	26. 46 ± 1. 10 35. 32	P)	20. 45 ± 0. 50	26. 62	_
la Par	堆肥后	$18.93 \pm 0.82$ $25.73$	27. 16	$21.73 \pm 1.66$	27. 63	-3.76
7.	堆肥前	87. 78 ± 4. 27 18. 69	_	$63.32 \pm 2.36$	13. 09	_
Zn	堆肥后	$50.45 \pm 1.89$ $10.63$	43. 10	$57.62 \pm 3.37$	12. 01	8. 23
Ni	堆肥前	$5.60 \pm 0.14$ 12.33	_	$4.21 \pm 0.31$	9. 79	_
	堆肥后	$2.82 \pm 0.08$ 7.67	37. 84	$3.56 \pm 0.16$	9. 90	-1.17

## 2.4 生物炭对施用堆肥后土壤重金属总量的影响

施用污泥堆肥后,土壤重金属总量变化见表5. 各处理重金属含量均高于 CK 组处理,可见施加污泥堆肥后,污泥中的重金属元素也随之进入土壤,导致土壤重金属含量增加. 随着施肥量的增加,土壤重金属总量有一定变化,但重金属含量与污泥堆肥施用量没有显著的相关关系(P>0.05),这可能是由于施肥时仅将堆肥产品均匀撒在土壤表面(0~5 cm),而下层土壤(>5 cm)中没有混入堆肥,导致土壤取样混合分析时重金属变化不显著. 施加污泥堆肥后,土壤中5种重金属除 Cd 外,其余均未超过国家土壤环境质量(GB 15618-1995)二级标准,试验组中 Cd 含量超过了土壤环境质量二级标准(pH<7.5 时,二级标准为 0.3 mg·kg<sup>-1</sup>; pH>7.5 时,二级标准为 0.6 mg·kg<sup>-1</sup>),这主要是由于供试土壤 Cd 含量较高(0.64 mg·kg<sup>-1</sup>),且原污泥

中 Cd 含量也较高(1.87 mg·kg<sup>-1</sup>),导致堆肥产品 Cd 含量依然较高,施入土壤后可能会对土壤造成 二次 Cd 污染.

本试验结束后,大部分处理土壤重金属含量均 呈降低趋势(降低量<14%),一方面可能是由于试 验田未隔绝水分下渗,在定期浇水过程中,土壤中 的重金属随着水分下渗向地下迁移;另一方面可能 是由于植物从土壤中富集了部分重金属,但是由于 作物生长周期较短,因而短期内富集土壤中的重金 属较少.

### 2.5 生物炭对施用堆肥后土壤重金属有效态的 影响

有研究表明,土壤中重金属的迁移性和生物毒性主要取决于重金属的形态分布,而不仅仅是其总量<sup>[24]</sup>.土地利用前后,土壤中各重金属有效态含量如表6和表7所示,随着施肥量的增加,土壤重金

表 5 不同处理组土壤重金属总量变化/mg·kg-1

Table 5	Heavy metal	contents in	coile with	different	treatments/mg·kg	<sub>v</sub> – 1
rable o	пеауу шегаг	contents in	SOHS WITH	amerem	Treatments/ mg • Kg	2

重金属	取样时期	CK 组	A1 组	A2 组	A3 组	B1 组	B2 组	B3 组
Cd	利用初期	0. $64 \pm 0.10$	0. 83 ± 0. 09	$0.89 \pm 0.74$	0. 87 ± 0. 08	0. $88 \pm 0.12$	$0.93 \pm 0.13$	0. 94 ± 0. 07
	利用结束	0. $66 \pm 0.06$	0. 78 ± 0. 06	$0.86 \pm 0.11$	0. 83 ± 0. 11	0. $76 \pm 0.04$	$0.95 \pm 0.10$	0. 92 ± 0. 06
Pb	利用初期 利用结束	19. $38 \pm 1.68$ 19. $80 \pm 2.67$	$24.35 \pm 0.92$ $23.49 \pm 2.16$	$26.61 \pm 1.31$ $24.50 \pm 0.80$	$27.48 \pm 2.13$ $23.90 \pm 1.53$	$22.39 \pm 2.35$ $20.63 \pm 1.54$	21. 94 ± 4. 60 20. 86 ± 2. 83	$24.36 \pm 3.04$ $22.26 \pm 0.70$
Cu	利用初期	$38.58 \pm 3.69$	42. 30 ± 4. 26	45. 86 ± 3. 00	44. 53 ± 2. 94	$43.77 \pm 2.88$	47. 36 ± 3. 02	48. 94 ± 2. 98
	利用结束	$38.20 \pm 2.03$	40. 84 ± 1. 43	42. 47 ± 1. 47	42. 47 ± 3. 01	$40.36 \pm 3.98$	44. 43 ± 2. 36	47. 83 ± 2. 87
Zn	利用初期	$58.50 \pm 3.14$	73. $76 \pm 2.38$	80. 69 ± 4. 97	$88.73 \pm 6.55$	$80.76 \pm 2.80$	90. 58 ± 5. 33	95. 56 ± 4. 49
	利用结束	$57.42 \pm 3.06$	74. $14 \pm 2.03$	83. 93 ± 4. 24	$80.71 \pm 1.33$	$75.30 \pm 5.40$	85. 11 ± 3. 74	82. 45 ± 6. 24
Ni	利用初期	$24.90 \pm 1.48$	$26.76 \pm 0.97$	$26.58 \pm 2.04$	27. 81 ± 1. 90	25. 79 ± 1. 65	26. 38 ± 2. 37	26. 14 ± 0. 76
	利用结束	$25.49 \pm 1.78$	$25.42 \pm 1.41$	$26.01 \pm 2.24$	26. 95 ± 1. 22	26. 94 ± 1. 04	25. 34 ± 2. 07	25. 81 ± 3. 08

#### 表 6 A 组处理土壤重金属有效态含量及分配率

Table 6 Available contents of heavy metals in soils from Group A

重金属	取样时期 -	A1 组	[	A2 组		A3 组	
里並馮	<b>以</b> 件的别	含量/mg·kg-1	分配率/%	含量/mg·kg-1	分配率/%	含量/mg·kg <sup>-1</sup>	分配率/%
Cd	利用初期	$0.082 \pm 0.004$	9. 88	$0.076 \pm 0.011$	8. 54	$0.088 \pm 0.006$	10.11
Cu	利用结束	$0.076 \pm 0.010$	9. 74	$0.070 \pm 0.007$	8. 22	$0.085 \pm 0.004$	10. 24
Pb	利用初期	$1.27 \pm 0.17$	5. 22	1. 43 ± 0. 29	5. 37	1.49 ± 0.16	5.42
1.0	利用结束	1. 17 $\pm$ 0. 31	4. 98	1. 28 ± 0. 31	5. 22	$1.30 \pm 0.27$	5. 44
Cu	利用初期	11. 63 $\pm$ 0, 30	27. 49	12. 42 ± 0. 21	27. 08	11. 98 ± 0. 48	26. 90
	利用结束	$10.66 \pm 0.92$	26. 10	10. 81 ± 0. 55	25. 45	10. 85 $\pm$ 0. 53	25. 55
Zn	利用初期	6. 98 ± 0. 17	9. 46	6. 58 ± 0. 08	8. 15	8. 13 ± 0. 31	9. 16
Zii	利用结束	$6.68 \pm 0.36$	9.01	6. 74 ± 0. 09	8. 03	7. $13 \pm 0.27$	8. 83
) <sub>Ni</sub>	利用初期	1. 33 ± 0. 05	4. 97	1. 34 ± 0. 04	5. 04	1. 36 ± 0. 10	4. 89
- 1/1	利用结束	1. 25 ± 0. 14	4. 92	1. 27 ± 0. 11	4. 88	1. 24 ± 0. 08	4. 60

表 7 B 组处理土壤重金属有效态含量及分配率

Table 7 Available contents of heavy metals in soils from Group B

<b>手</b> △目	114 Jt. V-J. +-11	B1 组	1	B2 组	1	B3 🕏	<u>E</u>
重金属	取样时期	含量/mg·kg <sup>-1</sup>	分配率/%	含量/mg·kg-1	分配率/%	含量/mg·kg-1	分配率/%
Cd	利用初期	$0.092 \pm 0.013$	10. 45	0. 087 ± 0. 006	9. 35	0. 092 ± 0. 015	9. 79
Cu	利用结束	$0.080 \pm 0.005$	10. 53	$0.091 \pm 0.012$	9. 58	$0.087 \pm 0.008$	9.46
Pb	利用初期	$1.36 \pm 0.23$	6. 07	$1.54 \pm 0.10$	7. 02	$1.53 \pm 0.20$	6. 28
1 D	利用结束	$1.23 \pm 0.11$	5. 96	1. 47 $\pm$ 0. 23	7. 05	1. $46 \pm 0.08$	6. 56
Cu	利用初期	13.89 ± 1.11	31.73	15. $25 \pm 0.71$	32. 20	$14.70 \pm 1.04$	30. 04
Cu	利用结束	12. $86 \pm 0.81$	31.86	14. 15 $\pm$ 0. 34	31. 85	13. $89 \pm 0.52$	29. 04
Zn	利用初期	$8.24 \pm 0.69$	10. 20	$9.00 \pm 0.50$	9. 94	10. 03 $\pm$ 0. 54	10. 50
Zil	利用结束	$7.62 \pm 0.61$	10. 12	$8.45 \pm 0.31$	9. 93	$8.90 \pm 0.20$	10. 79
Ni	利用初期	$1.25 \pm 0.06$	4. 85	$1.24 \pm 0.04$	4. 70	$1.33 \pm 0.10$	5. 09
111	利用结束	$1.28 \pm 0.07$	4. 75	$1.21 \pm 0.09$	4. 78	$1.30 \pm 0.07$	5. 04

属有效态含量有一定变化,但统计发现,污泥堆肥施用量与土壤重金属有效态含量也没有显著的相关关系(P>0.05).同时,各处理重金属有效态含量在试验前后没有显著变化.总体来看,在污泥堆肥利用初期与利用试验结束后,各处理土壤中5种重金属有效态含量占其总量的质量分数相差很小,也没有显著变化.生物炭对污泥堆肥土壤利用后的重金属有效态没有显著影响,表明污泥堆肥与土壤混

合后的重金属未被钝化.

有研究表明,在土壤中直接添加生物炭会钝化土壤中的重金属有效态.如刘晶晶等<sup>[25]</sup>在土壤中直接施加 5%的稻草炭,能有效降低土壤中 Cd、Cu、Pb和 Zn的有效态含量,几种重金属有效态降幅达 34.5%~52.5%. 王红等<sup>[26]</sup>发现水葫芦炭对土壤中 Zn和 Pb的吸附率分别达到 21.83%和44.57%.生物炭施入土壤可通过物理吸附、离子

交换、沉淀络合等交互作用机制<sup>[27,28]</sup>,显著降低土壤重金属的有效态,减少它们在环境介质中的迁移性,从而降低土壤重金属的生物毒性. Méndez等<sup>[29]</sup>的研究也发现生物炭能降低土壤中重金属含量,主要是由于有机质对土壤重金属的矿化作用,从而降低重金属的生物有效性.

而在本试验中,污泥堆肥处理阶段生物炭对重金属的钝化效果较好,而污泥堆肥后再进行短期土壤利用的过程重金属难以继续被钝化.这可能是由于在污泥堆肥过程中,添加的生物炭已经和污泥中的重金属发生物理吸附、离子交换等结合作用,钝化了污泥中的重金属,因而将污泥堆肥施入土壤后可利用的生物炭量减少;同时,由于土壤利用时间和试验周期较短,导致在土壤利用过程中生物炭难以再继续钝化重金属.因此,需要进一步研究其长期效应.

#### 3 结论

水稻生物炭对污泥堆肥及其利用过程中 5 种重金属(Cd、Pb、Cu、Zn、Ni)总量没有显著影响,但添加生物炭能显著降低堆肥过程重金属有效态含量,对污泥堆肥过程中的重金属具有较好的钝化效果,且对 Zn、Ni 的钝化作用尤为显著. 生物炭对后续土壤利用后的重金属有效态具有一定影响,但效果不显著. 总体来说,生物炭对城市污泥堆肥处理过程中的重金属具有较好的钝化效果,但在短期土壤利用过程难以继续钝化重金属.

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