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顺德水道土壤及沉积物中重金属分布及潜在生态风险 评价

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摘要:水源地周边土壤及河道沉积物的环境质量状况极大程度影响着河流饮用水安全.为调查顺德水道水源地重金属空间分布特征及其污染来源,本研究采集了顺德水道周边表层土壤及其主要支流入河口沉积物,并测定各样品中 Cd、Zn、Pb、Cu、Ni、Cr等6种重金属浓度,最后基于两种潜在生态风险评价方法对其生态风险进行评价.结果发现,顺德水道表层土壤中 Zn、Cr、Pb、Cu、Ni和 Cd 平均含量分别为 186.80、65.88、54.56、32.47、22.65 和 0.86 mg·kg⁻¹,除 Cu、Ni 外其它重金属均超过顺德土壤背景值;8个主要支流入河口间表层沉积物中6种重金属元素平均含量依次为:Zn(312.11 mg·kg⁻¹)>Cr(111.41 mg·kg⁻¹)>Pb(97.87 mg·kg⁻¹)>Cu(92.32 mg·kg⁻¹)>Ni(29.89 mg·kg⁻¹)>Cd(1.72 mg·kg⁻¹),除 Ni 之外其余均高于顺德土壤背景值.主成分分析结果发现表层土壤中 Cr、Ni 含量主要受自然母质影响,Zn、Pb、Cu和 Cd主要来源于该地区制造业的废水排放;沉积物中6种重金属均来源于外源输入,受顺德水道周边的工业活动影响.基于环境生物可利用态的潜在生态风险评价结果发现顺德水道周边表层土壤中 Cd 呈现轻微的潜在生态危害,而入河口沉积物中 Cd 呈现中度的潜在生态危害,土壤和沉积物中 Zn、Pb、Cu和 Ni 的潜在生态危害程度均表现为轻微.由于基于环境生物可利用态的潜在生态风险评价充分考虑了土壤理化性质及重金属形态,其结果低于 Hakanson 潜在生态风险评价结果,可避免对重金属的潜在危害程度的高估.

关键词:土壤; 沉积物; 重金属; 环境生物可利用态; 潜在生态风险评价; 顺德水道

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Spatial Distribution and Potential Ecological Risk Assessment of Heavy Metals in Soils and Sediments in Shunde Waterway, Southern China

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Abstract: Environmental quality of soils and sediments around water source area can influence the safety of potable water of rivers. In order to study the pollution characteristics, the sources and ecological risks of heavy metals Zn, Cr, Pb, Cu, Ni and Cd in water source area, surface soils around the waterway and sediments in the estuary of main tributaries were collected in Shunde, and ecological risks of heavy metals were assessed by two methods of potential ecological risk assessment. The mean contents of Zn, Cr, Pb, Cu, Ni and Cd in the surface soils were 186. 80, 65. 88, 54. 56, 32. 47, 22. 65 and 0. 86 mg·kg⁻¹ respectively, and they were higher than their soil background values except those of Cu and Ni. The mean concentrations of Zn, Cr, Pb, Cu, Ni and Cd in the sediments were 312. 11, 111. 41, 97. 87, 92. 32, 29. 89 and 1. 72 mg·kg⁻¹ respectively, and they were higher than their soil background values except that of Ni. The results of principal component analysis illustrated that the main source of Cr and Ni in soils was soil parent materials, and Zn, Pb, Cu and Cd in soils mainly came from wastewater discharge of local manufacturing industry. The six heavy metals in sediments mainly originated from industry emissions around the Shunde waterway. The results of potential ecological risk assessment integrating environmental bioavailability of heavy metals showed that Zn, Cu, Pb and Ni had a slight potential ecological risk. Cd had a slight potential ecological risk in surface soils, but a moderate potential ecological risk in surfaces sediments. Because the potential ecological risk sersellated of risks were lower than those of Hakanson methods, and it could avoid overestimating the potential risks of heavy metals.

Key words: soil; sediment; heavy metals; environmental bioavailability; potential ecological risk assessment; Shunde waterway

河道沿岸土壤是水陆生态系统的过渡与缓冲区域,可通过物理、化学和生物作用截留多种污染物质,但持久性污染物如重金属则将长期滞留其中^[1,2]. 水体沉积物作为水环境中重金属的主要蓄

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积介质,可以反映水体长期受重金属污染的状况.河道沿岸土壤及沉积物中重金属污染不仅对河道周边及其水体动植物产生危害,同样地,可在一定条件下进入水体导致"二次污染"^[3~6].而水源地周边土壤及河道沉积物的环境质量状况极大程度影响着河流饮用水安全.近年来,国内外学者对水环境重金属污染开展研究众多,但重点多为水体或河流沉积物中重金属的含量特征^[3,7,8]、形态赋存^[9~11]、来源分析^[12,13]及污染评价^[14]等,而对以河道沿岸土壤及河道沉积物相结合的重金属研究相对较少.

潜在生态风险评价法常用于评价水体、土壤、沉积物、街尘等介质中重金属对环境的风险.例如尚林源等[15]采用 Hakanson 潜在生态风险指数法对密云水库入库河流沉积物中的重金属的潜在生态风险进行评价,确定不同人库河流河段的风险等级,为水源地入库河流的污染物控制提供依据; Hu等[16]通过潜在生态风险评价法对珠江三角洲地区不同土地类型下的土壤风险进行排序.近年来,有学者提出应以重金属的环境生物可利用态代替重金属总量,或基于重金属不同形态来评估其潜在生态风险[17,18].这类方法不仅评估污染物总量,同时考虑到不同区域中土壤性质对重金属迁移和生物可利用性的影响,其评价结果也更为合理.

顺德水道位于珠江三角洲平原中部,是顺德区的重要饮用水源地,河道周边分布一定量的工业企业和农业用地,其生产活动对顺德水道水质、周边土壤及沉积物环境构成潜在威胁.本研究以顺德水道周边表层土壤及8个主要河道支流入河口沉积物为对象,通过采样分析 Cd、Zn、Pb、Cu、Ni 和 Cr 这6种重金属含量,揭示两种介质中重金属污染的空间分布特征,分析土壤及沉积物中重金属污染物来源,对环境生物可利用态的重金属进行潜在生态风险评价,并与传统的 Hakanson 潜在生态风险评价结果进行对比,以期为河道周边土壤及沉积物重金属污染生态风险研究提供科学依据.

1 材料与方法

1.1 采样点设置

顺德水道位于佛山市顺德区,是佛山市和广州市的重要饮用水源地.河道周边分布一定量的工业企业和农业用地,其生产活动对顺德水道水质、周边土壤及沉积物环境构成威胁.为研究顺德水道土壤和沉积物中重金属分布特征及其风险大小,沿河道东西走向共设置51个0~20 cm 的表层土壤采样

点,每一个土壤样品均由样点周围 100 m² 范围内采集的 5 个土样混合而成. 同时采集顺德水道 8 个主要支流入河口沉积物,自上游至下游依次为 A、B、C、D、E、F、G、H,如图 1.

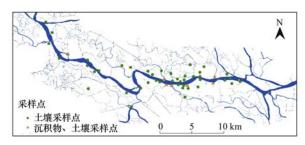


图1 采样点分布示意

Fig. 1 Distribution of sampling sites

1.2 样品分析与测定

土壤、沉积物室内风干后磨细过 100 目尼龙筛,重金属参照 USEPA 3050 建议的消煮方法改进,以 HCl-HNO₃-HF-HClO₄ 湿法消解^[17], Cu、Pb、Zn、Cr、Ni 采用 ICP-OES 测定, Cd 采用石墨炉原子吸收测定.分析过程中用土壤标样 GSS-16 进行质量控制,同时进行空白和重复实验(重复率为 10%).实验中所用试剂均为优级纯,溶液配制用水均为超纯水,所有玻璃器皿均在 10% 硝酸溶液中浸泡过夜后洗净使用.各元素回收率在 93.58% ~110.98% 之间,符合质量控制标准.

土壤、沉积物理化性质采用常规方法测定:pH测定采用电位法^[19](水土比为 5:1, DENVER UB-7 pH 计测定);有机碳采用 Vario ELⅢ元素分析仪测定(取 0.5 g 土壤,由 30 mL 1 mol·L⁻¹盐酸溶液振荡 24 h,去除无机碳后测定).

1.3 数据处理

常规数理统计分析采用 SPSS 12. 0 软件完成, 主成分分析及相关性分析采用 R 3.2.0 软件完成, 空间分布图通过 ArcGIS 9.3 及 Sigmaplot 12.0 软件 制作.

1.4 评价方法

本研究分别采用 Hakanson 潜在生态风险指数 法与基于环境生物可利用态的分层潜在生态风险评价法对研究区内土壤、沉积物进行重金属的总量及 环境生物可利用态进行风险评价,比较两种评价方法的风险程度.

1.4.1 Hakanson 潜在生态风险指数法

采用 Hakanson 潜在生态风险指数法^[20]对重金属污染进行生态风险评价,该方法以土壤/沉积物中重金属的元素背景值为基准,结合重金属的生物毒

性(毒性响应因子)、环境效应(污染指数)计算其 潜在生态风险系数,其生态风险等级判定依据见表 1,评价公式如下:

$$RI = \sum_{i}^{n} E_{r}^{i} = \sum_{i}^{n} (T_{r}^{i} \times C_{f}^{i}) = \sum_{i}^{n} (T_{r}^{i} \times C_{D}^{i} / C_{R}^{i})$$

式中, C_i 为重金属元素 i 的污染指数; C_D 为土壤中重金属元素 i 的实测含量; C_R 为参照值; T_i 为重金属元素 i 的毒性响应因子; E_i 为重金属元素 i 的潜在生态风险系数; RI 为综合潜在生态风险指数.

1.4.2 基于环境生物可利用态的分层潜在生态风险评价法

大量研究表明土壤中重金属的形态与农作物对重金属的吸收量存在良好的相关性,其中水溶态重金属可直接被植物利用,被称为环境生物可利用态.与 Hakanson 潜在生态风险评价指数法相比,基于环境生物可利用态的分层潜在生态风险评价法引入"环境生物毒性响应系数(T_b)×毒性响应因子(T_c)"代替"传统的毒性响应因子(T_c)",对土壤及沉积物的生态风险评价采用分层方法,以重金属的环境生物可利用性进行潜在生态风险评价[17.18,21],定量计算重金属的潜在风险指数,并根据表 1 判定

其生态风险等级,其评价公式如下[17,18]:

$$RI = \sum_{i}^{n} E_{r}^{i} = \sum_{i}^{n} (T_{b}^{i} T_{b}^{i} \times T_{e}^{i} \times C_{f}^{i})$$
$$= \sum_{i}^{n} \left(\sqrt{\frac{R_{b}^{i}}{P_{b}^{i}}} \times T_{e}^{i} \times \frac{C_{D}^{i}}{C_{p}^{i}} \right)$$

式中, T_b 为土壤重金属元素 i 的环境生物毒性响应系数; R_b 为样品中重金属元素 i 水溶态占总含量的质量分数,其中重金属元素 i 的水溶态含量参照文献[22],5 种重金属元素的水溶态预测模型如表 2 所示; P_b 为参照土壤的相应比值; T_e 为重金属元素 i 的毒性响应因子;其余系数与 Hakanson 潜在生态风险指数法相同.

其中以广东省顺德地区背景值^[23] 为参照值, T_r 、 T_e 分别参照文献[18, 24],依次为:Cd(30,15)、Zn(1,1)、Pb(5,4)、Cu(5,2)、Ni(5,3)、Cr(2,11). 研究区内 pH 范围为 6. 5 ~ 7. 5,均值 7. 0,有机碳质量分数范围为 1. 03% ~ 1. 93%,均值 1. 36%。本研究中 Cr 含量^[18] 低于国家二级标准(200 $mg \cdot kg^{-1}$,pH: 6. 5 ~ 7. 5),且缺少其水溶态预测模型,因此仅考虑 Cd、Zn、Pb、Cu、Ni 这 5 种重金属元素的潜在生态风险评价.

表 1 潜在生态风险指数等级划分表

Table 1 Classification of potential ecological risk index

The state of the s								
单一重金属潜在生	生态风险系数($E_{ m r}^i$)	综合潜在生态风险指数(RI)						
阈值区间	潜在生态危害程度	阈值区间	潜在生态危害程度					
$E_{\rm r}^i < 40$	轻微	RI < 150	轻微					
$40 \le E_{\rm r}^i < 80$	中度	$150 \leq RI < 300$	中度					
$80 \le E_{\rm r}^i < 160$	强度	$300 \le RI < 600$	强					
$160 \le E_{\rm r}^i < 320$	很强	RI≥600	很强					
$E_{\rm r}^i \geqslant 320$	极强							

表 2 5 种重金属元素的水溶态含量预测模型[22]

Table 2 Predicting models for dissolved concentrations of Cd, Zn, Pb, Cu and Ni

元素	回归曲线1)	R^2
Cd	$\log_{10}(DM) = (-0.47 \pm 0.02) pH + (1.08 \pm 0.02) \log_{10}(TM) - (0.81 \pm 0.05) \log_{10}(SOM) + (3.42 \pm 0.11)$	0. 884 * *2)
Zn	$\log_{10}\left(\mathrm{DM}\right) = (-0.55\pm0.04)\mathrm{pH} + (0.94\pm0.08)\log_{10}\left(\mathrm{TM}\right) - (0.34\pm0.12)\log_{10}\left(\mathrm{SOM}\right) + (3.68\pm0.31)$	0. 618 * *
Pb	$\log_{10} (\mathrm{DM}) = (-0.37\pm 0.04)\mathrm{pH} + (0.56\pm 0.07)\log_{10} (\mathrm{TM}) + (1.81\pm 0.22)$	0. 347 * *
Cu	$\log_{10}\left(\mathrm{DM}\right) = (-0.21\pm0.02)\mathrm{pH} +(0.93\pm0.05)\log_{10}\left(\mathrm{TM}\right) -(0.21\pm0.02)\log_{10}\left(\mathrm{SOM}\right) +(1.37\pm0.14)$	0. 611 * *
Ni	$\log_{10}\left(\mathrm{DM}\right) = (-1.05\pm0.09)\mathrm{pH} + (1.21\pm0.22)\log_{10}\left(\mathrm{TM}\right) - (0.85\pm0.21)\log_{10}\left(\mathrm{SOM}\right) + (7.02\pm0.62)$	0. 727 * *

1) DM(水溶态重金属, $mg \cdot L^{-1}$); TM(重金属总量, $mg \cdot kg^{-1}$); SOM(有机质,以 C% 计); 2) * *表示 P < 0.01

2 结果与讨论

2.1 顺德水道周边表层土壤重金属含量及空间分布顺德水道周边表层土壤中重金属元素含量的范围及平均值见表 3,各重金属平均含量大小依次为: Zn(186.8 mg·kg⁻¹) > Cr(65.88 mg·kg⁻¹) > Pb

 $(54.56 \text{ mg·kg}^{-1}) > \text{Cu}(32.47 \text{ mg·kg}^{-1}) > \text{Ni}$ $(22.65 \text{ mg·kg}^{-1}) > \text{Cd}(0.86 \text{ mg·kg}^{-1}).$ 与史瑾 瑾^[25]在顺德乐从的研究对比,发现 Cd 和 Zn 元素含量较为接近,Pb、Cu 和 Ni 元素含量略低于其研究结果;与刘子宁等^[26]在顺德区的研究结果相比,Zn、Cu、Pb和Cr元素含量近似,而Cd元素含量高

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王 2	顺德水道表层土壤及表层沉积物重金属元素含量

Table 3	Heavy metal	contents in	surface	soils and	sediments	of Shunde watery	vav

类型	统计值	Cd	Zn	Pb	Cu	Ni	Cr	文献
	最小值/mg·kg ⁻¹	0.12	59.11	11.91	5.06	2.99	42.78	本研究
表层土壤	均值/mg·kg ⁻¹	0.86	186.8	54.56	32.47	22.65	65.88	
14/4/二%	最大值/mg·kg ⁻¹	4.01	1408.3	324.75	99.6	42.78	100.86	
	变异系数/%	98.7	116.9	94.4	66.3	36.3	17.6	
	最小值/mg·kg ⁻¹	0.65	136.82	29.80	40.22	17.62	63.76	本研究
沉积物	均值/mg·kg-1	1.72	312.11	92.32	97.87	29.89	111.41	
06/2/193	最大值/mg·kg ⁻¹	2.18	481.84	163.70	236.03	39.64	321.61	
	变异系数/%	0.48	117.72	46.50	62.73	7.10	30.59	
表层土壤	顺德乐从	0.72	179.7	73.52	51.34	32.56	_	[25]
14/4/二%	顺德区	0.48	178.91	48.27	44.20	_	68.89	[26]
沉积物	佛山水道	3.76	503.61	236.09	192.31	_	317.04	[27]
二级标准/m	ng•kg ⁻¹	0.40	250	50	100	80	200	GB 15618-2008
顺德背景值	/mg•kg ⁻¹	0.44	109.6	50.1	43.75	32.6	62	[23]

于其结果,这可能因为本研究采样点主要集中在河道周边.

在空间分布上(图2),顺德水道周边表层土壤

中 Zn 和 Pb 的含量空间分布相似,均表现为上游含量较高,沿河道水流方向由西向东逐渐降低; Cd 在全研究区域的含量分布较为均匀,且 54.9%

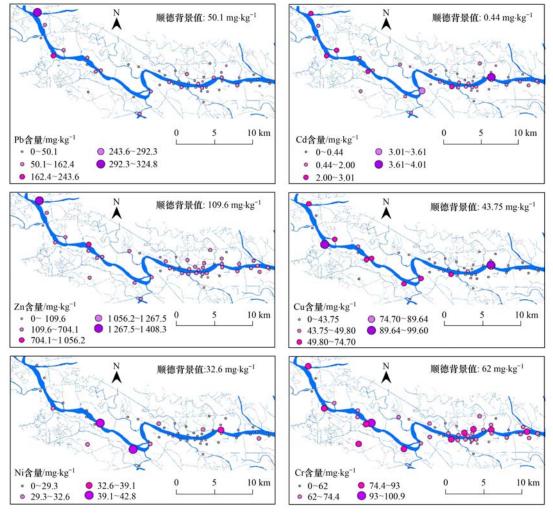


图 2 表层土壤各元素空间分布

Fig. 2 Spatial distributions of heavy metals in surface soils along Shunde waterway

以上的采样点 Cd 含量高于背景值含量; Cu 和 Ni 的空间分布较为相似,高值均出现在顺德水道"S型"回旋处, Cr 则在全区域分布均匀, Ni、Cr 这 2种元素的分布与该区域深层土壤(150~200 cm)中 Ni 和 Cr 含量的分布相似^[23],说明表层土壤 Ni 和 Cr 这 2 种元素受人为干扰小,主要反映土壤母质的含量特征.

2.2 表层沉积物重金属含量及空间分布

支流入河口是连接河涌与主河道的关卡,污染物可随河涌水体进入河道,并在沉积物中累积,因此河道沉积物中重金属含量可以反映较长时期内随水体进入河道重金属的含量高低^[28].顺德水道8个

主要支流入河口表层沉积物中重金属含量如图 3 所示. 数据表明各入河口间表层沉积物中重金属平均含量从大到小依次为: $Zn(312.11 \text{ mg} \cdot \text{kg}^{-1}) > Cr(111.41 \text{ mg} \cdot \text{kg}^{-1}) > Cu(97.87 \text{ mg} \cdot \text{kg}^{-1}) > Pb(92.32 \text{ mg} \cdot \text{kg}^{-1}) > Ni(29.89 \text{ mg} \cdot \text{kg}^{-1}) > Cd(1.72 \text{ mg} \cdot \text{kg}^{-1}). 与利锋等[27]2006 年在佛山水道沉积物的研究对比发现, 本研究区内表层沉积物的 Cd、Zn、Pb、Cu 和 Cr 这 5 种元素均低于佛山水道沉积物的平均含量,这说明水源地沉积物重金属累积程度低于普通城市水网河道. 但与顺德区土壤背景值对比(表 3),沉积物中重金属元素均存在不同程度的累积.$

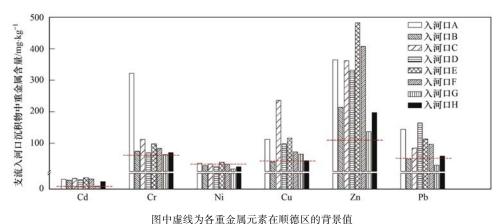


图 3 顺德水道主要支流入河口表层沉积物重金属含量分布特征

Fig. 3 Distribution of heavy metals in sediments collected from the main river gates of Shunde waterway

从图 3 可以看出:各支流入河口沉积物中 6 种 重金属含量差异明显. 位于上游的入河口 A 沉积物 中6种重金属含量均超过8个人河口沉积物平均含 量及顺德区土壤背景值,且 Cr 元素在 A 入河口沉 积物中累积程度最高,约为背景值的5.19倍;处于 中游的入河口 B、C、D 及 E 沉积物中 Cd、Cr 和 Ni 元素含量差异较小且与各支流入河口平均含量持 平,而 Cu、Zn 和 Pb 含量差异明显且超过整体平均 水平,其最大值分别为背景值的 5.39、2.85 及 3.27 倍;位于下游的入河口F、G、H沉积物中各重金属 含量均低于或持平于上中游入河口中沉积物含量. Cd、Ni 在各支流入河口沉积物中含量差异较小,而 Cr、Cu、Zn、Pb 均表现为上中游含量较高而下游含 量较低. 有学者在容桂工业区河涌沉积物的研究结 果表明重金属污染物含量总体上下游小于上游[23]. 本研究结果与其相似. 研究区中下游是广州与佛山 市集中式引用水取水范围,受到水源地政策保护,其 重金属污染源较少. 本研究结果表明顺德水道上游 可能存在重金属排放源.

2.3 土壤、沉积物重金属来源分析

表层土壤各重金属相关性分析结果表明:Zn与 Pb、Cu、Cr、Pb与Cu、Cr、Ni, Cu与Ni、Cr及Ni 与 Cr 呈正相关系(P<0.05), Cd 与其他 5 种元素均 呈极显著性正相关. 为减少变量维度,对顺德河道 表层土壤、主要支流入河口表层沉积物中6种重金 属进行主成分分析,结果如图 4 所示. 顺德河道周 边表层土壤6种重金属含量数据方差被分为了2个 因子,可解释总方差的85%.第一主成分因子中负 荷较大的为 Cd、Cu、Zn 和 Pb,解释了总方差的 44%,这4种重金属代表外源污染;第二主成分因 子为 Cr 和 Ni,这两种重金属累积程度较小,该主成 分代表成土母质. 顺德表层土壤重金属含量的空间 分布表明,土壤中 Zn、Pb 呈现上游至下游逐渐降低 趋势,Cd 在全研究区内均存在不同程度的累积. 河 道上游为龙江和乐从地区,是较为发达的制造业区, 其家具、塑料建材、小家电及纺织服装产业均为顺 德区支柱性产业. 这些制造工业污染源可能排放含 有重金属的污水进入河涌,经地表水运移传输而污

染周边土壤.同时,顺德道上游与北江干流相接,北江流域矿产资源丰富,有色冶金企业较多,其矿渣、尾矿及污废水易随雨水冲刷至下游,并在沿岸土壤沉积. Cd、Cu、Zn 和 Pb 作为典型的铅锌矿的组合

元素出现伴生累积,也证实顺德水道土壤重金属元素可能来自上游的矿区^[29]. 顺德水道河道宽敞,水网密布,故其航运较为发达,因此土壤中 Cd 和 Zn 也可能来源于水上运输及船只表面处理行业.

表 4 顺德水道周边表层土壤、支流入河口表层沉积物重金属相关系数矩阵1)

Table 4 Matrix of Pearson's correlation coefficients of heavy metals in surface soils and sediments samples collected from Shunde waterway

类型	重金属	Cd	Zn	Pb	Cu	Ni	Cr
	Cd	1	0. 508 * *	0.710 * *	0. 864 * *	0. 487 * *	0. 435 * *
	Zn		1	0. 863 * *	0. 515 * *	0. 179	0. 297 *
表层土壤	Pb			1	0. 715 * *	0. 307 *	0. 410 * *
47公上%	Cu				1	0. 694 * *	0. 680 * *
	Ni					1	0. 765 * *
	Cr						1
	Cd	1	0. 890 * *	0. 769 * *	0. 640 *	0. 602 *	0. 333
	Zn		1	0. 838 * *	0. 692 * *	0. 797 * *	0. 408
沉积物	Pb			1	0. 553 *	0. 612 *	0. 516
O LIVING	Cu				1	0. 570 *	0. 375
	Ni					1	0. 514
	Cr						1

1) * *表示在置信度(双测) 为 0.01 时,相关性是显著的; *表示在置信度(双测) 为 0.05 时,相关性是显著的

从表 4 可以看出顺德水道 8 个主要支流入河口沉积物中 Cd 与 Zn、Pb、Cu、Ni 呈正相关(P < 0.05). 这说明沉积物中 Cd、Zn、Pb、Cu 和 Ni 的来源具有一定共性,可能是由村级工业园区排放废水,经由顺德水道周边河涌进入主河道,与流经矿区、携带大量重金属的上游河水混合并沉积在河道底部. Cr 与其他 5 种重金属间不呈相关性,且仅在支流入河口 A 处沉积物中含量较高,其余沉积物采样点中较少存在 Cr 富集,这说明顺德水道周边 Cr 的排放源较少,且集中于河道上游. 图 4(b)为各支流入河口沉积物中 6 种重金属的主成分分析,分析表明只有一个主成分因子(可解释总方差的 68%),说明 6 种重金属

元素主要来源于外源污染,可能来源于顺德水道周边的工业活动及北江上游的携带沉积作用.

2.4 潜在生态风险评价结果分析

对河道周边表层土壤及入河口表层沉积物的理化性质分析结果表明:顺德水道周边表层土壤的pH介于6.17~7.99之间,均值为7.24,其有机碳质量分数约占1.36%(范围为0.691%~2.816%);入河口表层沉积物的pH范围为7.11~7.78,均值为7.45,有机碳含量略高于表层土壤,约为1.54%(范围为0.56%~2.69%).根据表2所列的预测模型计算5种重金属元素的水溶态含量,并以顺德元素背景值作为参比值,对顺德水道周边表层土壤及人

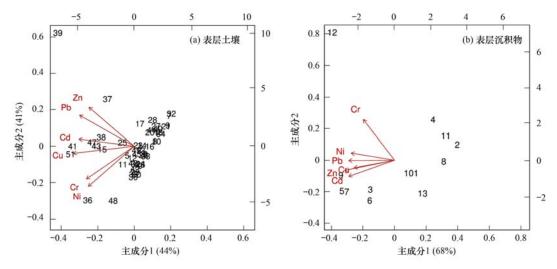


图 4 顺德河道表层土壤、支流入河口表层沉积物中 6 种重金属的主成分分析

Fig. 4 Principal component analysis of heavy metals in surface soils and sediments samples collected from Shunde waterway

河口表层沉积物重金属的环境生物可利用态进行潜在生态风险评价. 表 5 为顺德水道周边表层土壤及

入河口表层沉积物中各重金属的两种潜在生态风险 评价结果.

表 5 两种潜在生态风险评价方法结果对比

Table 5 Comparison of the results obtained with the two potential ecological risk assessment met
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评价方法	介质	番目	$E^i_{f r}$				RI	
计侧方法	介质	项目	Cd	Zn	Pb	Cu	Ni	- KI
		min	8.45	0.54	1.19	0.58	0.46	13.23
	表层土壤	average	58.87	1.70	5.45	3.71	3.47	73.20
	农压工场	90%	148.94	2.64	11.42	7.47	4.90	176.53
Hakanson 潜在		max	275.17	12.85	32.41	11.38	6.56	306.93
生态风险评价		min	44.56	1.25	2.97	4.60	2.70	58.87
	沉积物	average	117.99	2.85	9.21	11.19	4.58	145.82
	DL47(42)	90%	143.79	3.92	14.94	17.34	5.70	184.83
		max	149.76	4.40	16.34	26.97	6.08	185.14
	表层土壤	min	5.21	0.32	1.08	0.23	0.07	7.92
		average	26.69	1.39	3.65	1.41	1.88	34.96
		90%	67.12	2.27	5.97	2.65	3.61	79.28
基于生态环境可利		max	102.50	7.55	12.54	3.95	6.80	114.53
用态生态风险评价		min	14.42	0.79	2.02	1.67	0.62	19.52
	沉积物	average	46.57	2.06	5.30	3.78	1.54	59.25
	0.45(19)	90%	64.83	2.89	7.96	5.80	2.14	83.62
		max	67.94	3.51	8.11	8.33	2.56	90.45

两种评价方法结果均表明: Pb、Cu、Zn、Ni的潜在生态风险危害程度均表现为轻微(如表 5); 人河口表层沉积物的平均综合潜在生态风险指数(RI)高于河道周边土壤,其均值均低于 160,表现为轻微污染. 顺德水道表层土壤中 Cd 元素基于Hakanson潜在生态风险评价的单项生态风险系数均值的危害等级表现为中度,最大值则达到很强;而基于环境生物可利用态的潜在生态风险评价均值的危害等级则表现为轻度,最大值为强度(见表 5).人河口沉积物 Cd 元素的危害等级亦表现为Hakanson法高于环境可利用态法,前者表现为强度,后者为中度. 两种评价方法中 Cd 元素对该区域综合潜在生态风险指数 RI 的贡献率均超过 70%,说明 Cd 是造成区域重金属污染综合潜在生态风险最主要的因素.

将两种评价方法结果对比可知: Cd、Cu、Ni 基于环境生物可利用态的潜在风险系数均值约为 Hakanson 潜在风险系数均值的 33% ~54%,而 Zn、Pb 则为58% ~82%.5 种重金属基于环境生物可利用态的潜在生态风险系数均小于 Hakanson 潜在生态风险系数.由于 Hakanson 潜在生态风险系数.由于 Hakanson 潜在生态风险考虑的是重金属总量对环境的风险,而基于环境生物可利用态的潜在生态风险评价法考虑了土壤的理化性质及重金属形态,因此该评价方法可避免过高估计重金属的生态风险.此外本研究采用经验公式估算水溶态重金属含量,如有条件可使用实测值代替,其评

价结果将更准确.

3 结论

(2)对土壤及沉积物中重金属潜在生态风险评价结果表明:基于环境生物可利用态的潜在生态风险结果低于 Hakanson 潜在生态风险评价结果. 前者充分考虑了土壤理化性质及重金属形态,可避免高估重金属的潜在危害程度. 两种评价结果中除Cd 元素外其余 4 种元素的潜在生态危害程度均表现为轻微,Cd 元素是区域重金属综合潜在生态风险的主要影响因素.

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