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微囊藻毒素对水稻根系生长和抗氧化系统的影响

土壤重金属镉标准值差异比较研究与建议

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摘要:伴随我国工农业生产的快速发展,土壤污染物超标等问题逐年加剧,由土壤镉污染导致水稻出现"镉米"等食品安全事件频发,引起社会对土壤镉环境标准的广泛关注.建立土壤环境标准是开展土壤环境质量现状评价的基础,世界各国虽皆以镉对动植物和人体健康影响的最大允许值作为定值的基础依据,但定值方法、标准应用目标各有侧重,因而标准限值差异较大.通过对国内外土壤重金属镉标准限值和我国土壤镉背景值的研究,给出了全国各省级行政区域土壤表层镉背景含量范围值和部分土壤类型镉背景含量范围值;报告了我国现行土壤镉环境质量标准存在的突出问题,提出土壤环境质量标准中应设置土壤镉梯度标准限值、有效态与总量标准限值并存以及严格区分人为污染与高背景等建议.

关键词:土壤污染; 重金属镉; 标准限值; 背景值; 有效态

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Soil Heavy Metal Cadmium Standard Limit and Range of Background Value Research

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Abstract: With the rapid development of industrial and agricultural production in China, problems such as excessive soil pollutants worsen year by year, and soil cadmium pollution resulting from the emergence of "cadmium rice" and other food security incidents occur frequently. It causes the extensive concern of society on soil cadmium environmental standards. A soil environmental standard is the foundation for the evaluation of soil environmental quality. The maximum allowable value of cadmium that affects plant and animal and human health is regarded as the basis value all over the world; however, the certification methods and the goals of standard application are different, thus the standard limit has a great difference. Through the research on domestic and foreign soil cadmium standard limits and the soil cadmium background values in China, the soil cadmium background content range values of the provinces and of some types of soil are given. We report the outstanding problems existing in China's current environmental quality standard of soil cadmium. It is proposed that the soil environmental quality standard should allow the coexistence of gradient of soil cadmium standard limit, effective state and total standard limit, and anthropogenic pollution and background values exceeding the standard should be strictly distinguished.

Key words: soil pollution; heavy metal cadmium; standard limit; background value; effective state

我国土地资源短缺,建立土壤环境标准是开展土壤环境质量现状评价的基础,是指导土地利用规划和土壤污染治理,保护土地资源的重要手段^[1]. 伴随我国工农业生产的快速发展,土壤污染物超标等问题逐年加剧,由土壤镉污染导致水稻出现"镉米"等食品安全事件频发,引起社会对土壤环境标准的广泛关注^[2~5]. 如何正确认识和科学确定土壤污染物的标准限值,更好地管理和利用土地资源,实现保护土壤生态系统和人体健康的目标,已成为急需解决的重要课题.

1 发达国家土壤环境中镉的标准限值

土壤环境镉标准值通常是基于多种可能暴露途径的生态风险评价或健康风险评价而制定的^[6~11],由于土壤系统的复杂性,土壤环境标准的制定需要

一个不断完善的过程,世界各国在制定土壤环境镉标准时,多以保护土壤基本功能、生态系统安全、人体健康等为目标^[12],通常要考虑镉的化学毒性、环境背景值、影响迁移和暴露等环境因素,确定土壤镉各级别标准限值.各国制定土壤标准的原则依据基本相同,即以镉对动植物和人体健康产生影响的最大允许值作为定值的基础依据,但定值方法、标准应用目标各有侧重,标准限值差异较大(见表1).

(1)美国是国家制定导则,由各州自行确定土 壤镉标准限值,由环保署颁布用于推导保护生态环

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境和用于推导保护人体健康的土壤筛选导则(soil screening guidance, SSG)^[13~16],规定了生态土壤筛选值(ecological soil screening values-plant)为32 mg·kg⁻¹;人体健康土壤筛选值(human health soil screening values-ingestion)为70 mg·kg⁻¹;美国禁止杀虫剂中含有镉;规定工作环境空气中镉尘限值为200 mg·m⁻³.

- (2)英国环保署(Environment Agency, EA) 2002 年按照土地利用类型,考虑镉对人体健康和对植物的影响^[17],制定了土壤质量指导值,规定土壤污染"起始浓度"为 3.5 mg·kg⁻¹,但未给出土壤修复所需的"修复值".
- (3) 丹麦制定了由背景水平 0.03 ~ 0.5 mg·kg⁻¹、生态毒理学基准(eco toxicological benchmark) 0.3 mg·kg⁻¹、土壤质量基准(soil quality benchmarks) 0.5 mg·kg⁻¹构成的标准限值体系,以确保准确评判土地的利用目标和受污染程度^[18,19].
- (4)日本土壤标准以保护土壤和地下水为目标,综合考虑土壤镉的生物活性,在制定土壤标准中采用了浸出液(将土壤和10倍量的水混合,将污染物浸出)标准,实现了土壤镉有效态分析检测,规定了土壤试样溶出量镉的标准限值为0.01 mg·L⁻¹、大米为1.0 mg·kg^{-1[20,21]}.
- (5)荷兰国家土壤环境标准主要由土壤镉的目标值(target value)、调解值(intervention value)、最高允许浓度(maximum permissible concentration,

- MPC)等构成,荷兰国家土壤标准以保护生态系统和人体健康以及修复污染土壤为目标,综合考虑了土壤质地和有机质组成,构成了比较完善的标准体系[22].
- (6) 澳大利亚国家环境保护委员会(National Environmental Protection Council, NEPC)对于土壤镉设立了 A 级为土壤背景值 0. 04~2 mg·kg⁻¹和 B 级为土壤调查值 3 mg·kg^{-1[23]}. 一般来说,当测定区域土壤污染物浓度符合 A 级时,可认为未受到污染;如高于 B 级,表示该区域土壤环境需进行进一步调查. 由于澳大利亚不同地区的生态多样性,各地还可以自行建立区域生态调研值(regional ecologically investigation levels, REIL)
- (7)加拿大环境部长理事会(The Canadian Council of Ministers of the Environment CCME)制定的土壤环境质量标准名称单一,称作土壤质量指导值,其标准限值为 1.6 mg·kg⁻¹,但在标准制定时同时考虑了人体健康风险和生态风险两类受体,一般用于初步判断污染场地是否需要进一步的详细风险评估.
- (8)瑞士现行标准是总量和有效态标准并用. 土壤镉总量的指导值(guideline values)为 0.8 mg·kg⁻¹;临界值(the critical value)为 2 mg·kg⁻¹;农用地修复值(repair value)为 30 mg·kg⁻¹;土壤镉可溶态指导值为 0.02 mg·kg⁻¹;临界值为 0.02 mg·kg⁻¹;农用地修复值为 0.1 mg·kg^{-1[24]}.

表1 一些国家和地区土壤中镉的标准限值¹⁾/mg·kg⁻¹

Table 1 Cadmium standard limits of soil in some countries and regions/mg·kg⁻¹

序号	国家和地区	土壤镉限值 (1)	土壤镉限值 (2)	备注
1	欧共体	1~3		
2	美国	3. 56	32	(1)通用值;(2)生态筛选值
3	澳大利亚	0.04 ~ 2	3	(1)背景值;(2)环境调查值
4	中国台湾	2. 5	5	(1)农产地监测基准值;(2)农产地管制标准值
5	英国	3.5	1	(1)农副业产地;(2)公园和运动场
6	日本	0.01 (mg·L ⁻¹)) 1	(1)试样溶出标准;(2)大米标准
7	荷兰	0.8	12	(1)目标值;(2)调节值
8	加拿大	1.6	1.5	(1)国家标准;(2)北魁克省标准
9	丹麦	0.03 ~ 0.5	0.5	(1)背景水平;(2)土壤质量基准值
10	瑞士(总量)	0. 8	2	(1)指导值; (2)临界值
10	瑞士(可溶态)	0. 02	0. 1	(1)可溶态指导和临界值;(2)可溶态修复值
11	法国	2		
12	德国	3		
13	意大利	3		
14	前苏联	5		

1) 日本采用 1:10 试样水溶液标准;瑞士可溶态采用 0.1 mol·L-1 NaNO3 提取液 1:2.5 干重计算

时,所采取的暴露方式、摄入途径(如食物、水、空 气、尘土等)及其比例有所不同,计算确定的土壤最 大允许浓度就相差甚远:加之最高允许浓度往往作 为各国进行土壤污染修复的行动值,标准的制定势 必受到土壤修复技术和经济能力的影响. 另外,国际 上对土壤质量标准限值的命名各不相同, 如美国的 土壤筛选值、加拿大的土壤质量指导值、澳大利亚 土壤环境调查值、丹麦的土壤基准值、荷兰的土壤 目标值和调解值等,各国命名虽然不同,但实际上是 涵盖了两个层次的内容:一个是保护土壤生态功能, 另一个层次是保护人体健康. 如荷兰的目标值是指 当低于此值时,对环境的影响可忽略,土壤具有完善 的生态环境功能;调解值是指当高于此值时,有可 能引起严重的污染事件,存在"潜在危险性";一般 土壤背景值多为评价基本标准; 指导值主要以预防 土壤污染为目的,保护对象为土壤生态功能;临界 值代表可能存在风险的域值; 修复值是绝对的限 值,当土壤污染物总浓度或可溶态浓度超过修复值 时,必须立即采取必要措施以规避风险.

2 我国各类标准中镉的限值比较

标准限值的选择是土壤环境质量评价的关键, 现行《土壤环境质量标准》(GB 15618-1995)过于强 调全国标准限值统一性,由于我国不同区域地球化 学条件差异显著,土壤镉背景含量地区差异明显, 采用同一标准限值来评价全国不同地区土壤环境质 量显然不十分科学[25~29]. 我国相关产地环境行业标 准大多直接引用国标 GB 15618, 在指标选取、确定 标准限值方面大致相同,缺少各行业保护目标的暴 露量评估试验. 如:《食用农产品产地环境质量评价 标准》(HJ 332-2006)、《温室蔬菜产地环境质量评 价标准》(HJ 333-2006)、《无公害食品蔬菜产地环 境条件》(NY 5010-2002)、《畜禽养殖产地环境评 价规范》(HJ 568-2010)等(见表 2)均属于直接引用 国标 GB 15618. 在《污染场地风险评估技术导则》 (报批稿)中虽然给出了"住宅类用地土壤启动值和 工业及其他用地土壤启动值",但导则中缺少对各 地方政府制定相应地方标准值的强制要求.

表 2 我国土壤镉的相关标准限值/mg·kg-1

Table 2 Cadmium standard limits of soil in China/mg·kg⁻¹

标准	项目	镉标准限值
土壤环境质量标准(GB 15618-1995)	一级标准 二级标准	0. 2 0. 3 ~ 0. 6
,	三级标准	1
	pH 值 < 6.5	0.3
食用农产品产地环境质量评价标准(HJ 332-2006)	pH 值 6. 5 ~ 7. 5	0. 3
	pH 值 > 7. 5	0. 6
	pH 值 < 6.5	0. 3
温室蔬菜产地环境质量评价标准(HJ 333-2006)	pH 值 6.5~7.5	0. 3
	pH 值 >7.5	0. 4
	pH 值 < 6.5	0. 3
无公害食品蔬菜产地环境条件(NY 5010-2002)	pH 值 6.5~7.5	0. 3
	pH 值 >7.5	0.6
	pH 值 < 6.5	0. 3
畜禽养殖产地环境评价规范(HJ 568-2010)	放牧区 pH 值 6.5~7.5	0. 3
	pH 值 > 7.5	0. 6
	养殖场、养殖小区	1
污染场地风险评估技术导则(报批稿)	住宅类用地土壤启动值	6. 8
13 A SOME MENTILLIA A VI (IKININ)	工业及其他用地土壤启动值	13
农用污泥中污染物控制标准(GB 4284-84)	最大浓度(pH < 6.5)	5
水/1117/16 [17 木型 圧IIIII (OD 4204-04)	最大浓度(pH≥6.5)	20

3 土壤镉背景含量范围值

3.1 土壤镉背景含量范围值计算

土壤环境背景值是在不受或很少受到人类活动

影响下的土壤化学组成或元素含量水平,是判定土壤是否受到外源污染的基本依据.由于我国土壤环境背景值成因复杂,元素含量空间分布差异明显,因此,土壤环境背景值对于一定区域来讲应该以范围

值表示^[30~32],其范围值获得方法为:首先在调查单元内尽可能地采集较多的代表性样本,然后采用地球化学统计法对样品的原始分析数据进行顺序量统计,确定含量最小值、最大值和数据分散程度,计算出均值和标准差,然后根据样品数据分布情况,若是近似对数正态分布的调查单元,采用几何均值和几何标准差确定95%置信度范围值,即:"95%置信度范围值=几何均值/几何标准差的平方~几何均值

×几何标准差的平方";若是近似正态分布的调查单元,采用算术均值和算术标准差应用" \bar{X} ±2S"公式计算95%置信度范围值.全国各省(市、区)土壤表层镉样品数据均属于近似对数正态分布,因此,其95%置信度范围值应以几何均值和几何标准差计算,其结果见表3,土壤环境背景范围值可作为各省(市、区)土壤环境质量评价的基准,用以保护土壤自然背景状态.

表 3 全国各省(市、区)土壤表层镉背景含量范围值1)/mg·kg-1

Table 3 Surface soil cadmium background content range value of the provinces (city, area) in China/mg·kg⁻¹

地区	最小值	最大值	几何平均	标准差	95% 置信度 范围值	地区	最小值	最大值	几何平均	标准差	95% 置信度 范围值
全国	0.001	13. 430	0.0740	2. 118 0	0. 017 ~ 0. 332	江西	0.006	1.010	0.0696	2. 605 6	0. 010 ~ 0. 473
辽宁	0.001	0.412	0.0908	1.8290	0. 027 ~ 0. 304	湖北	0.016	8. 220	0. 113 7	2. 504 2	$0.018 \sim 0.713$
河北	0.002	0.474	0.0561	3. 498 1	0.005 ~ 0.687	湖南	0.002	4. 500	0.0787	2.6082	0. 012 ~ 0. 535
山东	0.021	0. 204	0.0745	1.6601	$0.027 \sim 0.205$	陕西	0.031	0. 249	0.0886	1. 435 9	0. 043 ~ 0. 183
江苏	0.008	2. 470	0.0501	3. 190 4	0.005 ~ 0.510	四川	0.010	0.150	0.0750	1. 368 7	$0.040 \sim 0.141$
浙江	0.004	0. 220	0.0472	2. 623 0	0. 007 ~ 0. 325	贵州	0.042	7.650	0. 208 6	3.7779	0.015 ~ 2.977
福建	0.016	0. 447	0.0538	2. 103 5	0. 013 ~ 0. 238	云南	0.009	3.409	0. 103 5	3.0434	0. 011 ~ 0. 959
广东	0.004	2. 286	0.0408	2. 214 3	0.008 ~ 0.200	宁夏	0.046	0. 254	0. 102 5	1. 530 4	0. 044 ~ 0. 240
广西	0.006	13. 430	0.0791	3. 995 6	0.005 ~ 1.263	甘肃	0.037	0. 236	0.1106	1. 361 5	0.060 ~ 0.205
黑龙江	0.036	0.400	0.0806	1.4422	0. 039 ~ 0. 168	青海	0.073	0. 264	0. 132 9	1. 266 5	0. 083 ~ 0. 213
吉林	0.013	0.429	0.0914	1.4962	0. 041 ~ 0. 205	新疆	0.016	1.634	0. 103 7	1.7113	$0.035 \sim 0.304$
内蒙古	0.004	0. 214	0.0374	2. 552 1	0.006 ~ 0.244	西藏	0.006	0. 583	0.0775	1. 347 3	0. 043 ~ 0. 141
山西	0.031	0. 358	0. 111 8	1.6986	0. 039 ~ 0. 323	北京	0.005	0.339	0.0534	2. 541 3	$0.008 \sim 0.345$
河南	0.039	0. 276	0.0726	1. 256 2	0.046 ~ 0.115	天津	0.054	0. 943	0.0870	1. 285 7	0. 053 ~ 0. 144
安徽	0.020	0. 344	0. 083 7	1. 702 0	0. 029 ~ 0. 243	上海	0.052	0. 331	0. 124 1	1. 597 6	0. 049 ~ 0. 317

1)根据《中国土壤元素背景值》中国环境科学出版社 1990年, 计算得出 95% 置信度范围值, 未包括中国台湾、香港、澳门地区, 下同

3.2 土壤镉背景含量区域分布不均

土壤成土母质类型是影响土壤镉背景值的重要因素之一,不同自然条件下发育的土类其土壤镉背景含量有明显的差异^[33~35],即使是同一土类由于发育母岩不同其背景元素含量也有很大不同,从全国20个主要土壤类型来看(见表 4),不同土壤类型镉背景含量差异显著,明显高于全国均值(0.074 mg·kg⁻¹)的土类有石灰土、水稻土、灰褐土、灰棕漠土;明显低于全国均值的土类有红壤、赤红壤、栗钙土等.

全世界多数土壤镉含量在 0. 01 ~ 2. 0 mg·kg⁻¹ 之间,中值为 0. 35 mg·kg⁻¹,我国土壤的自然镉含量在 0. 01 ~ 1. 80 mg·kg⁻¹之间,平均为 0. 163 mg·kg⁻¹,比世界多数土壤的平均镉含量低^[36].土壤中镉一般以可给态、交换态和难溶态存在,其对大气沉降和随水流迁移到土壤中的镉有较强的吸附力,吸附率在 85% ~ 95%,吸附的镉一般停留在表层 0~15 cm 土壤, 15 cm 以下含量显著减少. 根据全国"七五"土壤背景值调查结果,我国土壤表层(0

~20 cm) 镉的背景含量最小值为 $0.001 \text{ mg} \cdot \text{kg}^{-1}$,最大值为 $13.430 \text{ mg} \cdot \text{kg}^{-1}$,相差巨大. 各省(市、区) 土壤镉背景含量超过全国均值($0.074 \text{ mg} \cdot \text{kg}^{-1}$)的有 20 个省(市、区),表层镉含量范围值如:贵州($0.015 \sim 2.977 \text{ mg} \cdot \text{kg}^{-1}$)、广西($0.005 \sim 1.263 \text{ mg} \cdot \text{kg}^{-1}$)、云南($0.011 \sim 0.959 \text{ mg} \cdot \text{kg}^{-1}$)、湖北($0.018 \sim 0.713 \text{ mg} \cdot \text{kg}^{-1}$)等省区最大值远高出现行国家标准,但其属高背景值,并非外来污染所致,各省(市、区)辖区内土壤镉含量分布也十分不均.

由于土壤背景值缺少环境学和生态学意义,并不能真正体现污染对土壤生态环境及人体健康造成危害的程度,土壤自然背景高与人为活动污染超标有本质上的不同,在制定标准和各类总结报告中不可混为一谈. 我国台湾省在 2011 年颁布的《土壤污染监测标准》中第二条规定:"本标准所列土壤中物质(包含镉等 8 种)浓度,受区域土壤地质条件及环境背景因素影响,经具体科学性数据研判,属非因外来污染而达本标准值,经中央主管机关同意,不适用本标准".即:自然背景超标土壤不执行该标准.由

表 4 全国部分土壤类型(表层)镉背景含量范围值/mg·kg-1

Table 4	Cadmium background	content range va	lue of some types o	f soil (sur	face) in China	/ma.ka-1
rabie +	Gaummum Dackground	content range var	iue oi some types c	n son (sui	race, in Gillia	t/ Hig · Kg

土壤类型	最小值	最大值	几何平均	标准差	95% 置信度 范围值	土壤类型	最小值	最大值	几何平均	标准差	95% 置信度 范围值
石灰土	0.003	13. 430	0. 385	4. 183 5	0. 022 ~ 6. 745	紫色土	0.010	0.710	0. 075	2. 022 1	0. 018 ~ 0. 307
水稻土	0.008	3.000	0.108	2. 157 7	$0.023 \sim 0.502$	棕壤	0.001	0.485	0.078	1.7727	$0.025 \sim 0.245$
灰褐土	0.006	0.301	0. 127	1. 535 9	0.054 ~ 0.300	黄棕壤	0.008	8. 220	0.079	2. 203 9	0. 016 ~ 0. 382
棕钙土	0.005	0.589	0.072	2. 672 6	$0.010 \sim 0.515$	草甸土	0.005	0.300	0.074	1. 692 1	$0.026 \sim 0.211$
栗钙土	0.002	0.303	0.041	3. 377 5	0.004 ~ 0.463	黑土	0.004	0.165	0.073	1.4123	$0.037 \sim 0.146$
灰棕漠土	0.005	0. 257	0. 103	1. 429 5	$0.051 \sim 0.211$	灰钙土	0.026	0.172	0.083	1.4166	0. 041 ~0. 167
白浆土	0.032	0.429	0.093	1.6576	$0.034 \sim 0.255$	黄壤	0.005	4.500	0.064	2.0030	$0.016 \sim 0.258$
潮土	0.005	0.943	0.085	1. 937 5	$0.023 \sim 0.320$	绵土	0.006	0. 249	0.093	1. 346 5	0. 052 ~ 0. 169
黑钙土	0.005	0.393	0.087	2. 117 1	0.019 ~ 0.389	红壤	0.002	4.500	0.047	2. 204 0	0. 010 ~ 0. 229
褐土	0.002	0. 583	0.081	1. 957 1	0.021 ~ 0.310	赤红壤	0.005	0.505	0.033	2. 306 6	0.006 ~ 0.176

此可见,台湾省是将土壤自然背景超标与人为污染超标区别对待的一个例证.同时,为了进一步防控土壤的外源污染,保护土壤生态环境安全和人体健康,还必须在土壤环境背景基准值之上制定第二级和第三级标准限值.

4 我国土壤镉标准的主要问题

4.1 标准中镉含量定值偏严

我国现行《土壤环境质量标准》(GB 15618-1995)中规定耕地土壤执行二级标准,其中"镉"标准值为 0.3~0.6 mg·kg⁻¹. 而在国外的相应标准中荷兰和瑞士均为 0.8 mg·kg⁻¹、加拿大为 1.6 mg·kg⁻¹、台湾省食用作物农地管制基准值为 2.5 mg·kg⁻¹、管制标准值为 5 mg·kg⁻¹、澳大利亚环境调查值为 3 mg·kg⁻¹、欧共体规定为 1~3 mg·kg⁻¹,综合比较我国标准可能偏严. 应该根据我国各类土地利用目标进行土壤生态环境和人体健康风险评估,研究修订我国土壤镉的标准限值,以适应实际工作的需要.

4.2 土壤镉的超标与污染

土壤环境中镉含量的自然高背景值超标与人为污染超标有本质上的不同.目前,按照现行《土壤环境质量标准》(GB 15618-1995)评价土壤镉元素含量,会出现高背景值点位无外源污染超标情况.因为该国标中规定的镉元素标准值,是根据土壤中镉的平均背景含量及95%置信水平来确定的,该标准是我国首个土壤环境标准,在保护土壤环境方面起了重要作用.但是,我国地域辽阔,气候、植被、成土因素及土壤类型复杂多样,同一元素在不同区域、不同类型土壤中的背景含量存在较大差异,许多地区土壤自然背景值高于国家标准值,必然出现即使土壤没有受到外源污染其评价结果依然超标的现

象,应该使大家清楚地认识到以现行国标评价的结果超标不一定是污染.

4.3 土壤镉标准限值层次单一

现行《土壤环境质量标准》(GB 15618-1995)中规定了"镉"有 5 个标准限值,但均属于同一层次,只是区分了不同土地利用类型和不同酸碱度条件下分别执行不同的标准限值.荷兰国家土壤环境标准分为目标值、调解值;瑞士制定了土壤镉总量的指导值、临界值、农用地污染修复值;澳大利亚国家对于土壤镉设立了 A 级为土壤背景值和 B 级为土壤调查值;丹麦制定了由背景水平值、生态基准值、土壤质量基准值三位一体的土壤质量标准限值;美国规定了生态土壤筛选值;人体健康土壤筛选值等.以上国家均制定了土壤镉标准的梯度限值,分别用以保护土地的不同利用目标.

4.4 土壤镉缺少有效态标准值

我国现行国家标准中镉是以全量制定的标准值,全量只能代表该元素的总贮量,并不能真正表达镉的供给水平.土壤中镉一般以可给态、交换态和难溶态存在,不同存在形态的镉其迁移、转化、吸附、解吸能力各不相同,其受土壤 pH 值和有机质含量影响较大.在监测与评价中有效态标准值无疑优于土壤镉全量值,因为有效态含量可以避免因土壤母质类型和土壤理化性质的不同而带来的实际危害差异,更能表达土壤功能及土壤环境质量的优劣程度.

5 建议

(1) 建议制修订适合我国现阶段土壤镉的各级标准值. 以我国敏感人群对土壤重金属镉的每日允许摄入剂量(非致癌性污染物)或可接受风险水平(致癌性污染物)为基本出发点,根据不同土地利用

方式,确定暴露风险受体、暴露途径及迁移转化特性等参数,并进行模型计算,最后对计算的标准值再参考国外标准、结合国内经济技术水平进行必要的调整.

- (2)当实测土壤镉含量值超过国家土壤标准限值时,建议首先与该地域土壤背景值比较,区别是外来污染还是自然高背景值,在确认是高背景情况下,应与有明显外源污染土壤区别对待;确系有污染源污染要进一步进行环境质量综合评价和各类风险评估.
- (3)设立土壤镉的梯度标准限值.如:一级:土壤自然背景值(自然背景保护);二级:土壤环境指导值(受污染但可使用);三级:土壤污染修复值(高风险必须治理).分别作为保护土壤功能自然背景,保护土壤生态环境和人体健康安全,开展污染土壤修复的标准值.对于不同用途土壤如:食用农产品产地、生态环境用地、住宅用地、工业用地等分别制定不同的标准限值,同时考虑镉的有效态标准值.
- (4)将有效态标准限值与全量标准限值同时使用.在实际应用中,有效态标准限值的可靠性取决于有效态测定方法的可靠性,在常见的土壤有害元素中铜、镉、铅和锌的提取和测定方法比较成熟.总量标准值虽然不能反映元素的有效性,但它具有测定方法成熟、数据结果比较稳定等优点,因此,在现阶段将有效态标准限值与全量标准限值同时使用较为适宜.

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