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次 目

酸化蛭石的表面有机修饰及其对疏水性微污染物的吸附

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摘要: 插层法制备有机黏土存在比表面积低、负载不均等问题,限制了此类材料吸附性能的进一步提高. 基于此,通过选择性酸浸的方法预处理天然蛭石,以三甲基氯硅烷(CTMS)和三乙基氯硅烷(CTES)对其进行表面有机修饰,利用 FTIR、BET、SEM和热重等方法对材料进行表征. 结果表明,改性酸化蛭石的比表面积可达 361.0 m²·g⁻¹,而有机插层蛭石的比表面积仅为6.0 m²·g⁻¹,有机基团更稳定地以共价键形式负载于酸化蛭石表面. 以疏水性有机微污染物邻苯二甲酸二乙酯(DEP)为测试目标,考察材料的吸附性能. 在本实验条件下,测得 CTES 改性酸化蛭石、CTMS 改性酸化蛭石和有机插层蛭石对 DEP 的吸附量分别为 63.7、51.2 和 15.7 mg·g⁻¹,证明有机修饰后的酸化蛭石具有更强的疏水性吸附能力,有机负载的均匀性是决定吸附能力的关键因子. 动力学研究表明吸附行为遵循拟二级动力学方程;吸附等温线表现出线性特征,可由 Henry 和 Freundlich模型进行描述,表明分配作用是吸附过程的主要机制.

关键词:黏土;酸化蛭石;有机改性;吸附;邻苯二甲酸二乙酯

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Surface Organic Modification of Acid Vermiculite and Its Adsorption of Hydrophobic Micro Pollutants in Aqueous Solutions

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Key words: clay; acid vermiculite; organic modification; adsorption; diethyl phthalate (DEP)

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Abstract: To solve the problems of intercalated organoclay such as low surface area and inhomogeneous organic loading, natural vermiculite was activated by acid leaching and then modified by trimethylchlorosilane (CTMS) and triethylchlorosilane (CTES). The modified materials were characterized by FTIR, BET, SEM and TG. Experimental results indicated that the surface area of the modified acid vermiculite (361.0 m²·g⁻¹) was much larger than that of the intercalated organovermiculite (6.0 m²·g⁻¹), moreover, the organic groups were grafted onto the surface covalently. Diethyl phthalate (DEP), a typical hydrophobic micro-organic pollutant, was used to test the adsorption capacity of different adsorbents. The adsorption amounts of DEP were 63.7, 51.2 and 15.7 mg·g⁻¹ for CTES, CTMS and intercalated organovermiculite in this study, respectively. The high organic affinity of modified acid vermiculite was due to both the bigger surface area and the homogeneous organic loading. The adsorption kinetics was found to follow the pseudosecond-order model. The isotherms exhibited linear characteristics and could be described by Henry and Freundlich equations, indicating that the partition process is the main control mechanism of the removal of DEP.

黏土矿物是一类储量丰富、环境友好的天然材料,具有良好的吸附性能^[1,2],并因其独特的层状结构而成为应用广泛的无机载体材料^[3,4],其中以插层法最为普遍^[5]. 插层法通过液相离子交换实现功能化负载,其制备方法相对简单,然而所得材料比表面积(S_{BET})低^[6]、负载不均匀^[7]、接枝键较弱^[8]等问题难以克服. 如以最常见的十六烷基三甲基溴化铵改性,膨润土^[9]、蒙脱石^[10]、蛭石^[11]等改性黏土的 S_{BET} 均在 50 m²·g⁻¹以下,缺乏足够的孔结构,并且有机阳离子在水中受离子浓度的影响,容易从黏土层间脱落而进入溶液,由此导致其在水处理中的实用性大大降低^[12,13].

解决以上问题需从吸附过程的机制分析入手, 关键在于寻找新的改性技术路线,充分开发黏土的 载体功能. 本课题组利用分子模拟的方法进行理论 计算,证明了提高有机黏土吸附能力的关键取决于 可提供有机物的吸附位数量^[7]. 插层黏土的有机接 枝点为层间可交换阳离子,受天然黏土阳离子交换 容量(CEC)的限制,增加有机吸附位只能依靠采用 具有长碳链烷基的改性剂,由此导致有机阳离子的

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团聚效应而造成孔道堵塞,使得插层黏土在材料表面为非均匀负载而是分散的分子堆积,有机吸附位无法充分与污染物接触.由此推之,若在黏土表面开发出新的有机接枝点,并将更多的短链烷基接枝到黏土表面,形成均匀的表面有机修饰,则可使材料的比表面积和负载均匀性得到显著改善.基于这一思路,采取酸浸的方法对天然蛭石进行表面处理,利用蛭石层间八面体阳离子的选择性溶出,大幅提高其比表面积并形成均匀分布的(单位晶格)活性SiOH基团,以三甲基氯硅烷进行共价键型有机接枝,该材料对疏水性有机物的吸附能力为传统有机插层黏土的3.1倍[14],证实了理论预测.

在此基础之上,为了探究进一步优化改性酸化蛭石疏水性吸附能力的方法,解析有机负载量在提升吸附容量影响及吸附速率方面的影响,选择与三甲基氯硅烷结构一致、碳链长度加倍的三乙基氯硅烷作为改性剂,在等摩尔当量的条件下对酸化蛭石进行改性.通过对比不同吸附材料的疏水性吸附能力,重点考察改性剂碳链长度对吸附效果的影响,从而获取进一步提高改性酸化蛭石吸附性能的数据.测试的目标污染物为邻苯二甲酸二乙酯(diethylphthalate, DEP), DEP 是一种用途广泛的塑化剂^[15],也是废水生物降解后普遍存在的微量中间产物^[17],作为一种典型的内分泌干扰物对人的生殖系统具有严重危害^[18,19],可以作为需要重点控制的

1.3 有机插层蛭石的制备

作为对比,采用十六烷基三甲基溴化铵(CTAB)对天然蛭石进行插层改性,改性方法为溶液离子交换法^[25],改性剂用量为100%阳离子交换容量(CEC),样品经冷干燥24h,装入密封瓶备用,记为CTAB-V.

1.4 材料表征

FTIR 分析在美国 Thermo 公司 Nicolet 6700 红

有机微污染物[20~24].

1 材料与方法

1.1 实验材料

实验所用蛭石购自河北省灵寿县鼎盛矿业加工厂,由于天然蛭石矿物中包含一些杂质,如石英砂、方解石、细砂等,因此使用之前需进行提纯. 根据之前的处理方法^[25]对天然蛭石进行除杂和提纯,并用3 mol·L⁻¹的盐酸进行活化处理^[14],得到酸化蛭石(记为3M),密封储存备用.

所用三甲基氯硅烷(CTMS)、三乙基氯硅烷(CTES)、十六烷基三甲基溴化铵(CTAB)和邻苯二甲酸二乙酯(DEP)等试剂均为分析纯,购自阿拉丁药剂公司(上海).

1.2 酸化蛭石的疏水化改性

酸化蛭石的有机接枝反应可由式(1)和(2)表示. 取等摩尔量(0.06 mol)的 CTMS 和 CTES 液体,以甲苯为溶剂配制体积比为 20%的溶液 100 mL,加入 5 mL 吡啶作为生成 H⁺的中和剂. 将 2.0 g 酸化蛭石加入其中,反应置于温度为 40℃、转速为 170 r·min⁻¹的摇床中进行 24 h. 样品离心后,先用无水乙醇清洗残留在表面的硅烷偶联剂和有机溶剂,再用去离子水反复清洗,直至硝酸银反应无白色沉淀.最后在 105℃条件下烘干 24 h,密封储存备用,样品记为 CTMS-V 和 CTES-V.

外光谱仪上进行,采用 KBr 压片法,样品与 KBr 的质量比为 1:200,仪器测试的扫描范围:4000~400 cm $^{-1}$; N₂ 吸附/脱附实验在美国麦克仪器公司 ASAP 2020M 型物理吸附仪上完成,样品质量约 0.2 g,吸附-脱附温度为 -196°C,平衡时间为 30 s,数据处理在麦克公司的软件上进行,其中比表面积由 Brunauer-Emmett-Teller (BET)方程计算,孔径分布中:介孔分布由 Barret-Joyner-Halenda (BJH)方程计

算,微孔分布 Horvath-Kawazoe (HK)方程计算,平均 孔径由 BJH 脱附模型计算,总孔体积由单点吸附量 计算: SEM 测试由日本电子 JSM-6330F 电镜完成, 样品测试前进行喷金处理以增加导电率,测试电压 为 15.0 kV; TG 分析在德国耐驰 TG 209F1 热重分 析仪上进行,具体条件为:保护气 N,流量为 30 mL·min⁻¹, 升温速率 10℃·min⁻¹, 温度范围为 20~ 1030℃,样品质量约10 mg.

1.5 吸附实验

1.5.1 动力学实验

称取 25.0 mg 的蛭石样品于系列 100 mL 带聚 四氟乙烯垫片的螺口玻璃瓶中,分别移取 25.0 mL 初始浓度为99.6 mg·L⁻¹的 DEP 溶液,在25℃、170 r·min⁻¹的恒温摇床中进行吸附反应. 分别于不同 时间点取样离心,采用高效液相色谱法(HPLC)测 定 DEP 浓度. HPLC 操作条件: Agilent C18 色谱柱, 流动相为8:2的甲醇/超纯水体系,流速1.0 mL·min⁻¹,柱温40℃,UV 检测器波长228 nm. 所有 样品均做空白对照实验,以扣除器壁等非吸附方式 的损失,所得实验结果均为两次平行实验的均值. 用式(3)计算蛭石的动态吸附量 q_i .

$$q_t = (c_0 - c_t) / W_0 (3)$$

式中, $q_{\iota}(mg \cdot g^{-1})$ 为任意时间点上的吸附量, c_{Ω} 和 c_{ι} $(mg \cdot L^{-1})$ 分别为污染物起始浓度和时间 t 时液相 浓度, $W_0(g \cdot L^{-1})$ 为吸附剂浓度.

1.5.2 等温吸附实验

称取 25.0 mg 的蛭石样品于一系列 100 mL 带 聚四氟乙烯垫片的螺口玻璃瓶中,分别加入25.0 mL 不同浓度的 DEP 溶液,在 15、25 和 35℃条件下 置于 170 r·min -1 摇床中进行吸附反应,24 h 后离心 分离,测定溶液中 DEP 浓度. 根据平衡溶液中污染 物的浓度 c_{\circ} ,由式(4)计算蛭石的平衡吸附量 q_{\circ} .

$$q_{e} = (c_{0} - c_{e})/W_{0} \tag{4}$$

2 结果与讨论

2.1 材料表征

2.1.1 红外光谱(FITR)分析

不同蛭石材料的 FTIR 分析如图 1 所示. 天然 蛭石在1010 cm⁻¹的位置有一个最强的吸收峰,这 是硅氧四面体层中 Si-O 键所形成的伸缩振动 峰^[26],位于1630 cm⁻¹的吸收峰则为层间水分子 中—OH 的弯曲振动峰[27].

相比天然蛭石,酸化蛭石的表面基团发生了明 显改变. 位于1630 cm⁻¹的层间水分子中—OH 的

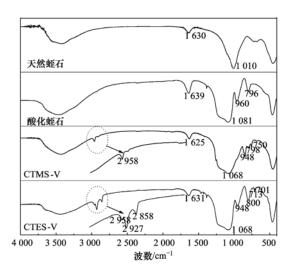


图 1 不同蛭石的红外光谱

Fig. 1 FTIR spectra of different vermiculites

弯曲振动峰强度增强,表明酸化蛭石表面的亲水性 增强,这是硅羟基增多的表现. 酸化蛭石中的 Si-O 键形成的伸缩振动峰(1010 cm⁻¹)也随之迁移至 1081 cm⁻¹,表明蛭石的层状结构发生了坍塌^[28]. 另外,酸化蛭石在 960 cm⁻¹和 796 cm⁻¹出现两个新 的吸收峰.

有机改性后,酸化蛭石的基本特征并未改变. 而是分别在2885、2921和2958 cm⁻¹出现了有机基 团的吸收峰,这些吸收峰分别为—CH。和—CH。的 伸缩振动峰[29],表明硅烷偶联剂被负载到酸化蛭石 表面. 从层间水峰强度来看,改性蛭石的强度明显 低于酸化蛭石,由此可知材料的疏水性得到提高.

2.1.2 比表面积与孔结构

表1列出了不同蛭石材料的比表面积和孔径参 数. 天然蛭石的比表面积 S_{BET} 仅为 14.4 $\text{m}^2 \cdot \text{g}^{-1}$,孔 径分布表明天然蛭石缺乏足够的微孔结构. 经过酸 浸处理后,酸化蛭石的 S_{BET} 得到了大幅提升,达到 528.0 m²·g⁻¹,平均孔径由7.9 nm 降到2.89 nm,孔 体积和微孔的比例升高,表明酸化处理起到了很好 的造孔作用. 经过硅烷化处理后的酸化蛭石的比表 面积有所下降,但孔径结构并未发生明显变化, CTMS-V 和 CTES-V 的比表面积与孔结构相差不大.

相比之下,传统有机蛭石的比表面积由 14.4 m²·g⁻¹降至6.0 m²·g⁻¹,这主要是由于大量有机改 性剂的堆积堵塞了层间域所致. 其孔容积也相应减 少,平均孔径则由 7.9 nm 增大至 8.97 nm,这是插 层后蛭石层间域被扩撑的结果[13]. 由此可知:插层 黏土虽然实现了有机负载,但材料不具有足够孔体 积,缺乏作为吸附剂的必要条件.

衣 1 小问题有例外的比较画标构扎住力	表1	不同蛭石材料的比表面积和孔径	分
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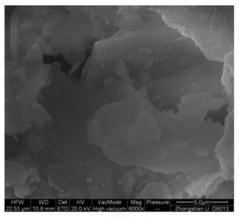
Table 1 SSA and PSD of different vermiculites

样品	$S_{\text{RFT}}/\text{m}^2 \cdot \text{g}^{-1}$	D_{av}/nm	$V_t/\text{cm}^3 \cdot \text{g}^{-1}$	气孔分布		
1十1日	S _{BET} / m · g	D _{av} / IIII	v _t / cm ⋅g	$V_{ m micro}/\%$	$V_{ m meso}/\%$	
天然蛭石	14. 4	7. 90	0. 029	0.004 (14)	0.025 (86)	
酸化蛭石	528. 0	2. 89	0. 349	0. 122 (35)	0. 227 (65)	
CTMS-V	358. 4	2. 83	0. 250	0.085 (34)	0.165 (66)	
CTES-V	361.0	2. 74	0. 247	0.087 (35)	0.160 (65)	
CTAB-V	6. 0	8. 97	0. 017	0.003 (18)	0.014 (82)	

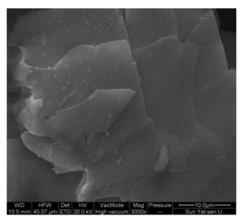
 $1)S_{\mathrm{BET}}$: 比表面积, D_{av} : 平均孔径, V_{ι} : 总孔体积, V_{micro} : 微孔体积, V_{meso} : 介孔体积

2. 1. 3 SEM

以 CTES 为例,图 2 给出了酸化蛭石经过有机 改性后,其表面形貌的变化.由 SEM 对比可知,酸 化蛭石仍具有一定的层状结构,以片状结构为主, 表面较为光滑;有机改性后,酸化蛭石的基本形貌未见明显变化,表明有机接枝的缩聚反应在酸化蛭石均匀发生,因为并未造成改性剂的堆积与团聚.



(a) 酸化蛭石



(b)改性酸化蛭石

图 2 酸化蛭石和改性酸化蛭石的 SEM 图

Fig. 2 SEM pictures of acid vermiculites before and after silanization

2.1.4 TG 分析

在影响有机黏土的吸附能力因素中,材料的有机负载量(f_{oc})也是重要的考察指标之一.目前,有机黏土吸附性能与 f_{oc} 呈正相关是普遍接受的观点[$^{[30]}$].进一步地,Fuller等[$^{[32]}$ 提出:短碳链改性黏土的吸附以分散的位点吸附为主,吸附能力较弱;而长碳链改性黏土的吸附以液/固相间的分配为主,吸附能力强.按照这一观点,使用具有长碳链烷基的改性剂成为提高有机黏土性能的手段.鉴于各种改性材料在高温条件下都会发生碳化分解,采用 TG的方法比较材料的 f_{oc} .

如图 3 所示,4 种材料的失重率依次为: CTAB-V > CTES-V > 酸化蛭石 > CTMS-V. 由此可明显看出,CTAB-V 的 f_{oc} 最大,这是由于其采用了长碳链改性剂的原因. 对于酸化蛭石,由于亲水性特点,表面吸附有较多的水分,因而在低温度区间(20~100°C)发生较大的失重. 对于 CTES-V 和 CTMS-V

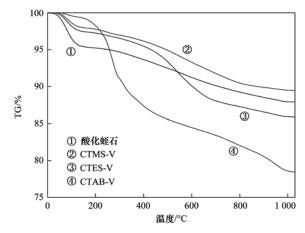


图 3 酸化蛭石和不同改性蛭石的 TG 曲线

Fig. 3 Thermal gravimetric curves of different vermiculites

两种材料,其失重曲线的规律基本相似,由于碳链长度加倍,CTES-V的 f_{∞} 大于 CTMS-V. 从曲线的趋势可以看出:在 150 ~ 400 $^{\circ}$ 之间,有机插层蛭石失重率迅速增大,这是因为 CTAB 不稳定导致的气化分

解;而酸化蛭石及有机改性蛭石失重较少,表明改性酸化蛭石的热稳定性强于传统有机黏土. 当温度超过 400° 时,有机改性蛭石表面上的有机官能团开始受热分解,失重速率迅速超过酸化蛭石. 由 100° 们的失重比可以推算 f_{oc} 大小,其中 CTAB-V 的 f_{oc} 最大(~12.5%),CTES-V 的 f_{oc} 其次(~3.69%),CTMS-V 的 f_{oc} 最小(~1.58%).

2.2 吸附动力学

图 4 给出了不同吸附剂对 DEP 的吸附动力学曲线. 可以看出,3 种材料的吸附量随时间的增加而增加,在初始阶段吸附速率很高,1 h 内的吸附量

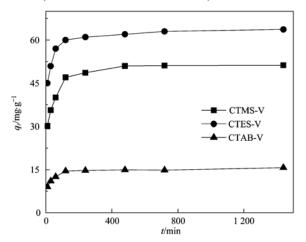


图 4 同一 DEP 浓度下不同改性蛭石的吸附动力学曲线

Fig. 4 Adsorption kinetics of DEP on different adsorbents

均已达到平衡吸附量的 70%. 随着时间的延长,吸附曲线逐渐平缓,在 4 h 左右吸附趋于平衡. 由图中曲线也可以看出,CTES-V 和 CTMS-V 的吸附能力明显高于 CTAB-V,证明改性酸化蛭石对疏水性有机物具有更强的亲和性. 在 c_0 = 99.6 mg·L⁻¹条件下,CTES-V、CTMS-V 和 CTAB-V 的平衡吸附容量分别为 63.7、51.2 和 15.7 mg·g⁻¹.

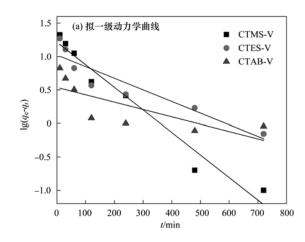
为了更好地分析 DEP 在不同改性蛭石上的吸附动力学,分别采用拟一级动力学模型^[33]和拟二级动力学模型^[34]对实验数据进行拟合,其方程式分别可由式(5)和(6)表示:

$$\lg(q_e - q_t) = \lg q_e - k_1 t \tag{5}$$

$$\frac{t}{q_t} = \frac{1}{k_2 q_e^2} + \frac{t}{q_e} \tag{6}$$

式中, $k_1(\min^{-1})$ 和 $k_2[g\cdot(\mathrm{mg}\cdot\mathrm{min})^{-1}]$ 分别为拟一级和拟二级吸附动力学常数, q_t 和 q_e 分别为在时间点 t 上的吸附量和平衡吸附量($\mathrm{mg}\cdot\mathrm{g}^{-1}$).

拟一级和拟二级吸附动力学的拟合曲线如图 5 所示. 从中可以看出:拟一级动力学模型对数据的拟合较差,不适合用来描述改性蛭石对 DEP 的吸附行为;而拟二级动力学模型和实验数据具有很高的相关性,相关系数大于 0.99,可以用来研究吸附动力学过程. 拟一级和拟二级吸附动力学的模型参数列于表 2.



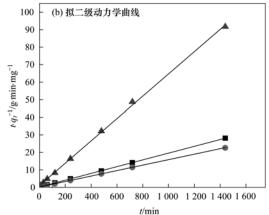


图 5 不同改性蛭石的拟一级动力学曲线和拟二级动力学曲线

Fig. 5 Pseudo-first-order and pseudo-second-order kinetic models for the adsorption of DEP onto adsorbents

表 2 不同改性蛭石的动力学拟合参数

Table 2 Parameters and correlation coefficients (\mathbb{R}^2) of two kinetic models

吸附剂	a ma.I -1	$q_{e,\mathrm{exp}}$	拟一级动力学			拟二级动力学		
吸附加	$c_{0/} \operatorname{mg} \cdot \operatorname{L}^{-1}$	/mg•g ⁻¹	$q_{\rm e}/{ m mg}\cdot{ m g}^{-1}$	k_1	R^2	$q_{\rm e}/{ m mg}\cdot{ m g}^{-1}$	k_2	R^2
CTMS-V	99. 6	51. 2	16. 3	0.0034	0. 954	51.7	1.62×10^{-3}	0. 999
CTES-V	99. 6	63.7	10. 5	0.0017	0.843	63. 9	1. 80×10^{-3}	0. 999
CTAB-V	99. 6	15.7	3. 43	0. 001 1	0.510	15. 7	3. 86×10^{-3}	0. 999

2.3 吸附等温曲线

为了评价改性蛭石材料对 DEP 的吸附性能,测试了不同温度条件下的吸附等温线,如图 6 所示. 从中可以看出,3 种材料的吸附等温线明显呈线性,这是典型的疏水性吸附特征,有机物在吸附剂表面的富集并非"位点吸附"过程,而是液固相的"分配过程"^[32]. 用于描述等温吸附过程的经典模型有Langmuir、Freundlich 和 Henry 模型,其中 Langmuir模型主要用于描述单层无分子间作用力吸附,Freundlich 模型为经验公式,而 Henry 模型用于描述

线性吸附过程,其线性表达分别如式(7)~(9)所示.

$$\frac{c_{\rm e}}{q_{\rm e}} = \frac{1}{q_{\rm max}K_{\rm I}} + \frac{c_{\rm e}}{q_{\rm max}} \tag{7}$$

$$\ln q_e = \ln K_F + \ln c_e / n \tag{8}$$

$$q_{\rm e} = K_{\rm d} c_{\rm e} \tag{9}$$

式中, $q_e(mg \cdot g^{-1})$ 为平衡吸附量, $c_e(mg \cdot L^{-1})$ 为液相平衡浓度; q_{max} 为单分子层饱和吸附量($mg \cdot g^{-1}$), K_L 为 Langmuir 吸附平衡常数($mg \cdot L^{-1}$); $K_F(L \cdot mg^{-1})$ 和n(无量纲)为Freundlich经验常数; K_d 为Henry分配系数($L \cdot g^{-1}$).

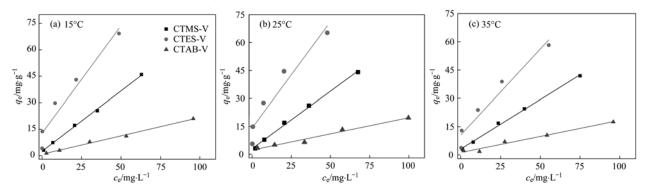


图 6 不同温度下改性蛭石对 DEP 的吸附等温曲线

Fig. 6 Adsorption isotherms of DEP onto different vermiculites

由表 3 列出的模型拟合参数可以看出: Henry 模型与吸附数据具有最高的拟合系数; Langmuir 模型通常用来描述有限的位点吸附,如离子交换 $^{[35]}$ 、 π - π 色散吸附 $^{[36]}$ 等过程,因而不适合用来模拟线性吸附行为; Freundlich 模型为经验公式,对数据也具有一定的拟合精度.由此,选择 Henry 分配系数 $K_{\rm d}$

作为参数衡量材料的吸附能力.可以看出, CTES-V和 CTMS-V的 K_d 值分别约为 CTAB-V的 6倍和 3倍, 表明改性酸化蛭石比有机插层蛭石具有更强疏水性吸附能力.对比不同温度下的吸附效果可以发现: 低温有利于 DEP 在吸附剂表面的吸附, 这表明吸附过程为自发进行的放热过程.

表 3 不同改性蛭石吸附 DEP 的等温线拟合参数

Table 3	Adsorption	isotherm	parameters	for	DEP	onto	different	vermiculite

吸附剂	温度/K	Henry 线	性方程		Langmuir 方程		F	Freundlich 方和	呈
177 [11] 7[1]	溫吳/K	$K_{\rm d}/{\rm L}\cdot{\rm mg}^{-1}$	R^2	$q_{\rm max}/{ m mg}\cdot{ m g}^{-1}$	$K_{\rm L}/{\rm L}\cdot{\rm mg}^{-1}$	R^2	$K_{\rm F}/{ m L} \cdot { m mg}^{-1}$	n	R^2
	288	0. 683	0. 998	71. 43	0. 0205	0. 628	2. 683	1. 560	0. 975
CTMS-V	298	0. 617	0.995	90. 91	0.0131	0.803	1.758	1. 326	0. 998
	308	0. 523	0. 996	62. 50	0. 0204	0.680	2. 627	1.661	0. 967
	288	1. 217	0. 925	71. 43	0. 187	0. 906	18. 38	3. 279	0. 979
CTES-V	298	1. 144	0.901	71. 43	0. 175	0.943	14. 44	2. 625	0. 959
	308	0. 911	0. 933	62. 50	0. 121	0.908	10. 23	2. 381	0. 926
	288	0. 208	0. 993	45. 45	0. 0073	0. 574	0. 580	1. 316	0. 985
CTAB-V	298	0. 174	0. 972	29. 41	0.0152	0.570	1.510	1. 961	0.890
	308	0. 170	0. 972	37. 04	0.0082	0. 227	1.081	1. 825	0. 758

2.4 有机负载均匀性对吸附过程的影响

对于传统有机插层蛭石和改性酸化蛭石,其在水溶液中对疏水性有机物的富集机制是相同的,即依靠材料负载的有机基团与有机物之间的疏水性作用力. 如 2.1.4 节所述, f_{∞} 是决定有机物吸附容量

的关键因子,根据 TG 的结果,有机插层蛭石的 f_{∞} 高于两种改性酸化蛭石,而其吸附能力却明显低于后两者.为了解析有机物在不同吸附剂上的吸附机制,以图 7 来表示 3 种改性蛭石材料对有机物的吸附过程.

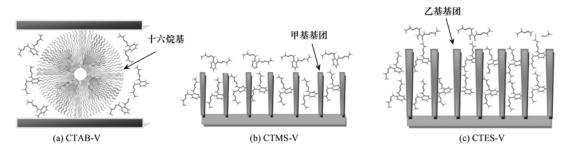


图 7 不同改性材料的吸附机制示意

Fig. 7 Adsorption mechanism of vermiculite-based adsorbents

图 7(a)给出了有机插层黏土材料的一般结构, 有机阳离子在黏土层间域内由疏水力驱动而形成胶 团(胶束)类结构[13]. 随着有机阳离子的聚集,在黏 土层间形成了大面积疏水区,由此使黏土材料具有 了对有机物的吸附能力. 然而,由于有机阳离子聚 集的无序性,导致疏水区分布不均,并堵塞了分子扩 散的基本通道,因此表现为污染物只能在少量被吸 附于疏水性的外表面,而内部的疏水性区域无法有 效暴露给污染物进行吸附. 相比之下,酸化蛭石不 仅具有更大比表面积,且活性 SiOH 平均分布于其 晶格骨架结构上[28],因此接枝后的有机基团(甲基、 乙基) 更均匀地分布于材料的表面, 如图 7(b) 和 7 (c)所示. 在这种材料表面,污染物分子可以充分地 获得与有机基团接触的机会而形成吸附,因此尽管 改性酸化蛭石的 f_{∞} 低于有机插层蛭石,但其对 DEP 的富集能力却更强. 对比 CTES-V 和 CTMS-V 的吸 附效果,前者的 K_a 值接近于后者的 2 倍. 这一结果 与之前的理论预测是一致的,即长碳链烷基可提供 给有机物更多的有效吸附位. 从 f_{oc} 值来看,CTES-V 也约为 CTMS-V 的 2 倍,这与吸附结果是相吻合的. 由此说明,改性酸化蛭石材料符合吸附能力与 f_{cc} 值 呈正相关的规律,但前提是有机基团应均匀负载于 载体表面,若如有机插层的不均匀负载,吸附能力不 能随有机负载物的增加而提高. 以上结果表明:考 察此类疏水性材料吸附能力的因素,除了材料比表 面积和有机负载量,还应包括有机基团在材料表面 负载的均匀性,即可与污染物分子接触的有效吸 附位.

3 结论

(1)通过盐酸的活化作用,将低表面积天然蛭石转变为富含活性硅羟基的多孔材料,以共价键的形式将两种有机基团(甲基和乙基)分别接枝到其表面,改性材料显示出良好的有机亲和性,对 DEP

的吸附能力 (K_d) 分别约为有机插层蛭石的 3 倍和 6 倍

(2)酸化蛭石的表面性质为实现均匀的有机修饰提供了条件,乙基改性材料的吸附能力为甲基的 2 倍,符合材料 f_{oc} 与疏水性吸附性能呈正相关的规律,并验证了分子模拟对改性剂碳链长度影响的理论预测,也为下一步开发酸化蛭石复合材料的吸附性能提供了参考数据.

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