

(HUANJING KEXUE)

ENVIRONMENTAL SCIENCE

第34卷 第7期

Vol.34 No.7

2013

中国科学院生态环境研究中心 主办

科学出版社出版



环龙科豆 (HUANJING KEXUE)

ENVIRONMENTAL SCIENCE

第 34 卷 第 7 期 2013年7月15日

次 目

铁屑-微生物协同还原去除水体中Cr(VI)研究

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摘要:考察了铁屑和微生物对受污染水体中Cr($ext{VI}$)的还原去除能力以及Cr($ext{VI}$) 去除效率的影响因素,分析了反应后铁屑表面的组成以及Cr($ext{VI}$)还原产物的形态特征. 结果表明,铁屑-微生物协同处理对水体Cr($ext{VI}$)的去除具有促进作用,在 18 h内 Cr($ext{VI}$) 去除率就可达到 100%. 在 25~42℃ 范围内,温度升高有利于Cr($ext{VI}$)的去除;Cr($ext{VI}$) 还原去除的最适宜初始 pH 为 5. 8. Cr($ext{VI}$) 去除效率随着铁屑投加量和微生物接种量的增大而增大,随着Cr($ext{VI}$) 初始浓度的增大而减小. $ext{Mn}^{2+}$ 、 $ext{Co}^{2+}$ 、 $ext{Co}^{2+}$ 、 $ext{Cu}^{2+}$ 和 $ext{Ni}^{2+}$ 离子对Cr($ext{VI}$)的还原去除都有一定的抑制作用,其中 $ext{Mn}^{2+}$ 的影响最小, $ext{Ni}^{2+}$ 的抑制作用最为明显. XPS 分析结果显示,铁屑表面吸附和沉积了 Cr 元素,且有Cr($ext{III}$)和Cr($ext{VI}$) 两种价态;Cr2 $p_{3/2}$ 轨道处的出峰由Cr($ext{III}$) 在(576.8 ± 0.1) eV 处的峰和Cr($ext{VI}$) 在(578.1 ± 0.1) eV 处的峰叠加而成,还原产物Cr($ext{III}$) 极有可能以 Cr($ext{OH}$) 。以及铁铬氧化水合物 [Fe_xCr_{1-x}(OH) 3]形式存在.

关键词:铁屑;微生物; Cr(VI);影响因素; XPS

中图分类号: X52 文献标识码: A 文章编号: 0250-3301(2013)07-2650-08

Removal of Cr(VI) by Iron Filings with Microorganisms to Recover Iron Reactivity

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Abstract: The reduction of Cr(VI) by iron filings and microorganisms was carried out during our investigation. Effects of factors (e. g. temperature, initial pH, iron filings loadings, inoculum size, initial concentration of Cr(VI) and other ions) on Cr(VI) reduction were studied, and the X-ray photoelectron spectroscopy (XPS) was applied to explore elements composition on the surface of iron filings. Experimental results revealed a promotion of Cr(VI) reduction by iron filings with the presence of microorganisms, and the Cr(VI) removal was complete within 18 h. Results showed the Cr(VI) reduction preferred a higher temperature within the range of 25-42°C and the optimum initial pH was supposed to be 5.8. The efficiency of Cr(VI) reduction was increased with increasing amounts of iron filings and microorganisms, and was decreased with the increasing initial concentration of Cr(VI). Cr(VI) reduction was increased with increasing amounts of iron sould cause varying degrees of inhibition of Cr(VI) reduction, and among these ions, the effect of Cr(VI) was the smallest and that of Cr(VI) as the Cr(VI) reduction with XPS indicated the deposition of Cr(VI) reduction or Cr(VI) and Cr(VI) as the Cr(VI) region could be decomposed into two peaks at Cr(VI) and Cr(VI) and Cr(VI) was most likely to be in the form of Cr(CII) or Cr(CIII) and Cr(VII) was most likely to be in the form of Cr(CIII) or Cr(CIII) and Cr(VII) was most likely to be in the form of Cr(CIII) or Cr(CIII) and Cr(VII) was most likely to be in the form of Cr(CIII) or Cr(CIII) and Cr(VII) was most likely to be in the form of Cr(CIII) or Cr(CIII) and Cr(VII) was most likely to be in the form of Cr(CIII) or Cr(CIIII) and Cr(VII) was most likely to be in the form of Cr(CIII) or Cr(CIIII) and Cr(VII) and Cr(VII) was most likely to be in the form of Cr(CIII) or Cr(CIIII) and Cr(VII) and Cr(VII) are Cr(VII) and Cr(VII) and Cr(VII) and Cr(VII) are Cr(VII) and Cr(

Key words: iron filings; microorganism; Cr(VI); influencing factors; XPS

格是冶金、电镀、印染、陶瓷、木材防腐及皮革制造等行业的重要原料,在工业生产中得到了广泛的应用.与此同时,大量含Cr(Ⅵ)废水排入生态环境,成为土壤、地下水及地表水中最常见的污染物之一. 铬的毒性与价态有关,且不同价态的化合物之间可以相互转化. 水体中铬主要以Cr(Ⅵ)和Cr(Ⅲ)两种形式存在. Cr(Ⅵ)具有很强的迁移性和毒性^[1],可诱发癌症,具有潜在的致畸和致突变作用. 相对来说,Cr(Ⅲ)性质稳定,易生成氢氧化物沉淀,是人体重要的微量营养元素^[2],对人体的毒性比Cr(Ⅵ)小100倍^[3]. 因此,将Cr(Ⅵ)还原为低毒性的Cr(Ⅲ),并利用其不易溶解的性质从水体中沉淀去除,成为水体Cr(Ⅵ)污染治理的重要选择.

目前国内外处理水体中Cr(VI)的研究主要集

中在吸附法、铁还原法、电化学处理法及微生物法. Singha 等^[4]对 6 种生物吸附剂去除Cr(VI)的动力学、吸附平衡和热力学等进行了研究;李克斌等^[5]利用荞麦皮对水体中Cr(VI)进行吸附,研究其吸附特性和机制. Liu 等^[6]利用铁还原菌驱动的生物电池还原去除水体中Cr(VI),得到了较好的Cr(VI)去除效果. 零价铁(Fe⁶)具有反应速度快、高还原势以及廉价等特点,自 Gillham 等^[7]提出将铁屑用于地下水的原位修复以来,Fe⁶ 还原修复污

收稿日期: 2012-10-06; 修订日期: 2012-12-10

基金项目: 国家自然科学基金项目(21277119); 浙江省科技计划项

目(2012C23061)

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染物便得到了广泛的研究和应用. 近年来,许多研究人员对 Fe^0 还原去除Cr(VI)的机制、动力学及影响因素等进行了深入研究 $[^{8^{-10}}]$,发现 Fe^0 对水体中的Cr(VI)具有还原去除能力. Mitra 等 $[^{11}]$ 对 Fe^0 还原去除 Cr(VI)的 动力学进行了研究,并建立了Cr(VI)去除速率模型. 然而单独用 Fe^0 还原去除水体中Cr(VI)时,往往由于 Fe^0 表面的钝化作用,即 Fe^0 颗粒表面的反应点位随着沉淀物的沉积而逐渐减少,而使Cr(VI)的去除效率逐渐降低, Fe^0 不能充分利用. Oh 等 $[^{12}]$ 研究发现二氧化硅对 Fe^0 去除水体中Cr(VI)起到促进作用,他们认为二氧化硅消耗了 Fe^0 -Cr(VI)的反应产物,从而减缓了 Fe^0 表面的钝化作用.

自然界中许多微生物,如 Goebacter 和 Shewanella 都被证实具有将Fe(Ⅲ)还原为Fe(Ⅱ)的能力^[13]. Mohatt 等^[14]曾利用铁还原菌还原Fe(Ⅲ)生成的Fe(Ⅱ),来修复土壤中磺胺甲基异噁唑(SMX)污染. 另一方面,许多微生物对高浓度的Cr(Ⅵ)表现出很强的耐受能力,并能直接将Cr(Ⅵ)还原成Cr(Ⅲ)^[15]. Dogan 等^[16]研究了 Pseudomonas对Cr(Ⅵ)的还原去除,发现胞外分泌物对Cr(Ⅵ)的还原起到主要作用. Escherichia coli 是常见的肠道细菌,具有还原Fe(Ⅲ)和Cr(Ⅵ)的能力,且其对Fe(Ⅲ)的还原能力大于其对Cr(Ⅵ)的还原^[17]. Guo 等^[18]研究在蒽醌磺酸盐(anthraquinone-2-sulfonate)存在下 Escherichia coli 对Cr(Ⅵ)的还原作用,得到了很好的Cr(Ⅵ)去除效率.

铁屑具有廉价易得,且可进行废物利用的特点,因此本研究采用铁屑作为 Fe^0 还原去除水体中Cr(VI).同时为了更有效地利用铁屑,保持较高的Cr(VI)去除效率,作者在铁屑-Cr(VI)体系中加入肠埃希氏菌($Escherichia\ coli$),主要研究铁屑-微生物-Cr(VI)体系对Cr(VI)的还原去除能力.

1 材料与方法

1.1 铁屑的预处理

铁屑原料来自浙江大学机械厂车床废料. 称取一定量颗粒大小接近且 > 10 目的铁屑,用 0.1 mol·L⁻¹的稀硫酸浸泡后,用丙酮洗涤 3 次,以去除铁屑表面的氧化层、油污、可能吸附的有机物及其他杂质. 最后用去离子水洗涤至中性,备用.

1.2 微生物的培养

实验菌种取自浙江大学环境工程研究所李伟教 授研究团队分离纯化的肠埃希氏菌(Escherichia coli FR-2)^[19]. 经驯化、富集后,取恒温摇床中在 32℃、150 r·min⁻¹条件下厌氧培养 2 d 的细菌,4000 r·min⁻¹离心 10 min 后,配成一定浓度的菌悬液待用.

基础培养基 ($\operatorname{mg} \cdot \operatorname{L}^{-1}$)组成:葡萄糖 1 000, $\operatorname{KH_2PO_4}$ 120, $\operatorname{Na_2SO_3}$ 28, $\operatorname{MgCl_2} \cdot \operatorname{6H_2O}$ 40, $\operatorname{NaHCO_3}$ 2 160, $\operatorname{NH_4Cl}$ 400.微量元素($\operatorname{mg} \cdot \operatorname{L}^{-1}$, 125 倍浓度)组成: $\operatorname{CoCl_2} \cdot \operatorname{6H_2O}$ 879, $\operatorname{MnCl_2} \cdot \operatorname{4H_2O}$ 1 980, $\operatorname{CuSO_4} \cdot \operatorname{5H_2O}$ 500, $\operatorname{NaMoO_4} \cdot \operatorname{2H_2O}$ 440, $\operatorname{NiCl_2} \cdot \operatorname{6H_2O}$ 380, $\operatorname{H_3BO_3}$ 28, $\operatorname{ZnCl_2}$ 200, $\operatorname{CaCl_2}$ 2 000.微生物培养液由基础培养基、微量元素以及 30 $\operatorname{mg} \cdot \operatorname{L}^{-1}\operatorname{Fe}(\operatorname{II})$ (以 $\operatorname{FeCl_3}$ 形式)和 10 $\operatorname{mg} \cdot \operatorname{L}^{-1}\operatorname{Cr}(\operatorname{VI})$ (以 $\operatorname{K_2Cr_2O_7}$ 形式)组成,并调节 pH 至 5.8 左右.实验所用试剂均为分析纯,采购自阿拉丁试剂(中国)有限公司和国药化学试剂有限公司.

1.3 铁屑-微生物协同还原Cr(VI)实验

配制由基础培养基、微量元素和 10 mg·L⁻¹ Cr(\overline{V} I) 组成的微生物培养液,调节 pH 后加入到 250 mL 培养瓶中. 通 \overline{N}_2 驱氧,并在 \overline{N}_2 保护下加入经预处理后的铁屑,接种一定量微生物. 最后用橡胶塞密封,放入恒温摇床中在 32°C、150 r·min⁻¹条件下培养. 每隔一定时间用注射器通入一定量 \overline{N}_2 后,从培养瓶中取出等量样品. 经8 000 r·min⁻¹ 离心 5 min 后,所得上清液用 0.45 μ m 滤膜过滤,留待分析.

实验主要考察温度 $(25、32、37、42、47^{\circ})$ 、 $pH(4.8 \times 5.8 \times 6.8 \times 7.8)$ 、铁形态 (可溶性三价铁、铁粉、铁屑)、铁投加量 $(0 \times 1 \times 2 \times 3 \times 5 \text{ g·L}^{-1})$ 、微生物量 $(0.025 \times 0.05 \times 0.1 \times 0.15 \text{ mg·mL}^{-1})$ 、(Cr((VI)) 初始浓度 $(10 \times 20 \times 30 \times 40 \times 50 \text{ mg·L}^{-1})$ 以及其他离子 (Cu²⁺ \times Co²⁺ \times Mn²⁺ \times Ni²⁺ \times Zn²⁺ \times 为 Cr((VI)) 还原效率的影响。每组实验不同条件值均使用同一批微生物同时进行实验,以确保比较的可靠性.

1.4 分析测试方法

Cr(Ⅵ)采用二苯碳酰二肼分光光度法,于 540 nm 波长下用上海棱光技术有限公司出产的Spectrumlab 23A可见分光光度计测定吸光度.采用原子吸收法测定样品中总 Cr 含量,原子吸收分光光度计型号为 AA-6300(日本 Shimadzu 公司). Cr(Ⅲ)含量通过计算,扣除总 Cr 中 Cr(Ⅵ)含量得到.

测定 Fe 含量的样品不经过离心、过滤,取样后直接测定. Fe(Ⅱ)含量采用邻菲罗啉分光光度法测

定,测定波长为510nm,仪器使用上海棱光技术有限公司出产的Spectrumlab 23A可见分光光度计.用10%盐酸羟胺将样品中Fe(Ⅲ)还原为Fe(Ⅱ)后,用同法测定总Fe含量.

微生物采用干重法定量,即将一定体积的菌悬液用电热恒温鼓风干燥箱(DHG-9053A,上海一恒科技有限公司)烘干至恒重,称量计算微生物接种量. SEM 分析用英国 Leica Cambridge 公司出产的S440型扫描电子显微镜完成. pH 值使用梅特勒-托利多仪器(上海)有限公司生产的 pH 计(SG2)测定.

对反应后的铁屑进行 X 射线光电子能谱 (XPS)分析,测定其表面组成. XPS 采用 VG ESCALAB MARK II 光电子能谱仪(ESCALAB,英国 VG 公司)测定,以 Mg Kα 作为射线源.

所有实验组均设置3个平行实验,且取其数据的平均值±标准偏差作为实验结果.

2 结果与讨论

2.1 铁屑-微生物协同还原去除六价铬

本研究所用微生物是以 Escherichia coli FR-2 为主要组成的混合细菌,对驯化后微生物进行扫描电镜测试,结果见图 1. 在放大4 680倍的扫描电镜下,微生物形态以杆状为主,另有一小部分的球状细菌.

铁屑-微生物协同还原去除Cr(VI)的实验在 32℃、150 r·min⁻¹的恒温摇床中进行,初始 pH 值调 节至5.8,设置未投加铁屑和微生物的实验组作为 空白,结果见图 2. 对比空白实验组结果可知,铁屑 和微生物均可单独还原去除溶液中Cr(VI); 其中微 生物在 18 h 内可还原去除约 60% 的Cr(VI),30 h 内Cr(VI)去除率达到77%;而铁屑在9h内使 Cr(VI)去除率达到约22%后,Cr(VI)去除率便逐步 趋于平稳,30 h后仍只有28%.分析认为铁屑还原 去除Cr(VI)是发生在铁屑表面的氧化还原反应,铁 屑表层 Fe⁰ 与水相反应被氧化成Fe(Ⅱ), 当水相中 Cr(VI)接触到铁屑表面并被吸附时,便立即与 Fe(Ⅱ)发生反应被还原,因此投加铁屑的实验组在 反应初始阶段就得到了较好的Cr(VI)去除率,如图 2 所示. Cr(VI) 在铁屑表面被还原后, 生成氢氧化 铬[Cr(OH),]沉淀或铁铬氧化水合共沉淀物 [Fe_xCr_(1-x)(OH),],沉积于铁屑表面,阻碍了铁屑 与水相的接触以及水相中剩余Cr(VI) 的还原. 因此 随着反应的进行, Cr(VI) 的去除效率便趋于稳定, 不再增大.

相对来说,同时投加铁屑和微生物的实验组, Cr(VI)还原去除最快,在18 h 内Cr(VI)去除率就可达到100%. 另外对只投加铁屑或微生物的实验组结果对应各取样时间进行加法计算,所得结果见图2 中的"投加微生物+投加铁屑"曲线,从中可知,同时投加微生物和铁屑对Cr(VI)的去除效率高于单独投加微生物或铁屑后Cr(VI) 去除效率的叠加. 实验结果表明铁屑和微生物在还原去除Cr(VI) 时具有促进作用,能够更有效的去除水体中Cr(VI).

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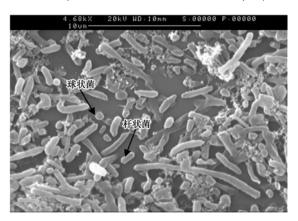
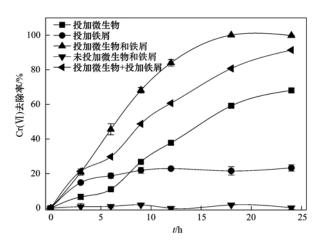


图 1 扫描电镜(SEM)下微生物形态特征

Fig. 1 SEM image of the bacteria



ρ_{Cr(VI)}: 10 mg·L⁻¹; pH: 5.8; 温度: 32℃; W: 150 r·min⁻¹

图 2 铁屑-微生物对Cr(VI)的还原去除作用

Fig. 2 Results of $Cr(\ V\!\!I)$ removal by iron filings and microorganisms

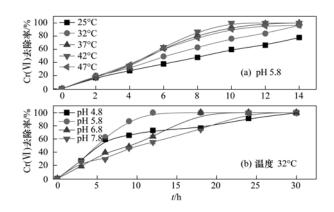
2.2 温度和 pH 对Cr(VI) 还原的影响

微生物的生长和活动受温度和 pH 影响较大,不同微生物最适宜的温度和 pH 各有不同. 而以往研究表明^[20~22],pH 的降低有利于铁对Cr(VI)的还原去除. 因此,温度和 pH 都将在一定程度上影响铁屑-微生物对Cr(VI)的还原去除.

图 3 所示为不同温度和初始 pH 对Cr(VI)去除率的影响. 从图 3(a)可以看出,在一定范围内,温

度的升高有利于Cr(VI)的去除. 温度为 32° 和 37° C时,Cr(VI)去除率在 14 h 内均可达到 96% 以上;温度为 42° C时,Cr(VI)在 10 h 内便可去除完全. 但当温度升高至 47° C时,Cr(VI)去除效率较 37° C和 42° C均有所下降,温度过高不利于Cr(VI)的还原去除. 在 4 h 以内,Cr(VI)去除率受温度影响并不明显,分析认为在反应初始 4 h 内,Cr(VI)的去除主要是铁屑对其具有还原作用,且温度对铁屑还原六价铬影响甚微,而微生物尚处于适应调整阶段,未能起到很好的还原作用.

从图 3 (b) 可以看出,铁屑-微生物还原去除 Cr(VI)的最佳 pH 为 5. 8. pH 值为 5. 8 时,Cr(VI) 在 12 h 内即可去除完全; 而 pH 为 4. 8、6. 8、7. 8 时,Cr(VI)完全去除时间分别需要 30、18、24 h. 表 1 是不同初始 pH 条件下,Cr(VI)完全去除后溶液的 pH 值. 结果表明在微生物的调节作用下,各实验组的 pH 值均在向 5. 4 靠拢. 由此可以推测,微生物生长和活动的最适宜 pH 值为 5. 4. 以下实验结果均在温度 32 ∞ 、pH 5. 8 条件下得到.



$$\begin{split} & \rho_{\rm Fe0} : 2 \ {\rm g \cdot L^{-1}} \ ; \ \rho_{\rm cell} : 0.1 \ {\rm mg \cdot mL^{-1}} \ ; \\ & \rho_{\rm Cr(W)} : 10 \ {\rm mg \cdot L^{-1}} \ ; \ W : 150 \ {\rm r \cdot min^{-1}} \end{split}$$

图 3 温度和 pH 对Cr(VI) 去除效率的影响

Fig. 3 Effects of different temperatures and initial pH on Cr(VI) removal

表 1 Cr(VI)去除前后溶液 pH 值变化情况

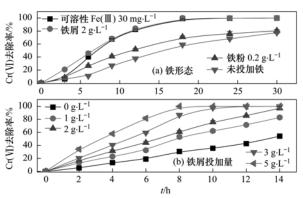
_	Table 1	Changes of pH	values before and	after Cr(VI)	reduction
	序号	1	2	3	4
	初始 pH	4.8	5. 8	6. 8	7. 8
	终止 pH	5, 35	5, 48	6. 39	6, 50

2.3 铁形态及铁屑投加量对Cr(VI)还原的影响

在保持其他条件一致的情况下,分别投加 30 mg·L⁻¹Fe(\blacksquare)(以 FeCl₃ 形式)、2 g·L⁻¹铁屑和 0. 2 g·L⁻¹铁粉,并以未投加任何形态铁作为对比,检验

不同形态铁对Cr(Ⅵ)的还原去除能力,结果见图 4 (a). 从中可知,铁的投加均能促进Cr(Ⅵ)的还原,且以溶解态Fe(Ⅲ)和铁屑的促进作用最为明显,在18 h 内便可将Cr(Ⅵ)去除完全. 投加铁粉的实验组对Cr(Ⅵ)的还原去除能力较差,30 h 后Cr(Ⅵ)去除率才达到80%,与对比实验组77%的Cr(Ⅵ)去除率相差不大. 因此,相比于铁粉,溶解态Fe(Ⅲ)和铁屑对Cr(Ⅵ)具有较好的还原去除作用.

图 4(b) 所示为不同铁屑投加量对Cr(VI) 去除 率的影响,Cr(VI)的还原去除效率随着铁屑投加量 的增加而增大. 铁屑还原去除Cr(VI)是发生在铁屑 表面的氧化还原反应,铁屑投加量的增加直接增大 了铁屑表面积,有利于Cr(VI) 的还原去除. 随着反 应进行,铁屑表面反应点位逐渐减少,铁屑钝化作用 越来越明显,但同时微生物的生长和活动逐渐频繁, 逐渐将 Fe(Ⅲ) 还原成 Fe(Ⅱ) 并再次与水相中 Cr(VI)反应,参与Cr(VI)的还原去除. 图 5 所示为 溶液中Fe(Ⅱ)含量随时间的变化情况,在未投加 Cr(VI)的实验组中, Fe(II)含量随时间逐渐增加; 而投加Cr(VI)的实验组中,当溶液中存在Cr(VI) 时,未能检测出Fe(Ⅱ),当溶液中Cr(Ⅵ)含量低于 检测限时,Fe(Ⅱ)含量逐渐增加. 对比只投加铁屑 的空白实验组,可以证实,微生物将Fe(Ⅲ)还原为 Fe(**II**) 并释放至溶液中,以及Fe(**II**) 对Cr(**VI**) 的还 原去除.



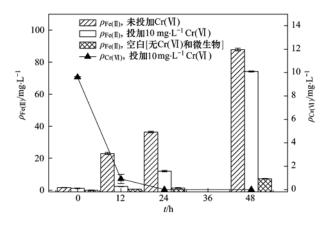
 $ho_{\rm cell}$: 0.1 mg·mL $^{-1}$; $ho_{\rm Cr(\,W\,)}$: 10 mg·L $^{-1}$; pH: 5.8; 温度: 32 $^{\circ}$ C; W: 150 r·min $^{-1}$

图 4 铁形态和铁屑投加量对Cr(VI)去除效率的影响

Fig. 4 Comparison of Cr(VI) removal by different iron forms and iron filings loadings

2.4 微生物量对Cr(VI)还原的影响

微生物接种量的测定采用干重法,即将一定体积的菌悬液烘干至恒重,称量计算微生物接种量.图 6 所示为不同微生物接种量对Cr(VI)还原去除

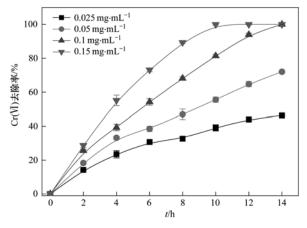


 $ho_{\rm Fe0}$: 2 g·L $^{-1}$; $ho_{\rm cell}$: 0.1 mg·mL $^{-1}$; pH: 5.8; 温度: 32 $^{\circ}$ C; W: 150 r·min $^{-1}$

图 5 溶液Fe(II)含量变化情况

Fig. 5 Changes of Fe($\rm II$) concentrations over time under different conditions

效率的影响,从中可知,Cr(VI)去除率随着微生物接种量的增加而逐渐增大.当微生物接种量为 0.15 mg·mL⁻¹时,Cr(VI)在 10 h 内便可去除完全;此时微生物接种量为 0.1、0.05、0.025 mg·mL⁻¹的实验组,Cr(VI)去除率分别达到 81%、55%、39%.在14 h 内,微生物接种量为 0.1 mg·mL⁻¹时,Cr(VI)去除率达 100%;而接种量为 0.05 mg·mL⁻¹和 0.025 mg·mL⁻¹的实验组,Cr(VI)去除率分别为 72% 和 46%.实验结果表明微生物接种量增加有利于Cr(VI)的还原去除,微生物接种量越大,Cr(VI)去除效率越大.但另一方面,Cr(VI)去除率的增量未能达到微生物接种量增量所对应的Cr(VI)去除效果,也就是说,随着微生物接种量增加,单位微生物接种量对Cr(VI)去除率的贡献却呈减小趋势.



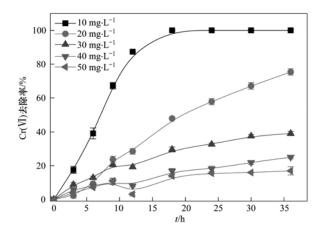
 $ho_{
m Fe0}$: 2 g·L $^{-1}$; $ho_{
m Cr(W)}$: 10 mg·L $^{-1}$; pH: 5.8; 温度: 32 $^{\circ}$ C; W: 150 r·min $^{-1}$

图 6 微生物接种量对Cr(VI)去除效率的影响

Fig. 6 Effects of different inoculum sizes on Cr(VI) removal

2.5 Cr(VI)初始浓度对其还原的影响

不同的Cr(VI)初始浓度,在改变Cr(VI)在溶液 中含量的同时,也对微生物的生长活动产生不同的 影响,并最终表现为对Cr(VI)去除率的影响. 实验 考察Cr(VI)初始浓度为 10~50 mg·L⁻¹时,Cr(VI) 去除率随时间变化的情况,结果见图 7. Cr(VI) 去 除效率随着Cr(VI)初始浓度的增大而显著减小;当 Cr(VI) 初始浓度为 10 mg·L⁻¹时, Cr(VI) 去除率在 12 h 内达到 87%, 在 18 h 内Cr(VI) 去除完全; 当 Cr(Ⅵ) 初始浓度为 20 mg·L⁻¹时,18 h 内Cr(Ⅵ) 去 除率只达到48%, Cr(VI)完全去除则需要72 h; 而 当Cr(VI) 初始浓度为 30、40、50 mg·L⁻¹时,18 h 内 和72 h内Cr(VI)去除率分别只有30%、17%、 14%和52%、31%、24%,且Cr(VI)去除率增加缓 慢. 结果表明,Cr(VI)初始浓度对Cr(VI)去除率的 影响较大, Cr(VI) 初始浓度在 20 mg·L-1 以下时 Cr(VI) 去除率随时间增加较快并最终可达到 100%,而 Cr(VI) 初始浓度在 30~50 mg·L-1 时 Cr(VI) 去除缓慢. 从Cr(VI) 去除量的绝对值来说, 随着Cr(VI)初始浓度的增加,相同时间内Cr(VI)的 去除量逐渐减少. 分析认为,随着Cr(VI) 初始浓度 增加,微生物受较高浓度毒性影响而减弱了还原 能力.



 $ho_{\rm Fe0}$: 2 g·L⁻¹; $ho_{\rm cell}$: 0.1 mg·mL⁻¹; pH: 5.8; 温度: 32°C; W: 150 r·min⁻¹

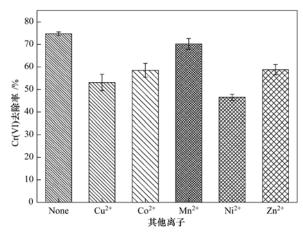
图 7 Cr(VI) 初始浓度对其去除效率的影响

Fig. 7 Effects of different initial Cr(\overline{VI}) concentrations on Cr(\overline{VI}) removal

2.6 其他离子的影响

水体中其他离子的存在往往会对Cr(VI)的还原去除产生影响,实验考察几种常见重金属二价阳离子存在情况下Cr(VI)的还原去除效率. 在保持其他条件相同的情况下,分别在微生物培养液中加入

0.1 mmol·L⁻¹的Cu²⁺、Co²⁺、Mn²⁺、Ni²⁺、Zn²⁺,10 h 内各实验组Cr(VI)去除率见图 8. 结果显示,添加 了 Cu²⁺、Co²⁺、Mn²⁺、Ni²⁺、Zn²⁺的实验组,在 10 h 内Cr(VI) 去除率分别为 53%、58%、70%、47%、 59%;而未添加其他离子的实验组,Cr(VI)去除率 达到 75%. 也就是说,添加的其他重金属离子对 Cr(VI)的还原去除均有一定的抑制作用,且影响从 小到大分别为 Mn²⁺、Zn²⁺、Co²⁺、Cu²⁺、Ni²⁺. He 等[23]研究了上述离子对 Ochrobactrum sp. 还原去除 Cr(VI)的影响,发现 Cu2+、Co2+、Mn2+具有微弱的 促进作用, 而 Zn2+、Ni2+具有抑制作用. Liu 等[24] 的研究显示 Escherichia coli 间接还原Cr(VI)的效率 在 Cu2+、Ni2+和 Co2+的存在下受到一定程度的抑 制,与本文结果一致. 分析认为 He 等[23]的研究中, 某些重金属离子的促进作用是由于增强了与 Cr(VI)还原相关的蛋白质的活性,或者参与了还原 酶或电子传递物质的合成,从而促进了微生物直接 还原 C(VI)的酶促反应过程. 本研究与 Liu 等[24]的 实验条件和结果较为相似,Cr(VI)的还原去除以微 生物的间接还原为主,添加的重金属离子对Cr(VI) 的间接还原去除过程产生了不利影响.



 $ho_{\mathrm{Fe}0}$: 2 g·L⁻¹; ho_{cell} : 0.1 mg·mL⁻¹; $ho_{\mathrm{Cr}(\mathrm{W})}$: 10 mg·L⁻¹; pH: 5.8; 温度: 32°C; W: 150 r·min⁻¹

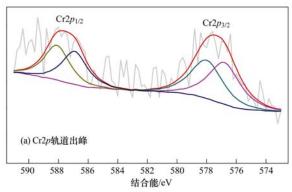
图 8 其他离子对Cr(VI)去除效率的影响

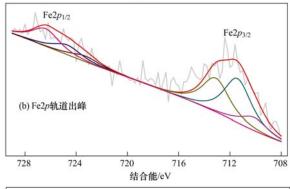
Fig. 8 Effects of other ions on Cr(VI) removal

3 Cr(VI)还原产物形态分析

为了验证Cr(VI)还原产物在铁屑表面的沉积,分析铁屑表面 Fe 和 Cr 的形态特征,本研究采用 VG ESCALAB MARK II 光电子能谱仪对反应后铁屑进行 XPS 分析,结果见图 9. 图 9(a)在结合能 576~579 eV 和 586~589 eV 处有两个明显的出峰,分别代表 $Cr2p_{1/2}$ 轨道和 $Cr2p_{1/2}$ 轨道. 根据 XPS 手册 [25],

 $Cr2p_{3/2}$ 轨道处的出峰由Cr(III)(CrOOH)在(576.8±0.1)eV 处的峰和Cr(VI)(CrO_3)在(578.1±0.1)eV 处的峰叠加而成. $Cr2p_{1/2}$ 轨道的结合能比 $Cr2p_{3/2}$ 轨道高9.9 eV 左右^[26]. 由此可知,铁屑表面有 Cr 元素存在,且 Cr 有 Cr(III) 和 Cr(VI) 两种价态. 由于各峰的峰面积对应了不同价态 Cr 的量,从图9(a)可知,铁屑表面的Cr(III) 和 Cr(VI) 含量相当. Cr 元素的 XPS 分析表明,铁屑-微生物-Cr(VI) 体系去除 Cr(VI) 时铁屑同时起到吸附和还原的作用.





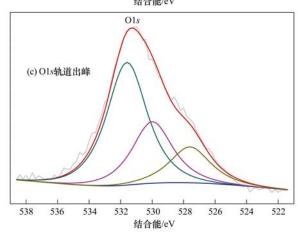


图 9 铁屑表面 XPS 分析

Fig. 9 X-ray photoelectron spectroscopy (XPS) for iron filings

图 9(b) 所示为 $Fe2p_{3/2}$ 轨道和 $Fe2p_{1/2}$ 轨道分别 在结合能 710 ~ 714 eV 和 723 ~ 727 eV 处的出峰, $Fe2p_{3/2}$ 轨道处的 3 个分峰表明铁屑表面的 Fe 以

Fe(Ⅲ)(711.5 eV ± 0.1 eV 和 713.2 eV ± 0.1 eV) 和Fe(Ⅱ)(710.2 eV ± 0.1 eV) 两种价态存在,且 Fe(Ⅲ)占 Fe 的绝大多数,Fe(Ⅱ)含量很少. O1s 轨道在结合能 527 ~732 eV 处有一个明显的出峰,由 3 个分峰叠加而成,其中(529.9 ± 0.1) eV、(531.5 ± 0.1) eV 两处的分峰经查[16]与 O^{2} 和 OH^{-} 有关.

铁屑表面 Cr、Fe 和 O 这 3 种元素的 XPS 分析证明了反应体系中Cr(VI)还原产物在铁屑表面的沉积,且还原产物极有可能以 $Cr(OH)_3$ 以及铁铬氧化水合物 $[Fe_xCr_{l,x}(OH)_3]$ 形式存在.

4 结论

- (1) 铁屑和微生物均能单独去除水体中Cr(W),但去除效率有限;而铁屑-微生物协同去除Cr(W)时两者具有促进作用,能够更有效的去除水体中Cr(W).
- (2)在 25~42℃内,Cr(VI)去除效率随着温度的升高而增大;当温度升高至 47℃时,Cr(VI)去除效率较 42℃有所下降. 实验结果表明,铁屑-微生物-Cr(VI)体系去除Cr(VI)的最适宜初始 pH 为 5.8.
- (3)铁屑-微生物协同还原去除Cr(VI)受到铁屑 投加量、微生物接种量和Cr(VI)初始浓度的影响; Cr(VI)去除效率随着铁屑投加量和微生物接种量的 增大而增大,随着Cr(VI)初始浓度的增大而减小.
- (4) Mn^{2+} 、 Zn^{2+} 、 Co^{2+} 、 Cu^{2+} 和 Ni^{2+} 离子对铁屑-微生物协同还原去除Cr(VI) 都有一定的抑制作用,其中 Mn^{2+} 的影响最小, Ni^{2+} 的抑制作用最为明显.
- (5)反应后铁屑表面的 XPS 分析表明, $Cr2p_{3/2}$ 轨道处的出峰由 $Cr(\mathbb{II})$ 在(576.8 ± 0.1) eV 处的峰和 $Cr(\mathbb{II})$ 在(578.1 ± 0.1) eV 处的峰叠加而成; $Fe2p_{3/2}$ 轨道处的 3 个分峰分别由 $Fe(\mathbb{II})$ (711.5 eV ± 0.1 eV 和 713.2 eV ± 0.1 eV)和 $Fe(\mathbb{II})$ (710.2 eV ± 0.1 eV)形成; O1s 轨道在(529.9 ± 0.1) eV、(531.5 ± 0.1) eV 两处的分峰与 O^{2-} 和 OH^{-} 有关.铁屑表面 Cr、Fe 和 O 这 3 种元素的 XPS 分析证明了反应体系中 $Cr(\mathbb{VI})$ 还原产物在铁屑表面的沉积,且还原产物极有可能以 $Cr(OH)_3$ 以及铁铬氧化水合物[$Fe_xCr_{1-x}(OH)_3$]形式存在.

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Environmental Science (monthly)

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(HUANJING KEXUE)

(月刊 1976年8月创刊)

2013年7月15日 34卷 第7期

ENVIRONMENTAL SCIENCE

(Monthly Started in 1976)

Vol. 34 No. 7 Jul. 15, 2013

主	管	中国科学院	Superintended	by	Chinese Academy of Sciences	
主	办	中国科学院生态环境研究中心	Sponsored	by	Research Center for Eco-Environmental Sciences, Chinese	
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		18号,邮政编码:100085)			KEXUE)	
电话:010-62941102,010-62849343					P. O. Box 2871, Beijing 100085, China	
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出	版	4 望出版社	Published	by	Science Press	
щ ///		北京东黄城根北街 16 号			16 Donghuangchenggen North Street,	
		邮政编码:100717			Beijing 100717, China	
印刷装	订	北京北林印刷厂	Printed	by	Beijing Bei Lin Printing House	
发	行	斜学出版社	Distributed	by	Science Press	
		电话:010-64017032			Tel:010-64017032	
		E-mail:journal@mail.sciencep.com			E-mail:journal@mail.sciencep.com	
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国外总发	行	中国国际图书贸易总公司	Foreign		China International Book Trading Corporation (Guoji	
		(北京 399 信箱)			Shudian), P. O. Box 399, Beijing 100044, China	

中国标准刊号: ISSN 0250-3301 CN 11-1895/X

国内邮发代号: 2-821

国内定价:90.00元

国外发行代号: M 205

国内外公开发行